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Vermicomposting biotechnology: recycling of palm oil mill wastes into valuable products

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Abstract

Background: Palm oil mill effluent and palm press fiber are problematic wastes generated by the palm oil mill industries in Malaysia. This study has endeavored to assess the possibility of the vermicomposting of residue from the palm oil mills using epigeic earthworms *Lumbricus rubellus* under laboratory conditions. The study was conducted over 50 days using four combinations in three replicates of each treatment as palm oil mill effluent: palm press fiber in 50:50 ratio (T_1), palm oil mill effluent/palm press fiber/cow dung in 50:25:25 ratio (T_2), palm oil mill effluent/palm press fiber/cow dung in 50:25:25 ratio (T_4). Twenty healthy adult *L. rubellus* with average weight of 3.92 g was introduced.

Results: Results showed that T_3 has a significant decrease in C/N ratio (14.81 ± 0.07) compared to the other treatments. The presence of cow dung and lawn clipping in the mixtures makes it more suitable for vermicomposting process as early compost productions were recorded in T_2 and T_3 .

Conclusion: The study showed that the major polluting problem in palm oil mills can be tackled through vermicomposting technique. Based on the results, vermicompost is found suitable for agriculture purposes as an organic fertilizer as well as soil conditioner.

Keywords: Vermicomposting; Palm oil mill waste; Lumbricus rubellus; Bioconversion

Introduction

The large amount of wastes generated from palm oil mill industry in Malaysia creates economic and environmental problems. The total oil palm plantation area was about 4,051,374 ha in 2005 and is expected to increase to 5.10 million ha in 2020 (Jalani et al. 2002). Meanwhile, the Malaysian Palm Oil Board (MPOB) reported that the global production of oil palm and the plantation area has increased to 41% of the world production (Rupani et al. 2010). Palm oil mills (381) in Malaysia generate about 26.7 million tonnes of solid biomass such as palm press fiber (PPF) and about 30 million tonnes of palm oil mill effluent (POME) (Yacob et al. 2005). Therefore, large area of land is needed to dispose POME and other by-products produced in large quantity during the oil palm plantation and production (Chan 2000). Yacob et al. (2006) estimated that about 0.5 to 0.75 tonnes of POME will be discharged from a mill for every tonne of fresh fruit bunch processed.

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POME is acidic in nature (pH = 3), but it is a major source of organic carbon (OC) and nitrogen.

PPF is another by-product being produced in abundant amount by palm oil mill as a solid biomass. PPF is usually left as solid waste after oil extraction (Sreekala et al. 1997). Part of the oil is lost in the fiber after screw press extraction of crude palm oil, and the rest of the oil remain in the fiber, making it suitable as a combustible material (Lim Siong et al. 2009). Although PPF is similar to rice straw, it contains a higher percentage of lignin which cannot be easily digested by animals (Prasertsan and Prasertsan 1996).

Since oil palm by-products decompose slowly under normal condition, soil micro- and macro-organisms are required to enhance the decomposition process. One of the possible bioconversion techniques is vermicomposting which is a cost effective solution to waste disposal problems (Lofs-Holmin 1986). Vermicomposting with earthworms produces humus-like product at a lower processing time, with low phytotoxicity, high in N content, greater fertilizer value, and additional production of earthworms (Lorimor et al. 2001). Therefore, vermicomposting can be considered

© 2013 Rupani et al.; licensee Springer. This is an Open Access article distributed under the terms of the Creative Commons Attribution License (http://creativecommons.org/licenses/by/2.0), which permits unrestricted use, distribution, and reproduction in any medium, provided the original work is properly cited. to be a suitable and an efficient technology to convert such industrial wastes into valuable resources.

Some species of epigeic earthworms such as *Lumbricus rubellus* can live in decaying organic waste materials and tolerate temperatures ranging from 0°C to 40°C at neutral or near neutral conditions, and the final product is rich in nitrogen content (Nagavallemma et al. 2004), odor-free, and has fine particle materials which are highly available nutrients for plants (Marsh et al. 2005; Suthar 2007a, b). The utility of epigeic earthworms for successful degradation of organic wastes is predictable for different industries such as paper and pulp (Elvira et al. 1998; Elvira et al. 1997), sugar processing (Kale 1998, Reddy and Shantaram 2005), and olive oil mill wastes (Macci et al. 2009; Macci et al. 2010).

The aim of this study was to evaluate the effectiveness of vermicomposting of POME and PPF produced from palm oil mill industries, and to determine the suitable combination of wastes in terms of rapid degradation and potential for plant establishment and growth. The experiment used PPF as a media for earthworms and to characterize the vermicompost quality.

Methods

POME and PPF were obtained from Malpom Industry, Pulau Pinang, Malaysia. Fresh cow dung (CD) was procured from a local cowshed around Penang. The earthworm species used was *L. rubellus* which is epigeic and considered as a potential organic-waste-composting worm (Edwards and Bohlen 1996). Lawn clipping (LM) was taken from the University campus. Cow dung and lawn clipping were used as amendment materials (bulking material) during this study. Chemical characteristics of POME and PPF are reported in Table 1.

Experimental layout

Four treatment groups with three replicates each were set up. Each mixture consisted of different ratio of POME, PPF, CD, and LM on weight basis. Small holes were drilled at the bottom of each plastic container ($34 \text{ cm} \times 36 \text{ cm} \times$

Table 1 Chemical characteristics of initial materials usedin the experiment

Parameters	POME	PPF
рН	3.9	7.18
C (%)	31.5	45.21
N (%)	3.8	0.5
C/N	8.3	91.45
BOD (mg/ L^{-1})	25,000	-
COD (mg/L ^{-1})	50,000	-

BOD biochemical oxygen demand, COD chemical oxygen demand, POME palm oil mill effluent, PPF palm press fiber.

Table 2 Different treatment composition

Treatment	Treatment composition
T ₁	POME (50%) + PPF (50%)
T ₂	POME (50%) + PPF (25%) + CD (25%)
T ₃	POME (50%) + PPF (20%) + CD (15%) + LM (15%)
T_4	PPF (100%) ^a

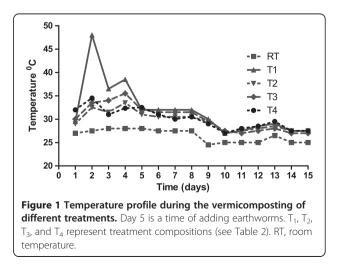
^aRepresents the percentage of initial substrate material. *CD* cow dung, *LM* lawn clipping, *POME* palm oil mill effluent, *PPF* palm press fiber.

11 cm) to drain away excess water. The composition of different feed substrates is illustrated in Table 2.

The vermicomposting unit was set up for 45 days as the sexual maturity period for this species is 4 to 6 weeks under favorable condition (Ismail 1997). The substrates were placed in the plastic container as vermireactors and placed in laboratory condition $(25^{\circ}C \pm 2^{\circ}C)$. After 120 h of pre-composting, 20 clitellate earthworms of *L. rubellus* having average live weight of 3.92 g were introduced to each setup containing 2,000 g (on wet weight basis) of substrate material (adapted from Kaviraj and Sharma (2003)).

Moisture was maintained between 70% and 80% throughout the study by sprinkling of adequate quantity of tap water. In order to prevent moisture loss, vermireactors were covered with jute bags. During precomposting, the waste material was allowed to compost, turned periodically in order to become palatable for the earthworms and eliminate volatile gases that may be toxic for the earthworms.

The worms were separated from the reactors by hand sorting, counted, and weighted weekly. The weighed earthworms were returned to their respective reactors. After 45 days, the earthworms were removed manually, and the total biomass of the earthworms was determined. Two days prior to termination, all treatments were not watered in order to sieve the compost easier.



Treatment ^a	рН	EC (ds m ⁻¹)	N (%)	OC (%)	C/N	Ca (mg L ⁻¹)	K (mg L ⁻¹)
T ₁	8.21 ± 0.02 a	2.20 ± 0.02 a	2.31 ± 0.36 b	39.01 ± 0.47 b	17.13 ± 0.07 b	$0.68 \pm 0.43 \text{ ab}$	11.97 ± 0.28 c
T ₂	8.31 ± 0.01 b	2.72 ± 0.03 b	2.53 ± 0.01 b	33.52 ± 0.27 a	$13.23 \pm 0.06 a$	0.54 ± 0.12 a	9.17 ± 0.33 b
T ₃	$8.40\pm0.01~\mathrm{c}$	$2.53 \pm 0.02 \text{ c}$	2.05 ± 0.05 b	30.44 ± 0.13 a	14.81 ± 0.07 a	$0.75 \pm 0.10 \ \mathrm{c}$	10.19 ± 0.82 b
T_4	$8.60 \pm 0.01 \text{ d}$	$2.9\pm0.04~d$	1.29 ± 0.01 a	42.21 ± 0.55 c	32.72 ± 0.33 c	0.73 ± 0.02c	7.79 ± 0.72 a

Table 3 Chemical composition of vermicompost from all treatments (mean \pm SD, n = 3)

^aTreatment composition (see Table 2). Different letters (a, b, and c) indicate statistically significant different values among treatments (ANOVA, P < 0.001).

Compost analysis

Chemical parameters were analyzed in all treatments before the introduction of earthworms and after every 15 days, up to 45 days. Compost was harvested after 45 days of vermicomposting. The following analyses were done: pH and electrical conductivity (EC) in water using a ratio of 1:10; the CNH analyzer was used to determine carbon and nitrogen. C/N ratio was calculated from the measured value of C and N. Nutrients such as Na, K, and Ca were determined after extracting the sample using ammonium acetate method (Simard 1993) and analyzed by atomic absorption spectrophotometer.

Statistical analysis

One way ANOVA was analyzed using PSAW Statistics 18, and all values are presented as the mean \pm standard deviation. The probability levels used for statistical significance were *P* < 0.05.

Results and discussion

The POME used in this study contains high level of organic matter which is indicated by high BOD (Table 1). Once the material with high organic matter is added into the PPF which is high in cellulosic material, heat is generated in the composting pile due to metabolic heat generation from biodegradation (Miyatake and Iwabuchi 2005). The temperature profiles of the first 15 days of the experiment are illustrated in Figure 1. The highest temperatures attained were at 48°C and 35°C in treatments 1 and 4, respectively, at the first 24 h of composting. Two more peaks occur at 35°C and 38°C in treatments 1 and 3 after the turning over of compost at day three. After 120 h of composting, the mean temperature of all the treatments was stable at $30^{\circ}C \pm 3^{\circ}C$, and the room temperature throughout the experiment was $25^{\circ}C \pm 2^{\circ}C$.

As presented in Table 3, the pH value in the vermicomposted material is lower than the initial value (Table 4) (P < 0.05). The reduction in pH was significant in all treatments. Maximum reduction of pH occurred in T₃, where the initial pH was at the 8.6 \pm 0.01 and reduced to 7.4 \pm 0.01 at the time of termination. The minimum reduction was in T₄, with pH 7.6 \pm 0.01 at the end of the experiment. These reductions of pH in vermicompost have also been recorded by other authors (Garg et al. 2006, Suthar and Singh 2008). According to Ndegwa et al. (2000), the reduction of pH during vermicomposting is due to the bioconversion of organic material into various intermediate types of organic acids. Also, higher mineralization of nitrogen and phosphorus into nitrites/nitrates and orthophosphate causes the lower pH (Ndegwa et al. 2000). Furthermore, it occurs due to the production of CO₂ and organic acids by microbial metabolism during decomposition of different substrates in the feed mixtures. EC increased significantly in all the treatments (Table 3). Minimum and maximum EC recorded at the end of vermicomposting experiment were in T_1 (2.20 ± 0.01) and T_4 (2.9 ± 0.04), respectively. Kaviraj and Sharma (2003) reported that the increase in EC at the end of the vermicomposting process is due to the loss of organic matter and the release of different mineral salts. Similar trends in increasing the EC at the time of termination are recorded by others (Warman and AngLopez 2002). The EC of all the treatments did not exceed the threshold value of 3 ds m⁻¹ which is recommended by Soumaré et al. (2002).

OC decreased in all the treatments at the final point of the vermicomposting process (Table 3). Maximum OC was recorded in T_4 treatment (42.21 ± 0.55), whereas minimum OC was observed in T_3 (30.44 ± 0.13). As compared to the initial substrates (Table 4), the vermicomposted material showed loss of organic carbon

Table 4 Initial chemical characteristics of different treatments (mean \pm SD, n = 3)

Treatment ^a	рН	EC (ds m ⁻¹)	N (%)	OC (%)	C/N	Ca (mg L ⁻¹)	K (mg L ⁻¹)
T ₁	7.21 ± 0.02 a	1.5 ± 0.02 c	0.78 ± 0.18 a	44.30 ± 0.20 d	58.68 ± 12.12 b	0.34 ± 0.04 b	5.65 ± 0.02 a
T_2	7.45 ± 0.08 b	1.24 ± 0.02 a	1.1 ± 0.10 b	44.39 ± 0.25 b	40.30 ± 0.58 a	0.37 ± 0.04 b	6.40 ± 0.12 b
T ₃	7.6 ± 0.011 c	1.22 ± 0.02 a	1.54 ± .02 c	45.60 ± 0.92 a	29.61 ± 0.27 a	0.21 ± 0.28 a	6.76 ± 0.17 c
T_4	7.71 ± 0.01 d	$1.43 \pm 0.02 \text{ b}$	0.49 ± 0.04 bc	45.21 ± 0.05 c	91.45 ± 7.68 a	0.22 ± 0.01 a	5.71 ± 0.23 a

^aTreatment composition (see Table 2). Different letters (a, b, and c) indicate statistically significant different values among treatments (ANOVA, P < 0.05).

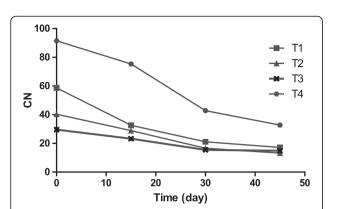


Figure 2 C/N ratio of all treatments during the

vermicomposting process.

as expected in a composting process (P < 0.01, for all the treatments). Through the muscular action of the earthworm gut, the ingested material gets fragmented and homogenized, and this increases the surface area for further microbial action, whereas the microorganisms within the earthworms' gut degrade the wastes and provide extra-cellular enzymes that are required for organic waste decomposition. Therefore, the combined action of earthworms and microorganisms causes organic carbon loss from the substrate through respiration (Edwards 1988). However, the difference in OC loss between the treatments was related to the different mixture of organic combinations in the vermireactors. The mineralization as observed in T₃ could be due to the suitable microbial condition in this treatment as compared to the other treatments.

There is a significant increase of nitrogen observed in all the treatments (Table 3). Final nitrogen in vermicomposted material was in the range 1.29 ± 0.01 to 2.31 ± 0.36 (percentage) in different treatments. Treatment 4 shows significantly less nitrogen content as compared to the other treatments; this is due to the cellulose content of fiber. As it was expected, due to the mineralization of organic carbon, the nitrogen percentage at the end of the vermicomposting process in the different treatments increased significantly. A similar trend has been observed by Kaushik and Garg (2004) who had reported a TKN increase of 2.0 to 3.2 times in textile mill sludge vermicompost. It is suggested that earthworms could increase N levels in vermicompost by the addition of their excretory products, mucus, etc. (Tripathi and Bhardwaj 2004). In general, different nitrogen pattern and mineralization activities depend on total amount of nitrogen in the initial waste and on the earthworm activity in the waste decomposition (Kale 1998; Suthar 2007b).

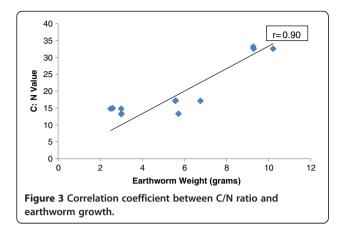
In the present study, C/N ratio offers an indication of degree of decomposition. The C/N of the vermicompost reflects the mineralization and stabilization of the substrate. The loss of carbon as carbon dioxide through microbial respiration and addition of nitrogen by earthworms in the form of mucus and nitrogenous excretory material causes the reduction of C/N ratios (Suthar 2008). Maximum and minimum C/N ratio was observed in treatment 4 (32.72 \pm 0.33) and treatment 2 (13.23 \pm 0.06) (Table 3). The C/N ratio in T_2 and T_3 did not show a significant result, whereas T₁ and T₄ showed significant results at the end of the vermicomposting process. The differences between all the treatments was significant (P < 0.001, for all treatments). Figure 2 shows that the initial C/N ratio ranges between 30 and 100 decreased to 13 and 32 in different treatments. All the treatments except treatment 4 show C/N ratio below 20, which is an indication of advanced degree of organic matter stabilization. Moreover C/N ratio below 20 reflects a satisfactory degree of organic waste maturity (Senesi 1989).

Nutrients such as Calcium (Ca) and potassium (K) concentration of vermicompost were presented in Table 3. Both Ca and K concentration in vermicompost was significantly higher than that in the initial mixture in all the treatments. Maximum Ca concentration in the final vermicompost was observed in T_3 (0.75 ± 0.1), while the maximum K concentration was in T_1 (11.97 ± 0.28) (Table 3). The minimum Ca concentration was recorded in T_2 (0.54 ± 0.12), and minimum K was in T_4 (7.79 ± 0.72) . When organic waste passes through the gut of the earthworm, calcium oxalate crystals get converted to calcium bicarbonate, which consequently enriches the worm cast with higher quality plant nutrient (Spiers et al. 1986). Similar results of calcium enrichments have been recorded by Gupta and Garg (2008b), Hartenstein and Hartenstein (1981), and Suthar (2007a,

Table 5 Growth of *L. rubellus* (mean \pm SD, n = 3) in different treatments

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Treatment ^a	lnitial weight (g)	Maximum weight achieved (g) during the experiment	Maximum weight achieved (week)	Final weight (g) at the end of the experiment	Biomass gain weight (%) at the end of the experiment
T ₁	3.72 ± 0.42 a	7.22 ± 0.19	After fourth week	5.97 ± 0.68 b	62.55
T_2	4.15 ± 0.24 a	6.82 ± 1.91	After fourth week	3.91 ± 1.56 a	-7.16
T_3	3.73 ± 0.43 a	4.946 ± 0.77	After fourth week	2.70 ± 0.26 a	-26.91
T_4	3.71 ± 0.40 a	10.46 ± 1.33	After third week	9.59 ± 0.53 c	161.24

^aFor treatment compositions, please refer to Table 2. Mean \pm standard deviation of the three replicates. Different letters (a, b, and c) indicate statistically significant different values among the treatments (ANOVA, P < 0.05).



b, c). Reports show inconsistent results regarding K in the final vermicompost of different industrial wastes. Higher potassium content has been observed in sewage sludge compost (Delgado et al. 1995), whereas Orozco et al. (1996) reported lower K in coffee pulp waste after vermicomposting. These differences in observation are generally due to the different chemical characteristics of initial feed mixtures.

Earthworm growth during vermicomposting process

A significant difference for growth and reproduction of L. rubellus is illustrated in Table 5. The biomass weight gain at the end of the experiment was significant in all treatments (P < 0.001). L. rubellus achieved the maximum weight after the fourth week of the vermicomposting process in T₁, T₂, and T₃, whereas in T₄, the maximum weight achieved was occurred on the third week of the vermicomposting process. Also, the maximum biomass weight gain at the end of the vermicomposting process (45 days) was higher in T_4 (1.61 ± 0.36), and it was significantly higher than the other treatments (P < 0.001) which could be due to the initial characteristics of the material (mainly lignin). The maximum weight gain was followed by weight loss by the end of the experiment which is followed in all the treatments particularly in the treatments 2 and 3. The biomass gain weight percentages were also calculated at the end of experiment. T₂ and T₃ show negative growth at the time of termination having -7.16% and -26.91%, respectively. A similar trend of weight loss at the time of termination of vermicomposting experiment was shown by other authors (Gupta and Garg 2008a; Suthar and Singh 2008). They described the weight loss caused by conversion of most of the used substrate to vermicompost that cannot support their growth. These results show that the addition of cow dung and lawn clipping can accelerate the vermicomposting process. Also, it is likely that earthworms digested the substrate with the addition of cow dung and/or lawn clipping before 45 days. This causes a decrease in their population due to insufficient food available to support further growth.

A positive correlation coefficient r = 0.90 was noted between the earthworm growth and the C/N ratio after 45 days of vermicomposting. The result indicated that high C/N ratio would positively increase the earthworm growth (Figure 3), as found by Gupta and Garg (2008a) and Suthar and Singh (2008). Ismail (2005) stated that the earthworms are not just sensitive to the pH of the material but also to the ingredients that are present there. While the pH range is acceptable, the components have to be auspicious. We can conclude here that the decomposition of POME and PPF mixed with cow dung and lawn clipping caused biological speed up in the vermicomposting process and reproduction of earthworms.

Conclusion

Bioconversion of palm oil mill effluent (pH 3.9) is possible through vermicomposting technology, employing epigeic earthworm *L. rubellus*. Cow dung and lawn clipping play a significant role in stabilizing the mixture and accelerate vermicomposting process. The study revealed that earthworm growth, biomass gain of *L. rubellus*, and vermicompost nutrients are higher when the POME–PPF is mixed with cow dung/lawn clipping. The study suggested that the vermicomposting is significantly effective in nutrient transformation in waste mixture. This study advocated the candidature of POME for vermicomposting operation to address the issue of sustainable industrial development.

Competing interest

The authors' declare that they have no competing interests.

Authors' contributions

PFR is the main author, carried out the vermicomposting studies and laboratory tasks, and drafted the manuscript. MHI and SAI gave supervisions and participated in the correction of the manuscript. All authors read and approved the final manuscript.

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