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Eco-biochemical responses, phytoremediation potential and molecular genetic analysis of *Alhagi maurorum* grown in metal-contaminated soils

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Abstract

Background: Alhagi maurorum Medik. (camelthorn) is a dominant desert plant indigenous in various habitats, including the Western Desert of Egypt. The plant is especially prevalent in and around economic iron ore deposits. Nutrient and heavy metal levels in *A. maurorum* tissues and soil samples were assessed to identify associations between heavy metal levels in plants and soil. The objective was to evaluate this species as an indicator of heavy metal pollution. Photosynthetic pigments, protein, proline, alkaloids, flavonoids, 2,2-diphenyl-1-picrylhydrazyls-cavenging, reduced glutathione, malondialdehyde, antioxidant enzymes, and stress-related gene expression were assessed to determine their functional roles in metal stress adaptation in ultra- and molecular structure. Additionally, the molecular genetic variation in *A. maurorum* samples was assessed using co-dominant sequence-related amplified polymorphism (SRAP) and inter simple sequence repeats (ISSR).

Results: A substantial difference in enzymatic and non-enzymatic antioxidants of *A. maurorum* was observed in samples collected from three sites. *A. maurorum* is suited to the climate in mineralized regions. Morphologically, the stem shows spines, narrow leaves, and a reduced shoot system. Anatomically, modifications included a cuticle coating on leaves and stems, sunken stomata, a compact epidermis, and a thick cortex. Significant anatomical-physiological differences were observed with varying heavy metal soil content, antioxidative enzyme activities increased as a tolerance strategy, and glutathione levels decreased in response to heavy metal toxicity. Heavy metal accumulation also affected the expression of stress-related genes. The highest levels of expression of *GST*, *G6PDH*, *6PGD*, *nitrate reductase* 1, and *sulfate transporter* genes were found in plants collected from site A1. However, *auxin-induced protein* exhibited its highest expression in plants collected from A2. Six SRAP combinations yielded 25 scoreable markers with a polymorphism rate of 64%, and 5 ISSR markers produced 11 bands with a polymorphism rate of 36.36% for three *A. maurorum* genotypes. The ME1xEM7 primer combinations provided the most polymorphic information content and resolving power, making it the most useful primer for differentiating *A. maurorum* genotypes. SRAP markers exhibited a higher diversity index (0.24) than ISSR markers (0.16).

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Conclusions: A. maurorum displayed adaptive characteristics for heavy metal sequestration from mining site soils and is proposed as a strong candidate for phytoremediation.

Keywords: Alhagi maurorum, Sequence-related amplified polymorphism, Inter simple sequence repeats, Heavy metals, Genetic variation, Antioxidants, gene expression, Eco-biochemical traits

Background

Heavy metals (HMs) are released by mining and metallurgical operations and thus, cause a threat to the environment. Fauna and flora sequester large concentrations of HMs in these environments [1]. Egypt has a long history of mining, which dates to predynastic times. The country has significant mineral wealth, with iron, phosphates, gold, and salts being the most important in terms of production volume. Mining of economic iron ore deposits at El-Wahat El-Bahariya extracts ores with an average iron content of 47.6% [2]. Groundwater at the Bahariya Oasis is used to help meet water demand for local agriculture, yet this resource shows iron levels as high as 40–60 mg/L on average. Such levels are substantially higher than the permissible limit for agricultural purposes (5 mg/L) [3].

Plants can be used for biological control of soil, air, and water contamination [4, 5]. Phytoremediation encompasses all plant-based bioremediation technologies [6]. Phytoremediation of soil polluted with HMs is recognized as a cost effective and environmentally sustainable clean-up technology. This approach employs plants to minimize, eliminate, immobilize environmental pollutants to restore sites for alternative private or public use [7]. Alhagi maurorum Medik., Fabaceae (commonly referred to as camelthorn, camelthorn-bush, Caspian manna, or Persian mannaplant) is an indigenous plant in the deserts of Persia, Egypt, Syria, Pakistan, and India but has been introduced to other parts of the world [8]. The plant notably grows well in iron mining areas and tolerates salty, sandy, rocky, and dry soils. A. maurorum is purgative, diaphoretic, and expectorant used in folk medicine to treat piles, migraines, warts, and rheumatism [9].

HMs accumulate in various plant parts depending on species, metals, and soil conditions [10, 11]. The availability of HMs for plant uptake is affected by soil parameters, such as organic matter, pH, and cation exchange capacity [12]. Some metals, such as copper (Cu), cobalt (Co), manganese (Mn), iron (Fe), zinc (Zn), and molybdenum (Mo) are essential for plant metabolism and growth but may be poisonous at supra-optimal levels [13]. HM phytotoxicity can be caused by a variety of cellular and molecular mechanisms, such as blocking functional groups of metabolic molecules, inactivating enzymes, displacing or substituting for essential components, and disrupting membrane integrity [14]. Hyperaccumulating plants exhibit metallic or metalloid elements in aerial tissues to levels that exceed usual physiological requirements for most plant species [15]. HM concentrations in tissues of such plants are commonly used to monitor environmental contamination resulting from iron mining [16].

Abiotic stress causes plants to produce reactive oxygen species (ROS), which can disrupt cell biomolecules [17]. Peroxidation of membrane lipids is a primary mechanism for ROS-induced toxicity and may be assessed by measuring malondialdehyde (MDA) levels. Tolerant plants develop antioxidant mechanisms involving antioxidant enzymes, such as catalase (CAT), superoxide dismutase (SOD), ascorbate peroxidase (APX), metabolites (phenolics and carotenoids), and expression of stress-tolerance genes to counteract the negative impacts of ROS [18]. Forest trees grown in lead-contaminated soil with different levels of water stress exhibited significantly increased peroxidase (POD) and SOD levels [19]. Secondary metabolites have a key role in plant growth and development under normal conditions and are also defense and tolerance mechanisms in response to environmental stress [20]. Further, glutathione (GSH) has a role in cellular protection against xenobiotics, oxyradicals, and metal cations [21]. GSH mitigates metal toxicity in cells by chelating metal ions and protects macromolecules by trapping free radicals [22]. ROS production in chloroplasts increases in response to environmental stress, and GSH involvement in antioxidative defenses justifies its use as a stress indicator [23].

HMs are often translocated and deposited in the root cell walls [24]. Exodermis and endodermis are important barriers to the absorption of metal ions [25]. Further, metal absorption and tolerance are affected by changes in leaf tissues and studying such changes may produce a more complete understanding of associated processes [26]. Awmack and Lock [27] found that the xylem of *A. maurorum* is well-formed, supporting the conduction of significant amount of water from root to stem. Also, the parenchymatous pith of roots retains moisture as a xero-phytic adaptation.

Various molecular markers are applied alone or in combination to evaluate genetic diversity and phylogenetics in plant species [28, 29]. The sequence-related amplified polymorphism (SRAP) approach is a highly repeatable DNA sequencing method [30], used in various applications, including genetic diversity assessment [31]. SRAP markers are selected over other molecular markers to take advantage of their simplicity, access to several co-dominant markers, and targeting of open reading frames [31, 32]. SRAP markers were successfully utilized to analyze genetic diversity and population structures of numerous species, including *Brassica juncea* (brown mustard) [33], *Melia* species [34], and *Carthamus tinctorius* L (Safflower) [35].

Dominant inter simple sequence repeat (ISSR) markers have several advantages, including the ability to detect high levels of polymorphism at a low cost and simple method with an outstanding stability and repeatability [36–38]. ISSR marker technology is successfully used in genetic diversity studies, DNA fingerprinting, and germplasm assessment in plants, including *Tuberaria major* [39], *A. maurorum* [40], and *Alhagi* sp. [38].

The present study evaluated molecular genetic variation and eco-biochemical and anatomical attributes of *A. maurorum* growing in iron mining areas at El-Gedida in Egypt. The assessment also focused on the possible utility of this species for phytoremediation. Findings provide a deeper understanding of HM accumulation in various tissues of *A. maurorum*.

Results

Analysis of soil and heavy metal availability

Table 1 indicates the mean values of the characteristics of topsoil samples from iron mining studied sites (A1, A2 and A3). pH of soils tended toward alkaline (7.73–8.98). Organic matter values ranged between 3.0 and 3.5%, indicating relatively poor content. EC ranged from 0.40 (A1) to 7.88 (A3) mmho/cm. The latter soil is from Harrah Oasis, where high salinity is observed. The highest value for K (1611.0 ppm) was recorded at A1, and the highest concentrations for total N (839.0 ppm) and total P (163.0 ppm) were recorded at A2. EC and pH are the most important factors since HMs are less available under alkaline conditions.

HM levels in the soil showed significantly higher levels of Zn, Ni, Cu, and Fe in soil from the iron ore site (A1). Bioavailability of HMs for composite soil samples from the iron mining site (A1) were, in descending order, Pb > Ni > Mo > Fe > Cu > Cd > Co > Mn > Cr > Zn > B. At A2, the order was Pb > Mo > Ni > Co > Mn > Cd > F e > Cr > Cu > Zn, and for sandy soil at A3, which exhibited the lowest levels of HMs, Cr > Mo > Ni > Fe > Mn > Cu > Pb > Co > B > Zn > Cd (Table 1).

Pearson's correlation coefficients were strongly positive among metals associated with iron ore, regardless of concentration: 0.99 for K-Fe, 0.91 for Fe-Zn and 0.99 for Fe-Cu. Correlation coefficients among particular metals, EC, and pH were strongly negative: -0.99 for pH-total nitrogen (N) and pH-Ni, -0.96 for EC-Cd and EC-Mo, -0.94 for EC-Zn, and -0.93 for EC-Pb. Total nitrogen (N) was strongly and positively correlated with several HMs: 0.98 for N-Ni, N-Cd and N-Mo; 0.96 for N-Pb; 0.88 for N-Zn; and 0.86 for N-B. All samples were poor in organic matter, and a weak correlation between organic matter and HM content might exist (Table 2).

Table 1 Geochemical characteristics of soil collected from the studied area

Parameters	A1	A2	A3	<i>F</i> -value
рН	7.73 ± 0.223^{a}	7.79 ± 0.270^{a}	8.98 ± 0.207^{b}	9.003
EC (mmho/cm)	0.40 ± 0.006^{a}	1.22 ± 0.027^{b}	$7.88 \pm 0.132^{\circ}$	2771.371
OC%	3.00 ± 0.069^{a}	3.50 ± 0.121^{a}	3.20 ± 0.185^{a}	3.543
N (ppm)	816.0 ± 28.267^{b}	839.0 ± 24.220^{b}	$224.0 \pm 9.053^{\circ}$	248.442
Ρ	87.00 ± 3.014^{b}	$163.00 \pm 4.705^{\circ}$	29.00 ± 0.837^{a}	424.377
К	1611.0 ± 46.506 ^c	794.0 ± 18.337^{b}	346.0 ± 9.988^{a}	474.922
Fe (mg/kg)	$24.50 \pm 0.566^{\circ}$	10.50 ± 0.303^{b}	6.34 ± 0.220^{a}	590.005
Zn	$3.06 \pm 0.088^{\circ}$	2.30 ± 0.066^{b}	1.26 ± 0.036^{a}	180.992
Mn	8.02 ± 0.185^{b}	$17.40 \pm 0.603^{\circ}$	5.38 ± 0.124^{a}	289.835
Cu	$17.09 \pm 0.493^{\circ}$	6.40 ± 0.222^{b}	3.84 ± 0.133^{a}	477.687
Co	8.60 ± 0.248^{b}	$18.30 \pm 0.423^{\circ}$	1.80 ± 0.0520^{a}	849.135
Ni	$32.10 \pm 0.927^{\circ}$	28.50 ± 0.987^{b}	6.90 ± 0.199^{a}	297.525
Cd	9.22 ± 0.213^{b}	$10.89 \pm 0.314^{\circ}$	1.12 ± 0.019^{a}	566.801
Cr	6.90 ± 0.199^{a}	7.91 ± 0.228^{a}	7.82 ± 0.316^{a}	4.892
Pb	33.00 ± 0.762^{b}	$45.00 \pm 1.299^{\circ}$	2.00 ± 0.058^{a}	650.193
Мо	29.90 ± 0.691^{b}	$34.50 \pm 0.797^{\circ}$	7.00 ± 0.162^{a}	572.109
В	$2.50\pm0.087^{\mathrm{b}}$	$3.60 \pm 0.083^{\circ}$	$1.40\pm0.032^{\rm a}$	234.840

The results are recorded as Mean of triplicates \pm Standard Error (SE). Different superscript letters refer to significant differences at (P < 0.05) level

Parameters	рН	EC	OC	Ν	Р	к	Fe	Zn	Mn	Cu	Co	Ni	Cd	Cr	Pb	Мо	в
pН	1																
EC	0.87**	1															
OC	-0.12	-0.01	1														
Ν	-0.91**	-0.99**	0.11	1													
Р	-0.66	-0.76*	0.5	0.83**	1												
к	-0.68*	-0.83**	-0.42	0.74*	0.28	1											
Fe	-0.6	-0.75*	-0.52	0.64	0.14	0.99**	1										
Zn	-0.79*	-0.94**	-0.21	0.88**	0.51	0.95**	0.91**	1									
Mn	-0.53	-0.59	0.65	0.68*	0.97**	0.04	-0.1	0.3	1								
Cu	-0.55	-0.72*	-0.51	0.61	0.11	0.98**	0.99**	0.89**	-0.13	1							
Co	-0.71*	-0.75*	0.49	0.83**	0.99**	0.25	0.12	0.48	0.97**	0.08	1						
Ni	-0.85**	-0.994**	0.019	0.976**	0.738*	0.83**	0.76*	0.96**	0.57	0.74*	0.72*	1					
Cd	-0.86**	-0.96**	0.17	0.98**	0.90**	0.66	0.55	0.81**	0.77^{*}	0.52	0.89**	0.94**	1				
Cr	0.62	0.41	0.31	-0.37	0.15	-0.71*	-0.73*	-0.59	0.31	-0.72*	0.12	-0.42	-0.23	1			
Pb	-0.81**	-0.93**	0.31	0.96**	0.94**	0.56	0.45	0.76*	0.85**	0.42	0.93**	0.92**	0.98**	-0.12	1		
Мо	-0.86**	-0.96**	0.17	0.98**	0.87**	0.66	0.56	0.81**	0.77^{*}	0.52	0.89**	0.94**	1.00**	-0.23	0.98**	1	
В	-0.67*	-0.81**	0.43	0.86**	0.99**	0.35	0.22	0.57	0.95**	0.19	0.98**	0.78^{*}	0.93**	0.12	0.96**	0.92**	1

 Table 2
 Pearson correlation between geochemical parameters of the study area

**Correlation is significant at the 0.01 level

*Correlation is significant at the 0.05 level

Metal concentrations and NPK

N, P, and K contents were significantly higher in plants obtained from site A1 compared to A2 and A3. *A. maurorum* roots at A1 accumulated trace metals as Fe > Ni > Cu > Al > Zn > Mn > Pb > Cr > Cd > Co > B. Similarly, shoots accumulated metals as <math>Fe > Ni > Al > Cu > Zn > Mn > Pb > Cr > Co > Cd (Table 3).

Phytoremediation efficiency of A. maurorum

HM accumulation and upward translocation were evaluated using BCFs, a simple method for quantitative characterization of available HM uptake from soil to plant. BCF was >1 for Fe, Zn, Cd, Co, Cr, and Pb, but <1 for B, Mn, and Ni at all sites. Further, BCFs were less than unity for Cu at A2 and for Mo at A1 and A2 (Table 4).

The TF, a measure of metal transport from roots to shoots, was greater than one for Mn, Pb, Co, Zn, Cr, Ni, Cd, and Cu, and less than unity for Fe, Mo, and B, at all sites.

HM accumulation was determined by SCF and exceeded unity at all sites for Zn, Mn, Cd, Pb, and Co, but was <1 for B. SCF was less than unity for Cu, Ni, and Mo at A1 and A2; for Fe at A2; for Cr at A3 (Table 4).

RCF exceeded unity for Mn, Pb, Cd, Co, and Zn at all sites and was >1 for Fe at A1 and A3, for Mo and Ni at A3, and for Cr at A1. The RCF for Cu and B was less than unity at all sites (Table 4).

Biochemical parameters of A. maurorum shoots

Analysis for bioorganic substances can detect important macromolecules, including pigments and primary and secondary metabolites, which could regulate various biological processes.

Non-enzymatic antioxidant activities

Chlorophyll a levels in leaves were 18.49 (A2) > 14.50 (A1) > 9.24 (A3) mg. g⁻¹ fresh wt. Chlorophyll b levels were 19.16 (A2) > 15.84 (A1) > 9.92 (A3) mg. g⁻¹ fresh wt. Carotenoids exhibited concentrations 0.27, 0.33, and 0.14 mg. g⁻¹ fresh wt. for A1, A2, and A3, respectively. Moreover, total pigments were calculated as 30.61, 37.97, and 19.30 mg. g⁻¹ fresh wt, for A1, A2, and A3, respectively (Fig. 1A). Significant differences in protein content of*A. maurorum*shoots were 61.38 mg. g⁻¹ fresh wt. at A1 and 11.19 mg. g⁻¹ fresh wt. at A3. Proline content was highest at A3, 42.21 mg. g⁻¹ fresh wt. (Fig. 1B).

Total flavonoid content gradually decreased as 8.68 (A1) > 5.71 (A2) > 4.84 (A3) mg. g^{-1} dry wt. Alkaloid content of shoot differed as 5.50 (1A) > 3.25 (A3) > 1.59 (A2) % dry weight (Fig. 1C).

Shoot extracts showed significant DPPH free radical scavenging impacts compared to ascorbic acid. The DPPH scavenging varied as 21.63, 57.68, and 74.63% at A1, A2, and A3, respectively (Fig. 1D). Lipid peroxidation levels in shoots gradually decreased as 13.85 at A1, 7.41 at A2 and 0.71 nM/L MDA at A3 (Fig. 1D). Reduced GSH analytes

appeared as major peaks at retention times of 2.62 and 3.50 min. The HPLC chromatogram suggested that reduced GSH was detected under HM stress at 3.50 min (Fig. 2). GSH concentrations in leaves from A1, A2, and A3 were 32.59, 69.67, and 47.14 μ g/mL, with retention times of 3.51, 3.53, and 3.53 min, respectively (Table 5 and Fig. 1D).

Enzymatic antioxidant stress modulators

Antioxidant enzyme activities for *A. maurorum* shoots were highest at site A1 for CAT ($5.50 \text{ mg} \cdot \text{g}^{-1}$ fresh wt.), POD ($2.34 \text{ mg} \cdot \text{g}^{-1}$ fresh wt.) and polyphenol oxidase (PPO) ($22.20 \text{ mg} \cdot \text{g}^{-1}$ fresh wt.) activities (Fig. 1E).

Semi-thin ultrastructure of A. maurorum shoots Description of leaf sections

Histological analysis of transverse sections of leaf blades showed typical three tissue differentiation: epidermis, mesophyll, and vascular. The outermost layer consisted of adaxial and abaxial thick-walled elongated epidermal cells covered with a thick layer of cuticle. Tanninferous cells or glands were distributed as a compact layer of small cells underneath the lower epidermis or large scattered cells in the periphery in the upper palisade parenchyma. Mesophyll tissue was bifacial and differentiated into palisade and spongy. The leaf was dorsiventral, and palisade mesophyll parenchyma was found on both adaxial and abaxial sides with a strip of spongy parenchyma in the middle portion of the lamina. Vascular bundles of the midvein were open collateral with narrow cambial tissue (Fig. 3).

The greatest leaf thickness was recorded at site A1 and the lowest at A2. The cuticular thickness of the adaxial epidermal surface was more than the abaxial surface (Fig. 4A–D). Moreover, tanninferous cells or tannin

Table 3 Heavy metal and macro-nutrient concentrations in different tissues of Alhagi maurorum collected from the study area

Element/ Ti	nent/Tissue A1		A2	A3	F- value
Micro-nutri	ents (Heavy metal	s) concentrations (mg/kg)			
Fe	Root	$225.45 \pm 11.524^{\circ}$	8.90 ± 0.208^{a}	137.60 ± 7.843^{b}	183.087
	Shoot	213.75±11.841°	7.18 ± 0.310^{a}	77.60 ± 4.322^{b}	208.079
Zn	Root	16.15 ± 0.566^{a}	19.73 ± 0.935^{b}	15.32 ± 0.750^{a}	9.376
	Shoot	18.15 ± 0.589^{ab}	19.31 ± 0.693^{b}	15.43 ± 0.704^{a}	8.995
Mn	Root	101.23 ± 5.324^{b}	161.80 ± 84.686^{a}	18.22 ± 0.808^{a}	174.263
	Shoot	195.70 ± 10.883^{a}	$190.30 \pm 9.925^{\circ}$	18.54 ± 0.885^{a}	3.638
Ni	Root	14.93 ± 0.530^{a}	15.32 ± 0.462^{a}	15.20 ± 0.872^{a}	0.352
	Shoot	15.43 ± 0.819^{a}	15.43 ± 0.398^{a}	$15.56 \pm 0.606^{\circ}$	0.003
Cr	Root	12.30 ± 0.398^{b}	6.53 ± 0.242^{a}	5.16 ± 0.103^{a}	100.914
	Shoot	$16.62 \pm 0.774^{\circ}$	13.09 ± 0.930^{b}	7.70 ± 0.243^{a}	70.143
Pb	Root	101.50 ± 4.619^{a}	100.40 ± 5.277^{a}	101.00 ± 0.462^{a}	0.015
	Shoot	102.00 ± 4.619^{a}	101.00 ± 5.600^{a}	101.40 ± 5.687^{a}	0.018
В	Root	$0.68 \pm 0.014^{\circ}$	0.16 ± 0.005^{a}	0.77 ± 0.009^{b}	2958.499
	Shoot	0.13 ± 0.004^{a}	0.15 ± 0.001^{a}	0.52 ± 0.017^{b}	442.859
Мо	Root	15.00 ± 0.572^{b}	15.16 ± 0.672^{b}	11.80 ± 0.427^{a}	9.935
	Shoot	13.40 ± 0.468^{ab}	10.22 ± 0.507^{a}	11.31 ± 0.522^{a}	21.977
Cu	Root	$4.59 \pm 0.087^{\circ}$	2.50 ± 0.017^{a}	2.12 ± 0.060^{a}	372.091
	Shoot	$12.75 \pm 0.710^{\circ}$	3.30 ± 0.035^{b}	5.60 ± 0.237^{b}	147.864
Со	Root	30.00 ± 1.155^{a}	30.93 ± 1.698^{a}	30.56 ± 0.883^{a}	0.053
	Shoot	30.31 ± 1.345^{a}	30.94 ± 1.703^{a}	31.89 ± 1.651^{a}	0.385
Cd	Root	15.40 ± 0.722^{a}	15.10 ± 0.606^{a}	14.16 ± 0.601^{a}	1.338
	Shoot	15.56 ± 0.768^{a}	15.50 ± 0.791^{a}	16.69 ± 0.722^{a}	0.669
Macro-nutr	ients (ppm)				
Ν	Root	$18,000.0 \pm 980.918^{\circ}$	5000.0 ± 248.838^{a}	6000.0 ± 302.53^{a}	102.481
	Shoot	$25,500.0 \pm 1385.06^{b}$	$21,000.0 \pm 1155.28^{b}$	9500.0 ± 501.717^{b}	93.509
К	Root	$4539.00 \pm 228.631^{\circ}$	890.0 ± 28.290^{a}	1068.00 ± 45.611^{a}	110.375
	Shoot	9612.0 ± 520.193^{b}	3738.00 ± 188.216^{b}	1691.0 ± 45.611^{a}	254.967
Р	Root	3676.59 ± 159.112^{b}	720.90 ± 55.945^{a}	865.08 ± 20.738^{a}	77.982
	Shoot	7785.72 ± 395.323^{b}	3027.80 ± 240.051^{b}	1369.71 ± 45.443^{a}	283.318

The results are recorded as Mean of triplicates \pm Standard Error (SE). Different superscript letters refer to significant differences at (P < 0.05) level

Element/ parar	neter	A1	A2	A3	<i>F</i> -value
Fe	BCF	17.98 ± 1.112 ^b	$1.54 \pm 0.098^{\circ}$	34.01 ± 0.999 ^c	352.465
	TF	0.95 ± 0.032^{b}	0.81 ± 0.012^{b}	0.56 ± 0.043^{a}	37.467
	SCF	8.75 ± 0.687^{b}	0.68 ± 0.049^{a}	$12.22 \pm 0.260^{\circ}$	193.831
	RCF	9.22 ± 0.427^{b}	0.85 ± 0.049^{a}	21.79 ± 1.257 ^c	188.881
Mn	BCF	0.24 ± 0.012^{b}	0.07 ± 0.003^{a}	0.19 ± 0.017^{b}	53.725
	TF	1.93 ± 0.064^{b}	1.18 ± 0.001^{a}	1.02 ± 0.078^{a}	69.680
	SCF	$24.49 \pm 1.926^{\circ}$	10.99 ± 0.700^{b}	3.46 ± 0.245^{a}	79.992
	RCF	$12.65 \pm 0.583^{\circ}$	9.34 ± 0.595^{b}	3.39 ± 0.020^{a}	95.100
Pb	BCF	6.17 ± 0.124^{a}	4.49 ± 0.202^{a}	101.31 ± 1.831 ^b	2702.955
	TF	1.01 ± 0.064^{a}	1.01 ± 0.081^{a}	1.01 ± 0.090^{a}	0.001
	SCF	3.09 ± 0.035^{a}	2.25 ± 0.191^{a}	50.62 ± 1.386^{b}	1173.761
	RCF	3.08 ± 0.159^{a}	2.23 ± 0.012^{a}	50.69 ± 3.217 ^b	222.388
Co	BCF	7.03 ± 0.306^{a}	3.37 ± 0.116^{a}	34.78 ± 1.531 ^b	361.254
	TF	1.02 ± 0.087^{a}	1.00 ± 0.002^{a}	1.04 ± 0.078^{a}	0.105
	SCF	$3.54\pm0.306^{\rm a}$	1.69 ± 0.058^{a}	17.80 ± 1.432^{b}	108.503
	RCF	3.49 ± 0.001^{b}	1.69 ± 0.061^{a}	$16.99 \pm 0.098^{\circ}$	15,744.450
Zn	BCF	11.21 ± 0.020^{a}	17.01 ± 0.685 ^b	$24.43 \pm 0.497^{\circ}$	184.221
	TF	1.12 ± 0.003^{a}	0.99 ± 0.095^{a}	1.01 ± 0.075^{a}	1.052
	SCF	5.93 ± 0.023^{a}	8.39 ± 0.058^{b}	$12.24 \pm 0.208^{\circ}$	643.149
	RCF	5.28 ± 0.003^{a}	8.62 ± 0.745^{b}	$12.20 \pm 0.705^{\circ}$	34.165
Cr	BCF	$4.19 \pm 0.061^{\circ}$	2.48 ± 0.026^{b}	1.65 ± 0.084^{a}	443.365
	TF	1.36 ± 0.104^{a}	2.01 ± 0.188^{b}	1.50 ± 0.121^{ab}	5.799
	SCF	$2.41 \pm 0.043^{\circ}$	1.65 ± 0.069^{b}	0.99 ± 0.018^{a}	216.698
	RCF	1.79 ± 0.104^{b}	0.83 ± 0.043^{a}	0.67 ± 0.064^{a}	65.785
Ni	BCF	0.59 ± 0.009^{a}	0.56 ± 0.017^{a}	4.46 ± 0.127^{b}	912.759
	TF	1.04 ± 0.081^{a}	1.01 ± 0.049^{a}	1.03 ± 0.116^{a}	0.021
	SCF	0.30 ± 0.006^{a}	0.28 ± 0.001^{a}	2.24 ± 0.064^{b}	935.757
	RCF	0.29 ± 0.017^{a}	0.28 ± 0.015^{a}	2.21 ± 0.191 ^b	101.071
Cd	BCF	3.39 ± 0.116^{a}	2.81 ± 0.032^{a}	27.56 ± 0.390^{b}	3599.182
	TF	1.03 ± 0.081^{a}	1.03 ± 0.020^{a}	1.19 ± 0.121^{a}	1.280
	SCF	1.72 ± 0.121^{a}	1.42 ± 0.032^{a}	14.93 ± 0.904^{b}	214.517
	RCF	1.67 ± 0.012^{a}	1.39 ± 0.001^{a}	12.62 ± 0.511^{b}	471.226
Мо	BCF	0.95 ± 0.006^{a}	0.74 ± 0.049^{a}	3.30 ± 0.095^{b}	525.331
	TF	$0.89\pm0.009^{\mathrm{b}}$	0.67 ± 0.006^{a}	$0.96 \pm 0.012^{\circ}$	283.136
	SCF	0.45 ± 0.001^{b}	0.30 ± 0.023^{a}	$1.62 \pm 0.038^{\circ}$	801.571
	RCF	0.50 ± 0.006^{a}	0.44 ± 0.029^{a}	1.68 ± 0.058^{b}	350.235
Cu	BCF	1.01 ± 0.020^{a}	0.91 ± 0.029^{a}	2.02 ± 0.121^{b}	70.260
	TF	2.79 ± 0.159^{b}	1.33 ± 0.069^{a}	2.65 ± 0.156^{b}	35.914
	SCF	0.75 ± 0.003^{a}	0.52 ± 0.006^{a}	1.47 ± 0.113^{b}	57.950
	RCF	0.27 ± 0.017^{a}	$0.39\pm0.026^{\text{b}}$	$0.55 \pm 0.009^{\circ}$	57.358
В	BCF	$0.32\pm0.020^{\mathrm{b}}$	0.09 ± 0.003^{a}	$0.92 \pm 0.012^{\circ}$	995.820
	TF	0.19 ± 0.002^{a}	0.94 ± 0.034^{c}	0.68 ± 0.003^{b}	358.633
	SCF	$0.05\pm0.003^{\text{a}}$	0.04 ± 0.001^{a}	0.37 ± 0.006^{b}	2355.250
	RCF	0.27 ± 0.017^{b}	$0.04\pm0.003^{\rm a}$	$0.55 \pm 0.006^{\circ}$	561.032

Table 4 Phytoremediation potential: bioaccumulation factor (BCF), translocation factor (TF), shoot coefficient factor (SCF) and root coefficient factor (RCF) for *A. maurorum* at three studied sites

The results are recorded as Mean of triplicates \pm Standard Error (SE). Different superscript letters refer to significant differences at (P < 0.05) level



polyphenol oxidase)

0.30 0.25

glands accumulated underneath the lower epidermis as small cells and scattered as large cells under the upper epidermis. The number and area of the tanninferous cells decreased as A3>A1>A2 sites. The palisade mesophyll

consisted of 3-4 layers of elongated parenchyma with

abundant chloroplasts and tanninferous cells toward the

adaxial epidermis. Mesophyll with palisade and spongy

thickness was maximum at site A1 > A3 and A2. Vascular

bundle area was highest at site A3 and lowest at A2. A. maurorum collected from A3 site showed the widest metaxylem and thickest phloem.

Description of stem sections

The results of the anatomical features of the A. maurorum stem between the three studied sites were detected in Fig. 4 (E and F). The epidermal cells were covered by



a thick layer of cuticle, except for the openings of the sunken stomata pores. The highest epidermal cell thickness was scored in A2 > A1 > A3 sites. The *A. maurorum* stem was considered a xerophytic type, with a maximum chlorenchyma thickness for A3 site and minimum for A2 site. Maximum vascular bundle thickness was recorded for A3 site, while the minimum was for A1. Furthermore, *A. maurorum* present at A3 site had thicker and stronger xylem and phloem fibres. Pith parenchymatous cell diameter was the widest in *A. maurorum* of A1, A2 and A3 sites.

Transverse stem sections of A. maurorum stem (Fig. 5) also showed three tissue types: epidermal, ground, and vascular bundles. The epidermis consisted of compact elongated parenchymal cells with several sunken stomata. One compact single layer was found underneath the epidermis, called the hypodermis, with several small tanninferous cells containing tannins. Ground tissue differentiated into cortex and pith. The former consisted of several layers of chlorenchyma, with thin-walled parenchyma beneath these layers. These layers consisted of parenchymatous cells occupying the central core of the stem. Stem vascular bundles were differentiated into small and large open collateral types. Medullary rays were located between large phloem fibers. Tanninferous cells in the A. maurorum stem were scattered in the cortex, pith, and medullary rays.

Intercorrelation between heavy metals and different responses of *A. maurorum*

The heatmap provides classification of all the studied biochemical and anatomical parameters of the A. maurorum shoot system using Pearson correlation (Fig. 6). Fe had significantly strong positive correlations with other HMs such as Mo and Cu, as well as with K, P, POD, protein, flavonoids and mesophyll thickness. On the other hand, Fe was negatively correlated with DPPH, reduced GSH and gland leaf area. Mn was positively correlated with the other metals (Zn, Cr and Cu). Similarly, Mn was positively correlated with PPO, MDA and pith area, but negatively correlated with Cd, B, proline, gland leaf area and leaf vascular bundle area. Co, Ni and Cd were positively correlated with Pb, but negatively correlated with Mo and photosynthetic pigments. Zn had a positive correlation with all other HMs except Cd, Cu and B. In addition, Zn was negatively correlated with proline, DPPH, alkaloids and POD. Moreover, Mo was positively correlated with Cu, K, P, protein, flavonoids, POD and mesophyll thickness.

Micronutrients (N, P and K) were positively correlated to each other and with Cr and Mn. They also had a positive impact on various parameters such as protein, flavonoids, MDA, POD and PPO. Conversely, changes in N, P and K were inversely correlated with changes in proline, DPPH, reduced GSH, shoot vascular bundle area and chlorenchyma thickness. Additionally, photosynthetic pigments (chlorophyll a, chlorophyll b and carotenoids) showed a positive relationship with each other and with other parameters such as protein, flavonoids, MDA, GSH and PPO, while they were inversely correlated with proline, DPPH, alkaloids, CAT, POD, leaf thickness, mesophyll and chlorenchyma thickness, shoot vascular bundle area and stem gland area. Flavonoids were significantly positively correlated with changes in total protein content. Cu and Fe were closely related to them. Flavonoids and protein content were positively correlated with MDA and enzymatic antioxidants. Additionally, proline was significant closely correlated with DPPH, while it had a negative correlation with protein, flavonoids and MDA. Furthermore, alkaloids were positively correlated with CAT (Fig. 6).

Molecular genetic analyses

Expression analysis of stress-related genes

In order to investigate the effect of heavy metal accumulation in the three studied sites at the molecular level, the expression levels of 6 genes (*GST*, *G6PDH*, *6PGD*, *nitrate reductase 1*, *sulfate transporter* and *auxin-induced protein*) in *A. maurorum* plants collected from the different studied sites were estimated and shown in Figs. 7 and 8. Notably, the highest levels of expression of *GST*, *G6PDH* and *6PGD* genes have been detected in plants collected from site A1, followed by that recorded in plants collected from A2 and A3 sites (Fig. 7).

Similarly, Fig. 8 shows considerable variations in the expression levels of *nitrate reductase 1, sulfate transporter* and *auxin-induced protein* in *A. maurorum* plants collected from different sites. The highest levels of expression of *nitrate reductase 1* and *sulfate transporter* have been detected in plants collected from site A1. However, *auxin-induced protein* exhibited its highest expression level in plants collected from site A2 (5 and 23% increases over that recorded in plants collected from A1 and A3 sites, respectively). These results indicate the heavy metal accumulation impacts on the expression of stress-related genes in plants collected from the studied sites.

Molecular genetic diversity analysis using ISSR and SRAP markers

Five ISSR primers were used and yielded 11 scoreable and reproducible fragments, 7 of which were monomorphic and 4 were polymorphic, with an average polymorphism percentage of 36.36%. MBF values varied from 0.83 in 49A, HB-10 and HB-11 primers to 1.00 in HB-9 primer. As shown in Table 6, each primer yielded distinct

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Sample code	Concentration (µg/ mL)	Rt (min)	Area
GSH	50	3.562	206,365
Standard	100	3.578	398,534
	150	3.581	564,721
	200	3.578	758,831
	250	3.596	966,297
A1	32.59128	3.510	137,455
A2	69.67321	3.526	276,895
A3	47.14322	3.528	192,175

 Table 5
 Concentration, area and retention time of reduced glutathione of standard of A. maurorum samples

amplification products with sizes varying from 200 to 900 bp. The polymorphism percentage for each primer ranged from 33.33% in HB-13 to 50% in 49A, HB-10, and HB-11. Levels of polymorphism and informativeness of ISSR markers were also expressed in some variables such as PIC, DI, and RP. The highest PIC value (0.22) was found in primers (49A, HB-10 and HB-11), while the lowest (0.00) was recorded in primer HB-9. The primers 49A, HB-10, HB-11, and HB-13 showed the highest Rp values (0.67), with an average of 0.53. All ISSR primers had a DI of 0.16 (Table 6).

SRAP analysis was also carried out to validate ISSR analysis results. Out of the 6 SRAP primers combinations screened, 25 amplicons in total were generated with 9 monomorphic and 16 polymorphic. MBF values in ME1xEM7 primer combinations and ME1xEM8 primer combinations varied from 0.5 to 0.83, respectively. The



Fig. 3 Fiscological analysis of transverse section of *A. maliforum* lear phenotype at 100X and magnified parts at 400X. **A** Lear intercostal region and margin of site A1 and its magnified part showing gland of site A1; **B** leaf midrib of site A2 and its magnified part showing gland of site A2; **C** Leaf intercostal region and margin of site A3 and its magnified part showing stomata. Black arrows indicated stomata openings; Figure abbreviations: Cu, cuticle; Epi, epidermis; PM, palisade mesophyll; SM, spongy mesophyll; VB, vascular bundle; Xy, xylem; Ph, phloem; SC, secretory cavity; Gl, gland; St, stomata; Stc, stomatal cavity. Bars = 10 μm



scorable fragments ranged in size from 100 to 1500 bp with an average polymorphism rate of 64%. ME1xEM7 primer showed the highest polymorphism (100%) followed by 71.42% in ME2xEM6, 66.66% in ME5xEM6, 50% in ME1xEM8 and lasted with 33.33% in ME2xEM7, with no percentage in ME5xEM8 (Table 6). Three out of the 7 unique bands were found in ME1xEM7 at the molecular sizes of 260, 490 and 1500 bp in *A. mauro-rum* at A1 site. ME2xEM6 primer has 2 unique bands of molecular sizes of 300 bp in *A. maurorum* at site A3 and



845 bp in *A. maurorum* at A1 site. Likewise, ME2xEM7 primer has one unique band of 190 bp in *A. maurorum* at A1 site and ME5xEM6 primer has a 140 bp unique band in *A. maurorum* at A1 site. Some parameters such as PIC, DI and RP indicated the ISSR markers effectiveness. ME1xEM7 primers showed the highest PIC (0.45), whereas ME5xEM8 primers had the lowest (0.00). The highest Rp value of the ME1xEM7 primers combination was 4.02, followed by 3.36 for ME2xEM6, and the lowest Rp value was 0.00 for the ME5xEM8 primers combination, with an average of 1.79. The DI of all SRAP primer combinations was 0.24 (Table 6).

Table 6 compared the performance between ISSR and SRAP analyses. Thirty-six amplification products were detected with 16 monomorphic, 20 polymorphic, and 7 unique bands generated by all ISSR and SRAP primers. SRAP primers had the maximum polymorphism percentage of 64%, compared to ISSR markers which had a polymorphism percentage of 36.36%. The total ISSR and SRAP polymorphism percentage was 55.55%. As indicated in Table 7, the maximum similarity index (0.939) was reported between *A. maurorum* at sites A2 and A3, followed by 0.652 at sites A1 and A2. However, the lowest similarity (0.578) was recorded at sites A1 and A3. To elucidate the genetic relationships among *A. maurorum* genotypes, the dendrogram based on SRAP and ISSR data was constructed (Fig. 9) and divided the samples into 2 main clusters. One cluster included *A. maurorum* at A1 site, while the other was sub-clustered into 2 groups separated by 0.3 genetic distance between *A. maurorum* plants at A2 and A3 sites.



Discussion

Metal bioaccumulation in plants is a major source of concern as the metal concentrations can surpass WHO-recommended safety levels, posing health and environmental risks to humans and environment [41]. Plants have developed sophisticated signalling systems, such as the 'universal' cascade, which can be used to monitor their reactions to HM stress [42]. Signalling channels are in charge of regulating these processes, which consist of a reception step (stimuli perception), a transduction step (intracellular and extracellular signal amplification) and, lastly, a response step (enzymatic or non-enzymatic) [43, 44]. Research studies should fine-tune these mechanisms to make plants safer for consumption through understanding the mechanisms used by plants to respond to HMs.

Fe is the most common element among the numerous metals examined in the groundwater samples of El-Wahat El-Bahariya [45]. In the present study, Fe and its related metals in the soils caused stress to the wild xerophytes growing in this deposit. Different tissues of *A. maurorum* showed different biochemical, ultrastructure and molecular genetic responses to HMs. Distinct variations were recorded in the mining area and other nearby locations in relation to environmental stresses, including aridity, HM toxicity and salinity.

EC and pH are important abiotic factors for determining plant establishment and colonisation in the zones degraded by mining operations, and could influence the metal ion biosorption process efficiency [46]. The soil water extracts were mildly alkaline at the studied sites, and there was a substantial salinity at A3. However, soil samples associated with plants at site A1 had a significantly lower pH than other sites. Higher soil pH levels result in an increased retention and lower HM solubility. A high level of HMs in the soil could suggest a similar concentration in plants, which could pose a significant risk to human health [47].

The presence of one HM in the soil and plant can affect the availability of another. In other words, HMs have both antagonistic and synergistic effects [48]. The interrelationship between soil HMs is very complex. However, in the present study, there was a synergistic relationship between Fe, Zn, Cu and Ni; Cd and Co; and Mo, B and Pb. Moreover, Cr demonstrated substantial antagonistic behaviour against most of the observed HMs.

The above-normal metal accumulation in various *A*. *maurorum* parts could suggest their resistance to HM



contamination in mining areas. As a result, creating a plant population around mining sites may assist in reducing the effects of mining by making soil more sustainable, either by lowering metal concentrations or by immobilising pollutants in the soil [49]. The concentrations of iron and other HMs were lower in the soil than that recorded in plant tissues. For example, cadmium and lead levels in sediments were lower than those in shoots. This result is similar to that recorded previously for the roots of Cyno*don dactylon* developing in the Nakivubo tributaries [50]. HMs are not only bio-accumulated in the roots of certain plants, but also translocated from the roots to the shoots of others [51]. Yangun et al. [52] found that A. maurorum accumulated above-normal levels of 5 mg/kg Pb and 10 mg/kg Cu in contrast to soils. A. maurorum accumulated HMs in plant parts more abundantly in shoots than in roots. For phytostabilisation, metal-resistant plants with a low metal concentration are preferred. Those plants accumulate HMs in their roots and are therefore weak translocators [53]. As metals are translocated to easily harvestable plant parts, phytoextraction can reduce HM levels in sediments to appropriate levels over time. In the present study, *A. maurorum* was chosen for phytostabilisation of Fe, B and Mo in its roots, while most HMs (Zn, Mn, Ni, Cr, Pb, Cu, Cd and other micro-elements) were translocated into shoot sections at threshold concentrations.

Differences in concentrations exist between organisms and plant components, suggesting their metal absorption capacities [54]. TF values (shoot/root quotient) less than unity indicate that metals have accumulated and are being processed in the root. For all HMs except Fe, Mo and B, TF values of >1 suggest preferential partitioning of metals in the shoots at the three sites studied. When



the TF value is greater than unity, further translocation occurs in the plant shoot, resulting in phytoextraction as a phytoremediation mechanism. Fast growth and a large root system are also preferred [55]. *Alhagi* species have the deepest root system of any plant in terms of proportion. It is a very hardy and destructive plant. Because of its extensive root system, it is very difficult to eradicate existing populations [56]. This was verified by our phytoextraction findings for most of the studied HMs.

The BCF, which is the ratio of metal concentration in plants to extractable metal concentration in the soil, is used to express accumulation quantitatively [57]. The current findings revealed that the BCF was > 1 for Fe, Pb, Co, Zn, Cr and Cd at all locations. It is concluded that *A. maurorum* could be used as bioaccumulative markers for these metals, and that their concentrations in the soil were well-represented by them. Similarly, *Plantago lanceolata* could be utilized as a bioaccumulative indicator not only for Pb and Zn, but also for Cd, according to Dimitrova and Yurukova [58] who studied this plant in polluted and non-polluted areas.

The SCF values assess the plant ability to accumulate HMs in the shoot biomass [59]. The shoots of *A. mauro-rum* will accumulate more Mn, Pb, Co, Zn, Cd, Ni, Mo and Cu than the roots. However, our findings as well as

those of Moreno-Jimenez et al. [60] support the hypothesis that there are gradients of plant-available metal levels in metalliferous soils, which are mirrored in the metaltolerant individuals gradient [61, 62]. *Cardaminopsis arenosa* was found to be unsuitable for phytostabilisation because it bioaccumulated high levels of Cd and Zn in its shoots, but had better growth cover than *Dudleya caespitosa* in the soil without the metal stabilising amendments [63, 64]. These findings are in agreement with those of *A. maurorum* shoots.

Plants can immobilise HMs by absorption and accumulation [65]. In *A. maurorum*, the RCF was greater than unity for Mn, Pb, Co, Zn and Cd. This indicates HM tolerance, adaptation to soil and environment characteristics, HM absorption capacity and root spatial fitting to pollutants.

Plants synthesize various secondary metabolites as protective mechanisms against environmental stresses such as salinity, drought, HMs and diseases [66]. The antioxidant protection mechanisms help plants defend themselves from damage when they are exposed to contaminated soils. Higher antioxidant enzyme activities and non-enzymatic constituent levels are essential for plants to withstand stress conditions such as metal toxicity. Originally, these were considered osmotic buffers, but in

Table 6	Primer	sequence,	band	length,	total	amplified	bands,	monomorphic,	polymorphic,	unique	bands	and	percentage	of
polymo	rphism a	ind efficienc	y of IS	SR and S	RAP a	nalysis of t	hree A. r	<i>naurorum</i> genot [,]	ypes					

Primer Name	Primer sequence (5`3`)	Band length (bp)	TAB	NMB	NPB	NUB	РРВ	MBF	PIC	RP
ISSR markers										
49A	CAC ACA CAC ACA AG	215-385	2	1	1	-	50%	0.83	0.22	0.67
HB-9	CAC CAC CAC GC	380-435	2	2	-	-	-	1.00	0.00	0.00
HB-10	GAG AGA GAG AGA CC	280-470	2	1	1	-	50%	0.83	0.22	0.67
HB-11	GTG TGT GTG TGT TGT CC	460-810	2	1	1	-	50%	0.83	0.22	0.67
HB-13	GAG GAG GAG C	470–975	3	2	1	-	33.33%	0.89	0.15	0.67
Total		-	11	7	4	-	-	-	-	-
Average		200–900	2.2	1.4	0.8	-	36.36%	0.88	-	0.53
DI of all ISSR pr	imers = 0.16									
SRAP markers										
ME1xEM7	F ME-1: GAGTCCAAACCGGATA R EM-7: GACTGCGTACGAATTCAA	275-1500	6	-	6	3	100%	0.50	0.45	4.02
ME1xEM8	ME-1: GAGTCCAAACCGGATA EM-8: CTGCGTACGAATTCAC	320-680	4	2	2	-	50%	0.83	0.22	1.36
ME2xEM6	ME-2: TGA GTC CAA ACC GGA GC EM-6: GAC TGC GTA CGA ATT CC	145–845	7	2	5	2	71.42%	0.66	0.32	3.36
ME2xEM7	ME-2: TGA GTC CAA ACC GGA GC Em-7: GACTGCGTACGAATTCAA	135–185	3	2	1	1	33.33%	0.78	0.15	0.66
ME5xEM6	Me-5: TGAGTCCAAACCGGAAG EM-6: GAC TGC GTA CGA ATT CC	110-200	3	1	2	1	66.66%	0.66	0.30	1.34
ME5xEM8	Me-5: TGAGTCCAAACCGGAAG EM-8: CTGCGTACGAATTCAC	100–250	2	2	-	-	-	1.00	0	0
Total		-	25	9	16	7	-	_	-	-
Average		100-1500	4.17	1.5	2.67	1.17	64%	0.74	-	1.79
DI of all SRAP p	rimers = 0.24									
Combined ISSR	and SRAP		36	16	20	7	55.55%	-	-	-

TAB Total amplified bands, NMB Number of monomorphic bands, NPB Number of polymorphic bands; NUB Number of unique bands, PPB Percentage of polymorphic bands, MBF Mean of band frequency, PIC polymorphism information content, Rp Resolving power and DI Diversity index

 Table 7
 Similarity index for three A. maurorum genotypes using ISSR and SRAP analysis

	A1	A2	A3
A1	1		
A2	0.652	1	
A3	0.578	0.939	1

addition to osmotic modification, they also tend to have an essential role in preserving the natural state of macromolecules, most likely through ROS scavenging [67]. This antioxidant defence mechanism includes enzymatic (SOD, CAT, and APX) and non-enzymatic (GSH, proline, alkaloids, carotenoids and phenols) antioxidants that function as scavengers of free radicals [68]. Some biomolecules involved in cellular metal detoxification might have chelating or antioxidant properties, as previously mentioned.



HM exposure elicited antioxidative responses, but the direction of these responses varied depending on the plant, tissue examined, metal applied and metal stress severity [69]. The present study also examined the nonenzymatic and enzymatic detoxification pathways in A. maurorum tolerance to HM stress. A. maurorum had higher CAT and POD, and a remarkable and drastic increase in PPO activity at site A1 compared with the other sites. H₂O₂ detoxification in plants is supported by CAT. Similarly, Sarker and Oba [70] discovered that an increase in CAT and SOD played a role in ROS detoxification in tolerant Amaranthus tricolour. There was a strong link between Cu and antioxidant enzymes (CAT, POD and PPO) in this research. In Carthamus tinctorius, Mazhoudi et al. [71] concluded that the stimulation of SOD activity in conjunction with CAT appeared to protect against membrane damage as Cu is especially toxic to membranes.

HMs have been linked to lower chlorophyll levels in several plant organisms. Compared with other sites, *A. maurorum* obtained from site A3 showed a substantial reduction in photosynthetic pigments. Similarly, reductions in the levels of photosynthetic pigments have been detected in different plants exposed to HMs [72]. Carotenoids act as antioxidants, protecting the plant from free radicals and photochemical damage. As a result, the fact that carotenoid levels are reduced may indicate that they play a protective role against oxidative stress [73]. According to Naidoo and Chirkoot [74], a decrease in the Chl (a+b) /Carotenoids ratio in *A. maurorum* at different coverage sites can be interpreted as evidence of a decrease in photosystem II photochemical production.

One of the main protective antioxidant mechanisms of plants exposed to HMs is the activation of biosynthesis of secondary metabolites such as flavonoids and alkaloids [75, 76]. Various studies indicate that polyphenolic compounds have a variety of biological impacts in plants [77]. Alkaloids in plants also have antioxidant properties [78]. In contrast to other sites, *A. maurorum* obtained from site A1 showed lower reductions in proline and increases in protein, flavonoids, alkaloids and tannins (measured by the number and area of shoot tanninferous cells). Previous phytochemical analysis of *A. maurorum* extracts revealed the existence of alkaloids, flavonoids, saponins, glycosides, steroids, tannins and anthraquinones as main components [79, 80].

Metal-detoxification ability of intracellular antioxidant enzymes can be improved by non-enzymatically synthesised compounds such as proline [81]. Proline functions as a stable and metabolic osmolyte, cell wall constituent, antioxidant, free radical scavenger and macromolecule stabiliser [82, 83]. Because of its sensitivity to triple stress categories such as salinity, drought and HMs, *A*. *maurorum* collected from site A3 had the highest proline content in the present study. Plants produce proline as a non-enzymatic reaction to stress imposed by a variety of abiotic and biotic factors such as drought, salinity, HMs and increased radiation [84].

The results showed that well-water irrigation in iron mining areas increased the protein and flavonoid content of A. maurorum, suggesting that they are the first to be exposed to HM contamination and could be the first line of protection. In like manner, Ozyazici [85] speculated that the rise in protein content following sludge treatment may be due to higher available soil nitrogen levels. During metabolism, plants develop complex secondary metabolites. Several of these metabolites could expel free radicals from the organism under stress, however, the metabolites ability to scavenge ROS decreases [86, 87]. A. maurorum collected from site A1 had a lower reduction in DPPH free radical scavenging effects than that collected from the other sites. A3 site subjected to various types of stress, had the highest percentage (74.63%). HMs (Cr, Cd and Pb) were applied to MS-medium and Brassica rapa seeds were allowed to germinate according to Siddigu et al. [88]. The in vitro plantlets were collected after germination and tested for DPPH scavenging activity. Control plants that had not been exposed to HMs had a substantially higher activity (87.06%). These HMs, especially Cd, have been found to not only restrict plant growth but also influence antioxidant activity. In contrast, Sulaiman [89] observed that leaf extracts of A. *maurorum* had a high free radical scavenging activity of about 95%, while flower extracts had a free radical scavenging activity of 82%. Moreover, Dhaniya and Parihar [90] showed high antioxidant potential, with 73.30% for the leaf and 88.1% for the stem extracts of *A. maurorum*.

Excess Cu has also been shown in numerous studies to encourage and stimulate the production of fenton-type ROS, resulting in an increase in MDA and dityrosine as oxidative damage biomarkers [91]. In *A. maurorum*, MDA showed a significant linear correlation with increasing metal concentrations, especially Cu. The effects of HMs on the content of MDA and photosynthetic pigments in bean seedlings (*Phaseolus vulgaris* L.) grown in Hoagland solution were investigated by Zengin [92]. MDA levels in HM-treated plants increased dramatically, while chlorophyll content in seedling leaves decreased.

Plants have developed a variety of mechanisms to combat HM toxicity. The main one is the chelation of metals at the intracellular and intercellular levels by forming phytochelatin or metallothionein metal complexes, followed by HM ions removal from sensitive sites and vacuolar sequestration of ligand-metal complexes [81]. GSH serves as a first line of protection against metal toxicity, complexing metals before induced phytochelatin synthesis reaches effective levels [93, 94]. *A. maurorum* showed a lower reduction of reduced GSH in the analysed region in this study. Ishikawa et al. [95] established that HM exposure results in significant depletion of GSH. This is a common response triggered by an increase in GSH consumption to produce phytochelatins. On the other hand, increases in the levels of non-enzymatic antioxidants in AMF-inoculated and calcium-treated plants suggested that they play a role in strengthening the antioxidant protection mechanism that results in continued development [96].

In the present study, most of the studied HMs consumed by the plant were mainly stored in the shoots, and A. maurorum had developed anatomical adaptations to cope with metal concentrations in its tissues. Anatomical adaptations were observed in the leaves and stems of A. maurorum, despite the presence of higher concentrations of HMs in leaf tissues. We may infer from our findings that A. maurorum is a xerophytic species with the thickest stem epidermis and thickened adaxial and abaxial epidermis of leaves. The number, location and distribution of tanninferous cells in the leaves and stem of A. maurorum collected from site A3 were increased. Increased tannin content is thought to be one of the non-enzymatic protective mechanisms against HMs in this case. The presence of sunken stomata in the epidermis of both leaves and stems is a water-saving technique. This research supports the findings of Awmack and Lock [27] who reported that anatomical tests of A. maurorum stems revealed a thick epidermis and a poorly constructed cortex. Vascular bundles are positioned radially from the centre outward. The xylem is well constructed, allowing significant amounts of water to be transported from the ground without pith. Gomes et al. [26] also stated that Brachiaria decumbens exhibited adaptive properties for survival in HM-polluted soil, implying that the species should be investigated further as a possible restorer. The thickened adaxial and abaxial epidermis found in the species may be a technique to reduce water loss by transpiration, which could explain the increased leaf blade turgor shown in the contaminated plants. Leaf curling is a technique for reducing surface transpiration and maintaining stomata in a humid microclimate to avoid dryness [26, 97].

It is important to report that the genes associated with the glutathione metabolism pathway (*6PGD*, *G6PDH* and *GST*) revealed higher expression levels. GST is a catalytic enzyme that uses glutathione to function in plant stress tolerance processes [98, 99]. Moreover, *6PGD* and *G6PDH* mediate NADP+ reduction to NADPH and play crucial role in the maintenance of glutathione under stresses [100, 101]. These findings were also in agreement with that reported by Wu et al. [99] who revealed higher expression levels of these genes in *Alhagi sparsifolia* plants grown under stress conditions. The results also indicated that these genes might synergistically regulate the glutathione metabolism in *A. maurorum* plants under stress conditions. The current study also revealed that heavy metal accumulation recorded in the different collection sites modulated the expression levels of *nitrate reductase 1, sulphate transporter* and *auxin-induced protein* in *A. maurorum* plants, indicating that *A. maurorum* plants respond to heavy metals accumulation via modulating their molecular mechanisms.

The use of molecular markers to investigate plant genetic homogeneity has been recommended since they would target various portions of the genome [40, 102]. One of the most significant advances in the field of molecular genetics is the use of molecular markers for the detection and assessment of DNA polymorphism [103]. In the present investigation, ISSR and SRAP markers have been successfully utilized to assess the genetic variation and fidelity in A. maurorum genotypes grown in HM-contaminated soil. Five ISSR primers produced 11 reproducible bands, 7 of which were monomorphic and 4 of which were polymorphic, whereas 25 amplicons in total with 9 monomorphic and 16 polymorphic with 7 unique bands were generated from 6 SRAP primers combinations. SRAP markers detected a high genetic differentiation in A. maurorum than ISSR markers. These results were in agreement with that recorded by Amirkhosravi et al. [38] who reported that 8 labeled inter simple sequence repeat (ISSR) primers generated a total of 243 bands used to screen 22 populations including 110 individuals of Alhagi species in Iran. Moreover, Abd Elhak et al. [104] studied the genetic diversity of Alhagi graecorum populations using 10 SCoT primers. A total of 140 fragments were amplified among the 25 individuals, with 37 monomorphic and 103 polymorphic fragments. Moreover, Jingade et al. [105] reported SRAP markers revealed a considerable polymorphism rate (67.83%) in Indian coffee (Coffea arabica L.). Even though, Agarwal et al. [40] reported that the amplification products of ISSR, SCoT and RAPD were found to be monomorphic across all A. maurorum samples.

In the current study, the DI and Rp parameters were utilized to estimate the level of polymorphism, genetic diversity, and informativeness of SRAP and ISSR markers. Furthermore, PIC values assist in establishing primers efficiency in genetic diversity analysis [106]. Their effectiveness was demonstrated by high polymorphism percentage and average number of polymorphic fragments per primer. PIC > 0.5, 0.5 > PIC > 0.25, and PIC < 0.25 are all associated with high, medium, or low loci polymorphisms, respectively [107]. The

mean of PIC value (0.16) in the present study suggested that ISSR markers might exhibit low loci polymorphism among A. maurorum genotypes, but SRAP markers exhibited a moderate mean of PIC value (0.24). Furthermore, ME1xEM7 primer combination, as well as 49 A, HB-10, and HB-11 primers, had the highest PIC and Rp values, making them the most informative primers for discriminating A. maurorum genotypes. The use of SRAP primers with moderate PIC values is a good strategy for assessing genetic diversity among A. maurorum genotypes. Furthermore, the average PIC value per primer in this investigation was close to that obtained by Soleimani et al. [108] who used SRAP markers which revealed a polymorphism rate of 53%. The average PIC value of ornamental pomegranate (Punica granatum L.) was 0.28. Furthermore, Pakhrou et al. [109] discovered that the PIC value (0.27) and marker index (MI=10.81) produced by IRAP markers were nearly identical to those produced by ISSRs (PIC=0.27 and MI=12) of the argan tree (Argania spinosa L.) in Morocco. Our findings corresponded with those of Kumar and Agrawal [110] who discovered that SRAP markers were more effective than ISSR markers in evaluating genetic diversity. As compared to ISSR markers (polymorphism percentage of 14.43%, PIC of 0.10) of Indian Simarouba glauca DC, SRAP markers generated higher polymorphism percentage (26.54%) and polymorphic information content (0.14). In the present study, the similarity matrices generated by SRAP and ISSR markers were the highest between A. maurorum from sites A2 and A3. Our findings corresponded with those of Amirkhosravi et al. [38] who employed ISSR markers to distinguish between Alhagi species in Iran, where *Alhagi* has retrieved two species; A. graecorum and A. maurorum.

Conclusion

HM-induced anatomical, biochemical and molecular genetic changes are direct indicators of the ecophysiological effects that these HMs have on the plant environment. A. maurorum is an ideal candidate for phytoremediation (phytoextraction and phytostabilisation) of Cr, Pb, Cd, Mn, Cu, Ni and Zn in iron mining ore deposits. A. maurorum obtained from site A1 (mining) had decreased levels in photosynthetic pigments, proline, DPPH and reduced GSH, and increased levels in antioxidant enzyme activities and encoding genes, MDA, protein, flavonoids, alkaloids and tannins. ISSR and SRAP markers are effective for assessing genetic variation and for differentiation of A. maurorum samples. A. maurorum demonstrated adaptive characteristics for survival in HM-contaminated soil, implying that A. maurorum should be studied further as a possible restorer of mining deposits contaminated by these HMs.

Materials and methods

Study area description

El-Wahat El-Bahariya is a vast topographic depression in the center of the Egyptian Western Desert, about 270 km southwest of Cairo and 180 km west of the Nile Valley. This depression is an interesting location in the Western Desert, particularly for rich iron ore deposits. These ores are found in the northern part of the depression and occupy an area of 11.7 km^2 , with a thickness ranging from 2 to 35 m, with an average of 9 m [2].

Three sites in this region were used in the research: A1, El-Gedida, an iron mining site at a longitude of 29°12′ E and a latitude of 28°27′ N; A2, located 5 km from iron mining operations at a longitude of 29°09′ E and a latitude of 28°29′ N; A3, El-Harra located 10 km from iron mining operations at a longitude of 29°04′ E and a latitude of 28°21′ N.

Collection and preparation of *A. maurorum* and soil samples

An A. maurorum sample was collected and pressed using firm cardboard sheets as an herbarium voucher specimen deposited in the botanical herbarium of the Botany Department at the Faculty of Science, Mansoura University, Egypt (deposition number 38). Confirmed sampling permissions needed to collect this plant material were obtained. The species was identified by Prof. Maha M. Elshamy (Professor of Plant Ecology and Flora at Botany Department, Faculty of Science, Mansoura University, Egypt) following Boulos [8]. Soil samples were collected from each of the three study sites, together with plant samples, in a rooting zone depth of 30 cm. Samples were put in plastic bags, placed on ice, transferred directly to the laboratory, and frozen before analysis. Shoot and root samples were preserved in triplicate for nutrient and HM content determination. Also, some shoot and root samples were fixed for ultrastructure analysis, and leaf samples were kept at -80 °C for later molecular analysis.

Determination of soil physicochemical properties

Organic matter was assessed in air-dried soil samples [111]. Electrical conductivity (EC) and pH were determined in soil suspensions (1:5 w/v dilution) with digital conductivity (Systronics-304) and pH meters (Labotronics-LT-1), respectively. Mixed-acid digestion was used to extract HMs from 0.5 g samples of dry soil. Total nitrogen (N) was assessed using the Kjeldahl method. Potassium (K) was determined using a flame photometer (CORN-ING M410), and phosphorous (P) was spectrophotometrically measured using the molybdenum blue method. Fe, Mn, Pb, Zn, Cd, Cu, Co, Ni, Mo, B, and Cr were digested with sulphuric acid (H_2SO_4) and estimated using inductively coupled plasma–optical emission spectroscopy (flame-AAS, GF-AAS, and ICP-AES) [112].

Bioavailability of heavy metals and nutrient concentrations in *A. maurorum*

HMs were extracted using a mixed-acid digestion process from 0.5 g samples of roots or shoots. Total nitrogen (N), potassium (K), and total phosphorus (P) were analyzed using the same techniques as for soil samples.

Phytoremediation capacity, accumulation, and upward translocation of HMs were assessed by calculating bioaccumulation factors (BCF), root coefficient factors (RCF), shoot coefficient factors (SCF), and translocation factors (TF) as [113, 114] from Eqs. (1), (2), (3) and (4): Total flavonoid content was estimated following the colorimetric aluminum chloride method [121]. The absorbance was read at 415 nm. A calibration curve was used to quantify total flavonoid content as quercetin at concentrations of 12.5–100 mg. mL⁻¹ in methanol. The antioxidant capacity of ethanolic extracts, with ascorbic acid as a standard, was calculated using scavenging of the stable 2,2-diphenyl-1-picrylhydrazyl (DPPH) free radical [122]. The solution absorbance was read at 515 nm per minute for 30 minutes. DPPH scavenging activity was determined using Eq. (5) as follows:

Scavenging activity (%) =
$$[(A - B)/A] \times 100$$
(5)

Where A represents the absorbance of control (only DPPH in solution), and B represents the absorbance of DPPH with sample.

BCF = HM concentration of plant root or shoot ($mg \cdot g^{-1}$)/HM in soil ($\left(\mathrm{mg}\cdot\mathrm{g}^{-1}\right)$) for each site	(1)
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$$RCF = HM \text{ conc. in root/HM conc. in soil}$$
 (2)

$$SCF = HM \text{ conc. in shoot/HM conc. in soil}$$
 (3)

$$TF = HM \text{ conc. in shoot/HM conc. in root}$$
 (4)

Biochemical analyses

Enzymatic antioxidant activities

CAT activity was measured spectrophotometrically as the decrease in absorbance of H_2O_2 at 240 nm [115]. POD was measured as previously described Zheng and Van Huystee [116]. Absorbance at 420 nm due to the formation of purpurogallin was used to measure polyphenol oxidase (PPO) activity [117].

Non-enzymatic antioxidant levels

Photosynthetic pigments were assessed by treating 2g of fresh tissue with 50% acetone (v/v) and refrigerating at 4°C overnight in full darkness. The extract absorbance was then measured spectrophotometrically at 453, 644, and 663 nm against an aqueous acetone blank [118]. Concentrations of pigments were determined, and values expressed as mg·g⁻¹ fresh wt. Total protein content of sample extracts was measured spectrophotometrically [119]. Proline content was estimated as reported by Bates et al. [120], and absorbance at 520 nm was read against a toluene blank. Proline content was estimated as 1 µmol proline per gram of fresh weight.

Alkaloid content was assessed following Harbone [123]. An orbital shaker was used to extract 2.0g of dry powders using 50 mL of 10% acetic acid in ethanol for 4hours at 200 rpm (Panasonic, MIR-S100, Japan). The filtered mixture extract was condensed to one-quarter of its original amount in a water bath. Concentrated ammonium hydroxide was mixed with the extract before the completion of precipitation. The whole suspension was left to settle; the precipitates were taken, washed with dilute ammonium hydroxide, and purified. A trace was then dried, and alkaloid content measured.

MDA was extracted from *A. maurorum* shoots [124] with slight modification. Briefly, MDA was collected from fresh shoot samples (0.5g) using 5 mL of 0.1% (w/v) trichloroacetic acid (TCA). Extracts were centrifuged for 25 minutes at $3000 \times g$. Next, 2 ml of 0.5% thiobarbituric acid (TBA) and 5% TCA were mixed with 3 mL of extract supernatant. TBA reacted with MDA in the acidic medium during incubation in a water bath (95°C) for 30 min. Reactions were stopped at room temperature, and a microplate reader was used to measure the absorbance of resulting pink solutions (TBA-MDA adducts or reaction products) at 534 nm. MDA content was determined following the formula introduced by Wang et al. [125].

GSH was estimated using high-performance liquid chromatography (HPLC). A. maurorum aerial tissues (leaves) were boiled in 3L of distilled water for 1 hour before drying under reduced pressure. Powder was stored in a clean bottle before use. Reduced GSH was measured in extracts using HPLC. Each extract was purified with a $0.22 \mu m$ syringe filter, and $10 \mu L$ of the filtrate was injected into a Waters 2690 Alliance HPLC system (Waters Inc., Milford,

CT, USA) with a Waters 996 photodiode array detector for sample analysis. A 1 mg/mL GSH aqueous stock solution was prepared, followed by serial dilution to obtain concentrations of 50, 100, 150, 200, and 250 µg/mL. These dilutions were also purified using a 0.22 µm syringe filter and 10 µL of filtrates were injected into the HPLC. Reduced GSH was separated on a C18 Xterra column, 4.6×250 mm, with a gradient sustained at a flow rate of 1 mL/min over a run time of more than 10 minutes. Buffer (0.2 M hexane sulphonic acid and 0.01 M KH₂PO₄ adjusted to pH3 with orthophosphoric acid) and methanol comprised the binary mobile phase. Isocratic elution was used for gradient elution from the column (80%:20%). An ultraviolet detector was used to measure reduced GSH at 210nm. Chromatographic peaks were established by comparing analytical retention times to the retention of reference compounds.

Semi-thin ultrastructure of A. maurorum shoots

One to two mm³ leaf or stem tissue samples were incubated overnight at 4°C in 0.1 M phosphate-buffered saline (PBS, pH7.4) with 2.5% (v/v) glutaraldehyde and 2% paraformaldehyde [126]. After washing with PBS for 15 minutes, samples were dehydrated using graded ethanol dilutions and embedded in Epon pure resin overnight at 4°C. Resin was polymerized at 70 °C for 3 days, and 0.5 and 0.1 µm section was cut with a knife using Leica EM UC7 ultramicrotome (Leica Mikrosysteme GmbH, Austria). Semi-thin 0.5 µm section was mounted on glass slides, stained with 1% (w/v) Toluidine Blue O with 1% (w/v) sodium borate for 5 minutes, viewed using an Olympus CX31RTSF model microscope (Olympus Corporation, Japan), photographed with ToupCam (X Cam Full HD camera), and examined using Image J (version 1.50i) software. Two representative images from a total of 10 images per species were digitized.

Stem parameters quantified stem cuticle thickness, chlorenchyma thickness, epidermis thickness, vascular bundle area, phloem area, xylem area, and pith cell diameter. The midrib was used as a reference point to measure leaf anatomical parameters. Epidermal thickness was measured on abaxial and adaxial margins, as was mesophyll. Gland size was also measured. Most measurements used the straight tool of Image J (https://imagej.nih.gov/ij/) on multiple consecutive sections (n = 25 aggregates/leaf). Vascular bundle area and gland size were measured using the freehand selection tool.

Molecular genetic analyses

Expression analysis of stress-related genes

Quantitative real-time PCR analysis was conducted to assess the expression levels of six genes (*GST*, *G6PDH*, *6PGD*, *nitrate reductase 1*, *sulfate transporter*, and *auxin-induced protein*) in *A. maurorum* plants collected from the study sites, A1, A2, and A3. Briefly, total RNA

and cDNA were prepared from plant tissues with RNeasy Plant Mini and Reverse Transcription kits (Qiagen), respectively. Using QuantiTect SYBR Green PCR kit (Qiagen), PCR reactions were prepared and performed in a total volume of 25μ l under previously described specific amplification conditions [99]: 50 °C for 2 min and 95 °C for 10 min; 40 cycles of 95 °C for 15 sec and 60 °C for 45 sec; 1 cycle of 95 °C for 15 sec, 60 °C for 15 sec, and 95 °C for 15 sec. Gene-specific primers were utilized for PCR amplification (Table 8) [99]. Actin was chosen as a reference gene. Relative expression levels were calculated using the method of Livak and Schmittgen [127].

Genetic diversity analysis using ISSR and SRAP markers

Five ISSR primers and six SRAP primer combinations were tested to assess polymorphism and genetic diversity. Five distinct ISSR primers (49A, HB-9, HB-10, HB-11, and HB-13) of 10–17 nucleotides were chosen. Forward and reverse SRAP primer combinations were also used (ME1xEM7, ME1xEM8, ME2xEM6, ME2xEM7, ME5xEM6, and ME5xEM8) (Table 6). DNA was extracted from young *A. maurorum* leaves using a DNeasy plant Mini Kit (http://www.biobasic.com) and stored at -20 °C for PCR amplification.

ISSR amplification reactions used a total volume of $30\,\mu$ l, consisting of $2\,\mu$ l of each primer, $25\,\text{mM}$ MgCl₂, 1U Taq polymerase, 2.5 mM of each deoxynucleotide, and 25 ng of genomic DNA. ISSR primer amplification was programmed in an automated thermal cycle (model Techno 512) for one cycle at 94 °C for 1 min, followed by 45 cycles of 1 min at 94 °C, a specific annealing temperature of 57 °C for 1 min, and a final extension for 2 min at 72 °C. Finally, the reaction was kept at 72 °C for 10 min.

 Table 8
 Primers used in the quantitative RT-PCR analysis [99]

Gene name	Primer sequence (5'-3')
GST	F: TCTTGGAGAACGCTCTTGGT
	R: GATGTCATGCTTGAACGCCTC
6PGD	F: GGGTTGTGGGGTTGGCTATT
	R: CCCTCTGAGCCTGAACAAGG
G6PDH	F: GGAGTCTCAAGGTGAAGCCT
	R: GGTGAAGTGCTTAGGGAGACA
Nitrate reductase 1	F: GCTCAAGCGCTGTGGAATTT
	R: GCCTGAGACAGCAAGGTACA
Sulfate transporter	F: TGCTTGGGTATATTCAGGCTGG
	R: GTGCTTCACCACTGCTACAAC
Auxin induced protein	F: TGGCTCTACCCCTCAGAGAT
	R: CATCGACAGAACACGGAAGC
Actin	F: GCGGGAAATTGTTCGTGACA
	R: AAGAACTTCTGGGCAACGGA

Amplification products were stained with ethidium bromide, separated on 1.5% agarose gels, and photographed under UV light.

Amplification of SRAP primers was completed using a DNA Thermal Cycler (model Techno 512, UK) with cycling parameters of 2 min of denaturation at 94°C, 5 cycles of 3 steps: 1 min of denaturation at 94 °C, 30 sec of annealing at 35°C, and 30 sec of elongation at 72°C. The annealing temperature was increased to 50°C for a subsequent 35 cycles, and extension used one cycle of 5 min at 72°C. PCR products were separated on 2% agarose gels in 1X TBE buffer (89mM Tris2, mM EDTA, 89mM Boric acid) at 115 V for 2.5-3 hrs. A 100-1500 bp standard DNA ladder was used for quantification. For binary data, only SPAP and ISSR amplified bands that were clear, strong, distinct, and reproducible were analyzed. Other parameters, such as the total number of polymorphic bands, diversity index, and polymorphic information content, were estimated using the equations of Gorji et al. [128]. The ability to distinguish among A. maurorum genotypes was assessed by determining the resolving power (Rp) using the formula of Prevost and Wilkinson [129].

Statistical analysis

Metal content and physiological, biochemical, anatomical, and molecular genetic data were collected in a randomized complete block design with three replications. Data distributions were tested for normality and equality of variance. ANOVA was used to determine the statistical significance of differences among all factors with post hoc tests set by SPSS (version 22.0, IBM Corp., Armonk, NY, USA); treatments were considered as independent variables. Data are expressed as means and standard errors. A *p*-value of ≤ 0.05 was considered statistically significant, and data are depicted in figures using SigmaPlot 14.0 (Systat Software, Inc., CA, USA). Pearson correlation coefficients were determined for metals in the soil at study sites and biochemical parameters from associated specimens of A. maurorum (Origin Lab Inc., Hampton, USA). Further statistical analyses, including a heatmap of Pearson correlation coefficients among metal concentrations, biochemical parameters, and the ultrastructure of A. maurorum shoot samples, were used to identify metals affected by biological activity in shoot samples [130]. Systat Software (Inc., San Jose, CA, USA) was used to analyze genotype correlations in the dendrogram.

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Authors' contributions

M.A.E.-E. and Y.M.H. performed the molecular genetic analyses (SRAP and ISSR markers analysis and gene expression analysis) and the elemental and anatomical experimental work. M.M.E. performed the eco-biochemical

experimental work. R.N. helped in data analysis. All authors have analyzed the data and wrote and approved the manuscript.

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Availability of data and materials

All data generated or analyzed during this study are included in this published article.

Declarations

Ethics approval and consent to participate Not applicable.

Consent for publication

Not applicable.

Competing interests

The authors declare that they have no competing interests.

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References

- Mishra VK, Tripathi BD. Concurrent removal and accumulation of heavy metals by the three aquatic macrophytes. Bioresour Technol. 2008;99(15):7091–7. https://doi.org/10.1016/j.biortech.2008.01.002.
- Said R. The Geology of Egypt. Rotterdam, Brookfield: A. A. Balkema. New York: Elsevier; 1990. p. 734.
- Food and Agriculture Organization. Miscellaneous problems: water quality for agriculture. 2017, From www.fao.org: http://www.fao.org/ docrep/003/T0234E/T0234E06.htm, Retrieved April 12.
- Peng K, Luo C, Lou L, Li X, Shen Z. Bioaccumulation of heavy metals by the aquatic plants *Potamogeton pectinatus* L. and *Potamogeton malaianus* Miq. And their potential use for contamination indicators and in wastewater treatment. Sci Total Environ. 2008;392(1):22–9. https://doi. org/10.1016/j.scitotenv.2007.11.032.
- Ruiz M, Velasco J. Nutrient bioaccumulation in *Phragmites austra*lis: management tool for reduction of pollution in the Mar Menor. Water Air Soil Pollut. 2009;205(1):173–85. https://doi.org/10.1007/ s11270-009-0064-2.
- Bhaduri AM, Fulekar MH. Assessment of arbuscular mycorrhizal fungi on the phytoremediation potential of *Ipomoea aquatica* on cadmium uptake. 3. Biotech. 2012a;2(3):193–8. https://doi.org/10.1007/ s13205-012-0046-86.
- Gratão PL, Prasad MNV, Cardoso PF, Lea PJ, Azevedo RA. Phytoremediation: green technology for the clean-up of toxic metals in the environment. Barz J Plant Physiol. 2005;17(1):53–64. https://doi.org/10.1590/ \$1677-04202005000100005.
- 8. Boulos L. Flora of Egypt, vol. 1. Cairo: Al Hadara Publishing; 1999. p. 340–1.
- Ahmad N, Bibi Y, Saboon Raza I, Zahara K, Idrees S, Khalid N, et al. Traditional uses and pharmacological properties of *Alhagi maurorum*: a review. Asian Pac J Trop Dis. 2015;5(11):856–61. https://doi.org/10.1016/ S2222-1808(15)60945-8.
- Barman SC, Kisku GC, Salve PR, Misra D, Sahu RK, Ramteke PW, et al. Assessment of industrial effluent and its impact on soil and plants. J Environ Biol. 2001;22(4):251–6.

- Espinoza-Quinones FR, Zacarkim CE, Palacio SM, Obregón CL, Zenatti DC, Galante RM, et al. Removal of heavy metal from polluted river water using aquatic macrophytes *Salvinia* sp. Brazilian J Phys. 2005;35(3b):744–6.
- Ali S, Abbas Z, Rizwan M, Zaheer IE, Yava I, Ünay A, et al. Application of floating aquatic plants in phytoremediation of heavy metals polluted water: a review. Sustainability. 2020;12(5):1927. https://doi.org/10.3390/ su12051927.
- Tangahu BV, Abdullah SR, Basri H, Idris M, Anuar N, Mukhlisin M. A review on heavy metals (As, Pb, and Hg) uptake by plants through phytoremediation. Int J Chem Eng. 2011;939161:31. https://doi.org/10. 1155/2011/939161.
- Rascio N, Navari-Izzo F. Heavy metal hyperaccumulating plants: how and why do they do it? And what makes them so interesting? Plant Sci. 2011;180(2):169–81. https://doi.org/10.1016/j.plantsci.2010.08.016.
- Baker AJM, Brooks RR. Terrestrial higher plants which hyperaccumulate metallic elements. A review of their distribution, ecology and phytochemistry. Biorecovery. 1989;1:81–126.
- Rizwan M, El Shamy MM, Abdel-Aziz HMM. Assessment of trace element and macronutrient accumulation capacity of two native plant species in three different Egyptian mine areas for remediation of contaminated soils. Ecol Indic. 2019;106:105463.
- Gill SS, Tuteja N. Reactive oxygen species and antioxidant machinery in abiotic stress tolerance in crop plants. Plant Physiol Biochem. 2010;48(12):909–30. https://doi.org/10.1016/j.plaphy.2010.08.016.
- Kelman D, Ben-Amotz A, Berman-Frank I. Carotenoids provide the major antioxidant defence in the globally significant N2-fixing marine cyanobacterium *Trichodesmium*. Environ Microbiol. 2009;11(7):1897–908. https://doi.org/10.1111/j.1462-2920.2009.01913.x.
- Zhou FR, Wang JX, Yang N. Growth responses, antioxidant enzyme activities and lead accumulation of *Sophora japonica* and *Platycladus orientalis* seedlings under Pb and water stress. Plant Growth Regul. 2015;75(1):383–9. https://doi.org/10.1007/s10725-014-9927-7.
- Ncube B, Van Staden J. Tilting plant metabolism for improved metabolite biosynthesis and enhanced human benefit. Molecules. 2015;20(7):12698–731. https://doi.org/10.3390/molecules200712698.
- 21. Meister A, Anderson ME. Glutathione. Annu Rev Biochem. 1983;52:711– 60. https://doi.org/10.1146/annurev.bi.52.070183.003431.
- 22. Freedman JH, Ciriolo MR, Peisach J. The role of glutathione in copper metabolism and toxicity. J Biol Chem. 1989;264(10):5598–605 PMID: 2564391.
- Tausz M, Šircelj H, Grill D. The glutathione system as a stress marker in plant ecophysiology: is a stress-response concept valid? J Exp Bot. 2004;55(404):1955–62. https://doi.org/10.1093/jxb/erh194.
- MacFarlane GR, Burchett MD. Cellular distribution of copper, lead and zinc in the grey mangrove, *Avicennia marina* (Forsk.) Vierh. Aquat Bot. 2000;68(1):45–59. https://doi.org/10.1016/S0304-3770(00)00105-4.
- Wójcik M, Vangronsveld J, D'Haen J, Tukiendorf A. Cadmium tolerance in *Thlaspi caerulescens*. Environ Exp Bot. 2005;53(2):163–71. https://doi. org/10.1016/j.envexpbot.2004.03.011.
- Gomes MP, Marques TCLL, Nogueira MG, Castro EM, Soares ÂM. Ecophysiological and anatomical changes due to uptake and accumulation of heavy metal in *Brachiaria decumbens*. Sci Agric. 2011;68(5):566–73. https://doi.org/10.1590/S0103-90162011000500009.
- Awmack CS, Lock JM. The genus Alhagi (Leguminosae: Papilionoideae) in the Middle East. Kew Bull. 2002;57(2):435–43. https://doi.org/10. 2307/4111121.
- Hajibarat Z, Saidi A, Hajibarat Z, Talebi R. Characterization of genetic diversity in chickpea using SSR markers, start codon targeted polymorphism (SCoT) and conserved DNA-derived polymorphism (CDDP). Physiol Mol Biol Plant. 2015;21:365–73. https://doi.org/10.1007/ s12298-015-0306-2.
- Etminan A, Pour-Aboughadareh A, Mohammadi R, Ahmadi-Rad A, Noori A, Mahdavian Z, et al. Applicability of start codon targeted (SCoT) and inter-simple sequence repeat (ISSR) markers for genetic diversity analysis in durum wheat genotypes. Biotechnol Biotechnol Equip. 2016;30(6):1075–81. https://doi.org/10.1080/13102818.2016.1228478.
- Li G, Quiros CF. Sequence-related amplified polymorphism (SRAP), a new marker system based on a simple PCR reaction: its application to mapping and gene tagging in *Brassica*. Theor Appl Genet. 2001;103:455–61. https://doi.org/10.1007/s001220100570.

- 31. Robarts DW, Wolfe AD. Sequence-related amplified polymorphism (SRAP) markers: a potential resource for studies in plant molecular biology. Appl Plant Sci. 2014;2(7):1400017. https://doi.org/10.3732/apps.1400017.
- Uzun A, Yesiloglu T, Aka-Kacar Y, Tuzcu O, Gulsen O. Genetic diversity and relationships within *Citrus* and related genera based on sequence related amplified polymorphism markers (SRAPs). Sci Hortic. 2009;121(3):306–12. https://doi.org/10.1016/j.scienta.2009.02.018.
- Singh A, Avtar R, Singh D, Sangwan O, Balyan P. Genetic variability, character association and path analysis for seed yield and component traits under two environments in Indian mustard. J Oilseed Brassica. 2013;4(1):43–8.
- Liao B, Wang F, Chen L, Li P, Ouyang K, Pian R, et al. Population structure and genetic relationships of *Melia* taxa in China assayed with sequencerelated amplified polymorphism (SRAP) markers. Forests. 2016;7(4):81. https://doi.org/10.3390/f7040081.
- Golkar P, Mokhtari N. Molecular diversity assessment of a world collection of safflower genotypes by SRAP and SCoT molecular markers. Physiol Mol Biol Plants. 2018;24(6):1261–71. https://doi.org/10.1007/ s12298-018-0545-0.
- Ng W, Tan S. Inter-simple sequence repeat (ISSR) markers: are we doing it right. ASM Sci J. 2015;9(1):30–9.
- Abate T. Inter simple sequence repeat (ISSR) markers for genetic diversity studies in *Trifolium* species. Adv life Sci Technol. 2017;55:33–7.
- Amirkhosravi A, Asri Y, Assadi M, Mehregan I. Genetic structure of *Alhagi* (Hedysareae, Fabaceae) populations using ISSR data in Iran. Mol Biol Rep. 2021;48(6):5143–50. https://doi.org/10.1007/s11033-021-06511-w.
- Trindade H, Sena I, Gonçalves S, Romano A. Genetic diversity of wild populations of *Tuberaria major* (Cistaceae), an endangered species endemic to the Algarve region (Portugal), using ISSR markers. Biochem Syst Ecol. 2012;45:49–56. https://doi.org/10.1016/j.bse.2012.06.028.
- Agarwal T, Gupta AK, Patel AK, Shekhawat NS. Micropropagation and validation of genetic homogeneity of *Alhagi maurorum* using SCoT, ISSR and RAPD markers. Plant Cell Tissue Organ Cult. 2015;120:313–23. https://doi.org/10.1007/s11240-014-0608-z.
- Ali H, Khan E, Ilahi I. Environmental chemistry and ecotoxicology of hazardous heavy metals: environmental persistence, toxicity, and bioaccumulation. J Chem. 2019;6730305:14. https://doi.org/10.1155/2019/ 6730305.
- Berni R, Luyckx M, Xu X, Legay S, Sergeant K, Hausman JF, et al. Reactive oxygen species and heavy metal stress in plants: impact on the cell wall and secondary metabolism. Environ Exp Bot. 2019;161:98–106. https://doi.org/10.1016/j.envexpbot.2018.10.017.
- Maslova SP, Golovko TK. Tropisms of underground shoots-Stolons and rhizomes. Biol Bull Rev. 2018;8:181–92. https://doi.org/10.1134/S2079 08641803009X.
- Keyster M, Niekerk LA, Basson G, Carelse M, Bakare O, Ludidi N, et al. Decoding heavy metal Stress Signalling in plants: towards improved food security and safety. Plants. 2020;9(12):1781. https://doi.org/10. 3390/plants9121781.
- 45. Abd El Wahed AG. Geoenvironmental assessment of Bahariya oasis, Western Desert, Egypt. PH D thesis. In: Geology Department, Faculty of Science: Ain Shams University; 2014.
- Plazinski W. Binding of heavy metals by algal biosorbents. Theoretical models of kinetics, equilibria and thermodynamics. Adv Colloid Interface Sci. 2013;197–198:58–67. https://doi.org/10.1016/j.cis.2013.04.002.
- Hu B, Jia X, Hu J, Xu D, Xia F, Li Y. Assessment of heavy metal pollution and health risks in the soil-plant-human system in the Yangtze River Delta, China. Int J Environ Res Public Health. 2017;14(9):1042. https://doi.org/10.3390/ijerph14091042.
- Chibuike GU, Obiora SC. Heavy metal polluted soils: effect on plants and bioremediation methods. Appl Environ Soil Sci. 2014;752708, 2:1–12. https://doi.org/10.1155/2014/752708.
- Palansooriya KN, Shaheen SM, Chen SS, Tsang DCW, Hashimoto Y, Hou D, et al. Soil amendments for immobilization of potentially toxic elements in contaminated soils: A critical review. Environ Int. 2020;134:105046. https://doi.org/10.1016/j.envint.2019.105046.
- Sekabira K, Oryem-Origa H, Mutumba G, Kakudidi E, Basamba TA. Heavy metal phytoremediation by *Commelina benghalensis* (L) and *Cynodon dactylon* (L) growing in urban stream sediments. Int J Plant Physiol Biochem. 2011;3(8):133–42.

- Baker AJM, McGrath SP, Reeves RD, Smith JAC. Metal Hyperaccumulator plants: A review of the ecology and physiology of a biological resource for phytoremediation of metal-polluted soils. In: Terry N, Banuelos G, editors. Phytoremediation of contaminated soil and water. Boca Raton: Lewis Publishers; 2000. p. 85–108.
- Yanqun Z, Yuan L, Schvartz C, Langlade L, Fan L. Accumulation of Pb, Cd, Cu and Zn in plants and hyperaccumulator choice in Lanping lead-zinc mine area. China Environ Int Jun. 2004;30(4):567–76. https://doi.org/10. 1016/j.envint.2003.10.012.
- Malik RN, Husein SZ, Nazir I. Heavy metal contamination and accumulation in soil and wild plants species from industrial area of Islamabad Pakistan. Pak J Bot. 2010;42(1):291–301.
- Abou-Shanab R, Ghanem N, Ghanem K, Al-Kolaibe A. Phytoremediation potential of crop and wild plants for multi-metal contaminated soils. Res J Biol Sci. 2007;3(5):370–6.
- Yanqun Z, Yuan L, Jianjun C, Haiyan C, Li Q, Schvartz C. Hyperaccumulation of Pb, Zn and cd in herbaceous grown on lead–zinc mining area in Yunnan, China. Environ Int. 2005;31(5):755–62. https://doi.org/10.1016/j. envint.2005.02.004.
- 56. Akbar S, Alhagi maurorum Medik. Fabaceae/Leguminosae. In: Handbook of 200 medicinal plants. Cham: Springer; 2020.
- González RC, González-Chávez MCA. Metal accumulation in wild plants surrounding mining wastes. Environ Pollut. 2006;144(1):84–92. https:// doi.org/10.1016/j.envpol.2006.01.006.
- Dimitrova I, Yurukova L. Bioindication of anthropogenic pollution with *Plantago lanceolata* (Plantaginaceae): metal accumulation, morphological and stomatal leaf characteristics. Phytologia Balcanica. 2005;11(1):89–96.
- Branquinho C, Serrano H, Pinto M, Martins-Loução M. Revisiting the plant hyperaccumulation criteria to rare plants and earth abundant elements. Environ Pollut. 2007;146(2):437–43. https://doi.org/10.1016/j. envpol.2006.06.034.
- Moreno-Jiménez E, Penălosa JM, Manzano R, Carpena-Ruiz RO, Gamarra R, Esteban E. Heavy metals distribution in soils surrounding an abandoned mine in NW Madrid (Spain) and their transference to wild flora. J Hazard Mater. 2009;162(2–3):854–9. https://doi.org/10.1016/j.jhazmat. 2008.05.109.
- 61. Ernst WHO. Evolution of metal tolerance in higher plants. For Snow Landsc Res. 2006;80(3):251–74 https://www.dora.lib4ri.ch/wsl/islan dora/object/wsl:15470.
- Majewska M, Kurek E, Słomka A. Effect of plant growth on total concentrations of Zn, Pb, and cd, and their distribution between operational fractions in the upper layer of 100-year-old zinc-lead waste heap. Pol J Environ Stud. 2011;20(3):591–7.
- Kucharski R, Sas-Nowosielska A, Małkowski E, Japenga J, Kuperberg JM, Pogrzeba M, et al. The use of indigenous plant species and calcium phosphate for stabilization of highly metal polluted sites in southern Poland. Plant Soil. 2005;273(1/2):291–305. https://doi.org/10.1007/ s11104-004-8068-6.
- Nadgórska-Socha A, Kafel A, Kandziora-Ciupa M, Gospodarek J, Zawisza-Raszka A. Accumulation of heavy metals and antioxidant responses in *Vicia faba* plants grown on monometallic contaminated soil. Environ Sci Pollut Res. 2013;20(2):1124–34. https://doi.org/10.1007/ s11356-012-1191-7.
- Taskila S, Tuomola M, Ojamo H. Enrichment cultivation in detection of food-borne salmonella. Food Cont. 2012;26(2):369–77. https://doi.org/ 10.1016/j.foodcont.2012.01.043.
- Khare S, Singh NB, Singh A, Hussain I, Niharika K, Yadav V, et al. Plant secondary metabolites synthesis and their regulations under biotic and abiotic constraints. J Plant Biol. 2020;63(3):203–16. https://doi.org/10. 1007/s12374-020-09245-7.
- Xiong L, Zhu J-K. Molecular and genetic aspects of plant responses to osmotic stress. Plant Cell Environ. 2002;25(2):131–9. https://doi.org/10. 1046/j.1365-3040.2002.00782.x.
- Hasanuzzaman M, Borhannuddin Bhuyan MHM, Zulfiqar F, Raza A, Mohsin SM, Al Mahmud J, et al. Reactive oxygen species and antioxidant defense in plants under abiotic Stress: revisiting the crucial role of a universal defense regulator. Antioxidants. 2020;9(8):681. https://doi.org/ 10.3390/antiox9080681.
- Shainberg O, Rubin B, Rabinowitch HD, Libal Y, Tel-Or E. Acclimation of beans to oxidative stress by treatment with sublethal iron levels. J

Plant Physiol. 2000;157(1):93–9. https://doi.org/10.1016/S0176-1617(00) 80141-8.

- Sarker U, Oba S. Drought stress enhances nutritional and bioactive compounds, phenolic acids and antioxidant capacity of *Amaranthus* leafy vegetable. BMC Plant Biol. 2018;18(1):258. https://doi.org/10.1186/ s12870-018-1484-1.
- Mazhoudi S, Chaoui A, Ghorbal MH, El Ferjani E. Response of antioxidant enzymes to excess copper in tomato (*lycopersicon esculentum* mill). Plant Sci. 1997;127:129–37. https://doi.org/10.1016/S0168-9452(97)00116-7.
- Liu L, Sun H, Chen J, Zhang Y, Li D, Li C. Effects of cadmium (cd) on seedling growth traits and photosynthesis parameters in cotton (*Gos-sypium hirsutumL*.). Plant Omics J. 2014;7(4):284–90 http://www.pomics. com/liu_7_4_2014_28.
- Sengar RK, Gautam M, Sengar RK, Grag SK, Sengar K, Chaudhary R. Lead stress effects on physiobiochemical activities of higher plants. Rev Environ Contam Toxicol. 2008;196:73–93. https://doi.org/10.1007/ 978-0-387-78444-1_3.
- Naidoo G, Chirkoot D. The effects of coal dust on photosynthetic performance of the mangrove, *Avicennia marina* in Richards Bay, South Africa. Environ Pollut. 2004;127(3):359–66. https://doi.org/10.1016/j. envpol.2003.08.018.
- Mithofer A, Schulze B, Boland W. Biotic and heavy metal stress response in plants: evidence for common signals. FEBS Lett. 2004;566(1–3):1–5. https://doi.org/10.1016/j.febslet.2004.04.011.
- Bhaduri AM, Fulekar MH. Antioxidant enzyme responses of plants to heavy metal stress. Rev Environ Sci Biotechnol. 2012b;11(1):55–69. https://doi.org/10.1007/s11157-011-9251-x.
- Tungmunnithum D, Thongboonyou A, Pholboon A, Yangsabai A. Flavonoids and other phenolic compounds from medicinal plants for pharmaceutical and medical aspects: an overview. Medicines. 2018;5(3):93. https://doi.org/10.3390/medicines5030093.
- Wink M. Evolution of secondary metabolites from an ecological and molecular phylogenetic perspective. Phytochemistry. 2003;64(1):3–19. https://doi.org/10.1016/S0031-9422(03)00300-5.
- Abdel Rahman SM, Abd-Ellatif SA, Deraz SF, Khalil AA. Antibacterial activity of some wild medicinal plants collected from western Mediterranean coast, Egypt: natural alternatives for infectious disease treatment. Afr J Biotechnol. 2011;10(52):10733–43. https://doi.org/10.5897/ AJB11.007.
- Samejo MQ, Memon S, Bhanger MI. Khan KM chemical composition of essential oils from *Alhagi maurorum*. Chem Nat Comp. 2012;48:898– 900. https://doi.org/10.1007/s10600-012-0417-8.
- Emamverdian A, Ding Y, Mokhberdoran F, Xie Y. Heavy metal stress and some mechanisms of plant defence response. Sci World J. 2015;756120:18. https://doi.org/10.1155/2015/756120.
- Pavlíkovà D, Pavlík M, Staszkovà L, Tlustoš P, Szàkovà J, Balík J. The effect of potentially toxic elements and sewage sludge on the activity of regulatory enzyme glutamate kinase. Plant Soil Environ. 2007;53(5):201–6. https://doi.org/10.17221/2204-PSE.
- Zarei S, Ehsanpour AA, Abbaspour J. The role of over expression of P5CS gene on proline, catalase, ascorbate peroxidase activity and lipid peroxidation of transgenic tobacco (*Nicotiana tabacum* L.) plant under in vitro drought stress. J Cell Mol Res. 2012;4(1):43–9. https://doi.org/10. 22067/jcmr.v4i1.18249.
- Szabados L, Savouré A. Proline: A multifunctional amino acid. Trends Plant Sci. 2010;15(2):89–97. https://doi.org/10.1016/j.tplants.2009.11.009.
- Ozyazici M. Effects of sewage sludge on the yield of plants in the rotation system of wheat-white head cabbage- tomato. Eurasian J Soil Sci. 2013;2(1):35–44 http://ejss.fess.org/10.18393/ejss.2013.1.035-044.
- Ahmad N, Fazal H, Abbasi BH. In vitro Larvicidal potential and antioxidative enzymes activities in *Ginkgo biloba*, *Stevia rebaudiana* and *Parthenium hysterophorous*. Asian Pac J Trop Med. 2011a;4(3):169–75. https:// doi.org/10.1016/S1995-7645(11)60063-1.
- Ahmad N, Fazal H, Ahmad I, Abbasi BH. Free radical scavenging (DPPH) potential in nine *Mentha* species. J Toxicol Ind Health. 2011b;28(1):83–9. https://doi.org/10.1177/0748233711407238.
- Siddiqui MM, Abbasi BH, Ahmad N, Ali M, Mahmood T. Toxic effects of heavy metals (cd, Cr and Pb) on seed germination and growth and DPPH-scavenging activity in *Brassica rapa* var. turnip. Toxicol Ind Health. 2014;30(3):238–49. https://doi.org/10.1177/0748233712452605.

- Sulaiman GM. Antimicrobial and cytotoxic activities of methanol extract of *Alhagi maurorum*. Afr J Microbiol Res. 2013;7(16):1548–57. https://doi. org/10.5897/AJMR12.1795.
- Dhaniya S, Parihar SK. Evaluation of antioxidant potential of *Dicoma* tomentosa and *Alhagi maurorum* leaf and stem powder. J Drug Deliv Ther. 2019;9(4-A):207–11. https://doi.org/10.22270/jddt.v9i4-A.3454.
- Lombardi L, Sebastiani L. Copper toxicity in *Prunus cerasifera*: growth and antioxidant enzymes responses of in vitro grown plants. Plant Sci. 2005;168(3):797–802. https://doi.org/10.1016/j.plantsci.2004.10.012.
- 92. Zengin F. Physiological behavior of bean (*Phaseolus vulgaris* L.) seedlings under metal stress. Biol Res. 2013;46(1):79–85. https://doi.org/10. 4067/S0716-97602013000100012.
- Mohamed BA, Huaxin D. Effects of Cr, cd, and Pb on ultrastructure, GSH and free cysteine in *Typha angustifolia*. Transcriptomics. 2015;3(2):5. https://doi.org/10.4172/2329-8936.1000115.
- Khanna K, Sharma A, Ohri P, Bhardwaj R, Elsayed F, Abd-Allah EF, et al. Impact of plant growth promoting Rhizobacteria in the orchestration of *Lycopersicon esculentum* mill. Resistance to plant parasitic nematodes: A Metabolomic approach to evaluate defense responses under field conditions. Biomolecules. 2019;9(11):676. https://doi.org/10.3390/ biom9110676.
- Ishikawa T, Li Z-S, Lu Y-P, Rea PA. The GS-X pump in plant, yeast, and animal cells: structure, function, and gene expression. Biosci Rep. 1997;17(2):189–207. https://doi.org/10.1023/a:1027385513483.
- Hashem A, Abd-Allah EF, Alqarawi AA, Malik JA, Wirth S, Egamberdieva D. Role of calcium in AMF-mediated alleviation of the adverse impacts of cadmium stress in *Bassia indica* [Wight] A. J Scott Saudi J Biol Sci. 2019;26:828–38. https://doi.org/10.1016/j.sjbs.2016.11.003.
- 97. Turner NC, Jones MM. Turgor maintenance by osmotic adjustment a review and evolution. In: Turner NC, Kramer PJ, editors. Adaptation of plant to water and high temperature stress. New York: Wiley; 1980. p. 87–103.
- Dixon DP, Edwards R. Glutathione transferases. In: Arabidopsis Book, vol. 8; 2010. p. e0131. https://doi.org/10.1199/tab.0131.
- 99. Wu H, Zhang Y, Zhang W, Pei X, Zhang C, Jia S. Transcriptomic analysis of the primary roots of *Alhagi sparsifolia* in response to water Stress. Plos One. 2015;10(3):e0120791. https://doi.org/10.1371/journal.pone. 0120791.
- Marino D, Dunand C, Puppo A, Pauly N. A burst of plant NADPH oxidases. Trends Plant Sci. 2012;17(1):9–15. https://doi.org/10.1016/j.tplan ts.2011.10.001.
- Wang X, Ma Y, Huang C, Wan Q, Li N, Bi Y. Glucose-6-phosphate dehydrogenase plays a central role in modulating reduced glutathione levels in reed callus under salt stress. Planta. 2008;227(3):611–23. https:// doi.org/10.1007/s00425-007-0643-7.
- 102. Lakshmanan V, Venkataramareddy SR, Neelwarne B. Molecular analysis of genetic stability in long-term micropropagated shoots of banana using RAPD and ISSR markers. Electron J Biotechnol. 2007;10(1):106–13. https://doi.org/10.4067/S0717-34582007000100010.
- Aneja B, Yadav NR, Chawla V, Yadav RC. Sequence-related amplified polymorphism (SRAP) molecular marker system and its applications in crop improvement. Mol Breeding. 2012;30(4):1635–48. https://doi.org/ 10.1007/s11032-012-9747-2.
- Abd El-hak GM, Ibrahim SD, Halem A, Hassan HZ, Rayan WA. Assessment of genetic diversity in different geographically isolated *Alhagi graecorum* Boiss. Populations using SCoT marker. Taeckholmia. 2019;39(1):44–66. https://doi.org/10.21608/TAEC.2019.19481.1008.
- Jingade P, Huded AK, Kosaraju B, Mishra MK. Diversity genotyping of Indian coffee (*Coffea arabica* L.) germplasm accessions by using SRAP markers. J Crop Improv. 2019;33(3):327–45. https://doi.org/10.1080/ 15427528.2019.1592050.
- Sivaprakash KR, Prashanth SR, Mohanty BP, Parida A. Genetic diversity of black gram (*Vigna mungo*) landraces as evaluated by amplified fragment length polymorphism markers. Curr Sci. 2004;86(10):1411–6. https://doi.org/10.1007/S00122-006-0380-5.
- Xie W, Zhang X, Cai H, Liu W, Peng Y. Genetic diversity analysis and transferability of cereal EST-SSR markers to orchardgrass (*Dactylis glomerata* L.). Biochem Syst Ecol. 2010;38(4):740–9. https://doi.org/10.1016/j. bse.2010.06.009.
- 108. Soleimani MH, Talebi M, Sayed-Tabatabaei BE. Use of SRAP markers to assess genetic diversity and population structure of wild, cultivated, and

ornamental pomegranates (*Punica granatum* L) in different regions of Iran. Plant Syst Evol. 2012;298:1141–9. https://doi.org/10.1007/s00606-012-0626-4.

- 109. Pakhrou O, Medraoui L, Yatrib C, Alami M, Filali-maltouf A, Belkadi B. Assessment of genetic diversity and population structure of an endemic Moroccan tree (Argania spinosa L.) based in IRAP and ISSR markers and implications for conservation. Physiol Mol Biol Plants. 2017;23(3):651–61. https://doi.org/10.1007/s12298-017-0446-7.
- 110. Kumar J, Agrawal V. Analysis of genetic diversity and population genetic structure in *Simarouba glauca* DC. (an important bio-energy crop) employing ISSR and SRAP markers. Ind Crop Prod. 2017;100:198–207. https://doi.org/10.1016/j.indcrop.2017.02.035.
- 111. Zaheer IE, Ali S, Saleem MH, Noor I, El-Esawi MA, Hayat K, Rizwan M, Abbas Z, El-Sheikh MA, Alyemeni MN, Wijaya L. Iron–Lysine Mediated Alleviation of Chromium Toxicity in Spinach (Spinacia oleracea L.) Plants in Relation to Morpho-Physiological Traits and Iron Uptake When Irrigated with Tannery Wastewater. Sustainability. 2020;12(16):6690. https://doi.org/10.3390/su12166690.
- 112. Aceto M, Abollino O, Bruzzoniti MC, Mentasti E, Sarzanini C, Malandrino M. Determination of metals in wine with atomic spectroscopy (flame-AAS, GF-AAS and ICP-AES); a review. Food Addit Contam. 2002;19(2):126–33. https://doi.org/10.1080/02652030110071336.
- 113. Kisku GC, Barman SC, Bhargava SK. Contamination of soil and plants with potentially toxic elements irrigated with mixed industrial effluent and its impact on the environment. Water Air Soil Pollut. 2000;120:121–37. https://doi.org/10.1023/A:1005202304584.
- 114. Ghosh M, Singh SP. A review on phytoremediation of heavy metals and utilization of its byproducts. Appl Ecol Environ Res. 2005;3(1):1–18. https://doi.org/10.15666/aeer/0301_001018.
- 115. Aebi H. Catalase in vitro. Methods Enzymol. 1984;105:121–6. https://doi.org/10.1016/s0076-6879(84)05016-3.
- Zheng X, Van Huystee RB. Peroxidase-regulated elongation of segments from peanut hypocotyls. Plant Sci. 1992;81(1):47–56. https://doi.org/10. 1016/0168-9452(92)90023-F.
- Cho YK, Ahn HK. Purification and characterization of polyphenol oxidase from potato: II. Inhibition and catalytic mechanism. J Food Biochem. 1999;23(6):593–605. https://doi.org/10.1111/j.1745-4514.1999. tb00588.x.
- 118. Allen SE. Chemical analysis of ecological materials. London: Blackwell Scientific Publications; 1989.
- 119. Bradford MM. A rapid sensitive method for the quantification of microgram quantities of protein utilizing the principle of protein-dye binding. Anal Biochem. 1976;72:248–54. https://doi.org/10.1006/abio. 1976.9999.
- 120. Bates L, Waldren RP, Teare ID. Rapid determination of free proline for water stress studies. Plant Soil. 1973;39:205–7.
- 121. Nabavi SM, Ebrahimzadeh MA, Nabavi SF, Hamidinia A, Bekhradnia AR. Determination of antioxidant activity, phenol and flavonoids content of *Parrotia persica* Mey. Pharmacol Online. 2008;2:560–7.
- 122. Brand-Williams W, Cuvelier ME, Berset C. Use of a free radical method to evaluate antioxidant activity. LWT Food Sci Technol. 1995;28(1):25–30. https://doi.org/10.1016/S0023-6438(95)80008-5.
- 123. Harbone JB. Phytochemical methods, a guide to modern techniques of plant analysis. London: Chapman and Hall; 1973. p. 185–6.
- Müller C, Kuki NK, Pinheiro DT, de Souza LR, Silva AlS, Loureiro ME, et al. Differential physiological responses in rice upon exposure to excess distinct iron forms. Plant Soil. 2015;391(1):123–38 http://rd.springer. com/journal/11104.
- 125. Wang YS, Ding MD, Gu XG, Wang JL, Yunli P, Gao LP, et al. Analysis of interfering substances in the measurement of malondialdehid content in plant leaves. Am J Biochem Biotechnol. 2013;9(3):235–42. https://doi.org/10.3844/ajbbsp.2013.235.242.
- 126. Reynolds ES. The use of lead citrate at high pH as an electron opaque stain in electron microscopy. J Cell Biol. 1963;17(1):208–12. https://doi.org/10.1083/jcb.17.1.208.
- Livak KJ, Schmittgen TD. Analysis of relative gene expression data using real-time quantitative PCR and the 2(-ΔΔC (T)) method. Methods. 2001;25(4):402-8. https://doi.org/10.1006/meth.2001.1262.
- 128. Gorji AM, Poczai P, Polgar Z. Efficiency of arbitrarily amplified dominant markers (SCOT, ISSR and RAPD) for diagnostic fingerprinting in

tetraploid potato. Am J Potato Res. 2011;88:226–37. https://doi.org/10. 1007/s12230-011-9187-2.

- 129. Prevost A, Wilkinson MJ. A new system of comparing PCR primers applied to ISSR fingerprinting of potato cultivars. Theor Appl Genet. 1999;98:107–12. https://doi.org/10.1007/s001220051046.
- Babicki S, Arndt D, Marcu A, Liang Y, Grant JR, Maciejewski A, et al. Heatmapper: web-enabled heat mapping for all. Nucleic Acids Res. 2016;44(W1):W147–53. https://doi.org/10.1093/nar/gkw419.

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