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Migration routes, population status and important sites used by the globally threatened Black-faced Spoonbill (*Platalea minor*): a synthesis of surveys and tracking studies

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Abstract

Background: The Black-faced Spoonbill (*Platalea minor*) is a globally threatened species, nesting mainly in western Korea with smaller numbers breeding in Liaoning Province, China, and Far East Russia. Recent winter field surveys to estimate the species' population size were almost totally conducted in coastal areas, but tracking studies showed that some individuals now winter inland. To ensure its long-term survival, we need a more comprehensive assessment of the current distribution and abundance of the species.

Methods: We combined the most recent count data and satellite tracking information to update existing information about the population abundance and distribution of the Black-faced Spoonbill at all stages of its annual life cycle, and how these have changed during 2004–2020.

Results: Black-faced Spoonbills mainly breed on the west coast of the Korean peninsula, while immature birds show a wider summer distribution throughout Yellow Sea coastal areas, when a few remain on wintering sites in the south. Combined tracking results and mid-winter counts confirmed known wintering sites on the east and south coasts of China, but showed that the species also winters on wetlands in the Yangtze River floodplain and in Southeast Asia. During 2004–2020, counts of wintering birds in coastal habitats increased from 1198 to 4864, with numbers wintering on the island of Taiwan contributing most to the overall increase. Latest counts found 5222 in 2021. We also identify key wintering and stopover sites as well as their current conservation status.

Conclusions: This study revised the known summering and wintering ranges of the Black-faced Spoonbill and assessed the conservation status of key sites based on a combination of field survey and satellite tracking data. We recommend prioritisation of further field research to identify and survey inland wintering areas in the Yangtze River floodplain and summering areas of immature birds. More tracking of adult individuals and birds during spring migration is necessary to fill these information gaps. We also suggest establishing a Black-faced Spoonbill monitoring platform to store, share and show real-time distribution range and population abundance data.

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Keywords: Black-faced Spoonbill, Breeding distribution, Coastal mudflats, Conservation status, *Platalea minor*, Population abundance, Winter distribution, Yangtze River floodplain

Background

The Black-faced Spoonbill (*Platalea minor*) is a globally threatened waterbird, which breeds mainly on the west coast of the Korean Peninsula (Kang et al. 2016), with small numbers breeding in Liaoning Province, China (Ding et al. 2000) and Far East Russia (Shibaev 2010). It winters in coastal areas of south and southeast Chinese mainland, on Taiwan and Hainan in China, western Japan, and Southeast Asia (BirdLife International and Handbook of the Birds of the World 2019), constituting just one biogeographical population (Wetlands International 2020).

Black-faced Spoonbills were considered relatively common prior to the 1950s, but the population declined during the Korean War, because of the destruction of breeding sites and subsequent pollution of habitats, thought to be caused by pesticide use (Yeung et al. 2006). In 1980s, the global population numbered less than 300 individuals (Kennerley 1989), and the species was designated as “Critically Endangered” by International Union for Conservation of Nature (IUCN) in 1994 (BirdLife International and Handbook of the Birds of the World 2017). Conservation efforts, including considerable investment in site protection during 1997–2014, increased the population of Black-faced Spoonbills from 535 to 2726 individuals (Sung et al. 2018), and the threatened category was down-listed to “Endangered” in 2000 (BirdLife International and Handbook of the Birds of the World 2017). During the 2021 global synchronous census coordinated by the Hong Kong Bird Watching Society (HKBWS), the number of wintering Black-faced Spoonbills reached a new maximum of 5222 individuals (Yu et al. 2021).

Besides the census, the first satellite tracking research was also initiated in 1998 in order to locate previously undiscovered breeding grounds, migration routes and stopover sites (Ueta et al. 2002) and further studies were undertaken in the 2010s (Wood et al. 2013; Jung et al. 2018; Jia et al. 2020; Son et al. 2020). Results from these latter studies revealed that immature individuals could remain on inland habitats during winter, indicating that the midwinter count focusing almost entirely on coastal areas could miss some individuals wintering inland.

Black-faced Spoonbills are sensitive to human disturbance (Ding et al. 2000; Zhang et al. 2010), while climate change and habitat loss together also threaten the recovery of the population (Pickett et al. 2018). Although recent studies have revealed valuable information about

population size and trends, distribution, and the migration routes of the Black-faced Spoonbills, we still need more knowledge relating to the important sites used by the species throughout the annual cycle and their conservation status. It is especially important for us to (1) understand the distribution of the Black-faced Spoonbill throughout the annual cycle, and (2) confirm whether current levels of site protection are adequate to secure the population for the future. Here, we attempt to provide new information in the current analysis to fulfil these two objectives.

Methods

Information sources

Count data

We derived census information from recent waterbird field surveys and from published information to establish the full distributional range and abundance of the Black-faced Spoonbill at different points in its annual cycle. The information included figures from the synchronized surveys conducted by the Hong Kong Bird Watching Society (HKBWS) during 2004–2020 (Sung et al. 2018; Yu et al. 2020), synchronized Yangtze waterbird surveys in 2004, 2005, 2016, 2018, 2019 and 2020 (Cao et al. 2010), and records in published scientific papers. The synchronized HKBWS surveys presented only count figures from major wintering sites (which are known to comprise at least 90% of the total numbers recorded during the census) from 2004 to 2020. Synchronized Yangtze waterbird surveys were undertaken during mid-January to early February, which are considered to cover the majority of the possible wintering sites of Black-faced Spoonbills in inland China (Yu and Swennen 2005; Cao et al. 2010). Other field survey results and satellite tracking results were obtained from published articles by searching the internet (www.cnki.net; apps.webofknowledge.com; sci-hub.org.cn) with key words (“black-faced spoonbill” or “black-faced spoonbills” or “*Platalea minor*”) (Additional file 1: Table S1).

Telemetry data

Telemetry data analyzed and presented here from all available studies derive from 34 different tagged individuals. These include 24 complete autumn migration routes from 24 birds caught at Chinese and South Korean breeding sites and 10 complete spring migration routes from 10 birds caught at two key wintering sites on Taiwan and Hong Kong in China (see Additional file 1: Table S2 for

detail information on individuals and Additional file 1: Table S3 for data summaries from tracking studies).

Analysis

We created new distribution maps based on the information from recent field surveys and tracking studies, overlaying these upon the distribution map generated by BirdLife International and Handbook of the Birds of the World (2019) within an ArcGIS 10.6 environment.

We used the telemetry results and data from field surveys on the wintering areas (Jin et al. 2009), breeding areas (Lee et al. 2001; Kang et al. 2016) and stopover areas (Yuan and Zhang 2006) to define different periods of the annual cycle (see “Results” below). Based on the information from the literature, we assigned birds using sites during the defined wintering period (December–February, see “Results” below) to wintering sites. Breeding sites were defined based upon direct observation of birds’ breeding behavior or movements to and from colonies interpreted from tracking data. Other sites occupied by birds during the period defined as summer (between May and September; see “Results” below), in the absence of evidence of breeding were noted as summering sites. We also present a tentative map of possible migration routes based on the combined results from the telemetry studies.

We attempted to assess the total current population size and the trends in abundance within some larger geographical units in recent years based on the HKBWS’s synchronized survey results from mid-winter counts carried out during 2004–2020, and supplemented by additional information from other sources. Population abundance and trends were analyzed by grouping individual sites as follows: Yangtze River Delta (including Shanghai, Zhejiang, and Jiangsu Province), Fujian Coast, Pearl River Delta (including Hong Kong, Macau and Guangdong Province), Hainan, on the island of Taiwan and in Japan, Korea, Philippines, Thailand and Red River Delta of Vietnam.

Internationally important wintering sites were determined based on the average of total winter counts from 2016 to 2020 (the past 5 years, according to the Ramsar criterion; Ramsar 2007), for identifying important key sites (defined as supporting >5% of the population) and regular key sites (>1% of the population; Damba et al. 2020). We arbitrarily defined key stopover sites as those used by more than 0.25% of the total wintering counts or that supported staging tagged individuals for more than 7 days. In five cases [two reported by Liu (2006) and three reported by Son et al. (2020)], it was not clear precisely how long birds had remained at stopover sites, so these were estimated in Table 1. To assess the current protection status of these key sites based on the definitions

above, we downloaded the boundary information from the World Database of Protected Areas (WDPA, <https://www.protectedplanet.net/en/thematic-areas/wdpa>) and National Nature Reserves (NNRs, <http://www.resdc.cn/data.aspx?DATAID=272>) with permissions. We then overlaid all the key wintering sites using ArcGIS 10.6 to identify which sites were located within WDPA and NNR boundaries. However, during this process we also became aware that the boundary information held in WDPA did not reflect the true protected area boundaries in a very few cases. For this reason, we asked the opinion of local experts to define the precise geographical extent of protection status for a few sites, e.g., the Deep Bay area of Hong Kong.

Results

Distribution

From the results presented here, we can add a breeding site in South Korea (Chilsando Islet; Kang et al. 2016) revealed by field surveys, and summering sites revealed by telemetry tracking, including Yancheng NNR in Jiangsu Province, China (Ueta et al. 2002). These results confirm that all the known summering sites are exclusively coastal and restricted to South Korea, China, and Russia (Fig. 1).

Tracking results from 11 individuals showed birds wintering in other previously less known areas. These include: Yancheng NNR in Jiangsu Province (Jia et al. 2020); Hangzhou Bay and Taizhou Bay in Zhejiang Province (Jung et al. 2018); Yangtze River floodplain [Poyang Lake in Jiangxi Province and Nanyi Lake in Anhui Province (Jia et al. 2020), Yongzhou in Hunan Province (Son et al. 2020), Wang Lake in Hubei Province (Wood et al. 2013), Shijiu Lake in Jiangsu Province (Jung et al. 2018)]; Funing Bay of Fujian Coast (Son et al. 2020) and Luoyuan Bay and Xiamen (Jung et al. 2018), all in Fujian Province; and Tra-sur Flooded Forest on the border of Vietnam and Cambodia (Wood et al. 2013). Based on HKBWS field survey results, Black-faced Spoonbills also winter on Batanes Island and Panpang in Philippines and Phetchaburi in Thailand (Yu et al. 2020). With the exception of the inland freshwater Yangtze River floodplain and the Tra-sur Flooded Forest, all other wintering areas are maritime coastal East Asian habitats (Fig. 1).

The results suggest the inclusion of the south coast of South Korea and the coastal area east to the Han River in South Korea (Kim and Kim 2005) as part of the staging areas, as well as parts of Shandong Province (Yellow River Delta; Shan et al. 2005) and Zhejiang Province (Hangzhou Bay; Jin et al. 2017) in China. Telemetry tracking data also revealed stopover sites on the southeast coast of Shandong Province (Rongcheng and Jiaozhou Bay; Jia et al. 2020) and Jiangsu Province (Linhong Estuary; Jia

Table 1 Migration schedules from 34 tracked Black-faced Spoonbills (*Platalea minor*)

ID	Summering site	Departure time	Wintering site	Arriving time	Duration (days)	Resource
1	Chilsando Islet, South Korea	2014/11/3	Hangzhou Bay, Yangtze River Delta	2014/11/6*	3*	Jung et al. (2018)
2	Chilsando Islet, South Korea	2017/10/13	Yongzhou, Hunan	2017/10/15	2	Son et al. (2020)
3	Chilsando Islet, South Korea	2017/10/30	Funing Bay, Fujian Coast	2017/11/1	2	Son et al. (2020)
4	Chilsando Islet, South Korea	2014/11/4	Xiamen, Fujian Coast	2014/11/12*	8*	Jung et al. (2018)
5	Chilsando Islet, South Korea	2017/10/28	Hong Kong, Pearl River Delta	2017/11/2	5	Son et al. (2020)
6	Chilsando Islet, South Korea	2017/10/19	Hong Kong, Pearl River Delta	2017/10/23	4	Son et al. (2020)
7	Chilsando Islet, South Korea	2014/11/3	Taijiang National Park, Taiwan	2014/11/5*	2*	Jung et al. (2018)
8	Chilsando Islet, South Korea	2017/10/19	Taijiang National Park, Taiwan	2017/10/30	11	Son et al. (2020)
9	Gyeonggi Bay, South Korea	2009	Hakata Bay, Japan	2009	NA	Wood et al. (2013)
10	Gyeonggi Bay, South Korea	2014/11/14	Shimabara Bay, Japan	2014/12/7*	23*	Jung et al. (2018)
11	Gyeonggi Bay, South Korea	2014/10/24	Shijiu Lake, Inland Jiangsu	2014/11/27*	34*	Jung et al. (2018)
12	Gyeonggi Bay, South Korea	2009	Wang Lake, Hubei	2009	NA	Wood et al. (2013)
13	Gyeonggi Bay, South Korea	2014/10/28	Hangzhou Bay, Yangtze River Delta	2014/10/30*	2*	Jung et al. (2018)
14	Gyeonggi Bay, South Korea	2014/11/2	Hangzhou Bay, Yangtze River Delta	2014/11/4*	2*	Jung et al. (2018)
15	Gyeonggi Bay, South Korea	2014/11/8	Funing Bay, Fujian Coast	2014/11/13*	5*	Jung et al. (2018)
16	Gyeonggi Bay, South Korea	2014/11/3	Taijiang National Park, Taiwan	2014/11/6*	3*	Jung et al. (2018)
17	Gyeonggi Bay, South Korea	2014/11/5	Taijiang National Park, Taiwan	2014/11/29*	24*	Jung et al. (2018)
18	Gyeonggi Bay, South Korea	2010	Tra-sur Flooded Forest, Cambodia/ Vietnam border	2010	NA	Wood et al. (2013)
19	Xingren Island, Liaoning	2017/11/8	Nanyi Lake, Anhui	2017/11/17	9	Jia et al. (2020)
20	Xingren Island, Liaoning	2017/10/22	Poyang Lake, Jiangxi	2017/10/30	8	Jia et al. (2020)
21	Xingren Island, Liaoning	2018/10/30	Yancheng NNR, Yangtze River Delta	2018/11/2	3	Jia et al. (2020)
22	Xingren Island, Liaoning	2018/11/5	Yancheng NNR, Yangtze River Delta	2018/12/5	30	Jia et al. (2020)
23	Xingren Island, Liaoning	2017/10/22	Hangzhou Bay, Yangtze River Delta	2017/11/19	28	Jia et al. (2020)
24	Xingren Island, Liaoning	2018/10/16	Taijiang National Park, Taiwan	2018/10/20	4	Jia et al. (2020)
25	Gyeonggi Bay, South Korea	1999/3/23	Hong Kong, Pearl River Delta	1999/4/10	18	Ueta et al. (2002)
26	Gyeonggi Bay, South Korea	1999/3/27	Hong Kong, Pearl River Delta	1999/4/9	13	Ueta et al. (2002)
27	Gyeonggi Bay, South Korea	1999/4/13	Hong Kong, Pearl River Delta	1999/4/22	9	Ueta et al. (2002)
28	Gyeonggi Bay, South Korea	1999/4/13	Hong Kong, Pearl River Delta	1999/5/4	21	Ueta et al. (2002)
29	Gyeonggi Bay, South Korea	1999/3/28	Taijiang National Park, Taiwan	1999/4/10	13	Ueta et al. (2002)
30	Gyeonggi Bay, South Korea	1999/3/15	Taijiang National Park, Taiwan	1999/3/31	16	Ueta et al. (2002)
31	Yancheng NNR, Yangtze River Delta	1999/4/23	Hong Kong, Pearl River Delta	1999/5/21	28	Ueta et al. (2002)
32	Yancheng NNR, Yangtze River Delta	1999/5/14	Hong Kong, Pearl River Delta	1999/6/16	33	Ueta et al. (2002)
33	Nampho, North Korea	1999/5/18	Taijiang National Park, Taiwan	1999/6/11	24	Ueta et al. (2002)
34	Yilan, Taiwan	1999/5/6	Taijiang National Park, Taiwan	1999/5/10	4	Ueta et al. (2002)

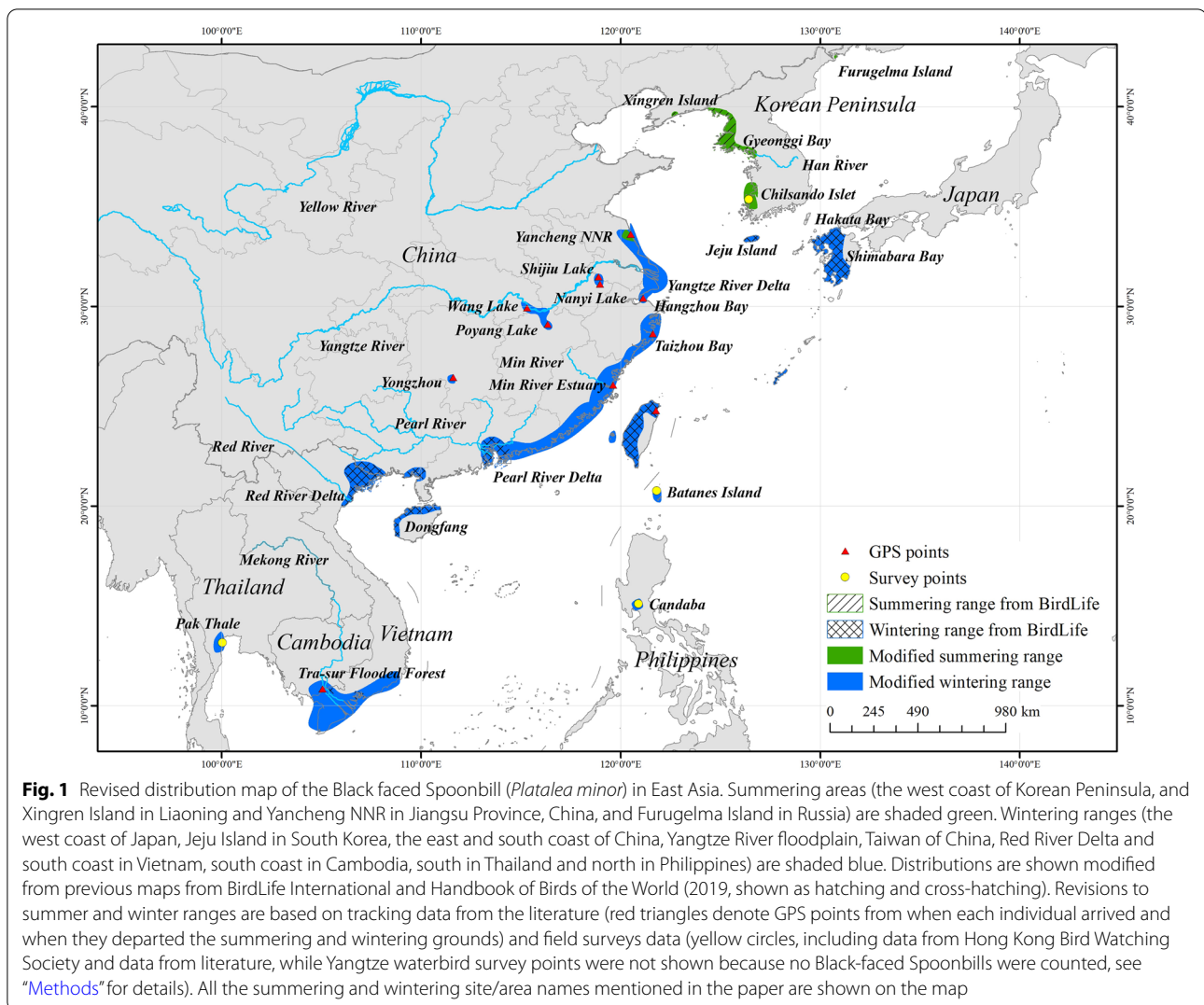
ID 1–24 were immature birds captured at breeding sites, all of which finished one complete autumn migration; ID 25–34 were the birds captured in wintering sites and all finished one spring migration, of which ID 25–30 were adults, ID 31–34 were immature birds. *Indicates occasions where the date or the duration were estimated (see “Methods” for more details)

et al. 2020) and inland Jiangsu (Jia River; Jung et al. 2018) in China. With the exception of sites on the Jia River, all other newly located stopover sites were in marine coastal areas (Fig. 2).

Migration routes and timing

We present a tentative map of possible migration routes based on the combined results of the telemetry studies in Fig. 3, based on 34 migration routes from 34 different individuals. Black-faced Spoonbills usually migrate

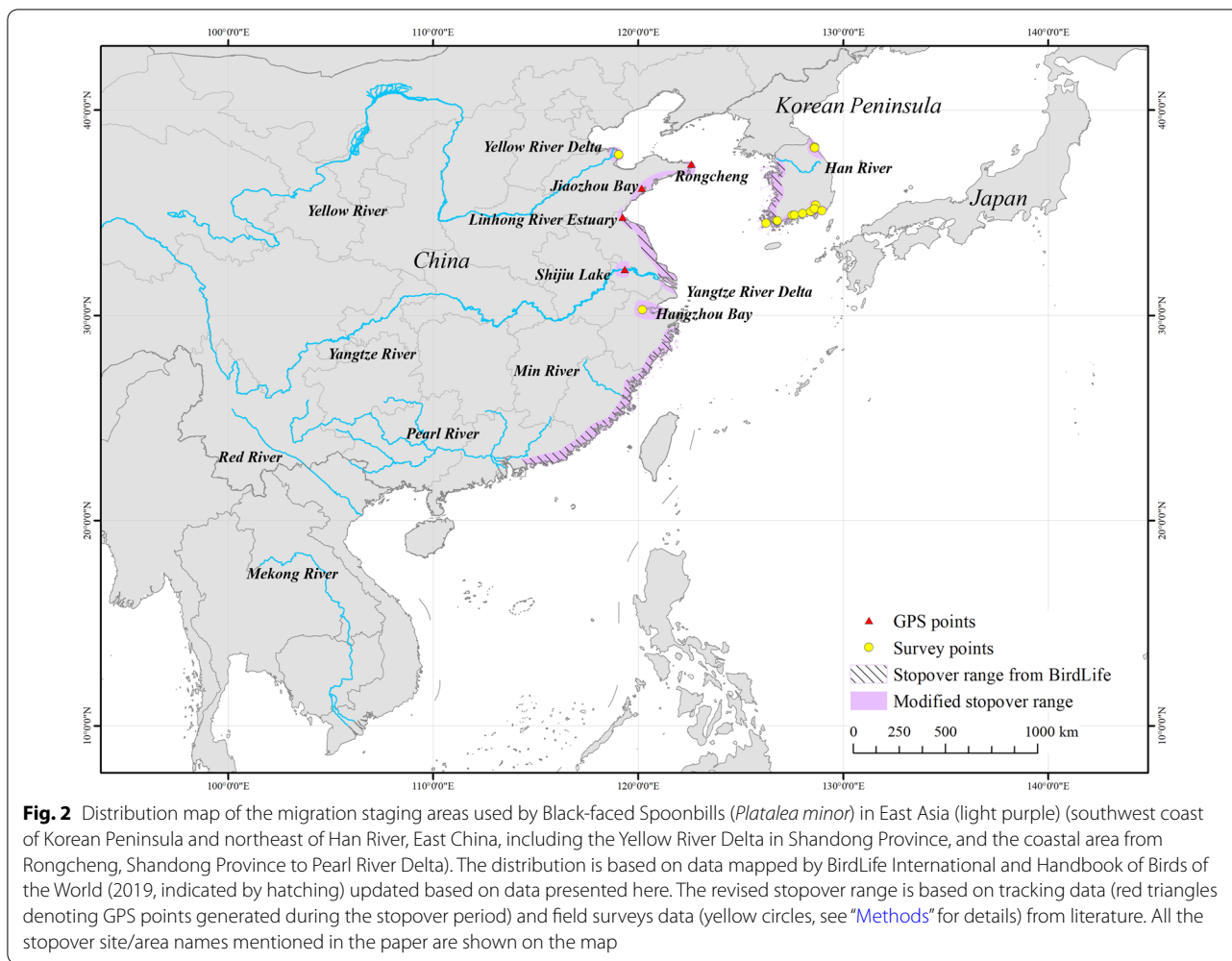
along the coast of southwest South Korea and in East China, occasionally along river floodplains, as in the case of the Yangtze River. According to telemetry data from the literature, first-year birds initiated their autumn migration from mid-October to mid-November (October 13–November 14; mean = October 29; SD = 8 days) and arrived to winter quarters in late October to early December (October 15–December 7; mean = November 8; SD = 14 days). The average duration of autumn migration was 9 days (SD = 10 days, range 2–34, $n = 21$ birds;



Jung et al. 2018; Jia et al. 2020; Son et al. 2020). Based on these tracking results and field survey results (Additional file 1: Table S1), we defined the autumn migration period of Black-faced Spoonbills as October to November inclusive, although a very few individuals started their movements before and continued after this period.

Adult Black-faced Spoonbills started spring migration from mid-March to mid-April (March 15 to April 13; mean = March 30; SD = 12 days) and arrived from late March to early May (March 31 to May 4; mean = April 14; SD = 12 days). The average migration duration was 16 days (SD = 6 days, range 9–24, $n = 6$ birds; Ueta et al. 2002). Juveniles started spring migration from late April to late May (April 23 to May 18; mean = May 7; SD = 11 days), and arrived from mid-May to mid-June (May 10 to June 16; mean = May 30; SD = 17 days). Average migration duration was 22 days (SD = 13 days, range 4–33, $n = 4$ birds; Ueta et al. 2002). Based on these

tracking results and field survey results (Additional file 1: Table S1), we defined the spring migration in adults as March and April, but in juveniles March to May, for whom the summering period was May to September. The wintering period was defined as December to February inclusive and we thus defined the summer, wintering range and migration staging range according to the distribution of where tracked birds were during these periods. The migration routes of the Black-faced Spoonbills connected the summering sites (Korean Peninsula), stopover sites (east coast of China) and wintering sites (from the southeast coast of China to the south of Philippines and Vietnam) and are based on the results of tracking studies. Twenty-four immature birds were tracked from breeding sites, including 18 from the main breeding sites in South Korea, and 6 from those in China, although no tracking data were available from Russia. Six of them wintered in inland China, mainly in Yangtze



River floodplain; 18 wintered in coastal areas, including Yangtze River Delta, Fujian Coast, Pearl River Delta, on the island of Taiwan, and Japan. From these tracking, the Yangtze River Delta and Yangtze floodplain and the wintering site on the Vietnam/Cambodia border were previously unknown as regular wintering sites (Fig. 1).

From wintering sites in China, four birds (two adults, two juveniles) were tracked from Taiwan and six from Hong Kong (four adults, two juveniles). All six adults summered at the Gyeonggi Bay breeding site, but among the four juveniles, only one summered in the breeding range (Namph in North Korea), while the other three summered at known non-breeding sites (one on Taiwan and two in Yangtze River Delta). The lack of segregation among these tracking results connecting breeding and wintering ranges supports the current hypothesis that there is only one biogeographical population of this species, until further tracking data can show to the contrary.

Abundance estimates and trends

According to the HKBWS survey results, the population of the Black-faced Spoonbill increased by 300%, from about 1200 to 4864 during 2004 to 2020 (Yu et al. 2020). Since 2004, over 50% of the Black-faced Spoonbills wintered on Taiwan (with the exceptions of 48% in 2007 and 47% in 2011) and these numbers contributed most to the increase of the overall population total. However, the period 2017–2020 also saw a rapid increase in numbers wintering on the Min and Yangtze River Deltas, while numbers wintering on the Pearl River Delta and in Japan have also maintained a steady increase (Table 2; Figs. 4 and 5).

Key sites during the annual life cycle

Summering period

From the literature search, we derived a total of 19 records of Black-faced Spoonbills from 9 sites during the summering period (Additional file 1: Tables S1, S3)

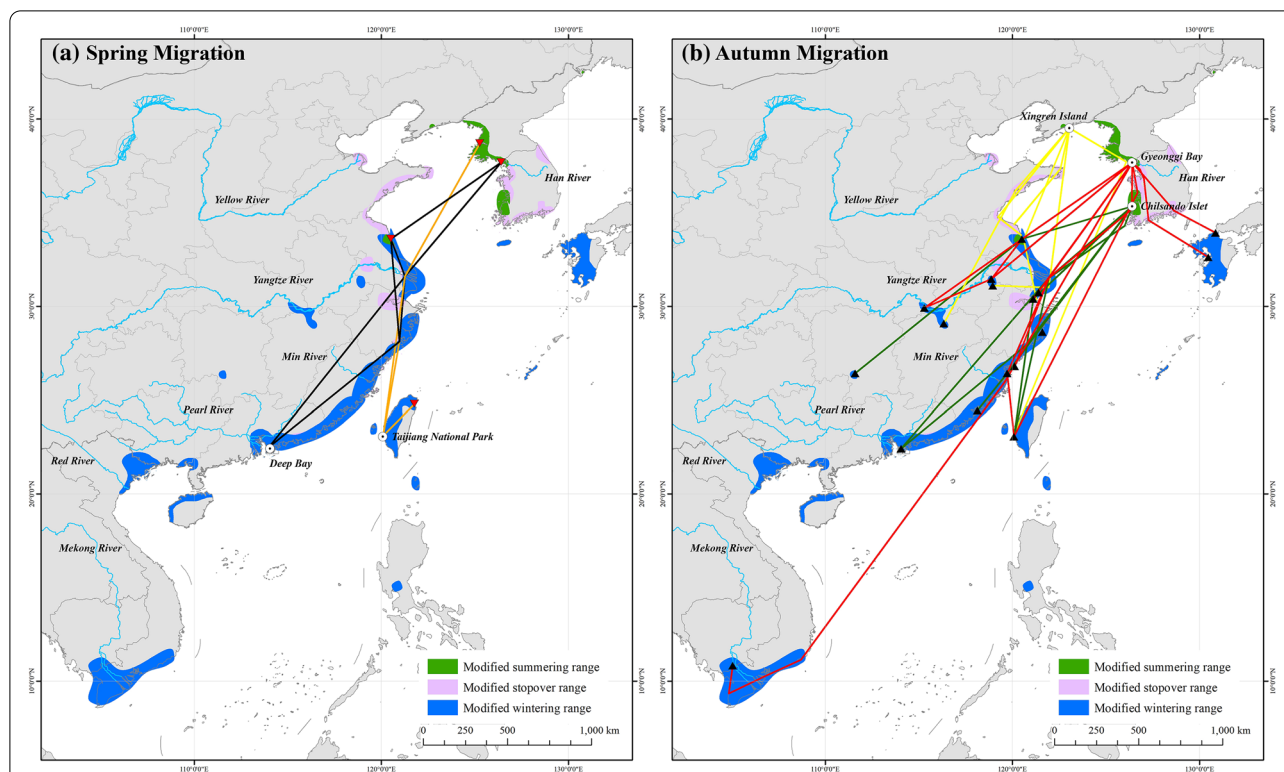


Fig. 3 The migration routes taken by tracked Black-faced Spoonbills (*Platalea minor*), based on telemetry data from 34 individuals reported in the literature (see “Methods” for details). **a** Spring migration, including 10 tagged individuals captured at 2 wintering sites: Hong Kong (Deep Bay, $n = 6$, black) and Taiwan (Taijiang National Park, $n = 4$, orange) in China. **b** Autumn migration, including 24 tagged individuals captured at three breeding sites: Chilsando Islet ($n = 8$, dark green) and Gyeonggi Bay ($n = 10$, red) in South Korea and Xingren Island in Liaoning Province, China ($n = 6$, yellow). Dot-centred circle = the capture sites. Revised summering, wintering, and stopover ranges are shown in green, blue and light purple, respectively. The red inverted and black regular triangles = summer and wintering sites, respectively, used by tracked individuals. The names of all the capture sites and main rivers are shown on the map. All lines are shown as the shortest distances between two sequential fixes, which may not represent the precise routes taken by the birds concerned

including 8 from South Korea, 8 in China, 2 in Russia and 1 in North Korea. Eight survey records revealed 5 breeding sites: Gyeonggi Bay and Chilsando Islet in South Korea, Xingren Island and Yuanbao Island in China, and Furugelma Island in Russia, whereas Gyeonggi Bay had the highest number of breeding individuals of 652 birds (Table 3). Of these sites, Gyeonggi Bay is within WDPA; sites used in Chilsando Islet are partially protected; Xingren Island and Yuanbao Island are protected by provincial government, while Furugelma Island lies outside the known reserve boundaries (Fig. 6a).

Wintering period

Key wintering sites were defined on 5% and 1% thresholds derived from the overall population estimate of 4113 birds, based on average total counts from the annual census during 2016–2020 (Yu et al. 2020). Nineteen internationally important wintering sites (based on the criterion that they supported more than 1% population at least once since 2004, rounded to a threshold of 40 birds) were

identified based on the maximum counts during 2004–2020 censuses. These comprised 5 important key sites (supporting >5% of the population, 200 birds) and 14 regular key sites (1–5%, 40–200 birds) (Fig. 6b). These 5 important key sites were: Taijiang National Park, Chiayi and Kaoshiung on Taiwan, Xinghua Bay in Fujian, and Deep Bay in Hong Kong and the adjacent Shenzhen. Of all these 19 key wintering sites, 12 sites are currently protected. Seven sites are within WDPA boundaries or NNR boundaries. The Deep Bay area of Hong Kong is partially protected. Haifeng in Guangdong Province, Dongfang in Hainan Province, Taipa-coloane in Macau, and Chiayi on Taiwan are located within a state protected area (Table 4).

Autumn and spring migration periods

There were 47 records providing some dated information about stopover sites (Table 5). Chongming Island in Shanghai (Yangtze River Delta) and Xinghua Bay in Fujian Province were the most intensively used stopover

Table 2 Annual counts of Black-faced Spoonbills (*Platalea minor*) in ten regions in January during 2004–2020

Year	Korea	Japan	Yangtze River Delta	Fujian Coast	Taiwan	Pearl River Delta	Hainan	Thailand	Philippines	Total
2004	23	147	22	1	633	293	62	2	NA	1198
2005	21	92	29	63	757	389	77	NA	NA	1484
2006	21	148	9	69	822	453	6	NA	NA	1602
2007	20	191	14	86	790	476	8	NA	NA	1630
2008	28	192	5	115	1030	527	11	5	NA	1962
2009	25	166	0	99	1104	460	8	NA	1	1926
2010	12	198	0	103	1280	577	7	1	0	2224
2011	14	220	3	53	834	536	57	1	0	1767
2012	17	196	57	135	1561	536	31	2	0	2570
2013	8	223	42	47	1624	485	32	1	NA	2501
2014	8	276	23	33	1658	466	36	1	0	2541
2015	14	293	13	32	2034	622	54	NA	NA	3102
2016	23	247	4	124	2060	623	77	0	1	3167
2017	18	305	54	118	2601	608	61	NA	NA	3827
2018	14	305	93	295	2191	626	66	1	3	3659
2019	23	324	188	281	2403	708	83	1	3	4079
2020	24	339	196	338	2783	726	86	NA	3	4555

The ten regions were identified based on the distribution of the surveyed wintering sites during 2004–2020 conducted by HKBWS. Counts of Yangtze River Delta includes counts of Shanghai, Zhejiang, and Jiangsu Province in China; counts of Fujian Coast includes counts of Fujian Province in China; counts of Pearl River Delta includes counts of Hong Kong, Macau, and Guangdong Province in China

sites. Numbers staging at the Yellow River Delta from 2002 to 2005 might have decreased during these years, but long-term regular monitoring activities at this site were lacking. The most recent records (110 birds in October and 142 birds in December) were 2 total counts from 7 sites along the Zhejiang coast in 2016. Twenty-three key sites were identified by meeting the threshold of 0.25% population and staging for 7 days from the field survey results and satellite tracking data. Among these sites, 4 were located within NNR boundaries: Yellow River Delta, Yancheng NNR, Chongming Island and Min River Estuary (Fig. 6c, d).

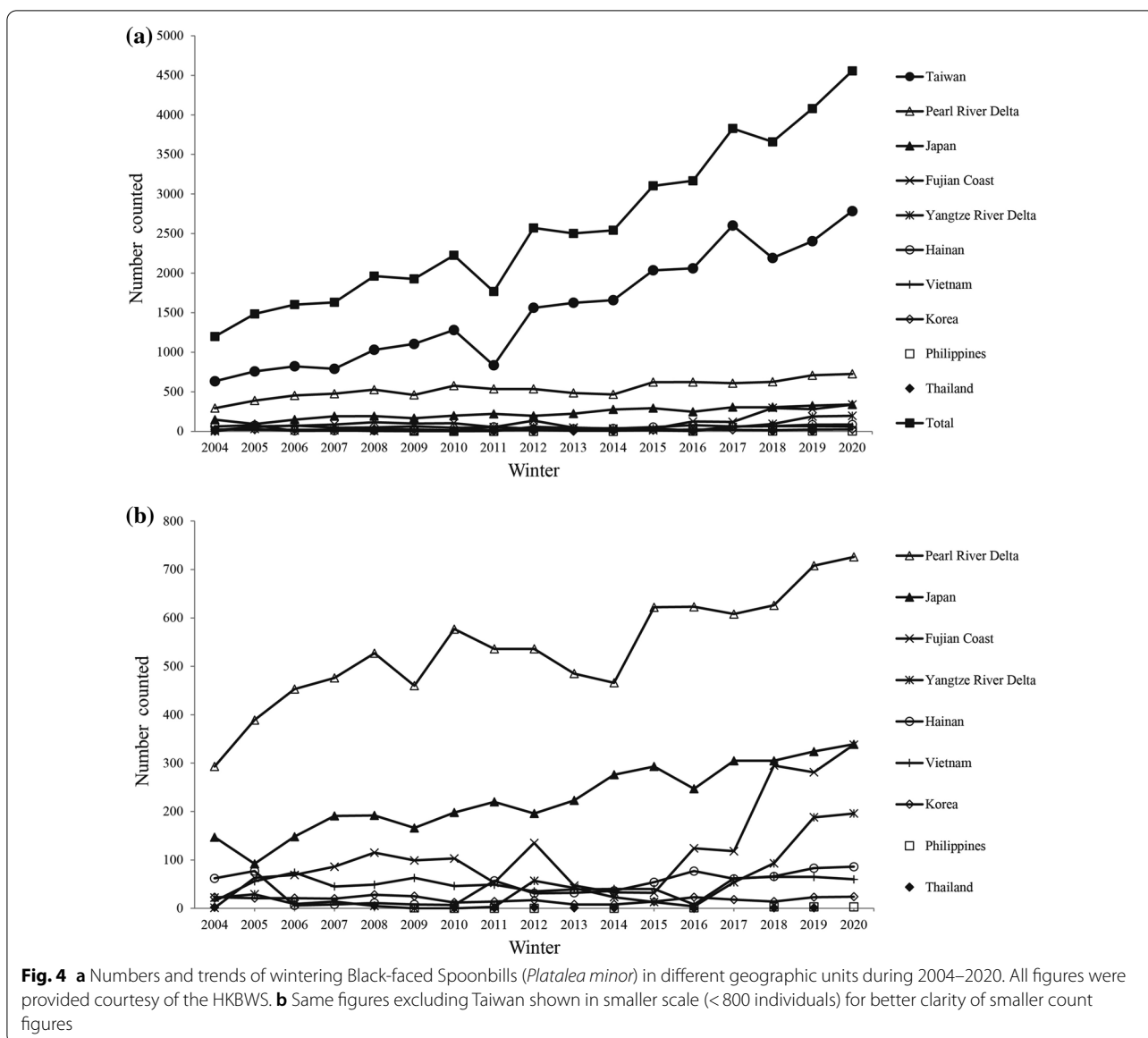
Data from 24 tracked individuals revealed 26 autumn stopover sites (Fig. 7), of which 7 sites were used for 8–31 days (based on maximum possible stopover duration), 9 sites for the staging of 2–7 days, and 10 sites for staging of one day only. Chongming Island of Shanghai recorded the highest number of tracked individuals (9), followed by 3 individuals at Yancheng NNR of Jiangsu Province and 2 at Luoyuan Bay of Fujian Province.

Discussion

In this paper, we integrated data from field surveys and satellite tracking studies to refine the existing distribution range of the Black-faced Spoonbill, identified some new key sites and reviewed the conservation status of these sites. We recommend including Yellow Sea coastal areas (e.g., the Yancheng NNR in Jiangsu Province; Fig. 1)

within the summer range of this species. We also found that Black-faced Spoonbills were reported summering in the Min River Estuary (several immature birds; Liu 2006) and Yilan, Taiwan (one tracked immature; Ueta et al. 2002), but consultation with local experts confirmed that these records were too few and irregular to justify including these sites as summering sites. Historically, Longtan Reservoir (Liaoning Province) was reported to be used by a dozen Black-faced Spoonbills in March 1993 (Qiu 2007a), but as this site lies well away from the known migration routes and the information is relatively old, this site was not included in the analysis of this study. Further survey is needed to confirm whether Black-faced Spoonbills still occur regularly there in the absence of recent records.

We also suggest the inclusion of wetlands on the east coast of China within the regular wintering area of this species. The Yangtze River floodplain wetlands need more surveys to elucidate the true status of the wintering Black-faced Spoonbills. Satellite tracking results in conjunction with a few isolated ground observations (e.g., Yu and Swennen 2005) showed that Black-faced Spoonbills could be found in freshwater inland habitats during the winter (Table 6, Fig. 3). With the increasing trend in the wintering population of this species, it could be speculated that new wintering flocks could be found there in the future (Wood et al. 2013). Records in these inland freshwater habitats gives us a hint that some



fledging Black-faced Spoonbills could mix with the congeneric Eurasian Spoonbills (*Platalea leucorodia*) wintering in this area which were counted in large numbers (Additional file 1: Table S4), as both these species have already been reported together using freshwater habitats in inland wintering sites (Yu and Swennen 2005; Wood et al. 2013; Sullender et al. 2016). Public awareness of the Black-faced Spoonbill should be raised among the waterbird surveyors, wetland managers and birdwatching communities in these areas to alert them to the possibility of their occurrence in these previously unusual wintering areas. Local ornithologists might not be aware that the Black-faced Spoonbill, thought to be a coastal species, could also be found in these freshwater habitats,

and that single individuals have the potential to be mixed in with the larger wintering flocks of Eurasian Spoonbills. In addition, hybridization between the two species makes it possible that a certain proportion of hybrid individuals exist in Eurasian Spoonbill colonies on the mainland. Since observed hybrid individuals appear very similar to Black-faced Spoonbills (Kwon et al. 2017; Tiunov 2021), careful monitoring is necessary in both breeding and wintering areas to determine their numbers and distribution throughout the range. Local observers and researchers should also be encouraged to report sightings of Black-faced Spoonbills to the international coordinator of the annual Black-faced Spoonbill winter census, as this will extend census coverage and improve our assessment

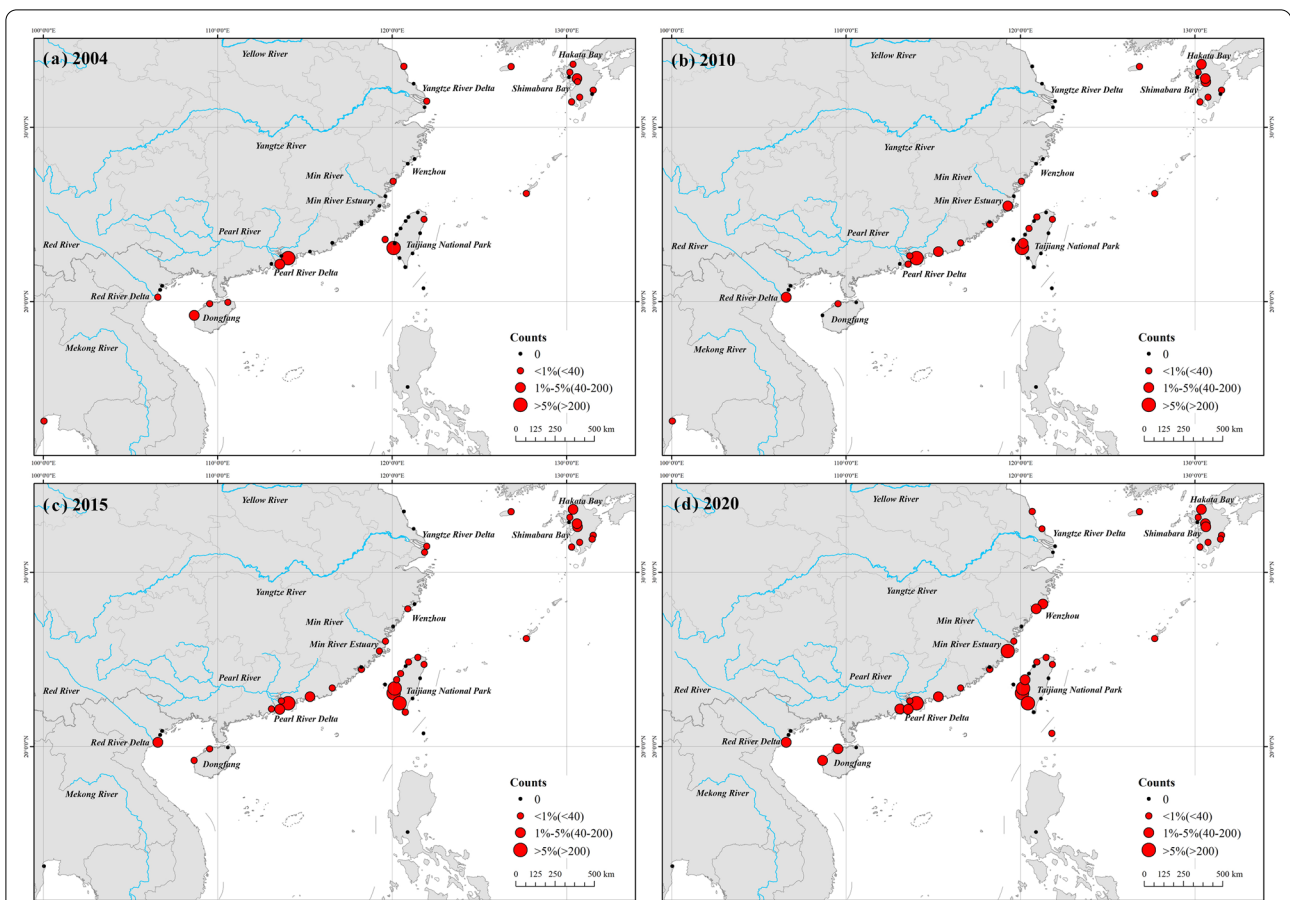
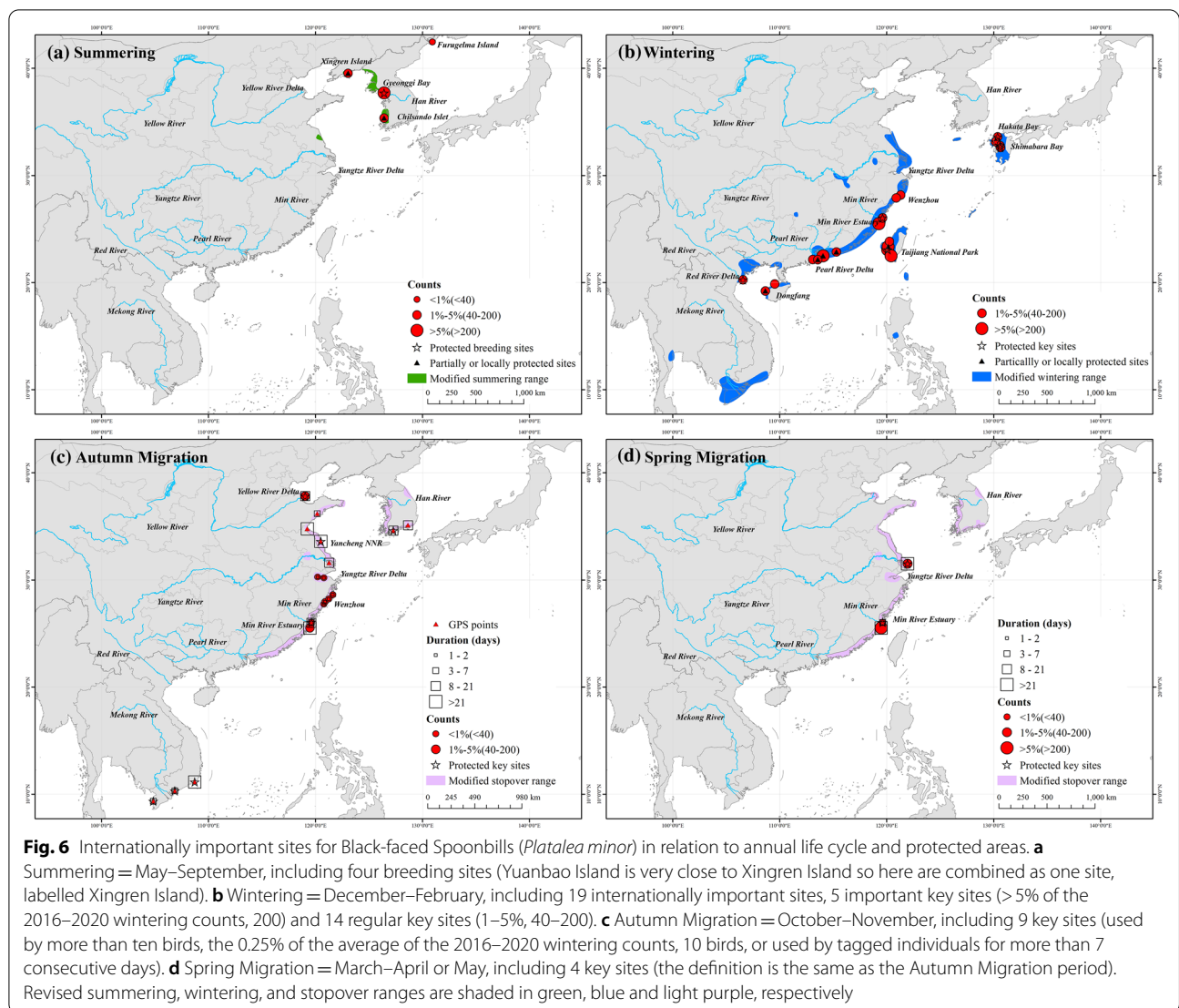


Fig. 5 Changes in distribution and abundance of Black-faced Spoonbills (*Platalea minor*) counted at internationally important sites in 2004, 2010, 2015 and 2020. **a** January 2004 (four internationally important sites, including two with 1%–5% of the average of the 2016–2020 wintering counts, 40–200, and two with > 5%, 200). **b** January 2010 (seven internationally important sites, including five with 1%–5% of birds counted, 40–200, and two with > 5%, 200). **c** January 2015 (nine internationally important sites, including six with 1–5% of birds counted, 40–200, and three with > 5%, 200). **d** January 2020 (fourteen internationally important sites, including ten with 1%–5% of birds counted, 40–200, and four with > 5%, 200). In recent years, there has been a rapid increase in wintering numbers on the Min and Yangtze River Deltas. Black circles indicate sites that were visited, but where no Black-faced Spoonbills were present. The size of the red circles indicates the numbers counted as a proportion of individuals, in relation to the total wintering counts in 2020

Table 3 Breeding sites of the Black-faced Spoonbill (*Platalea minor*) revealed by field survey records, as well as the maximum numbers of birds recorded in summer

Site name	Region	Longitude (°E)	Latitude (°N)	Number of birds	Year	Conservation status	Resource
Gyeonggi Bay area	Gyeonggi, South Korea	126.40	37.69	652	2006	Within WDPA	NRICH (2007)
Chilsando Islet	South Jeolla, South Korea	126.41	35.35	100	2014	Partially protected	NRICH (2014)
Yuanbao Island	Liaoning, China	123.09	39.50	14	2007	Locally protected	Qiu (2007a)
Xingren Island	Liaoning, China	123.05	39.53	49 nests	2019	Locally protected	Jia et al. (2020)
Furugelma Island	Primorsky, Russia	130.92	42.46	12 pairs	2018	Unprotected	Tiunov and Katin (2020)

In South Korea, Gyeonggi Bay area currently falls within existing protected area boundaries and sites used in Chilsando Islet are partially protected. In China, Yuanbao Island and Xingren Island are shown as provincial protected areas. The partially and locally protected areas were not within the current national protected area boundary



of the total population size. Given that the Black-faced Spoonbill population is expanding (based on the results of the winter census), further expansion of its wintering range into the wetlands along the Yangtze River remains highly likely in the future (Wood et al. 2013; Takano et al. 2014). Migratory waterbirds are known to show exploratory behavior during migration and colonize new staging and wintering areas (Baker 1980; Jansen 1981; Galvan et al. 2003). Black-faced Spoonbills have been reported migrating in age-specific groups (Ueta and Higuchi 2002; Wood et al. 2013), and since all birds tracked on autumn migration were immature birds (Wood et al. 2013; Jung et al. 2018; Jia et al. 2020; Son et al. 2020), their migration (thought independent of adults) may explain such exploratory patterns.

One major drawback of recent tracking studies of Black-faced Spoonbills is that results are heavily biased towards those from trapped and tracked fledglings, because they were more easily caught than the adults on the breeding sites (e.g., Wood et al. 2013; Son et al. 2020). Hence, information about the movements of individuals of other age groups, i.e., subadults and adults are still lacking, despite Ueta et al. (2002) tracking six adults to study their spring migration in 1999. Data retrieved from immature birds showed them not returning to the breeding sites, but remaining on their wintering grounds, or at least in southern parts of the normal summer distributional range (Ueta et al. 2002). Therefore, tracking of adult or sub-adult Black-faced Spoonbills could provide more information about their migration schedules,

Table 4 Internationally important wintering sites for the Black-faced Spoonbill (*Platalea minor*) showing maximum counts and current conservation status

Site name	Region	Longitude (°E)	Latitude (°N)	Maximum counts	Year	Conservation status
Ariake Sea North	Japan	130.19	33.14	43	2019	Within WDPA
Hakata Bay	Japan	130.37	33.60	66	2007	Within WDPA
Shimabara Bay	Japan	130.60	32.78	102	2020	Within WDPA
Yatsushiro Sea North	Japan	130.63	32.59	95	2017	Within WDPA
Xuanmun Bay	Zhejiang, China	121.29	28.17	97	2020	Unprotected
Wenzhou	Zhejiang, China	120.89	27.90	89	2019	Unprotected
Min River Estuary	Fujian, China	119.62	26.03	71	2017	Within NNRS
Xinghua Bay	Fujian, China	119.27	25.48	323	2020	Unprotected
Haifeng	Guangdong, China	115.29	22.86	198	2020	Locally protected
Jiangmun	Guangdong, China	113.08	22.16	90	2020	Unprotected
Lingao	Hainan, China	109.53	19.87	42	2020	Unprotected
Dongfang	Hainan, China	108.65	19.21	68	2005	Locally protected
Deep Bay	Hong Kong, China	114.04	22.49	462	2010	Partially protected
Taipa-coloane	Macau, China	113.56	22.14	61	2016	Locally Protected
Yunlin	Taiwan, China	120.26	23.84	102	2020	Unprotected
Chiayi	Taiwan, China	120.14	23.33	870	2016	Locally protected
Taijiang National Park	Taiwan, China	120.08	23.07	1839	2020	Within WDPA
Kaoshiung	Taiwan, China	120.41	22.49	257	2018	Unprotected
Xuan Thuy National Park	Vietnam	106.57	20.25	74	2006	Within WDPA

The key wintering sites (> 1% Criterion: 40 birds, $n = 14$) and important key sites (> 5% Criterion: 200 birds, shown in bold, $n = 5$) were identified based on the maximum counts during 2004–2020 from the Hong Kong Bird Watching Society (HKBWS). Seven sites currently fall within existing protected area boundaries. In addition, areas used in Deep Bay are partially protected and the Haifeng, Dongfang, Chiayi and Taipa-coloane sites are shown as provincial or state protected area, but the areas used by birds were not within the current national protected area boundary

staging site use and the distribution range of this segment of the population. This also suggests that the immature birds do not usually undertake the long non-stop flight of crossing the Yellow Sea to reach the Korean Peninsula for summer. By doing so, young individuals could also avoid competition with breeding birds (Kokko 1999).

We can synthesize information from tracking studies to generate distribution map for the Black-faced Spoonbill's stopover sites of autumn migration (Fig. 7), but data were insufficient to do the same for spring migration. The literature review found three important sites in spring migration: Yancheng NNR in Jiangsu Province (three birds); Shanghai (four) and Yueqing in Zhejiang Province (seven) (Ueta et al. 2002). This information is similar to the results from autumn migration, suggesting that these three areas could be important stopover sites of the Black-faced Spoonbills during both migration episodes. However, field surveys of these sites were still inadequate to record the actual frequency of use by Black-faced Spoonbills of these sites in spring (Fig. 6; Table 5). For this reason, we urge systematic monitoring activities during the migration period especially from mid-October to mid-November and from mid-March to mid-April at these sites, to generate information on abundance,

turnover and stopover ecology of this species, as well as the relative importance of the stopover sites.

On the breeding areas, field surveys were regularly conducted in South Korea (Kwon et al. 2015; Kang et al. 2016), China (Ding et al. 2000; Qiu 2007a, b; Jia et al. 2020) and Russia (Tiunov and Katin 2020; Tiunov 2021), but no survey or regular monitoring was conducted at sites within North Korea. As the current wintering population is more than 5200 individuals (Yu et al. 2021) and non-systematic summer counts recorded less than 1000 individuals (Additional file 1: Table S1), this discrepancy suggests that some breeding concentrations and aggregations of non-breeding immature individuals in summer remained undiscovered. These sites could be of conservation importance for the Black-faced Spoonbill based on the numbers of these missing individuals, so more field surveys here are urgently needed.

Black-faced Spoonbills are sensitive to human disturbance and benefit from protection of the sites (Wei et al. 2005; Zhang et al. 2010, 2012; Henmi and Takano 2012; Sung et al. 2018). The Black-faced Spoonbill takes five years to reach sexually maturity (Yu 2005), which means the protection of the sites for these segments of the population during staging and wintering period are

Table 5 Stopover sites with available arrival/departure dates of the Black-faced Spoonbill (*Platalea minor*) revealed by field survey records and telemetry data

Stage	Site name	Region	Longitude (°E)	Latitude (°N)	Number of birds	Arrival time	Departure time	Duration (days)	Resource
Autumn migration (Oct–Nov)	Gyeonggi Bay area	South Korea	126.40	37.69	1 (tracked)	2018/10/17*	2018/10/19*	2	Jia et al. (2020)
	Ansan	South Korea	126.82	37.31	1 (tracked)	2014/11/15*	2014/11/17*	2	Jung et al. (2018)
	Cheonan	South Korea	127.03	36.80	1 (tracked)	2014/11/17*	2014/11/18*	1	Jung et al. (2018)
	Seocheon County	South Korea	126.69	36.07	1 (tracked)	2014/11/4*	2014/11/5*	1	Jung et al. (2018)
	Yeonggwang County	South Korea	126.38	35.19	1 (tracked)	2014/11/6*	2014/11/7*	1	Jung et al. (2018)
	Goheung-gun	South Korea	127.27	34.61	1 (tracked)	2014/11/18*	2014/12/6*	18	Jung et al. (2018)
	Hajo-do	South Korea	125.99	34.30	1 (tracked)	2014/11/4*	2014/11/5*	1	Jung et al. (2018)
	Nagasaki	Japan	129.82	32.73	1 (tracked)	2014/12/6*	2014/12/7*	1	Jung et al. (2018)
	Yellow River Delta	Shandong, China	119.04	37.82	49	2002/10/27	2002/11/13	17	Shan et al. (2005)
	Yellow River Delta	Shandong, China	119.04	37.82	15	2003/10/23	2003/11/10	18	Shan et al. (2005)
	Yellow River Delta	Shandong, China	119.04	37.82	11	2004/11/2	2004/11/15	13	Shan et al. (2005)
	Rongcheng NNR	Shandong, China	122.58	37.36	1 (tracked)	2018/10/31*	2018/11/2*	3	Jia et al. (2020)
	Rongcheng NNR	Shandong, China	122.58	37.36	3	2002/11/20	NA	NA	Zhao et al. (2003)
	Jiaozhou Bay	Shandong, China	120.17	36.19	1 (tracked)	2017/10/22*	2017/10/29*	7	Jia et al. (2020)
	Linhong Estuary	Jiangsu, China	119.23	34.78	1 (tracked)	2018/11/6*	2018/12/4*	28	Jia et al. (2020)
	Yancheng NNR	Jiangsu, China	120.51	33.60	1 (tracked)	2014/10/26*	2014/11/27*	31	Jung et al. (2018)
	Yancheng NNR	Jiangsu, China	120.51	33.60	1 (tracked)	2017/10/25*	2017/11/18*	23	Jia et al. (2020)
	Yancheng NNR	Jiangsu, China	120.51	33.60	1 (tracked)	2017/10/13*	2017/10/14*	1	Son et al. (2020)
	Dafeng	Jiangsu, China	120.50	33.20	1 (tracked)	2014/10/25*	2014/10/26*	1	Jung et al. (2018)
	Jia River	Jiangsu, China	119.34	32.23	1 (tracked)	2014/11/27*	2014/11/28*	1	Jung et al. (2018)
	Yangtze River Estuary	Shanghai, China	121.27	31.60	1 (tracked)	2014/11/8*	2014/11/29*	21	Jung et al. (2018)
	Yangtze River Estuary	Shanghai, China	121.27	31.60	1 (tracked)	2014/11/6*	2014/11/7*	1	Jung et al. (2018)
	Yangtze River Estuary	Shanghai, China	121.27	31.60	1 (tracked)	2017/11/9*	2017/11/11*	2	Jia et al. (2020)
	Yangtze River Estuary	Shanghai, China	121.27	31.60	1 (tracked)	2017/10/30*	2017/10/31*	1	Son et al. (2020)
	Yangtze River Estuary	Shanghai, China	121.27	31.60	1 (tracked)	2017/10/28*	2017/11/1*	4*	Son et al. (2020)
	Yangtze River Estuary	Shanghai, China	121.27	31.60	1 (tracked)	2017/10/19*	2017/10/29*	10*	Son et al. (2020)
	Chongming Island	Shanghai, China	121.95	31.52	1 (tracked)	2014/10/29*	2014/10/30*	1	Jung et al. (2018)
Chongming Island	Shanghai, China	121.95	31.52	1 (tracked)	2014/11/9*	2014/11/11*	2	Jung et al. (2018)	
Chongming Island	Shanghai, China	121.95	31.52	1 (tracked)	2014/11/3*	2014/11/4*	1	Jung et al. (2018)	
Seven Wetlands in Zhejiang	Zhejiang, China	120.92	28.84	142	2016.12.1	NA	NA	Jin et al. (2017)	

Table 5 (continued)

Stage	Site name	Region	Longitude (°E)	Latitude (°N)	Number of birds	Arrival time	Departure time	Duration (days)	Resource
	Seven Wetlands in Zhejiang	Zhejiang, China	120.92	28.84	110	2016.10.23	NA	NA	Jin et al. (2017)
	Taizhou Bay	Zhejiang, China	121.61	28.63	1 (tracked)	2014/11/4*	2014/11/5*	1	Jung et al. (2018)
	Ou River Estuary	Zhejiang, China	120.74	27.94	1 (tracked)	2017/10/19*	2017/10/22*	3*	Son et al. (2020)
	Yueqing	Zhejiang, China	120.97	28.12	1 (tracked)	2014/11/5*	2014/11/6*	1	Jung et al. (2018)
	Funing Bay	Fujian, China	120.13	26.81	1 (tracked)	2014/11/12*	2014/11/14*	1	Jung et al. (2018)
	Luoyuan Bay	Fujian, China	119.72	26.43	1 (tracked)	2014/11/30*	2014/12/1*	1	Jung et al. (2018)
	Luoyuan Bay	Fujian, China	119.72	26.43	1 (tracked)	2014/11/6*	2014/11/11*	5	Jung et al. (2018)
	Min River Estuary	Fujian, China	119.62	26.04	22	2004/10/27	2004/10/30*	3*	Liu (2006)
	Xinghua Bay	Fujian, China	119.48	25.52	136	2007/11/3	2007/12/12	39	Jin et al. (2009)
	Changhua County	Taiwan, China	120.43	24.04	1 (tracked)	2014/11/6*	2014/11/7*	1	Jung et al. (2018)
Spring migration (Mar–April or May)	Longtan Reservoir	Liaoning, China	120.45	41.81	15*	1993/3/15*	NA	NA	Qiu (2007a)
	Yellow River Delta	Shandong, China	119.04	37.82	5	2005/4/2	2005/4/13	11	Shan et al. (2005)
	Chongming Island	Shanghai, China	121.95	31.52	61	2002/4/2	2002/5/7	35	Yuan and Zhang (2006)
	Chongming Island	Shanghai, China	121.95	31.52	135*	2003/4/1	2003/5/15	44	Yuan and Zhang (2006)
	Min River Estuary	Fujian, China	119.62	26.04	15	2003/5/9	2003/5/12*	3*	Liu (2006)
	Min River Estuary	Fujian, China	119.62	26.04	12	2003/4/17	NA	NA	Liu (2005)
	Xinghua Bay	Fujian, China	119.48	25.52	226	2008/3/8	2008/4/20	43	Jin et al. (2009)

*Indicates that the date and/or duration were estimated. Stopover sites with more than ten birds (0.25% of the average wintering counts in 2016–2020) or used for more than seven days are shown in bold

also important. However, many habitats used by Black-faced Spoonbills are close to developed areas, including major cities and harbors, and are experiencing severe habitat loss and human disturbance, especially on their wintering areas (Henmi and Takano 2012; Chen 2014; Choi et al. 2014; Wang et al. 2018). In addition, many of the internationally important sites for the Black-faced Spoonbill identified in this study currently lie outside of the existing WDPA or NNRs protection areas. Although some of sites could still be protected locally by provincial or state governments, this status might not provide adequate protection to the birds and their habitats against national development plans (Zhang et al. 2015). Therefore, we recommend that those sites qualifying as internationally important sites based on count data should be given national protection. Pickett et al. (2018) also suggested that the recent increase in population size was due to recent low numbers falling well below potential

population carrying capacities of their habitat, but climate change and on-going habitat loss are predicted to have adverse impacts upon this species, causing ultimate reductions in the size of their global population before 2050. Thus, we would argue that the future of the Black-faced Spoonbill is not totally secure, despite the recent increase in the global population.

To conclude, we urge continued field surveys and tracking studies to provide further information about the distribution and abundance of the Black-faced Spoonbill particularly in summering period and more tracking of adult individuals as the population increases. We also recommend the strengthening of protection and sympathetic management of wetlands along the China coast, based on the important sites identified in this study in Zhejiang and Fujian Province, as well as in the traditional wintering areas in Guangdong and Hainan Provinces. After identifying

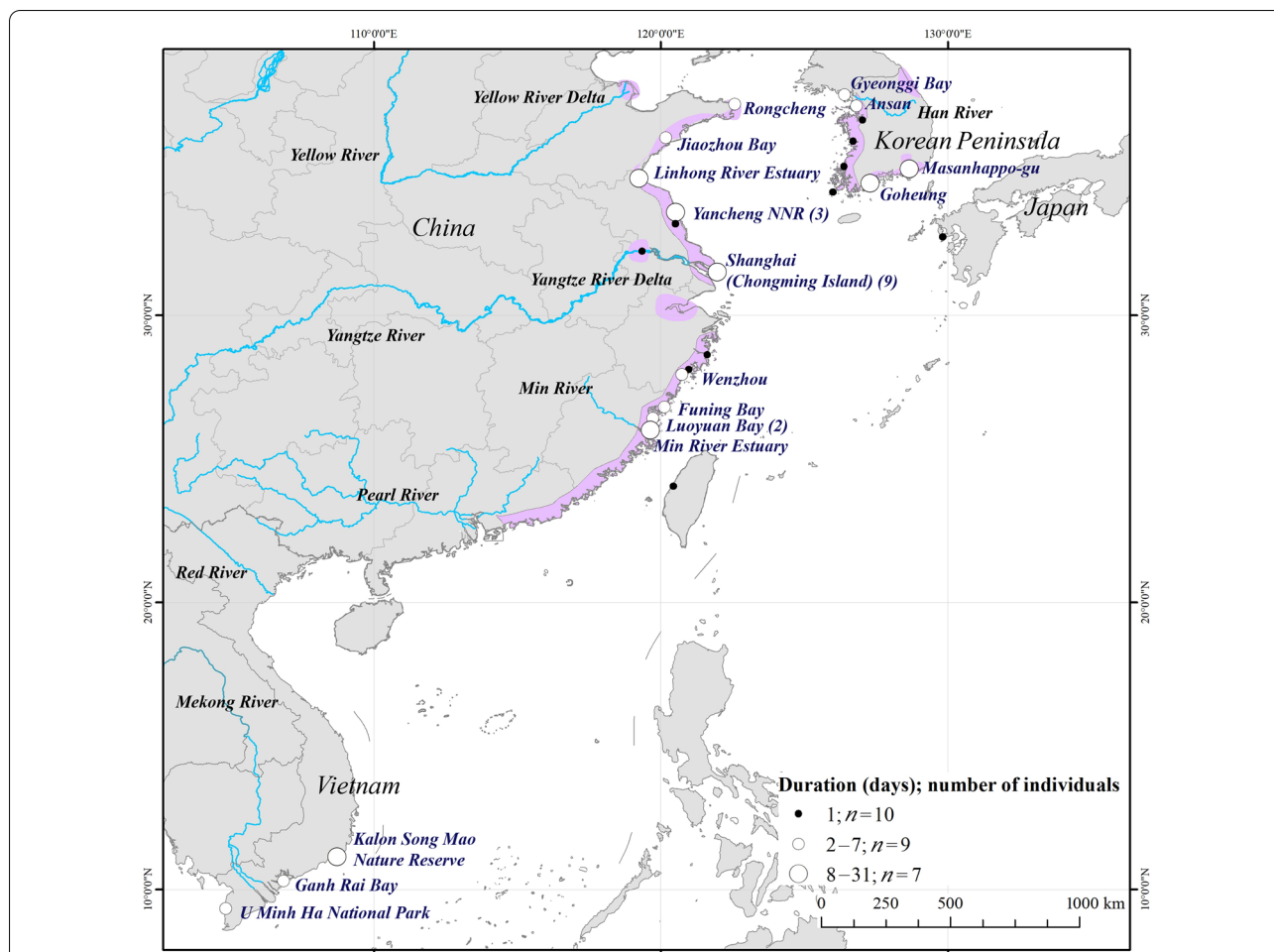


Fig. 7 The distribution of stopover sites of the 24 tracked Black-faced Spoonbills (*Platalea minor*) showing the longest stopover duration and the number of birds using the 26 sites during autumn migration. Numbers in brackets after site names indicate the number of different tagged birds recorded using that staging site. Among the sites, 10 sites were used for 1 day (black dots), 9 sites were used for 2–7 days (small white circles), and 7 sites were used for 8–31 days (large white circles). Sites used for at least 2 days were marked with their names on the map. Three sites were used more than once: Yancheng NNR in Jiangsu Province, 3 times; Shanghai (including Chongming Island), 9 times; Luoyuan Bay in Fujian Province, twice, all in China. Revised stopover range is shaded in light purple

key stopover sites along the southwest coast of Korea and east coast of China, regular monitoring activities should be established there to track the abundance and turnover rate of Black-faced Spoonbills during passage. Following this study, we also suggest the creation of an open platform to deposit count and census data, tracking studies results and regularly updated information on the distribution range and abundance of the Black-faced Spoonbill. This platform will be useful for the long-term

scientific research, conservation planning and raising awareness of this iconic waterbird in East Asia. Through this participation process, we hope to engage more researchers and conservationists locally and globally joining hands to protect this species and their wetland habitats, which could also be beneficial for other waterbirds.

Table 6 Number (and proportion) of all the tracked individuals using each of the wintering survey areas, compared to the numbers detected at each during the 2020 winter counts of Black-faced Spoonbill (*Platalea minor*)

ID	Wintering site	Number of tracked birds	Corresponding survey area	Number of surveyed population
1	Japan	2 (8%)	Japan	339 (7%)
2	Yangtze River Delta	6 (25%)	Jiangsu & Zhejiang & Shanghai	196 (4%)
3	Fujian Coast	3 (13%)	Fujian	338 (7%)
4	Pearl River Delta	2 (8%)	Hong Kong/Shenzhen & Guangdong & Macau	726 (16%)
5	Taiwan	5 (21%)	Taiwan	2783 (61%)
6	Other coastal area	0 (0%)	Hainan & Philippines & Thailand & Vietnam & Korea	173 (4%)
7	Inland China	5 (21%)	NA	NA
8	Inland Southeast Asia	1 (4%)	NA	NA
	Total	24		4555

The information on tracked birds was based upon data from 24 birds with complete autumn migration from the published literature (see “Methods” for details). The wintering counts were derived from field survey results in 2020 coordinated by the HKBWS

Supplementary Information

The online version contains supplementary material available at <https://doi.org/10.1186/s40657-021-00307-z>.

Additional file 1: Table S1. Sites used by Black-faced Spoonbills (*Platalea minor*) identified according to field survey data from published literature. **Table S2.** Parameters recorded for 34 individual Black-faced Spoonbills (*Platalea minor*) fitted with solar-powered satellite telemetry devices that completed at least one spring or autumn migration during 1999–2018 from published literature. **Table S3.** Sites used by Black-faced Spoonbills (*Platalea minor*) identified according to satellite tracking data from published literature. **Table S4.** Wintering counts of waterbirds in Poyang Lake, Wang Lake and Shijiu Lake during synchronized Yangtze waterbird surveys in 2004, 2005, 2016, 2018, 2019 and 2020.

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Authors' contributions

LC, YC, YTY and ADF conceived the ideas and designed methodology; YTY and YC collected the data; YC and XD analysed the data. YC led the writing of the manuscript, with contributions from FM, YY and ADF. All authors contributed critically to the drafts. YC and YTY contributed equally to the manuscript. All authors read and approved the final manuscript.

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Availability of data and materials

The data generated or analyzed during this study are included in this published article and its additional information files.

Declarations

Ethical approval and consent to participate

This study was examined and approved by the Animal Ethics Committee, Research for Eco-Environmental Sciences, Chinese Academy of Sciences, and was conducted under permission No. rcees-dlll-001.

Consent for publication

Not applicable.

Competing interests

The authors declare that they have no competing interests.

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