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# Nanoscale zero-valent iron-impregnated agricultural waste as an effective biosorbent for the removal of heavy metal ions from wastewater

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## Abstract

A novel, nanoscale zero-valent iron-impregnated cashew nut shell (NZVI-CNS) was synthesized towards the removal of Ni(II) ions from aqueous solution using impregnation procedure. The factors affecting Ni(II) ion adsorption in a batch mode were studied including the initial metal ion concentration, solution pH, temperature, adsorbent dosage, and contact time. The adsorption isotherm and kinetics could be described well with Freundlich and pseudo first-order, respectively. The maximum monolayer adsorption capacity for the removal of Ni(II) ions was found to be 70.05 mg/g. The calculated thermodynamic parameters showed that the removal of Ni(II) ions by the NZVI-CNS was spontaneous, feasible, and exothermic in nature. The amount of adsorbent needed to treat the known volume of the effluent was calculated by using single-stage batch adsorber design. The experimental results specifies that the NZVI-CNS have a high adsorption capacity for the removal of Ni(II) ions from aqueous solution.

**Keywords:** Adsorption, NZVI-CNS, Freundlich model, Pseudo first-order kinetics, Exothermic, Batch adsorber design

## Background

In recent decades, industrialization and urbanization have grown exponentially to meet the human needs. As a result, the environmental impact has been increasing dramatically due to the direct discharge of toxic effluent into the water bodies. The toxic substances are mostly hazardous to human health which includes heavy metals, pharmaceuticals, pesticides, and dyes. Among which, heavy metals received a special attention by many researchers because they are persistent nature due to lack of biodegradability (Martins et al. 2013; Singh and Das 2013; Gong et al. 2013). In recent years, these heavy metals play a vital role in most of the industries. Particularly, nickel is widely used in diverse metal products and processes. Few of their applications are in industries such as electroplating, batteries manufacturing, mining, metal finishing, porcelain enameling, and paint formulations. The toxicity of nickel ions will cause severe health issues such as dermatitis, allergies, renal disturbances,

hepatitis, infertility, lung cancer, stomatitis, gingivitis, insomnia, nausea, and different poisoning degrees to the kidney and cardiovascular system (Paulino et al. 2007; Fu et al. 2015; Jeon and Cha 2015). Hence, the effluents with nickel ions should be treated before letting them into the receiving water bodies. According to the US EPA, the permissible limit of nickel in wastewater effluent is 2 mg/L for short-term effluent reuse and 0.2 mg/L for long-term effluent reuse (El-Sadaawy and Abdelwahab 2014; U.S EPA 2004). According to the Bureau of Indian Standards, the maximum permissible limit for Ni(II) ions in drinking water has been fixed at 0.02 mg/L (BIS 1994). Therefore, many technologies such as chemical precipitation (Purkayastha et al. 2014), ion exchange, electrochemical technologies (Fu and Wang 2011), adsorption (Bilal et al. 2013), and membrane filtration process such as microfiltration (Tashvigh et al. 2015), reverse osmosis (Wei et al. 2014), ultrafiltration (Tanhaei et al. 2014), nanofiltration (Alzahrani and Mohammad 2014), and electrodialysis (Dermentzis 2010) have been employed for both removal and recovery of Ni(II) ion from aqueous environment. Most of these technologies

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are expensive and incompatibility to remove the trace level of heavy metal ions. Among these technologies, many researchers focused on adsorption technology which offers simplicity, technical feasibility, economical viability, and social acceptability and also got high removal efficiency of heavy metal from aqueous solution. On the other hand, the main advantage of this process is reversible, because the adsorbents used for metal ions removal can be regenerated by employing appropriate desorption process (Barakat 2011; Zhou et al. 2015; Mahmoud et al. 2015). The commercial adsorbent such as activated carbons (Rajkumar et al. 2014), polymeric materials (Anitha et al. 2015), clays (Lee and Tiwari 2012), biosorbents (Kumar et al. 2015), and other adsorbents have been commonly used for this adsorption process. Many adsorbents usually have lower adsorption capacity, higher cost, longer equilibrium time, poor regeneration abilities, and separation problems (Prabu et al. 2015, 2016). Hence, adsorbents qualifying the aforementioned problems were synthesized and employed for removal of metal ions from aqueous environment.

Nanoscale zero-valent iron (NZVI) is a promising technology because many researchers focused on this process and implemented in the treatment of hazardous contaminants from wastewater. This increase in interest on the use of NZVI is owing to its higher surface area, lower cost, non-toxic, and higher reactivity (Uzum et al. 2009; Boparai et al. 2011; Arshadi et al. 2014; Prabu et al. 2016). The removal mechanism is directional transfer of electrons from NZVI to the contaminants and then the contaminant is transformed into non-toxic or less-toxic species. In addition, it also degrades and oxidize the organic compounds in the presence of dissolved oxygen (DO) (Fu et al. 2014; Prabu et al. 2015, 2016). Mostly, direct use of NZVI is restricted due to its lack of stability, easy aggregation, and facing difficulties in separating NZVI from treated effluents. To mitigate these issues, NZVI is impregnated with supporting material such as cashew nut shell (CNS) with the help of sonication operation. This does not only provide better operation but enhances the reduction ability of NZVI (Fu et al. 2014; Prabu et al. 2016). Finally, the efficient removal of nickel ions in aqueous environment can be achieved by utilizing the synthesized nanoscale zero-valent iron-impregnated cashew nut shell (NZVI-CNS). In this research, the batch adsorption studies for the removal of Ni(II) ions were investigated by using the several parameters such as initial Ni(II) concentration, solution pH, temperature, adsorbent dose, and contact time. The adsorption isotherm, kinetics, design, and thermodynamic studies have been discussed to explain the adsorption process.

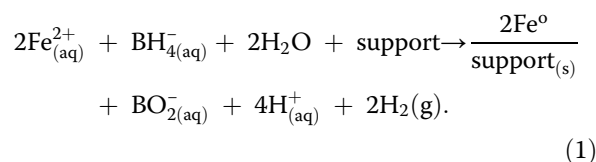
## Methods

### Chemicals and equipment

The chemical nickel(II) sulfate hexahydrate [NiSO<sub>4</sub>·6H<sub>2</sub>O] was used to prepare the stock solution of Ni(II) ions. The nickel salt was taken in a measured amount and diluted in double distilled water. To make the desired Ni(II) ions concentration solution (25–150 mg/L), the stock solution was further diluted with distilled water. The pH of the different ion concentration solution was measured with a pH meter (Elico Limited, India) and was adjusted using 0.1 N NaOH and 0.1 N HCL solutions. The concentration of Ni(II) ions onto the ions solution before and after adsorption process was measured with an atomic adsorption spectrophotometer (AAS, SL 176 Model, Elico Limited, Chennai, India).

### Preparation of the adsorbent

The adsorbent was prepared by simple liquid-phase reduction process, which is that NZVI was impregnated with the adsorbent material—CNS. One gram of CNS powder was washed with water to remove the impurities and then soaked in saturated FeSO<sub>4</sub>·7H<sub>2</sub>O solution (6.5 g in 25 mL with two drops of concentrated sulfuric acid) for 30 min. After that, the soaked CNS along with the saturated FeSO<sub>4</sub>·7H<sub>2</sub>O solution was continuously sonicated with the ultrasonic working frequency of 24 kHz and mechanically agitated with the speed of 500 rpm/min for 60 min. During the sonication process, the CNS particles were broken into small pieces under the thermodynamic condition. After the sonication, 0.1 mol/L NaBH<sub>4</sub> was slowly added to the solution at optimum pressure, temperature, and atmosphere. Finally, the ferrous ion was impregnated into the CNS particles, and the ferrous ion was reduced to ZVI as per the following reaction.



The CNS system was completely washed using distilled water followed by a wash with methanol to prevent the corrosion formation. At last, the CNS system was dried and stored in an oxygen-free nitrogen environment. This prepared material was abbreviated as NZVI-CNS (NZVI-impregnated CNS). During the synthesis of ZVI, the particles of boron were formed. These particles are toxic in nature, so special care must be needed for separating the boron particles.

### Batch adsorption studies

The batch adsorption studies were performed to optimize the parameters such as initial Ni(II) concentration,

solution pH, temperature, adsorbent dose, and contact time on the effective removal of the Ni(II) ions from the aqueous solution using the NZVI-CNS. One hundred milliliters of the Ni(II) ion solution with varying concentrations (25–150 mg/L), adsorbent dosages (0.5–3.0 g/L), solution pH (2.0–8.0), contact times (5–60 min), and temperatures (30–60 °C) was taken in 100 mL Erlenmeyer flasks and then the mixture was kept in a temperature-controlled rotary shaker. The mixture was withdrawn at different time intervals (0–60 min). The concentration of the residual metal ions in the solutions was measured using an atomic adsorption spectrophotometer (AAS) (SL 176 Model, Elico Limited, Chennai, India). The percentage removal of the Ni(II) ion was calculated by using the following equation:

$$\% \text{ Removal of metal ions} = \frac{C_o - C_e}{C_o} \times 100, \quad (2)$$

where  $C_o$  (mg/L) and  $C_e$  (mg/L) are the initial and equilibrium concentrations of Ni(II) ion solution, respectively.

#### Adsorption equilibrium experiments

The adsorption isotherm studies were performed in a series of 100 mL Erlenmeyer flasks by adding 0.2 g of adsorbent material (NZVI-CNS) into 100 mL of the Ni(II) ion solution with the various initial Ni(II) ion concentrations in the range of 25 to 150 mg/L at an optimum condition, and then, the mixtures were kept in a temperature-controlled shaking incubator until its equilibrium time. Once the equilibrium time is reached, the flasks were withdrawn from the incubation shaker and the adsorbent material NZVI-CNS was centrifuged to separate the supernatant and the spent adsorbent. The concentration of Ni(II) ion in the supernatant was examined by using atomic adsorption spectrophotometer (AAS) (SL 176 Model, Elico Limited, Chennai, India), and the values are substituted in the following formula to know the amount of Ni(II) ion adsorbed onto the NZVI-CNS at equilibrium,  $q_e$  (mg/g).

The amount of Ni(II) ions adsorbed at equilibrium  $q_e$  (mg/g)

$$q_e = \frac{(C_o - C_e)V}{m}, \quad (3)$$

where  $C_o$  and  $C_e$  are the concentrations of Ni(II) ions in the solution initially and at equilibrium (mg/g), respectively,  $V$  is the volume of the Ni(II) ion solution (L), and  $m$  is the mass of the adsorbent (g).

The two-parameter and the three-parameter isotherm models were used to explain the Ni(II) ions adsorption towards the selected adsorbent. The different parameters,

$R^2$ , and the error values are analyzed by the MATLAB R2009a software.

Langmuir (1918) and Freundlich (1906) are the two-parameter isotherm models used in this study.

The non-linear form of Langmuir adsorption isotherm model is as follows (Langmuir 1918):

$$q_e = \frac{q_m K_L C_e}{1 + K_L C_e}, \quad (4)$$

where  $q_e$  is the adsorption capacity at equilibrium (mg/g),  $q_m$  is the maximum monolayer adsorption capacity (mg/g),  $C_e$  is the concentration of the metal ions at equilibrium (mg/L), and  $K_L$  is the Langmuir equilibrium constant (L/mg).

The Freundlich adsorption model as follows (Freundlich 1906):

$$q_e = K_F C_e^{1/n}, \quad (5)$$

where  $C_e$  is the concentration of the metal ions at equilibrium (mg/L),  $K_F$  is the Freundlich constant [(mg/g)(L/mg)<sup>(1/n)</sup>] related to the bonding energy, and  $n$  is a measure of deviation from the linearity of adsorption (g/L). The significance of “ $n$ ” is as follows:  $n = 1$  (linear),  $n > 1$  (physical process), and  $n < 1$  (chemical process).

Redlich and Peterson (1959) and Sips (1948) are the three-parameter isotherm models used in this study.

The Redlich-Peterson isotherm model is given as follows (Redlich and Peterson 1959):

$$q_e = \left( \frac{K_{RP} C_e}{1 + \alpha_{RP} C_e^{\beta_{RP}}} \right), \quad (6)$$

where  $C_e$  is the concentration of metal ions at equilibrium (mg/L),  $\beta_{RP}$  is the exponent which lies between 0 and 1,  $K_{RP}$  is the Redlich-Peterson isotherm constant (L/g),  $\alpha_{RP}$  is the Redlich-Peterson isotherm constant (L/mg)<sup>1/β</sup>. The significance of  $\beta$  is as given as follows:  $\beta = 1$  (Langmuir adsorption isotherm model is a preferable adsorption isotherm model) and  $\beta = 0$  (Freundlich adsorption isotherm model is a preferable adsorption isotherm model).

The Sips adsorption isotherm model is given as follows (Sips 1948):

$$q_e = \left( \frac{K_S C_e^{\beta_S}}{1 + \alpha_S C_e^{1/\beta_S}} \right), \quad (7)$$

where  $K_S$  is the Sips model isotherm constant (L/g)<sup>β<sub>S</sub></sup>,  $\alpha_S$  is the Sips model constant (L/g)<sup>1/β<sub>S</sub></sup>, and  $\beta_S$  is often regarded as the heterogeneity factor, with values close to 1 indicating a homogeneous binding site and values greater than 1 indicating a heterogeneous adsorption system.

### Adsorption kinetic experiments

The adsorption kinetic studies were performed in a series of 100 mL Erlenmeyer flasks by varying the contact time from 0 to 60 min of the adsorbent with the different initial Ni(II) ion solutions (25–150 mg/L) along with the adsorbents and kept in a temperature-controlled shaking incubation. At predetermined time intervals, the flasks were withdrawn from the incubation shaker, the adsorption mixture was centrifuged to separate the supernatant and the spent adsorbent. The concentration of Ni(II) ion in the supernatant was examined by using atomic adsorption spectrophotometer (AAS) (SL 176 Model, Elico Limited, Chennai, India), and the values are substituted in the following formula to know the amount of Ni(II) ion adsorbed onto the NZVI-CNS at different time intervals,  $q_t$  (mg/g).

$$q_t = \frac{(C_o - C_t) V}{m}, \tag{8}$$

where  $q_t$  is the amount of Ni(II) ions adsorbed by the adsorbent at any time (mg/g),  $C_t$  is the concentration of Ni(II) ions at particular time (mg/L),  $V$  is the volume of the metal ion solution (L), and  $m$  is the mass of the adsorbent (g).

The pseudo first-order, pseudo second-order, and Elovich kinetic models are used to determine the rate of the adsorption process.

Pseudo first-order kinetic model is given as follows (Lagergren 1898):

$$q_t = q_e(1 - \exp(-k_1t)), \tag{9}$$

where  $q_t$  is the adsorption capacity at any time (mg/g),  $q_e$  is the equilibrium adsorption capacity (mg/g),  $k_1$  is the pseudo first-order kinetic rate constant ( $\text{min}^{-1}$ ), and  $t$  is the time (min).

Pseudo second-order kinetic model is given as follows (Ho and McKay 1999):

$$q_t = \frac{q_e^2 k_2 t}{1 + q_e k_2 t}, \tag{10}$$

where  $k_2$  is the pseudo second-order kinetic rate constant (g/mg min) and  $t$  is the time (min).

Elovich kinetic model is given as follows (Low 1960):

$$q_t = (1 + \beta_E) \ln(1 + \alpha_E \beta_E t), \tag{11}$$

where  $\beta_E$  is the desorption constant related to the activation energy of chemisorption (g/mg) and  $\alpha_E$  is the initial adsorption rate (mg/(g min)).

### Adsorption thermodynamic study

The adsorption thermodynamics studies were performed by varying the temperature of the batch adsorption system from 303 to 333 K. The adsorption process is

explained by the thermodynamic parameters like Gibbs free energy ( $\Delta G^\circ$ ,  $\text{kJ mol}^{-1}$ ), enthalpy ( $\Delta H^\circ$ ,  $\text{kJ mol}^{-1}$ ), and entropy ( $\Delta S^\circ$ ,  $\text{kJ mol}^{-1}$ ) which was calculated from the following equations:

$$K_c = \frac{C_{Ae}}{C_e} \tag{12}$$

$$\Delta G^\circ = -RT \ln K_c \tag{13}$$

$$\Delta G^\circ = \Delta H^\circ - T \Delta S^\circ \tag{14}$$

$$\log K_c = \frac{\Delta S^\circ}{2.303R} - \frac{\Delta H^\circ}{2.303RT}, \tag{15}$$

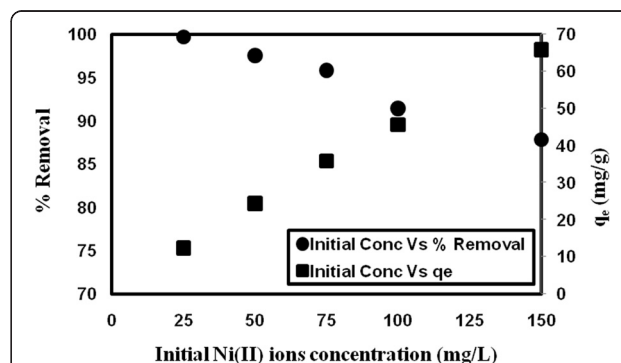
where  $C_{Ae}$  is the amount of Ni(II) ions adsorbed onto the adsorbent per liter of solution at equilibrium (mg/L),  $C_e$  is the equilibrium concentration in solution (mg/L),  $T$  is the temperature (K),  $R$  is the gas constant (8.314 J/mol/K), and  $K_c$  is the equilibrium constant. The values of  $\Delta S^\circ$  and  $\Delta H^\circ$  were calculated from the slope and the intercept of the plot of  $\log K_c$  versus  $1/T$ .

## Results and discussions

### Equilibrium studies

#### *Influence of initial Ni(II) ion concentration on the adsorption of Ni(II) ions and adsorption isotherms*

The effect of initial Ni(II) ions concentration on the adsorption of Ni(II) ions is the main important parameter for resolving the adsorption capacity of the adsorbent. In this study, the effect of Ni(II) ion concentration was studied with the range of nickel ion concentration = 25–150 mg/L, solution pH = 5.0, adsorbent dose = 2 g/L, volume of sample = 100 mL, equilibrium time = 30 min, temperature 30 °C, and the results were shown in Fig. 1. The experimental results showed that the percentage removal of Ni(II) ion was decreased from 99.72 to 87.89 % with an increase in initial Ni(II) ion concentration from 25 to 150 mg/L. Adsorption is a surface phenomenon,



**Fig. 1** Effect of initial Ni(II) ion concentration on the removal of Ni(II) ions by NZVI-CNS (nickel ion concentration = 25–150 mg/L, solution pH = 5.0, adsorbent dose = 2 g/L, volume of sample = 100 mL, equilibrium time = 30 min, temperature 30 °C)

the surface saturation of metal ions strongly depends on the initial metal ion concentrations. The initial concentration affords an essential driving force to conquer all mass transfer resistance of metal ions between the liquid phase and the solid phase. At low metal ion concentration, the availability of surface active sites might be high, i.e., the ratio of active sites to the metal ions in the aqueous solution is higher, therefore all the metal ions may be occupied on the active sites of the adsorbent (Li et al. 2013). The drop off of removal percentage might be due to the saturation of available active sites at higher concentration. The adsorption capacity of NSZV-CNS was increased from 12.46 to 65.92 mg/g, due to the higher adsorption rate and the exploitation of all available binding sites for adsorption at higher Ni(II) ion concentration.

The obtained equilibrium data from the effect of Ni(II) ion concentration studies were evaluated by using several adsorption isotherm model which can be used for the design of an adsorption system. In this study, Langmuir (Langmuir 1918), Freundlich (Freundlich 1906), Redlich-Peterson (Redlich and Peterson 1959), and Sips (Sips 1948) isotherm model were used to calculate the maximum monolayer adsorption capacity of NSZV-CNS and the equilibrium data for the adsorption of Ni(II) ions from aqueous solution; the results were shown in Fig. 2. The Langmuir constant ( $K_L$ ) and the monolayer adsorption capacity ( $q_m$ ), the Freundlich isotherm constant ( $K_L$ ) and  $n$ , the Redlich-Peterson isotherm constant ( $K_{RP}$  and  $\alpha_{RP}$ ) and  $\beta_{RP}$  and the Sips model constant ( $K_S$  and  $\alpha_S$ ) and  $\beta_S$  were determined using the plot  $C_e$  and  $q_e$ . The adsorption isotherm parameters, determination of coefficient values ( $R^2$ ), and error values (sum of squared errors (SSE), root mean squared values (RMSE)) were attained by fitting the adsorption isotherm data to the adsorption isotherm models are shown in Table 1. As it can be seen in Table 1, Freundlich adsorption isotherm models have high correlation coefficient value ( $R^2 = 0.9901$ ) and low error values (SSE = 27.77, RMSE = 2.635) compared to other isotherm

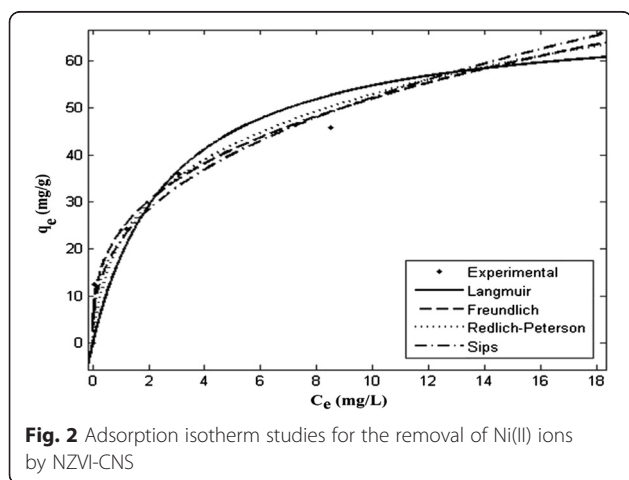


Fig. 2 Adsorption isotherm studies for the removal of Ni(II) ions by NSZV-CNS

models. The adsorption isotherm data for the removal of Ni(II) ions onto the NSZV-CNS was constantly fitted with the Freundlich isotherm model which indicates that the adsorption process was the heterogeneous nature of active sites on the surface of the NSZV-CNS, a multilayer adsorption occurred between the adsorbent material NSZV-CNS and the Ni(II) ions in the liquid solution (Jiang et al. 2012). The values of  $n$  that were determined from the Freundlich isotherm model were found to be 2.956 for NSZV-CNS; the value of  $n$  lies between 1 and 10, which indicates that the adsorption of Ni(II) ions onto the NSZV-CNS was physical adsorption. Based on the correlation coefficient values, the experimental data gave estimable fits within the following isotherm order: Freundlich > Sips > Redlich-Peterson > Langmuir.

**Influence of solution pH on the adsorption of Ni(II) ions**

The pH value of the aqueous solution plays a significant role in the adsorption process which affects the binding sites and chemical interaction between the metal ions and the adsorbents via changing the ionization state of the functional groups present on the surface of the adsorbent. In order to find out the optimum pH, the effect of solution pH for the removal of Ni(II) ions by NSZV-CNS was deliberated over the range of solution pH = 2.0–8.0, adsorbent dose = 2 g/L, volume of sample =

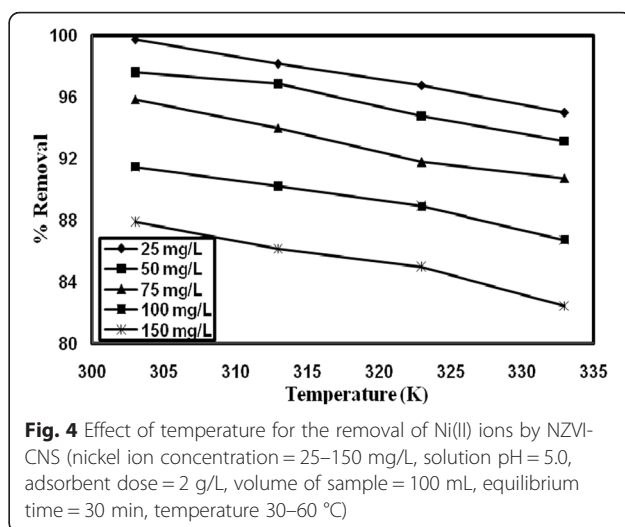
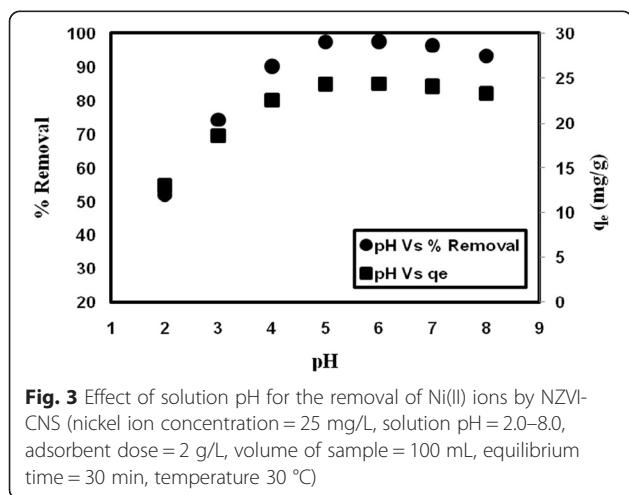
Table 1 Isotherm constants of the two-parameter and three-parameter models for the removal of Ni(II) ions by NSZV-CNS

	Parameters	Values	$R^2$
Two-parameter models			
Langmuir model	$q_m$ (mg/g)	70.05	0.9276
	$K_L$ (L/mg)	0.3577	
	SSE	203.4	
	RMSE	7.131	
Freundlich model	$K_F$ [(mg/g)(L/mg) <sup>(1/n)</sup> ]	23.88	0.9901
	$n$ (g/L)	2.956	
	SSE	27.77	
	RMSE	2.635	
Three-parameter models			
Redlich-Peterson model	$K_{RP}$ (L/g)	64.64	0.9613
	$\alpha_{RP}$ (L/mg) <sup>(1/<math>\beta_{RP}</math>)</sup>	2.011	
	$\beta_{RP}$	0.7487	
	SSE	108.7	
	RMSE	5.214	
Sips model	$K_S$ (L/g) <sup><math>\beta_S</math></sup>	21.76	0.9848
	$\alpha_S$ (L/mg) <sup>(<math>\beta_S</math>)</sup>	1.798	
	$\beta_S$	0.3806	
	SSE	40.31	
	RMSE	3.174	

100 mL, equilibrium time = 30 min, temperature 30 °C with a nickel ions concentration = 25 mg/L. Figure 3 shows that the percentage removal of Ni(II) ion was increased with the increase of pH value from 2.0 to 5.0. Beyond the pH value, the adsorption capacity remains constant due to the occurrence of hydroxyl ions which causes the arrangement of metal hydroxide complexes, which decline the adsorption capacity of NZVI-CNS. Figure 3 shows that the maximum removal of Ni(II) ions was attained at the solution pH of 5.0 which indicated that the adsorption capacity of NZVI-CNS was higher at acidic condition. At acidic condition due to the electrostatic interaction, the negatively charged Ni(II) ions was adsorbed by the positively charged adsorbent material NZVI-CNS.

**Influence of temperature on the adsorption of Ni(II) ions and thermodynamic studies**

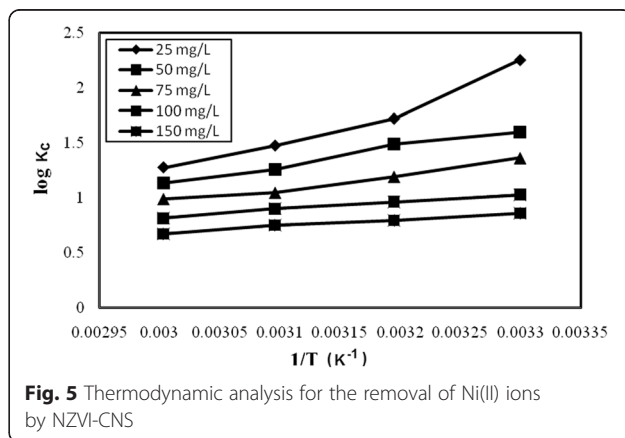
The effect of temperature for the removal of Ni(II) ions onto the NZVI-CNS was studied by allowing the Ni(II) ion concentration = 25–150 mg/L, solution pH = 5.0, adsorbent dose = 2 g/L, volume of sample = 100 mL, equilibrium time = 30 min, temperature 30–60 °C, and the results were shown in Fig. 4. The results showed that the percentage removal of Ni(II) ions was decreased with an increase in the temperature from 30 to 60 °C for the Ni(II) ion concentration (25–150 mg/L), which indicates that the adsorption process was exothermic in nature. At higher temperature, the kinetic energies of Ni(II) ions was higher. Consequently, the adsorption of high energy metal ions is not easier compared with low energy ions. This is probably due to the fading of adsorptive forces between the metal ions and the binding sites on the surface of the adsorbent material—NZVI-CNS, which means that the adsorption process is inversely proportional to the temperature (Karami 2013). The thermodynamic behaviors of the adsorption of Ni(II) ions onto



the NZVI-CNS were analyzed by plotting the graph between  $\log K_c$  and  $1/T$  (Fig. 5), and the calculated thermodynamic parameters such as Gibbs free energy ( $\Delta G^\circ$ ), enthalpy change ( $\Delta H^\circ$ ), and entropy change ( $\Delta S^\circ$ ) for the adsorption of Ni(II) ions are shown in Table 2. It was found that the negative value of Gibbs free energy was increased with an increase the temperature from 30 to 60 °C which confirms that the adsorption of Ni(II) ions onto the NZVI-CNS was more positive at low temperature and the adsorption process was feasible and spontaneous in nature (Li et al. 2013). The negative value of changes in enthalpy ( $\Delta H^\circ$ ) confirm that the adsorption of Ni(II) ion onto the NZVI-CNS was exothermic in nature. The negative value of changes in entropy ( $\Delta S^\circ$ ) indicated that the adsorption of Ni(II) ion onto the NZVI-CNS was enthalpy driven.

**Influence of adsorbent dosage on the adsorption of Ni(II) ions**

Temperature is the significant factor that can influence the adsorption process. The effect of temperature for the



**Table 2** Thermodynamic parameters for the adsorption of Ni(II) ions onto NZVI-CNS

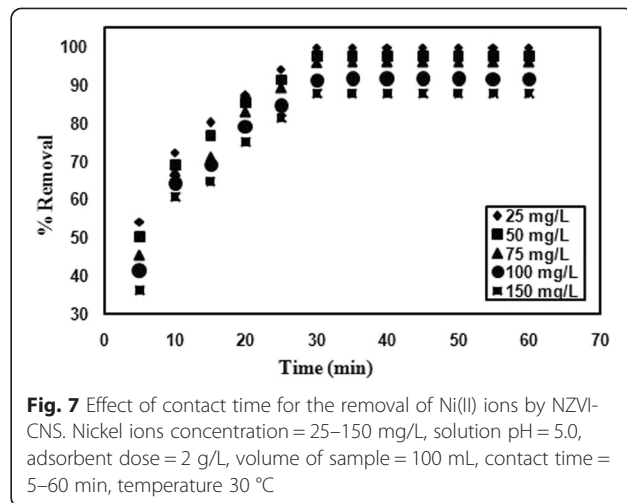
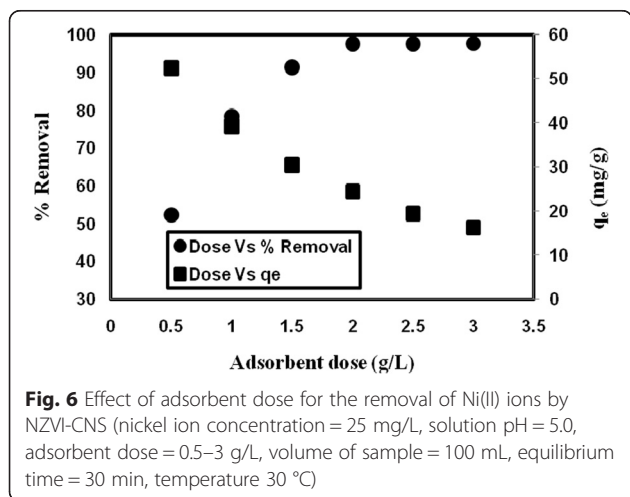
Metal ions	Concentration of Ni(II) ion solution (mg/L)	$\Delta H^\circ$ (kJ/mol)	$\Delta S^\circ$ (J/mol/K)	$\Delta G^\circ$ (kJ/mol)			
				30 °C	40 °C	50 °C	60 °C
Ni(II)	25	-61.998	-162.94	-14.846	-10.327	-9.134	-8.143
	50	-31.688	-73.52	-9.300	-8.936	-7.790	-7.226
	75	-24.699	-55.76	-7.920	-7.152	-6.479	-6.312
	100	-13.630	-25.12	-5.979	-5.782	-5.586	-5.200
	150	-11.719	-22.15	-4.994	-4.761	-4.653	-4.288

removal of Ni(II) ion was studied by using the optimum parameters such as Ni(II) ion concentration = 25 mg/L, solution pH = 5.0, volume of sample = 100 mL, equilibrium time = 30 min, temperature 30 °C, and the varied adsorbent dosage of 0.5–3 g/L, and the results were shown in Fig. 6. From Fig. 6, it was observed that the percentage removal of Ni(II) ions was increased with an increase in the adsorbent dosage from 0.5 to 3 g/L. This may be possible due to the increase in available binding sites of the adsorbent with the increase in adsorbent dosage (Fan et al. 2008). The maximum percentage removal was attained at 2 g/L, beyond that the removal percentage was almost constant. The equilibrium adsorption capacity ( $q_e$ ) was decreased with the increase in the adsorbent dosage. The reason for this may be due to the more availability of binding sites and increase in the surface area at lower adsorbent dose; therefore, the metal uptake capacity will be high at lower dosage. Conversely, overlying or aggregation of the adsorption sites at higher dosage (Lata et al. 2008).

**Influence of contact time on the adsorption of Ni(II) ions and kinetic studies**

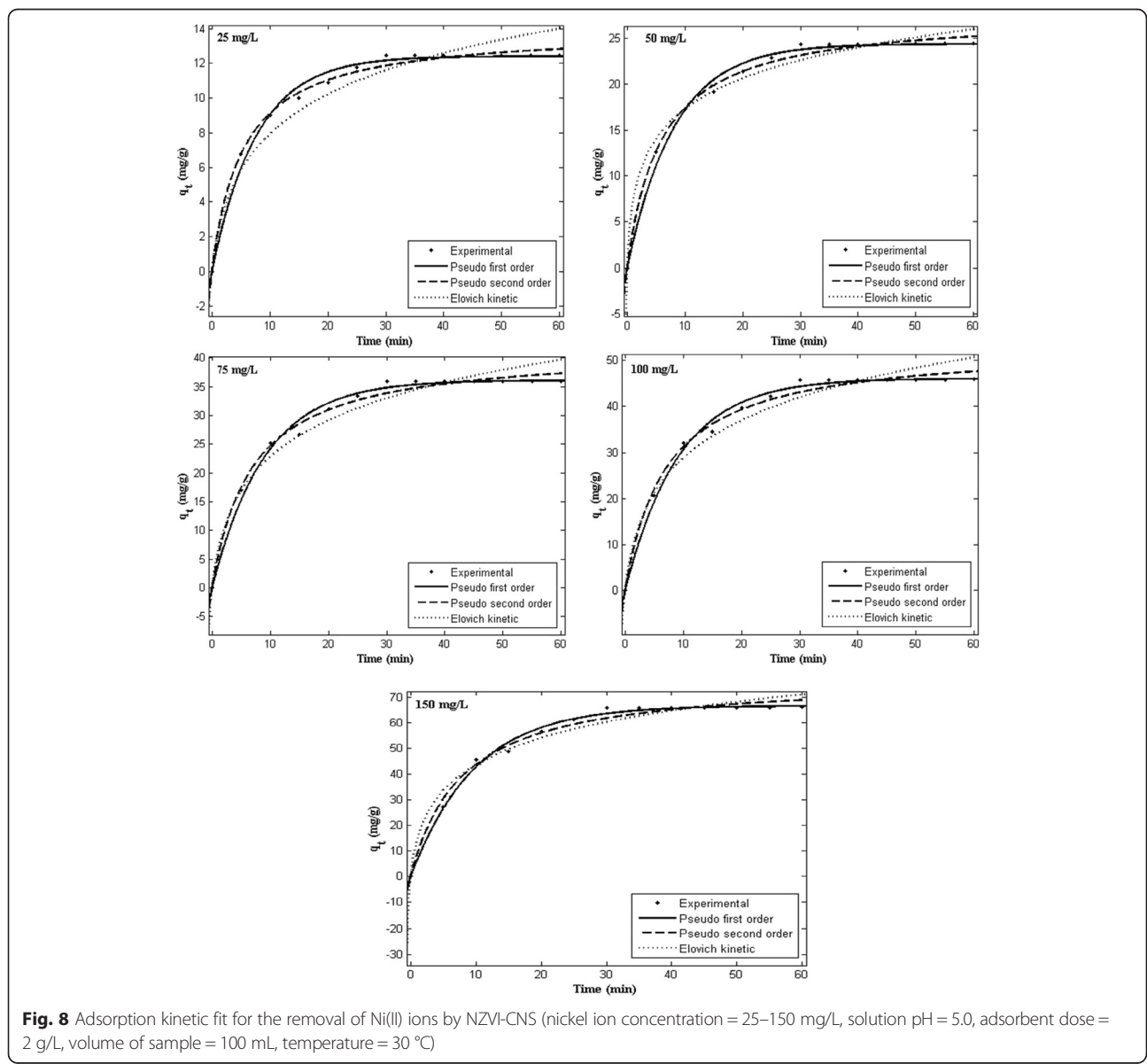
The effect of contact time for the removal of Ni(II) ions onto the NZVI-CNS was studied with the range of Ni(II)

ion concentration = 25–150 mg/L, solution pH = 5.0, adsorbent dose = 2 g/L, volume of sample = 100 mL, temperature 30 °C, and contact time = 5–60 min, and the results were shown in Fig. 7. From the Fig. 7, it was observed that the percentage removal of Ni(II) ion was increased with an increase in the contact time from 5 to 60 min. The result affirmed that the percentage removal of Ni(II) ions onto the NZVI-CNS was increased quickly at the first 10 min and then increased gradually and reached adsorption equilibrium at 30 min. The rapid increase of the adsorption of Ni(II) ions at the initial stage of contact time due to the accessibility of huge amount of binding sites on the surface of the adsorbent material. At lesser contact time, the residential time offered for the adsorption of Ni(II) ion onto the NZVI-CNS is less, for that reason, the percentage removal of Ni(II) ion is less. At higher contact time, the rate of adsorption slowly increased until equilibrium, this might be possibly due to the destabilization of driving force and the available binding sites were gradually tapered and taking long time to reach the adsorption equilibrium. The adsorption mechanism and the rate at which the Ni(II) ion was removed from the aqueous solution were estimated by using adsorption kinetic models. In these study, three kinetic models such as pseudo first-order (Lagergren



1898), pseudo second-order (Ho and McKay 1999), and Elovich kinetic model (Low 1960) were used to examine the adsorption kinetics for the removal of Ni(II) ion onto the NZVI-CNS, and the results were shown in Fig. 8. The estimated parameters such as pseudo first-order kinetics constant ( $k_1$ ) and  $q_e$ , the pseudo second-order kinetic constant ( $k_2$ ) and  $q_e$ , and the Elovich kinetic model constant ( $\alpha_{RP}$ ) and  $\beta_{RP}$  are listed in Table 3. The comparison was made between the calculated adsorption capacity ( $q_e$ , cal) and the experimental adsorption capacity ( $q_e$ , exp). It can be seen from Table 3 that the pseudo first-order kinetic were very closer to the experimental adsorption capacity value ( $q_e$ , exp) compared with other kinetic

models. Additionally, the determination of correlation coefficient values for the pseudo first-order kinetic was higher compared with pseudo second-order and Elovich kinetic model. The adsorption kinetic data for the removal of Ni(II) ions onto the NSZV-CNS were constantly fitted with pseudo first-order kinetic model which indicates that the physical adsorption is the rate-determining step (Aliabadi et al. 2013). The prediction of the rate-limiting step is important to understand the adsorption mechanism. As it can be seen in Table 2, the values of rate constant “ $k$ ” decreases as the concentration of Ni(II) ion increased; this might be possibly due to the lower competition for the binding sites at lower concentration. At higher





**Table 3** Adsorption kinetic data for the removal of Ni(II) ions by NZVI-CNS

Adsorption kinetic model	Parameters	Values (mg/L)				
		25	50	75	100	150
Pseudo first-order	$q_e$ (mg/g)	12.41	24.39	36.13	46.06	66.66
	$q_{e, \text{exp}}$ (mg/g)	12.47	24.49	36.14	45.95	65.99
	$k_1$ ( $\text{min}^{-1}$ )	0.1305	0.1221	0.1105	0.1089	0.1019
	$R^2$	0.9941	0.9934	0.991	0.9935	0.9933
	SSE	1.654	5.199	13.02	14.52	32.25
	RMSE	0.3877	0.6875	1.088	1.149	1.712
Pseudo second-order	$q_{e, \text{cal}}$ (mg/g)	13.97	27.68	41.56	53.19	77.84
	$k_2$ (g/mg min)	0.0135	0.006159	0.00353	0.00267	0.00164
	$R^2$	0.9912	0.9907	0.99	0.9907	0.9871
	SSE	0.9014	3.86	13.63	20.74	62.02
	RMSE	0.2863	0.5924	1.113	1.373	2.374
Elovich kinetic model	$\alpha_E$ mg/(g min)	1.478	2.958	0.272	0.181	0.2491
	$\beta_E$ (g/mg)	0.5539	1.128	3.303	4.596	5.872
	$R^2$	0.9378	0.9793	0.962	0.9608	0.9679
	SSE	9.666	12.59	51.87	87.75	154.1
	RMSE	0.8975	1.07	2.079	2.704	3.742

concentration, the competition for the binding sites will be high and accordingly lower adsorption rates are attained (Gao et al. 2013).

**Design of a single-stage batch adsorber**

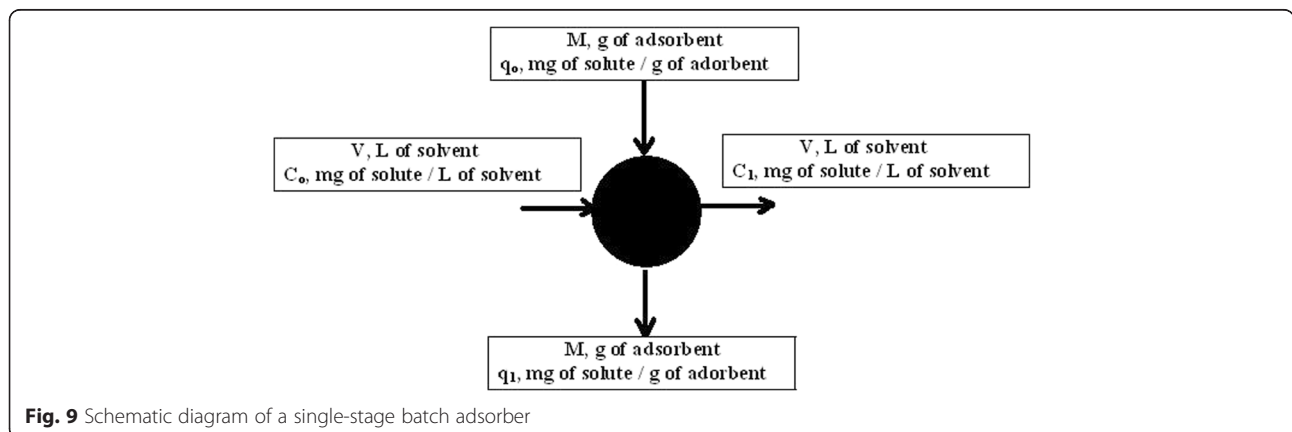
The best fitted adsorption isotherm model was used to design the single-stage batch adsorber for the removal of Ni(II) ions from aqueous solution. A schematic diagram of a single-stage batch adsorber is shown in Fig. 9. The main objective of the adsorber was to estimate the amount of adsorbent needed for treating the known volume and concentration of Ni(II) ions from the aqueous solution at optimum condition. The mass balance for the single-stage batch adsorption system at equilibrium condition can be given as follows (at time  $t = 0$ ,  $q_0 = 0$ ):

$$V(C_o - C_e) = M(q_e - q_o), \tag{16}$$

where  $M$  is the mass of the adsorbent (g),  $V$  is the volume of Ni(II) ion solution (L),  $C_o$  is the initial Ni(II) ion concentration (mg/L),  $C_e$  is the Ni(II) ion concentration at equilibrium (mg/L),  $q_e$  is the equilibrium adsorption capacity (mg/g), and  $q_o$  is the adsorption capacity at time  $t = 0$  (mg/g). In this study, the fresh adsorbent was used for the adsorption system, therefore  $q_o = 0$ , rearranging Eq. 16 gave

$$M = \frac{(C_o - C_e)}{q_e} V. \tag{17}$$

The Freundlich adsorption isotherm model was best fitted for the present adsorption system; this model was included in Eq. 17 and the equation can be written as



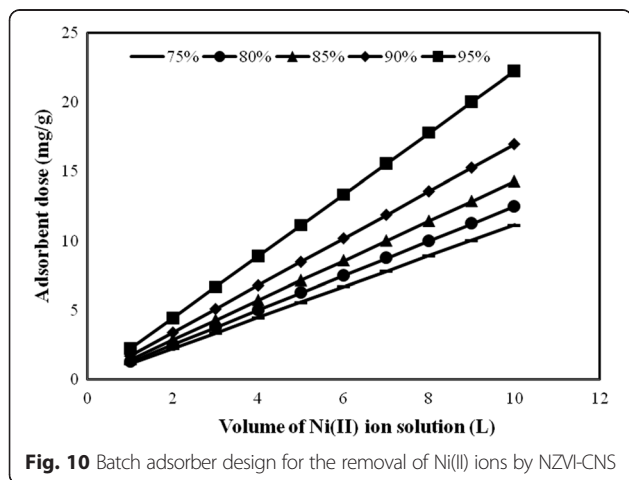
**Fig. 9** Schematic diagram of a single-stage batch adsorber

$$M = \frac{(C_o - C_e)}{K_F C_e^{1/n}} V. \quad (18)$$

Figure 10 shows the plot of mass of adsorbent versus volume of the solution of initial Ni(II) ion concentration of 75 mg/L for 75, 80, 85, 90, and 95 % of Ni(II) ion removal at different solution volumes (1–10 L) for a single-stage batch adsorber. This design was used to calculate the amount of adsorbent dose needed for the treatment of known volume of the contaminated water. For example, the amount of optimum adsorbent dose needed to treat 10 L of the Ni(II) ion solution with 95 % removal of 75 mg/L Ni(II) ion solution was found to be 22.27 g of adsorbent.

## Conclusions

In summary, we reported that the NZVI-CNS adsorbent material was successfully synthesized through the impregnated method which have been effectively used for the removal of Ni(II) ions from aqueous solution. The adsorption was studied kinetically using diverse adsorption kinetic models such as pseudo first-order, pseudo second-order, and Elovich kinetic models. The results showed that the adsorption process followed the pseudo first-order model based on higher correlation coefficient with low error values. The adsorption of Ni(II) ions onto NZVI-CNS was influenced by several operating parameters such as initial Ni(II) ion concentration, solution pH, contact time, temperature, and adsorbent loading. The maximum monolayer adsorption capacity for the removal of Ni(II) ions was found to be 70.05 mg/g at an optimum condition. The equilibrium isotherm has been analyzed using different isotherm models such as Langmuir, Freundlich, Redlich-Peterson, and Sips model. Among these, the Freundlich isotherm model was identified as a well-suitable adsorption isotherm model based on the high coefficient of determination values with low error values. The adsorption thermodynamics illustrated



**Fig. 10** Batch adsorber design for the removal of Ni(II) ions by NZVI-CNS

that the electrostatic interaction plays an important role in the adsorption process. The values of thermodynamic parameters such as Gibbs free energy, entropy, and enthalpy indicated that the adsorption process was spontaneous, feasible, and exothermic in nature. A batch adsorber was designed to determine the amount of adsorbent dose needed for the treatment of a known volume of the desired Ni(II) ion concentration. Based on the experimental studies, it can be concluded that the NZVI-CNS materials has more potential for the removal of Ni(II) ions, and correspondingly for many other heavy metal ions from aqueous solution, with the advantages of being low cost, recyclable, efficient, and economically friendly.

## Competing interests

The authors declare that they have no competing interests.

## Authors' contributions

PSK have made substantial contributions to the conception and design, or acquisition of data, or analysis and interpretation of data and also have given final approval of the version to be published. AS, PSR, and MY have been involved in drafting the manuscript and revising it critically for important intellectual content. All authors read and approved the final manuscript.

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