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Bacterial biosurfactant from *Citrobacter freundii* MG812314.1 as a bioremoval tool of heavy metals from wastewater

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Abstract

Background: The problem of heavy metal contamination is a global issue, and the challenge is to develop methods to remove heavy metals from soil and water. Recently, biosurfactants are one of the compounds that provide an attractive eco-friendly alternative to the physicochemical process in alleviating the heavy metals.

Results: Sixty bacterial isolates were isolated from Al-Rahawy drain sediments and screened for biosurfactant production. Only 10 isolates were recoded as biosurfactant producers by the oil spreading and emulsifying assays in addition to the ability of biosurfactant on heavy metal removal in wastewater. The most potent isolate was identified using morphological, cultural, biochemical characteristics, antibiotics susceptibility, and 16 s rRNA technique as *Citrobacter freundii* MG812314.1.

Conclusions: The produced biosurfactant was found to be more effective in removing heavy metals from wastewater, viz 80, 67, 66, 55, 45, 44, and 41% of aluminum, lead, zinc, cadmium, iron, copper, and manganese, respectively, under two inoculum potentials and two contact time. The interaction of heavy metals with biosurfactant was monitored using scanning electron microscope (SEM), energy dispersive X-ray spectra (EDX), and Fourier-transform infrared spectroscopy (FTIR) analyses.

Keywords: Heavy metals, Biosurfactant, Wastewater, Bioremediation, *Citrobacter freundii*

Introduction

Biosphere pollution due to heavy metals is a confused problem that causes negative effects on various environments. In some cases, the levels of heavy metals in all environments (air, water, and soil) are increasing to toxic levels with contributions from a wide variety of agricultural, industrial, and domestic sources. Metal-contaminated environments pose a risk on health and ecosystems. Moreover, the most abundant pollutants in waste water are heavy metals (Vijayanand and Divyashree 2015). Nowadays, the major problem that facing Egypt is the rise in the pollution level caused by the Nile's low water level, especially after completing the Ethiopia Dam building. Heavy metals that mainly derived from multiple anthropogenic sources are considered one of the important pollutants in the Nile River

(Ezzat et al. 2012). These metals cause serious ecological problems because they are non-degradable and can bioaccumulate through the food chain (Goher et al. 2014).

There are many remediation techniques used to remove metals from soils and wastewater such as chemical precipitation, oxidation, adsorption immobilization, and phytoremediation (Patel and Vashi 2010). However, in general, these conventional treatment methods are expensive and not environmentally acceptable as they can themselves produce other waste disposal problems. Also, the low bioavailability of those metals in soils may limit the efficiency of their industrial applications (Sheng et al. 2008). Therefore, these disadvantages have prompted the search for effective, cheap, and eco-friendly alternatives.

Biosurfactants are one of the compounds that help in alleviating the bad influences of heavy metals (Singh and Cameotra 2013); there are reports on the use bacterial biosurfactant for remediation of heavy metals from aqueous solution (Ramani et al. 2012). Biosurfactants are

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surface-active degradable organic compounds produced by microorganisms (Bachmann et al. 2014). Generally, their structure includes two moieties, hydrophilic and hydrophobic, the first moiety consisting of amino acids, peptides, anions or cations, and mono-, di-, or polysaccharides, and the second moiety consisting of unsaturated or saturated fatty acids (Basak and Das 2014). Also, biosurfactants can be classified into several groups, viz. glycolipids, lipopeptides, lipopolysaccharides, phospholipids, and fatty acids/neutral lipids. Some of the advantages of biosurfactants over synthetic ones are cost-effectiveness, lower toxicity, higher biodegradability, better environmental compatibility, high selectivity and specific activity at extreme temperatures, pH and salinity, and the ability to be synthesized from renewable feed-stock (Adamu et al. 2015; Lazarkevich et al. 2015). Moreover, biosurfactants can be tailor-made to suit different applications by changing the production conditions or by modifying the genes involved in their biosynthesis (Thernmozhi et al. 2011).

This study aimed to isolate, screen, and identify of the biosurfactant-producing bacteria from different heavy metals which contaminated the water in Egypt, then optimize the biosurfactant production from the most potent isolate, and finally, examine the ability of the produced biosurfactant to remove heavy metals from wastewater.

Materials and methods

Collected water sample locations

Two contaminated water samples were collected during spring 2016 from two different locations along Al-Rahawy drain, Giza Governorate, Egypt (Fig. 1). Location no. (1) was downstream delta barrage-upstream Al-Rahawy drain (5 km) (temperature 23.0 °C). Location no. (2) was downstream Al-Rahawy drain-upstream Sabal drain (flows into Rosetta Branch from the west side about 9 km north of Delta Barrage) (temperature

25.0 °C). Samples were collected in sterile glass bottles and stored at 4 °C until use.

Isolation of biosurfactant-producing bacteria

One milliliter of each water sample was inoculated into 50 ml nutrient broth (pH 7 ± 0.2) (HIMEDIA Co., Germany). All flasks were incubated at 32 ± 2 °C for 48 h in shaking incubator (150 rpm/min); then, 1 ml of each culture was serially diluted up to 10^{-6} . After that, 1 ml of each dilution was transferred to sterile Petri dish containing the same solidified medium and incubated at 32 ± 2 °C for 48 h. At the end of the incubation period, colonies were purified and maintained for further experiments (Sneha 2012).

Screening of the isolated bacteria for biosurfactant production

Each bacterial isolate was cultured in 100-ml Erlenmeyer flasks containing 50 ml of nutrient broth medium, inoculated with 10% inoculum, and then, incubated at 32 ± 2 °C for 72 h. At the end of the incubation period, the cultures were centrifuged at $7000 \times g$ for 10 min. The cell-free supernatant was analyzed for the presence of surface-active compounds by the following assays (Vijayanand and Divyashree 2015).

Oil spreading assay

Biosurfactant production was estimated using oil spreading assay as described by Youssef et al. (2004). One hundred microliters of crude sun flower oil was added to the surface of 50 ml of distilled water in a Petri dish to form a thin oil layer. Then, 10 μ l of cell-free supernatant was gently placed on the center of the oil layer. The diameter of the clearing zone on the oil surface would be visualized under visible light and measured after 30 s, which correlates to the surfactant activity, also known as oil displacement activity (Morikawa et al. 1993), using sodium dodecyl sulfate (SDS), a chemical surfactant compound, as control.

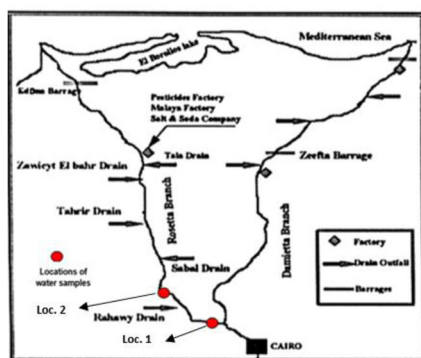


Fig. 1 Al-Rahawy drain, Giza Governorate, Egypt

Emulsifying assay

Two milliliters of each cell-free supernatant was mixed with 2 ml of toluene in clean glass tubes. The mixture was shaken vigorously for 2 min and then left to stand. Emulsification index (EI %) and emulsification stability (ES %) were measured in intervals up to 24 h using the following equations (Arieche and Guechi 2015).

$$\text{EI (\%)} = \frac{\text{Emulsion height (mm)}}{\text{Total height (mm)}} \times 100$$

$$\text{ES (\%)} = \frac{\text{Emulsion height (mm) after 24 h}}{\text{Emulsion height (mm) at zero time}} \times 100$$

Effect of biosurfactant on heavy metal removal in wastewater

The biosurfactants producing isolates which gave positive results in the previous screening assays were used in this experiment. One hundred milliliters of Erlenmeyer flasks containing 50 ml of wastewater were treated with 5% cell-free supernatant of the selected isolates and incubated for 48 h under shaking conditions. Then, the residual concentrations of heavy metal were detected using plasma atomic emission spectrometry (ICP-AES), and the bioremoval activity was measured using the equation by Basak and Das (2014):

$$\% \text{bioremoval activity} = \frac{C_i - C_f}{C_i} \times 100$$

where

C_i is the initial concentration of heavy metal in wastewater before treatment

C_f is the final concentration of heavy metal in wastewater treated with bacterial supernatant

Bacterial identification**Morphological, cultural, and biochemical characterization**

A light microscope was used to check for Gram stain reaction, shape, sporulation, and motility procedures, while hanging drop method was used to determine the motility of the most biosurfactant-producing bacteria (Bertrand et al. 2001). Spore formation was determined according to Harrigan and MacCance (1976). The cultural characteristics were recorded as colony morphology, i.e., color, shape, size, nature of the colony, and pigmentation, according to the method described by Hucker and Conn (1923). Additionally, biochemical characteristics (oxidase, catalase, and amylase enzymes, indole, methyl red, and Voges Proskauer tests as well as sugar fermentation) were estimated as according to the key of Bergey's Manual of Systematic Bacteriology (Janda et al. 1994).

Antibiotics susceptibility test

Susceptibility test of the potent biosurfactant-producing isolate was carried out using disk diffusion method (Matuschek et al. 2014). The antibiotics used were amphenicols (chloramphenicol), tetracyclines (tetracycline), quinolones (nalidixic acid), polymyxins (polymyxin b), macrolides (erythromycin 15 mg), and glycopeptide (vancomycin 500 mg).

DNA and phylogenetic analysis

The potent biosurfactant-producing bacterial isolate was identified by 16S rRNA sequencing analysis according to the method described by Abdel-Rahman et al. (2017). Sequence analysis was performed using Basic Local Alignment Search Tool for Nucleotides (BLASTN) with sequences at the National Center for Biotechnology Information (NCBI), USA, database.

Optimization of biosurfactant production

In addition to nutrient broth medium, three broth media, viz. mineral salt with glucose (MSG), LB, and peptone yeast extract glucose (PYG), were used in this experiment. All these media were obtained from HIME-DIA Co., Germany. Moreover, four incubation periods (3, 4, 5, and 6 days) and three inoculum potentials 10^6 CFU/ml (5, 10, and 15%) were examined according to Ibrahim (2018) and Basak and Das (2014), respectively.

Biosurfactant extraction and partial purification

Biosurfactants were extracted by adjusting the pH of cell-free supernatant to 2.0 using 6 N HCl and keeping it at 4 °C overnight. The precipitate thus obtained was pelleted by centrifugation at 7000×g for 20 min and dried (Suganya 2013). For partial purification, the crude surfactant was dissolved in distilled water at pH 7.0 and dried at 60 °C. The dry product was extracted with mixture of chloroform:methanol (65:15) and then filtered, and the solvent was evaporated.

Factors affecting the removal of heavy metals from wastewater

Two hundred milliliters of Erlenmeyer flasks containing 100 ml of wastewater were treated with the produced biosurfactant at various concentrations, viz. 5 and 10% (w/v), and incubated at 32 ± 2 °C for two different periods, namely 3 and 6 days. Then, seven heavy metal, namely aluminum (Al^{2+}), cadmium (Cd^{2+}), copper (Cu^{2+}), iron (Fe^{2+}), lead (Pb^{2+}), manganese (Mn^{2+}), and zinc (Zn^{2+}), concentrations were determined using plasma atomic emission spectrometry (ICP-AES) as described by Becker (2005). Wastewater without any amendments was used as control. The bioremoval activity was measured as the equation previously mentioned.

Metal chelating activity

In this experiment, three assays using scanning electron microscope (SEM), energy dispersive X-ray spectra (EDX), and Fourier-transform infrared spectroscopy (FTIR) analyses were done to study the ability of the produced biosurfactant to chelate heavy metals from wastewater as described by Basak and Das (2014). The morphology of the native biosurfactant and that interacted with heavy metals was investigated. After completely dried and grounded into powder form, samples were gold coated using SPI-Module sputter coater and examined under scanning electron microscope (JEOL, JSM-5500 LV) at 20 kV. EDX analysis of the samples was performed using X-ray microanalyzer (Module oxford 6587 INCA x-sight) attached to SEM at the Regional Center of Mycology and Biotechnology, Cairo, Egypt. Additionally, FTIR was conducted using Perkin Elmer spectrophotometer (Jasco-6100, Japan) to show the differences between native and heavy metals interacted biosurfactant. The dried materials were ground with KBr pellets and measured in the wavelength range 4000 to 500 cm^{-1} .

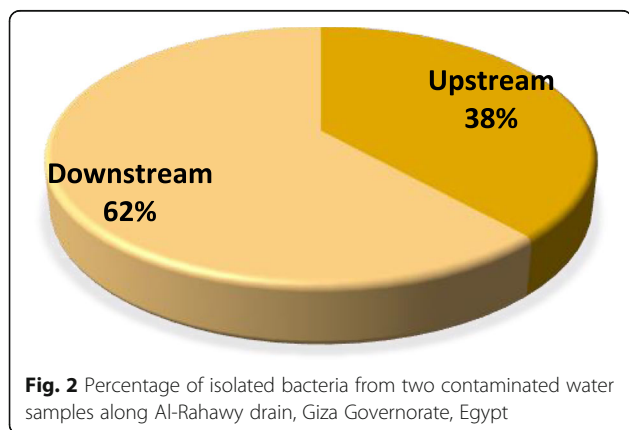
Statistical analysis

Analysis of variance in one- and two-way ANOVA was carried out using the SAS procedure guide (SAS 2004). Significant differences among means were tested using the Duncan multiple range test (Duncan 1955).

Results

Isolation of biosurfactant-producing bacteria

In this experiment, two contaminated water samples were collected from two different locations along Al-Rahawy drain, Giza Governorate, Egypt. A total of 60 bacterial isolates were recovered on nutrient agar medium. Twenty-three isolates (38%) were isolated from location (1) upstream Al-Rahawy drain sediment (BS1-BS23), and 37 isolates (62%) were isolated from location (2) downstream Al-Rahawy drain sediment (BS24-BS60) (Fig. 2).



Screening of the isolated bacteria for biosurfactant production

Oil spreading assay

The results in Table 1 showed that among all 60 tested bacterial isolates, only 10 isolates (BS3, BS9, BS12, BS17, BS33, BS34, BS37, BS42, BS43, BS55) were able to displace oil and form a clear zone. This may be due to the presence of biosurfactant which degrades the oil and form a clear zone. The diameter of clear zone ranged between 7 and 30 mm. The isolate BS37 showed higher zone of displacement compared to other isolates (Fig. 3), while the least zone of displacement was found by the isolate BS55.

Emulsifying assay

The results recorded in Table 1 and illustrated by Fig. 4 showed that the isolate BS37 showed the maximum emulsification index (89.0%) followed by the isolate BS3 (70.0%) and the least activity was shown by the isolate BS55 (31.4%). Also, the data proved that all tested isolates showed relative emulsification stability after 24 h. The isolate BS37 gave higher emulsion stability (90.9%) compared to other isolates.

Effect of biosurfactants on heavy metal removal in wastewater

Data presented in Table 2 indicated that all ten isolates were able to decrease all estimated heavy metals in wastewater (Al^{2+} , Cd^{2+} , Cu^{2+} , Fe^{2+} , Pb^{2+} , Mn^{2+} , and Zn^{2+}) with various percentages. Generally, Al^{2+} and Cu^{2+} were considered as the highest and the lowest removed heavy metals by all tested isolates, respectively. Moreover, the biosurfactant produced by the isolate BS37 was able to remove the highest amounts of all tested heavy metals from wastewater. Therefore, this isolate was selected for further experiments.

Identification of the most potent bacterial isolate

Bacterial isolate BS37 showed the highest biosurfactant production and exhibited the greatest heavy metal removal efficiency which was identified as follows:

Morphological and biochemical characterization

The isolate BS37 was straight short rods, Gram-negative, motile, and non-spore forming. Also, it was a facultative anaerobe, and its colonies were smooth, mucoid, and light gray in color but do not able to produce soluble pigments in broth medium (Table 3).

Phenotypic characterization

The characteristics of the isolate BS37 presented in Table 3 showed that it is not able to grow on nitrogen-free medium, produces hydrogen cyanide (H_2S), and is catalase-positive and oxidase-negative. It utilized citrate as

Table 1 Screening of the isolated bacteria for biosurfactant production

Assays	SDS	Bacterial isolates										MSE
		BS3	BS9	BS12	BS17	BS33	BS34	BS37	BS42	BS43	BS55	
Oil spreading (mm)	25 ^{ab}	20 ^{bc}	21 ^{bc}	11 ^{ef}	12 ^{ef}	17 ^{cde}	18 ^{cd}	30 ^a	9.0 ^f	11 ^{ef}	7.0 ^f	1.81
EI %	69.1 ^b	70.0 ^b	62.1 ^c	59.1 ^{cd}	32.1 ^f	41.8 ^e	41.8 ^e	89.0 ^a	33.5 ^f	55.6 ^d	31.4 ^f	1.0
ES %	71.4 ^{bc}	71.4 ^{bc}	70.9 ^{bc}	60.9 ^e	71.0 ^{bc}	59.8 ^e	69.4 ^c	90.9 ^a	65.7 ^d	64.7 ^d	73.2 ^b	1.0

Mean having similar letters in each column are not significantly different ($P < 0.05$)
 EI emulsification index, ES emulsification stability

a sole carbon source and is positive in indole and methyl red tests, but negative in Voges-Proskauer test. Additionally, it was able to utilize glucose, xylose, arabinose, maltose, and mannitol as sole carbon source and the acid derived from them which caused indicator color change, but do not able to use lactose.

Antibiotic susceptibility test

Antibiotic susceptibility patterns of the obtained BS37 isolate demonstrated a susceptibility profile characteristic (Table 3). The isolate was resistant to polymyxin b and erythromycin. It remained susceptible to chloramphenicol, tetracycline, nalidixic acid, and vancomycin.

DNA and phylogenetic analysis

Amplification of 16S rRNA of the isolate BS37 and its sequencing were carried out, and the BLAST analysis of the complete sequences of 16S rRNA revealed the isolate to be *Citrobacter freundii* (GenBank accession No. MG812314.1). Distribution of the various strains on the phylogenetic tree showed that the isolate was closely related to the reference standard strains (Fig. 5a, b). The bacterial classification is phylum, *Proteobacteria*; class, *Gammaproteobacteria*; order, *Enterobacteriales*; family, *Enterobacteriaceae*; genus, *Citrobacter*.

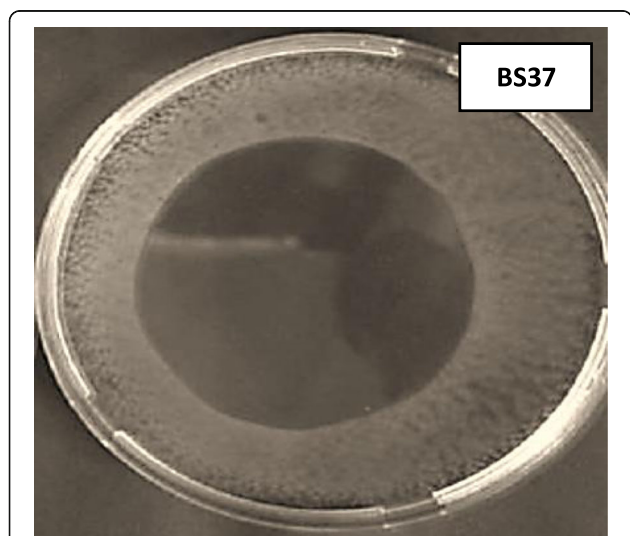


Fig. 3 Oil spreading assay of bacterial isolate BS37

Optimization of biosurfactant production by *C. freundii* MG812314.1

Citrobacter freundii could produce biosurfactant using all the tested media. It was observed that nutrient broth medium yielded the maximum biosurfactant production by *C. freundii* MG812314.1, and the emulsification index reached 89% (Fig. 6). In contrast, the lowest biosurfactant production (EI = 5%) was recorded when *C. freundii* MG812314.1 is cultured on LB medium. Regarding the interaction between the inoculum potential (%) and the incubation period (days) on biosurfactants production by *C. freundii* MG812314.1., data graphically illustrated by Fig. 7a, b, and c showed that inoculum potential of 10% and incubation period of 4 days were found to yield maximal amount of biosurfactant (EI = 95.8%).

Factors affecting the removal of heavy metals from wastewater

Data presented in Table 4 indicated that higher removal of heavy metals was recorded when wastewater was treated with biosurfactant at 10% than 5%. Results also showed higher removal efficiency with increasing the contact time when wastewater was treated with 5% biosurfactant. In contrast, when treated with 10%, the removal efficiency was decreased with the increasing of the contact time. Generally, the highest removal of all heavy metals was observed when wastewater was treated with 10% and incubated for 3 days. Interestingly, the highest and the lowest removed metals from wastewater by biosurfactant were Al^{2+} and Cu^{2+} , respectively. This trend of results was true when wastewater was treated with any biosurfactant concentration and incubated for any contact time. This result agrees with the results in



Fig. 4 Emulsifying assay of the tested bacterial isolates

Table 2 Efficiency of biosurfactants of the ten selected isolates to remove heavy metals from wastewater

Heavy meals	Biosurfactant-producing bacterial isolates										MSE
	BS3	BS9	BS12	BS17	BS33	BS34	BS37	BS42	BS43	BS55	
	Heavy metals removal (%)										
Aluminum (Al ²⁺)	82.0 ^{bc}	78.5 ^d	83.3 ^b	79.8 ^{cd}	83.4 ^b	83.4 ^b	87.4 ^a	81.5 ^{bcd}	81.3 ^{bcd}	82.2 ^{bc}	0.98
Cadmium (Cd ²⁺)	13.9 ^f	19.9 ^{de}	30.0 ^b	20.9 ^d	19.9 ^{de}	19.5 ^{de}	39.9 ^a	22.7 ^c	18.9 ^e	24.3 ^c	0.56
Copper (Cu ²⁺)	10.3 ^{de}	11.0 ^{bc}	8.10 ^g	10.7 ^{cd}	9.80 ^{ef}	2.10 ^h	19.2 ^a	11.3 ^b	9.70 ^f	9.30 ^f	0.19
Iron (Fe ²⁺)	19.3 ^f	21.3 ^d	31.3 ^b	24.3 ^c	18.3 ^g	11.3 ^h	33.8 ^a	20.3 ^e	11.0 ^h	21.1 ^d	0.25
Lead (Pb ²⁺)	51.0 ^b	6.70 ⁱ	23.8 ^h	41.7 ^d	32.5 ^f	46.4 ^c	56.8 ^a	36.9 ^e	29.2 ^g	16.4 ⁱ	0.29
Manganese (Mn ²⁺)	15.8 ^e	17.9 ^d	20.8 ^b	19.3 ^c	11.5 ^f	15.0 ^e	25.3 ^a	20.1 ^{bc}	ND	ND	0.27
Zinc (Zn ²⁺)	43.1 ^c	20.5 ⁱ	28.7 ^g	41.8 ^d	45.1 ^b	43.2 ^c	49.2 ^a	38.0 ^e	33.5 ^f	28.0 ^h	0.20

Mean having similar letters in each row are not significantly different ($P < 0.05$)

Table 2 which proved that Al²⁺ was removed more than other metals by all tested isolates.

Metal chelating activity

Scanning electron microscopy

SEM images showed the morphology of native biosurfactant (Fig. 8a) and the sequestered heavy metals onto biosurfactant (Fig. 8b). The spherical nodules in SEM image (b) confirmed the anchoring of heavy metal ions with biosurfactant molecule.

EDX analysis

EDX of native biosurfactant (Fig. 9a) and heavy metals interacted biosurfactant (Fig. 9b) showed that there were constant metals in either native or interacted biosurfactant, viz. O, Na, Al, P, S, and Cl, whereas Fe²⁺, Cu²⁺, Zn²⁺, and Pb²⁺ appeared in interacted biosurfactant with wastewater only.

FTIR analysis

The FTIR spectrum of native biosurfactant (Fig. 10a) showed the presence of O–H stretching at 3444.21 cm⁻¹. The absorption band noted at 1630.30 cm⁻¹ was due to carbonyl stretching of C=O group. A band at 1452.56 cm⁻¹ indicated C–O stretch. FTIR spectrum of biosurfactant interacted with heavy metals present in wastewater showed that the recorded peaks at 3444.21, 1630.30, and 1452.56 cm⁻¹ were shifted respectively to 3446.12, 1635.82, and 1456.87 cm⁻¹. Also, new bands appeared at 2842.11 and 1540.54 cm⁻¹ (Fig. 10b).

Discussion

Biosurfactants are amphiphilic nature compounds with massive diversity and broad spectrum of functions and environmental applications which make them good bio-products. Contaminated soil and water are known as continuous sources of different microorganisms which are able to emulsify and solubilize hydrophobic compounds and have advantages over their competitors;

Table 3 Morphological, cultural, biochemical characteristics and antibiotics susceptibility of the highest BS-producing bacterial isolate (BS37)

Characteristics	SB37			Characteristics	SB37		
Shape	Short rod			Grow on nitrogen free medium	-		
Gram staining	Negative			H ₂ S production	+		
Motility	Motile			Catalase production	+		
Respiration	Facultatively anaerobic			Oxidase production	-		
Spore formation	Non-spore forming			Citrate utilization	+		
Pigmentation	of colony	Light gray		Indole production	+		
	of medium	-		V.P. test	-		
Colony	Smooth and mucoid			Methyl Red	+		
Carbon source and acid derived	Glucose	Xylose	Arabinose	Maltose	Mannitol	lactose	
	+	+	+	+	+	-	
Antibiotic susceptibility	Chlora.	Tetra.	Nalidixic	Polymyx.	Erythro.	Vanco.	
	S	S	S	R	R	S	
(S): sensitive		(R): resistance		(+): positive		(-): negative	

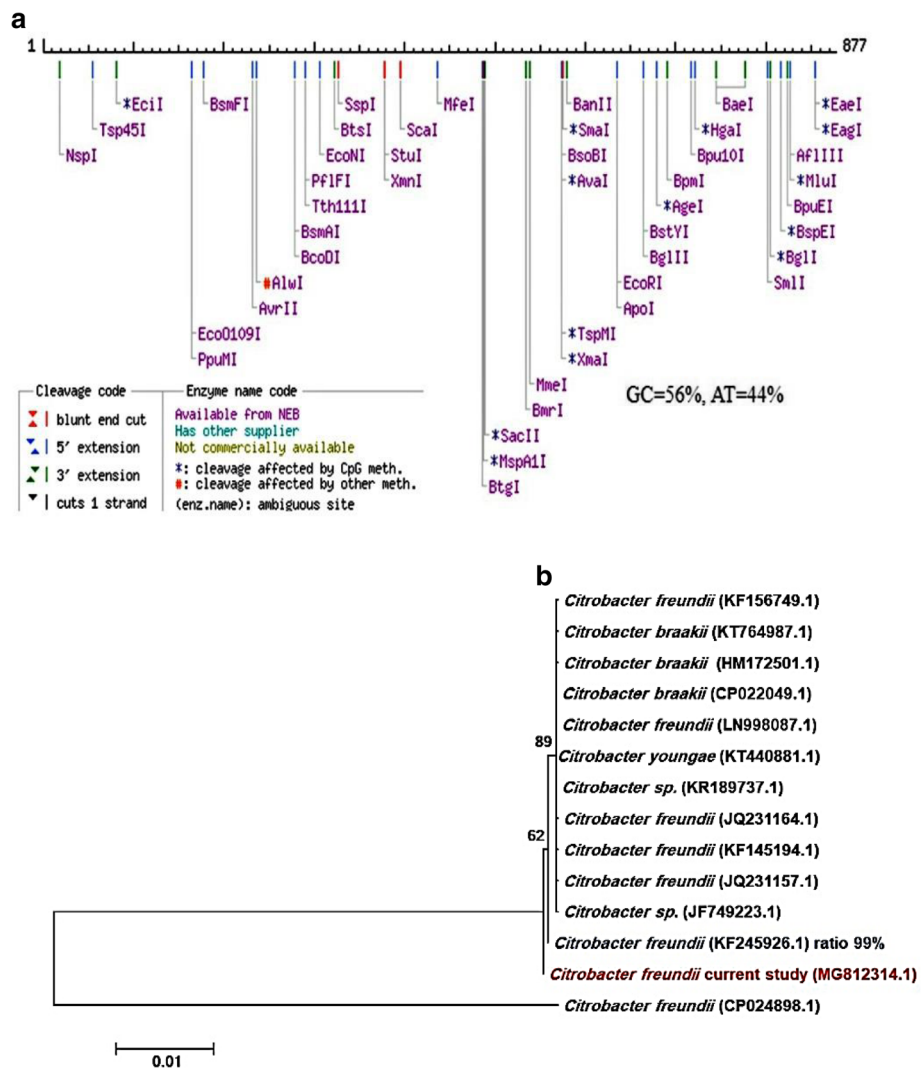


Fig. 5 a Restriction map of the partial sequence of 16S rRNA gene of interest with available commercially restriction enzymes. b Phylogenetic tree showing interrelationships of isolate (BS37) based on 16S rRNA sequences

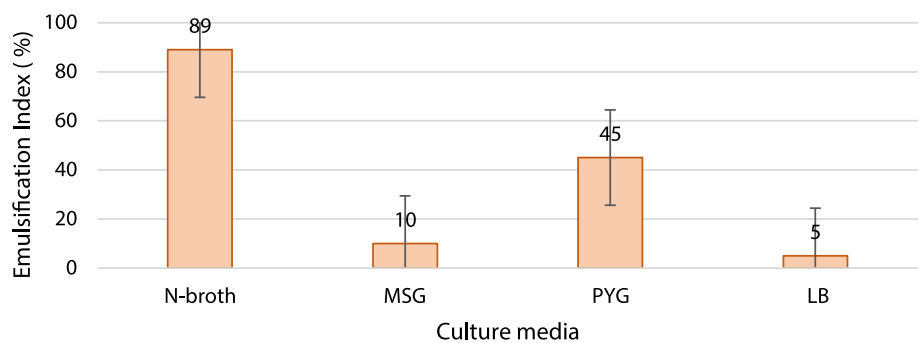
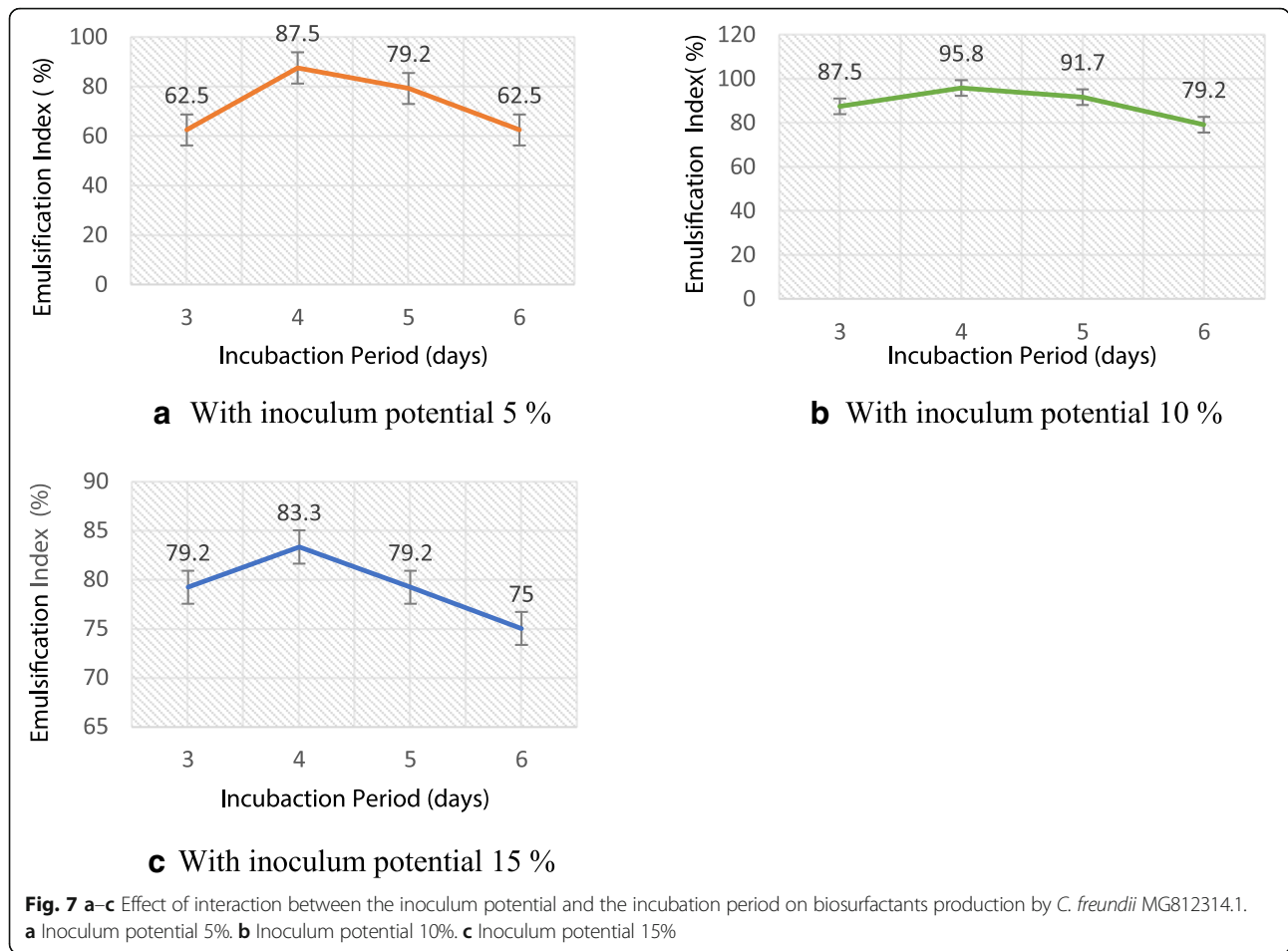


Fig. 6 Biosurfactant production by *C. freundii* MG812314.1 using different media



collecting samples from these environments provides a rich source of microorganisms with desirable properties. In this study, pure biosurfactant-producing bacterial isolates were recovered from contaminated water samples collected from two locations along Al-Rahawy drain, Giza Governorate, Egypt, and it was clear that larger number of isolates was obtained from downstream than upstream Al-Rahawy drain water samples. This result was logical and may be attributed to location no. 1 receiving pollutants from about 66 drains for 1000 km

passing through governorates from Aswan to Cairo, but with different degrees of pollution, whereas location no. (2) is extended after Al-Rahawy drain and carries the combined discharge from the agricultural drainage system and the two sewerage pump stations (Zenein and Abo-Rawwash) that flows into Rosetta Branch from the west side about 9 km north of Delta Barrage, in addition to the industrial facilities that drainage in Al-Rahawy drain or in its branched drains. It is worth mentioning that due to the steady population increase and the large urban expansions in the areas served by these two stations, the quantities of the entered sewage increased up to 180% of its capacity which led to the conversion of this sewage water without treatment to the drain. Also, untreated sewage from the group of population blocs and parcels spread throughout the drain and its branches in Giza Governorate (El Bedawy 2014; Egypt State of Environment 2015; Wolters et al. 2016). In the same trend, Elazzazy et al. (2015) isolated 23 biosurfactant-producing bacterial colonies from soil and sea water samples, which were collected from Jeddah region, Saudi Arabia.

Table 4 Factors affecting removal of heavy metals from wastewater

Biosurfactant amount (%)	Contact time (day)	Heavy metals removal (%)						
		Al ²⁺	Cd ²⁺	Cu ²⁺	Fe ²⁺	Pb ²⁺	Mn ²⁺	Zn ²⁺
5	3	47.0 ^b	25.0 ^c	50.0 ^c	30.0 ^b	20.0 ^c	30.0 ^c	55.0 ^c
	6	50.0 ^b	30.0 ^b	52.0 ^c	35.0 ^b	23.0 ^c	40.0 ^b	62.0 ^b
10	3	66.0 ^a	41.0 ^a	67.0 ^a	45.0 ^a	44.0 ^a	55.0	80.0 ^a
	6	54.0 ^b	39.0 ^a	61.0 ^b	33.0 ^b	37.0 ^b	53.0	77.0 ^a
MSE		2.17	0.91	0.91	2.34	1.77	1.47	1.38

Mean having similar letters in each column are not significantly different ($P < 0.05$)

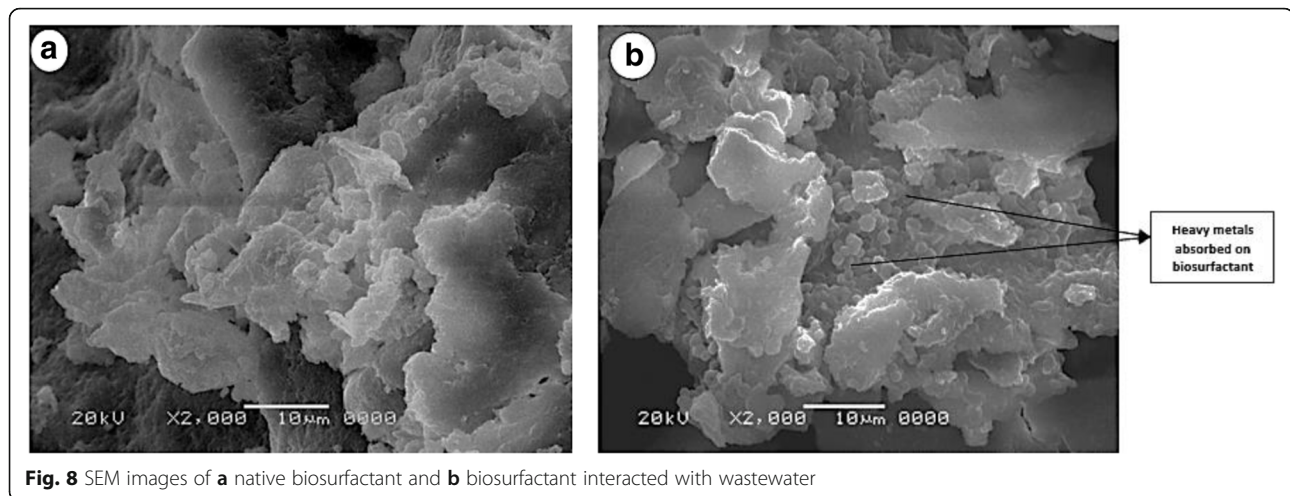


Fig. 8 SEM images of **a** native biosurfactant and **b** biosurfactant interacted with wastewater

The screening for biosurfactant production was carried out using two techniques, oil spreading and emulsion activity, as reported by Satpute et al. (2008) who affirmed that more than one screening method should be used in the primary screening for the biosurfactant producers. Further, these methods have many advantages, viz. simplicity, low cost, and quick application. Supplement to this trend, many researchers used other methods in this primary screening such as the cultivation of bacteria on blood agar medium and the drop collapse method as a sensitive and easy method; dark blue halo zone in the methylene blue agar plate supplemented with CTAB confirmed the presence of an anionic biosurfactant and surface tension decrease (Elazzazy et al. 2015). The oil displace test is indicative of surface wetting activity (Youssef et al. 2004) which depends on a drop of the bacterial supernatant containing a biosurfactant which collapses and spreads over the oily surface. There is a direct relation between the diameter of oil spread and biosurfactant's concentration, and in contrast, the bacterial supernatant lacking biosurfactant cannot spread oil due to the hydrophobicity of the oil surface that causes aggregation of droplets (Gautam and Tyagi 2006). The results of our experiment indicated that the selected 10 isolates were able to spread oil and form a clear zone; then, these isolates were exposed to complementary screening which included surface emulsion activity measurements. This emulsification index is one of the important methods to support the selection of potent biosurfactant producers. In this regard, these 10 isolates were also tested for their abilities for emulsification of toluene to confirm their biosurfactant production as reported by Chen et al. (2007) who evaluate biosurfactant-producing microbes using emulsification capacity as a simple screening method. Also, Vijayanand and Divyashree (2015) recorded the emulsification index of six bacterial strains ranged between 62.3 and 92.4%.

Additionally, Arieche and Guechi (2015) reported that between the six strains, only two gave emulsification stability after 24 h with 80 and 54.4%. The emulsifier action depends on the affinity of bioemulsifier for hydrocarbon molecules which include a direct interaction with hydrocarbon itself rather than its effect on surface tension of the environment (Amiriyani et al. 2006). Extracellular metabolites produced by bacteria, viz. biosurfactants and siderophores, can precipitate heavy metals (Rajkumar et al. 2010); these compounds efficiently restrict heavy metals and detoxify them by complex formation or by forming an effective barrier surrounding the cell (Pulsawat et al. 2003). The application of biosurfactants in the remediation of heavy metals targeted the removal of these ions by chemical interactions between the amphiphiles and the metal ions. This application seems to be more environmentally compatible and more economical than using modified clay complexes or metal chelators.

The removal of heavy metal contaminants from the environment is one of the potential areas in which the usefulness of biosurfactants has not been thoroughly explored (Luna et al. 2016). In this regard, the efficiency of the biosurfactants produced by the selected isolates to remove heavy metals from wastewater was studied. Our results generally indicated that Al^{2+} and Cu^{2+} were considered as the highest and the lowest removed heavy metals by all tested isolates, respectively. Moreover, the biosurfactant produced by isolate BS37 was able to remove the highest amounts of all tested heavy metals from wastewater. The bacterial biosurfactants can capture the metal ions through electrostatic interactions or complexation. The molecular nature of biosurfactants offers the possibility of interaction with the metals in solution, aiding in their subsequent removal and/or recovery. The anionic biosurfactants create complexes with metals in a nonionic form by ionic bonds. These

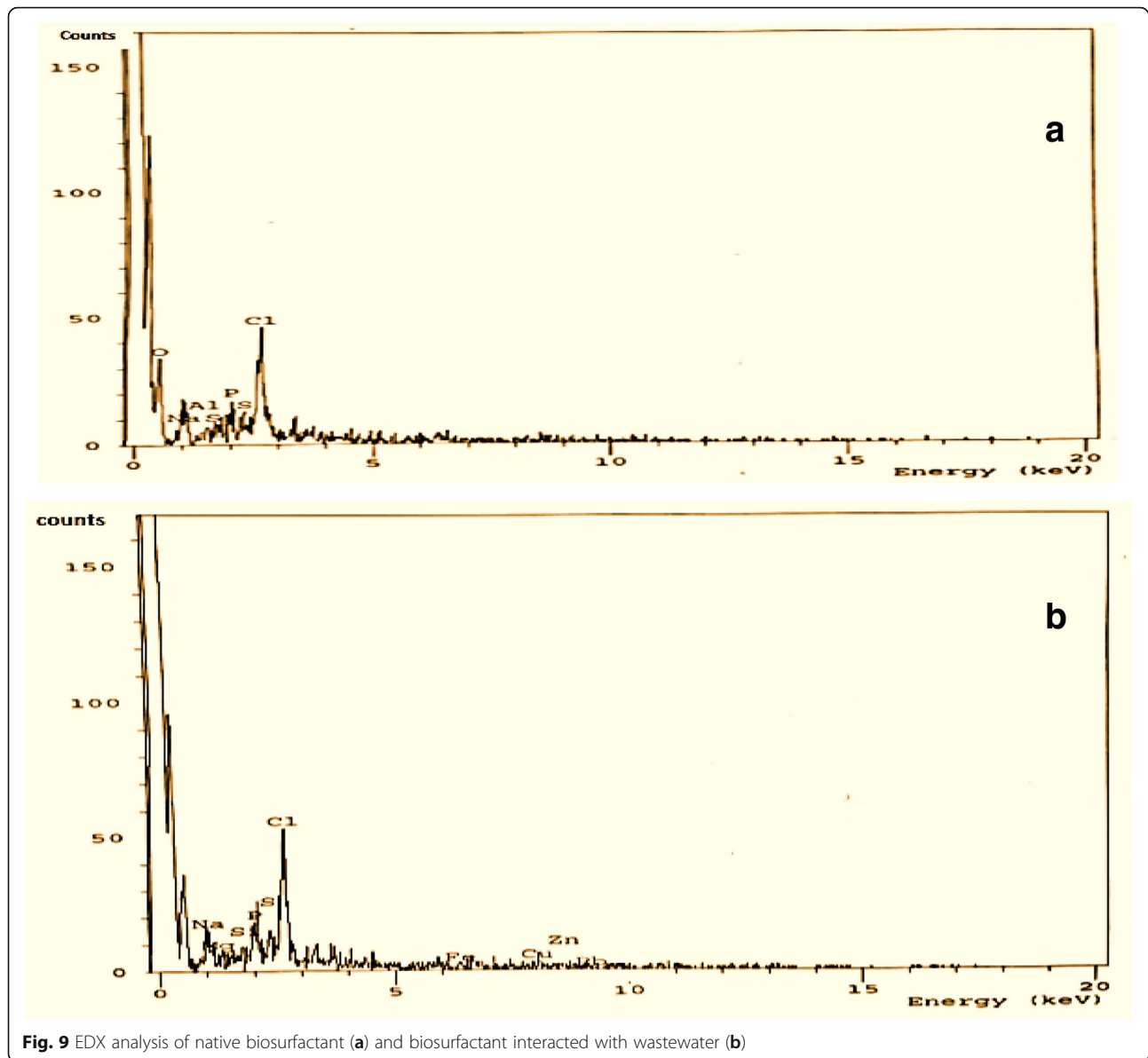


Fig. 9 EDX analysis of native biosurfactant (a) and biosurfactant interacted with wastewater (b)

bonds are stronger than the metal's bonds with the soil, and metal-biosurfactant complexes are desorbed from the soil matrix to the soil solution due to the lowering of the interfacial tension. The cationic biosurfactants can replace the same charged metal ions by competition for some but not all negatively charged surfaces (ion exchange). Metal ions can be removed from soil surfaces also by the biosurfactant micelles. The polar head groups of micelles can bind metals which mobilize the metals in water (Diaz et al. 2015).

The superior isolate BS37 showed the highest biosurfactant production and exhibited the greatest heavy metal removal efficiency which was identified according to the morphological and biochemical characterization

of colony, and its antibiotic susceptibility were tested. Its identification was accomplished by combining the alignment results of 16S rRNA sequence analysis with biochemical and physiological characteristics. The final identification of strain was *Citrobacter freundii*. Sharma and Fulekar (2009) confirmed this trend of results and documented *C. freundii* as a potential microorganism for remediation of copper. This potential organism can be used for bioremediation of heavy metals to clean up the environment (Macaskie et al. 2006). Additionally, Puranik and Paknikar (2012) and Al-Garni (2005) demonstrated *Citrobacter freundii* as a bioremoval agent for heavy metals such as lead, cadmium, and zinc from contaminated sites. *C. freundii* was able to produce biosurfactant using all the tested media, while it was observed

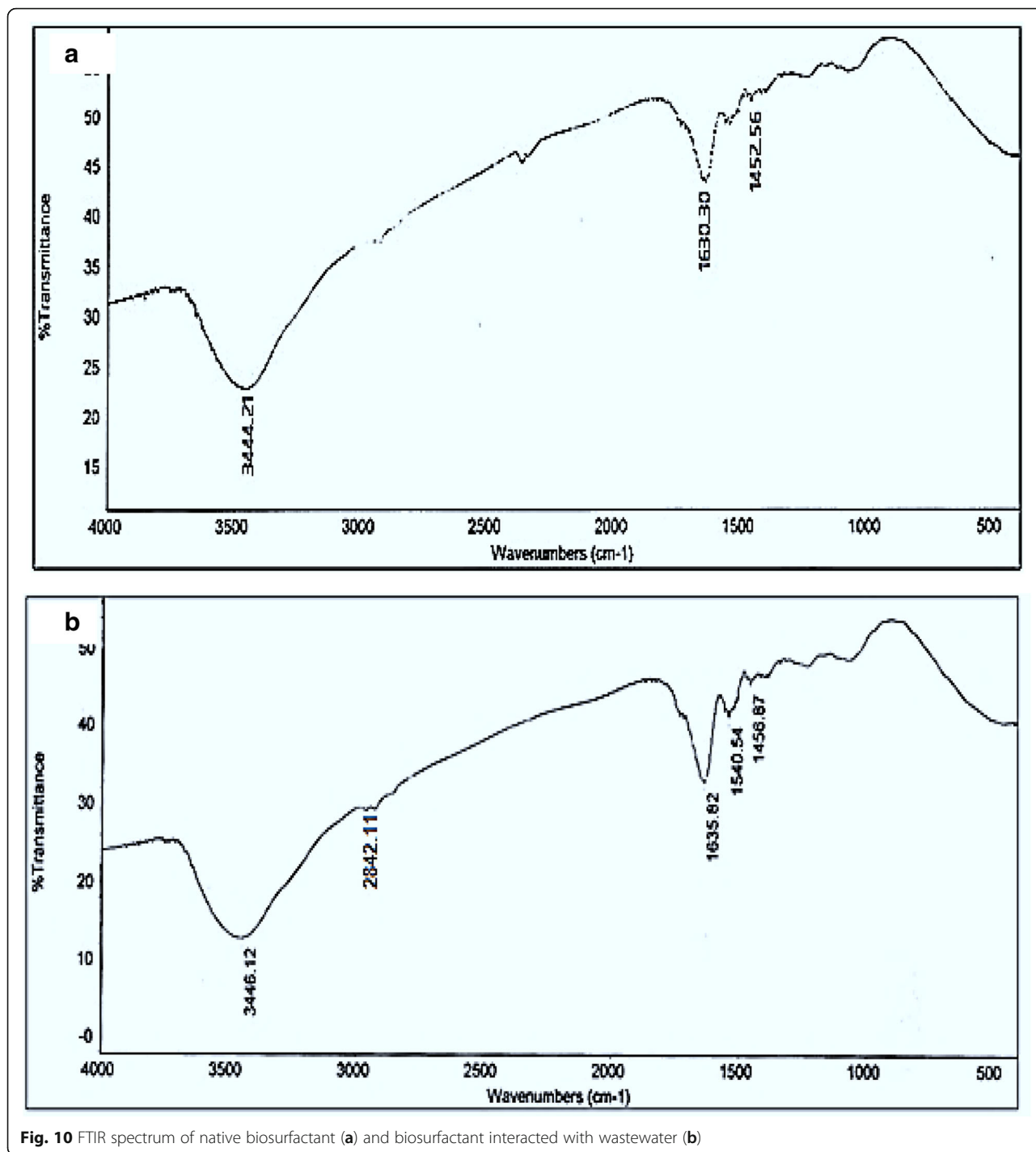


Fig. 10 FTIR spectrum of native biosurfactant (a) and biosurfactant interacted with wastewater (b)

that nutrient broth medium, inoculum potential of 10%, and incubation period of 4 days were found to yield maximal amount of biosurfactant by *C. freundii* MG812314.1. In this regard, Ibrahim (2018) reported that the biosurfactants from *C. freundii* HM-2 were produced in MSM supplemented with 2% WFO as the sole carbon source; also, this strain was able to produce biosurfactant under different incubation periods (12, 24, 36,

48, 60, 72, 84, 96 and 108 h), and it was found that the production of biosurfactant was gradually increased with the increasing of incubation period until reach maximum at 96 h then decreased.

From these results, biosurfactant has proven its ability as a washing agent in heavy metal removal, but more information is needed to optimize the process of heavy metal removal (Hidayati et al. 2014). Factors affecting

the efficiency of the produced bacterial biosurfactant for the removal of heavy metals from wastewater were investigated in this experiment. Results showed that higher removal of heavy metals was recorded when wastewater was treated with biosurfactant at 10% than 5%. This trend of results was done with all estimated heavy metals. Enhancement of metal remediation capacity with increasing amount of biosurfactant was reported by many researchers (Kim and Vipulanandan 2006; Basak and Das 2014). Results also showed higher removal efficiency with increasing contact time when wastewater was treated with 5% biosurfactant. In contrast, when treated at 10%, the removal efficiency was decreased with the increasing of the contact time. SEM images showed the morphology of native biosurfactant (Fig. 8a) and the sequestered heavy metals onto biosurfactant (Figs. 8b). The spherical nodules in SEM image (b) confirmed the anchoring of heavy metal ions with biosurfactant molecule. Basak and Das (2014) reported that removal of zinc ion from the aqueous solution was further confirmed using SEM. EDX of native biosurfactant (Fig. 9a) and heavy metals interacted biosurfactant (Fig. 9b) showed that there were constant metals in either native or interacted biosurfactant, viz. O, Na, Al, P, S, and Cl, whereas Fe^{2+} , Cu^{2+} , Zn^{2+} , and Pb^{2+} appeared in interacted biosurfactant with wastewater only. This reflected the efficiency of biosurfactant to absorb heavy metals from wastewater. These results were confirmed by those obtained by Basak and Das (2014) who reported that EDX analysis served as a direct proof of metal attachment to the micellar structure of the biosurfactant. FTIR analysis was performed for confirmation of removal of heavy metals from wastewater. FTIR spectrum of native biosurfactant (Fig. 10a) showed the presence of bands at 3444.21, 1630.30, and 1452.56 cm^{-1} . Moreover, FTIR spectrum showed that the recorded peaks at 3444.21, 1630.30, and 1452.56 cm^{-1} were shifted respectively to 3446.12, 1635.82, and 1456.87 cm^{-1} . Also, new bands appeared at 2842.11 and 1540.54 cm^{-1} (Fig. 10b). Similar results were reported by Basak and Das (2014) who recorded bands at 3439.08, 1629.85, and 1400.32 cm^{-1} in native biosurfactant produced by *Cryptococcus* sp. Additionally, Ibrahim (2018) used FTIR analysis to show the relation between the purified biosurfactants and the standard rhamnolipid. The bands recovered at wavenumber 3463 cm^{-1} and 3457 cm^{-1} were refer to the $-\text{OH}$ group (free hydroxyl groups of rhamnose rings), while, the bands recovered between 2926 and 2855 cm^{-1} due to the protraction vibrations of the $-\text{CH}_2-$ and $-\text{CH}_3$ groups of acyl chains. The bands observed at 2732 cm^{-1} could be attributed to the C–H hydrocarbon chains, while the carbonyl (ester) band was found at 1743 cm^{-1} . Also, the bands observed at 1377 and 1463 cm^{-1} confirm the presence of alkyl

groups. Additionally, he suggested that the bands recorded at 1377 cm^{-1} , 2732 cm^{-1} , and 1463 cm^{-1} may have chemical structures identical to those of glycolipids. The C–O stretching bands at 1463 and 1377 cm^{-1} demonstrated the presence of bonds between the carbon atoms and hydroxyl groups in the chemical structures of the glycoside portion. As compared to previous reports (Lan et al. 2015) these main chemical structure groups were in line with the structure characteristics of rhamnolipid (Ferhat et al. 2011; Noparat et al. 2014; Lan et al. 2015). Spectra analysis of native and interacted biosurfactant clearly indicated that the hydroxyl group ($-\text{OH}$), carboxyl ($-\text{COOH}$), and carbonyl ($\text{C}=\text{O}$) groups were the predominant contributors in heavy metal removal by biosurfactant of *C. freundii* MG812314.1. Similar results were observed by Huang and Liu (2013) who reported the involvement of hydroxyl and carboxyl groups in the removal of cadmium and lead from aqueous solution using bacterial biosurfactants.

Conclusion

One of the most promising fields of technology is the biological techniques employing biosurfactants as heavy metal removal tools. In the current study, *Citrobacter freundii* MG812314.1 isolated from Al-Rahawy drain was found to be the most potent biosurfactant producer. The produced biosurfactant exhibited a great capacity to remove the heavy metals depending on the factors like time and concentration of inoculums used. The results opened new perspectives for the use of this strain as a promising biosurfactant producer for efficient heavy metal removal. Further research on structural characterization, gene regulation, and commercial production of biosurfactant is needed to be feasibly used towards in situ bioremediation of industrial wastes.

Acknowledgements

The authors would like to thank Prof. Hamed E. Abou-Aly professor of Agricultural microbiology, Department of Agricultural Microbiology, Faculty of Agriculture, Benha University, Egypt for his valuable reviewing this manuscript.

Funding

Not applicable

Availability of data and materials

Data are available upon request from the authors.

Authors' contributions

The both authors contributed to the design and implementation of the research, the analysis of the results, and the writing of the manuscript. Also, both authors read and approved the final manuscript.

Ethics approval and consent to participate

Not applicable

Consent for publication

Not applicable

Competing interests

The authors declare that they have no competing interests.

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Received: 13 January 2019 Accepted: 12 March 2019

Published online: 24 April 2019

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