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Sorption behavior of strontium OPEN and europium ions from aqueous solutions using fabricated inorganic sorbent based on talc

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Sorption of Sr(II) and Eu(III) from aqueous solutions was studied using tin molybdate talc sorbent synthesized by the precipitation technique. The synthesized sorbent was characterized using diferent analytical tools, such as; FT-IR, SEM, XRD, XRF, TGA, and DTA. The sorption studies applied to Sr(II) and Eu(III) include the efects of shaking time, pH, concentrations, and saturation capacity. The sorption of Sr(II) and Eu(III) depends on pH, reaction kinetics obey the pseudo-2nd-order model, and the Langmuir model is better suited for the sorption isotherm. The thermodynamic parameters refect an endothermic and spontaneous sorption process. Desorption studies showed that 0.1 M HCl was the best desorbing agent for the complete recovery of Sr(II) (96.8%) and Eu(III) (92.9%). Finally, the obtained data illustrates that the synthesized sorbent can be applied and used as an efcient sorbent for the sorption of Sr(II) and Eu(III) from aqueous solutions and can be used as a promising sorbent to remove Sr(II) and Eu(III).

Keywords Sr(II), Eu(III), SnMoT sorbent, Reaction kinetics, Isotherm, Desorption

Various processes in power plants, recycling facilities, research centers, and using radioisotopes in industry and diagnostic medicine produce radwaste¹. The development and use of nuclear energy are accompanied by the generation of significant amounts of radioactive waste that cannot be ignored^{[2](#page-11-1)}. These streams may also contain other harmful and poisonous substances, such as heavy metals, organic material from decontamination processes, and radioisotopes. Radioactive nuclides that were emitted found their way into the soil, homes, trees, plants, water, and other buildings^{4-[8](#page-11-3)}. The radioisotope concentrations in these wastes must be reduced to acceptable levels before being released into the environment¹.

Owing to its high fssion yield (4.5%), medium half-life (28.9 y), emission of high-intensity beta rays, and high water solubility, radioactive strontium (^{90}Sr ^{90}Sr ^{90}Sr) is one of the most dangerous radionuclides 9 . In addition, ^{90}Sr readily accumulates in human bones through the food chain and has a chemistry similar to calcium. This can result in blood and bone malignancies^{10,11}. The most crucial step in achieving the upper radioactivity limit set by the Environmental Protection Agency for ⁹⁰Sr in drinking water is removing very low concentrations of the metal from natural water that contains high concentrations of competing cations (Ca^{2+} and Mg^{2+}). This limit is 0.3 Bq/L, or roughly 0.057 ppq (parts per quadrillion)^{[12](#page-11-7)}. ¹⁵²Eu (t_{1/2} = 13.54 y) and ¹⁵⁴Eu (t_{1/2} = 8.67 y) are mostly created as fssion products from radioactive waste. At the same time, they can be made by neutron irradiating isotopically enriched $152\text{Sm}_2\text{O}_3$ and neutron activating nuclear reactor control rods, respectively^{[13](#page-11-8)}. Europium is used in several medical diagnoses to detect renal and cardiac diseases and as a test to control blood hormone levels¹⁴.

Sr(II) and Eu(III) have been removed from the contaminated water using various techniques including adsorption¹⁵, chemical precipitation^{16–18}, solvent extraction¹⁹, membrane processes²⁰, and ion exchange^{21[–23](#page-12-5)} etc. Among these techniques for treating radioactive wastewater, considering its ease of handling, cheap cost, high efficiency, and high treatment capacity, adsorption may hold greater promise than many other approaches for converting trash into a more stable solid form with less volume. Adsorption is one of the main methods for eliminating radionuclides with great availability and economic viability. Consequently, creating numerous stable adsorbents is essential for decontaminating Eu(III) and Sr(II), particularly inorganic ones with strong thermal, chemical, and radiation stability, high adsorption affinity and efficiency, and good selectivity²⁴. Strontium and europium have been removed from aquatic environments using a variety of organic and inorganic adsorbents,

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including natural zeolites, porous carbon composites, graphene oxides, ammonium molybdophosphate composites, hydroxyapatites, and sodium titanates $25-27$.

Talc is branded by its 2:1 sheet structure, consisting of 2 layers of the Si–O tetrahedron and a layer of $Mg(OH)_2^{28}$. The talc (T) block is broken into T-powder, then milled and separated. It is extensively utilized in sectors including paper, plastic, rubber, food, medicine, cosmetics, and ceramics^{[29](#page-12-10)}. Talc can also be utilized as an adsorbent to remediate radioactive wastewater due to its high porosity, large specifc surface area, rod-like shape, and abundant hydrophilic Si-OH^{[30](#page-12-11)-32}.

Talc composites modified with other elements such as modified talcum³³, talc phosphogypsum ferri-silicate sorbent³⁰, Fe₃O₄/Talc³⁴, P(AA-AN)-talc^{[28](#page-12-9)}, and iron phosphate talc (IPT)³⁵. The novelty of this study includes the impregnation of tin and molybdate groups inside talc layers for the high-efficiency sorption of Sr(II) and Eu(III) from aqueous solutions that have not yet been previously studied by scholars. In this study, SnMoT sorbent was synthesized using the precipitation technique and used to remove Sr(II) and Eu(III) from aqueous solutions by batch methods^{36[,37](#page-12-17)}

This work aimed to synthesize tin molybdate talc (SnMoT) sorbent. The synthesized sorbent was characterized by utilizing various instruments. The possible uses of this sorbent in solid-phase extraction of $Sr(II)$ and $Eu(III)$ under various batch experiment conditions were assessed using their aqueous solutions.

Experimental

Materials and instruments

The main reagents synthesizing SnMoT sorbent were $SnCl₂·2H₂O$ and $Na₂Mo₄·2H₂O$, obtained from Sigma-Aldrich and Loba Chemie (India), respectively. SrCl₂ (Merck, Germany), Eu₂O₃, HNO₃, and HCl (Merck, Germany), as well as NaOH and NH₃ (El-Nasr Co, Egypt). Na₅P₃O₁₀ (Goway, China). Both chemicals and components used in this article possess analytical grades devoid of additional purifcation. For all experimentations, composites, and solutions were prepared using demineralized water. Bruker D2 Phaser II, Germany, and Alpha II Bruker, Germany, were used to evaluate SnMoT sorbent using X-ray difraction (XRD) and Fourier-transform infrared spectroscopy (FT-IR), respectively. Diferential thermal analysis (DTA) and thermogravimetric analysis (TGA) of the SnMoT sorbent were conducted with a Shimadzu DTG-60H instrument. For the TGA/DTA analysis, 20 mg samples were heated from room temperature up to 700 °C at a rate of 10 °C min−1 under a nitrogen atmosphere using alumina powder as the reference. The elemental analysis of SnMoT sorbent was detected using Philips sequential X-ray spectrometer-2400. The % of SiO_2 , MgO, MoO₃, SnO₂, Fe₂O₃, and Al₂O₃ was calculated based on the quantitative application procedure for Super-Q. The morphology of SnMoT sorbent was determined using the scanning electron microscopy (SEM) model Philips XL 30.

Preparation

Preparation of reagents

A talc-dispersed solution was prepared by dissolving 30 g talc powder in 300 mL DDW for 1 h in the presence of 0.3 g Na₅P₃O₁₀ as a dispersing agent. Na₂MoO₄·2H₂O solution (0.3 M) was prepared by dissolving 14.51 g Na₂MoO₄·2H₂O powder in 200 mL DDW for 1 h. SnCl₂·2H₂O solution (0.3 M) was prepared by dissolving 13.54 g $SnCl₂·2H₂O$ powder in 200 mL HCl (4 M) for 2 h.

Preparation of SnMo sorbent

The preparation of the SnMo sorbent was done using a co-precipitation technique. In this method, $SnCl₂·2H₂O$ solution (0.3 M) dropwise to Na_2MoO_4 :2H₂O solution (0.3 M) by a volumetric ratio equal unity at constant stirring. Afer complete addition, a brownish-red color was obtained. Ammonia solution (10% v/v) was dropwise to a mixed solution until a precipitate formed at pH (7.2), and the reaction mixture was diluted to one litter and allowed to settle through one day. The residue was washed several times to remove free chloride ions. The residue was dried for 24 h at 55 \pm 1 °C, sieved for different mesh sizes, and then stored at 25 \pm 1 °C.

Preparation of talc (T) sorbent

A talc-dispersed solution was precipitated using $(10\% \text{ v/v})$ ammonia solution dropwise. The residue was dried for 24 h at 55 ± 1 °C, sieved for different mesh sizes, and then stored at 25 ± 1 °C.

Preparation of SnMoT sorbent

The SnMoT sorbent was prepared as follows: $Na_2MoO_4.2H_2O$ solution (0.3 M) and SnCl₂·2H₂O solution (0.3 M) were added dropwise to the talc-dispersed solution by volumetric ratio Sn:Mo:T equal to 1:1:2 at constant stirring for 2 h. SnMoT was precipitated by adding (10% v/v) ammonia solution dropwise. The residue was washed, dried at 55 ± 1 °C for 24 h, grained, sieved for various mesh sizes, and stored at 25 ± 1 °C.

The best adsorbent selection

To select the best sorbent for % sorption, the % sorption of Sr(II) and Eu(III) onto produced sorbents by various volumetric ratios was carried out by shaking 0.1 g solid with 10 mL of Sr(II) and Eu(III) (100 mg/L) and $V/m = 100$ mL/g at 25 ± 1 °C for 24 h. After this time, the shaker is turned off, and the solution and solid are immediately separated. The initial and final concentrations (C_0 and C_f) of Sr(II) and Eu(III) used were measured using an atomic absorption spectrophotometer (Buck Scientifc, VGP 210) and Shimadzu UV–visible Recording Spectrophotometer (UV-160A) manufactured and supplied by Shimadzu Kyoto, Japan. The % sorption can be calculated by using $(Eq. 1)$ $(Eq. 1)$ $(Eq. 1)$ ^{[38](#page-12-18)[,39](#page-12-19)}:

2

$$
\% \text{sorption} = \left(\frac{C_{o} - C_{f}}{C_{o}}\right) 100\tag{1}
$$

The results in Table [1](#page-2-1) indicate the sequence order for the sorption of $Sr(II)$ and Eu(III) sorption onto different prepared sorbents: Sr(II) ˃ Eu(III). Also, sorbent no. 3 (SnMoT) is the best sorbent and is used for all experimental work.

Chemical stability

To test the sorbent's stability to various solvents, 50 mg of SnMoT was shaken with 50 mL of H₂O, HNO₃, HCl, and NaOH at varying concentrations [1–4 M]. Shake the appropriate solution periodically for about a week at 25 ± 1 °C. Infrared lamps were used to dry the filtrates before being examined gravimetrically^{37[,40](#page-12-20),41}.

Sorption studies

Many parameters like pH (1–8), concentration (50–1000 mg/L), agitating time (2–270 min), and temperature (25–65 °C) were checked to determine the ideal state for sorption. Batchwise contact was made between the sorbent and the sorbate solution; the samples were filtered out of the solution following sorption. All equilibrium measurements were carried out by shaking 0.1 g of SnMoT sorbent with 10 mL of Sr(II) and Eu(III) of the initial concentration of 100 mg/L with $V/m = 100$ mL/g in an agitator thermostat (Kottermann D-1362, Germany). The average of two duplicate experiments constituted all of the provided experimental results in this inquiry. The adsorption capacities at equilibrium $(q_e, mg/g)$ of Sr(II) and Eu(III) retained on the SnMoT sorbent were determined utilizing the next equation, respectively⁴²⁻⁴⁴:

$$
q_e = (C_o - C_e) \frac{V}{m}
$$

where C_0 and C_e are the initial and equilibrium concentrations of $Sr(II)$ and Eu(III) in the aqueous solution (mg/L); V is the volume of the solution (L), and m is the mass of the dried adsorbent (g).

Kinetic analysis

To elucidate the workings of the adsorption process, the pseudo-1st-order (PFO) (Eq. [3](#page-2-2)) and pseudo-2nd-order (PSO) (Eq. [4\)](#page-2-3). The PFO model, which represents a solid–liquid system, is based on the adsorbent's capacity for adsorption⁴⁵. The solid-phase adsorption capacity and the number of active centers on the adsorbent surface serve as the PSO model's foundation⁴⁶.

$$
q_t = q_e \left[1 - \exp(-K_1 t) \right] \tag{3}
$$

$$
\mathbf{q}_t = \frac{\mathbf{q}_e^2 \mathbf{K}_2 t}{1 + \mathbf{q}_e \mathbf{K}_2 t} \tag{4}
$$

t: time (min), K_1 and K_2 : the rate constants of the PFO (min⁻¹) and PSO model (g/mg_.min), respectively. Initial rates for the PFO and PSO adsorption models were computed utilizing Eqs. [\(5\)](#page-2-4) and ([6](#page-2-5)), respectively.

$$
H_1 = K_1 q_e \tag{5}
$$

$$
H_2 = K_2 q_e^2 \tag{6}
$$

 $H₁$ and $H₂$: the initial PFO and PSO adsorption rates (mg/g.min), respectively.

Isotherm modeling

The concentration data obtained to acquire the isotherms of the $Sr(II)$ and $Eu(III)$ loaded onto SnMoT sorbent were examined using nonlinear versions of the Langmuir (Eqs. [7](#page-3-0) and [8\)](#page-3-1) and Freundlich (Eq. [9\)](#page-3-2) models. Sorption isotherm measurements were made in the presence of initial concentrations (50–1000 mg/L) and pH 6 and 4 for Sr(II) and Eu(III), respectively. The Langmuir model postulates that the adsorbent surface's active adsorption centers are uniformly distributed^{[47](#page-12-26)}. The Freundlich model explains adsorbent surface heterogeneity, which also offers information on hyperbolic adsorption behavior⁴⁸.

				% sorption	
Samples	$SnCl2·2H2O$ (0.3 M)	$Na2MoO4·2H2O (0.3 M)$	Talc $(10\% \text{ w/v})$	Sr(II)	Eu(III)
$S-1$			200 mL	65.8	40.8
$S-2$	100 mL	100 mL		78.9	60.5
$S-3$	100 mL	100 mL	200 mL	99.2	76.6

Table 1. Conditions for the synthesis of diferent sorbents and their % sorption of Sr(II) and Eu(III) $(100 \text{ mg/L}, \text{V/m} = 100 \text{ mL/g}, \text{and sharing time } 24 \text{ h})$ at room temperatures.

$$
q_e = q_m \frac{K_L C_e}{1 + K_L C_e} \tag{7}
$$

$$
R_{L} = \frac{1}{1 + K_{L}C_{o}}\tag{8}
$$

$$
q_e = K_F C_e^{\frac{1}{n}}
$$
 (9)

The maximum adsorption capacity (mg/g) , Langmuir isotherm parameter, and the separation factor are denoted by q_m , K_1 , and R_1 , respectively. K_F and 1/n are Freundlich constant and adsorbent surface heterogeneity, respectively.

To estimate the degree of difference (χ^2) between the experimental data and the calculated data chi-square analysis was applied, which is calculated by the following equation⁴⁹.

$$
\chi^{2} = \sum_{i=1}^{n} \frac{\left(q_{exp.} - q_{cal.}\right)^{2}}{q_{cal.}}\tag{10}
$$

where q_{cal} and q_{exp} (mg/g) are the amount of ion adsorbed and the experimental equilibrium uptake amount, respectively. A smaller χ^2 value indicates a better-fitting isotherm.

Efect of temperature

Calculating thermodynamic parameters can help determine whether or not the adsorption process is spontaneous. Furthermore, using thermodynamic parameters at diferent reaction temperatures (298, 313, and 338 K), we can easily show the temperature effect on the Sr(II) and Eu(III) sorbed onto SnMoT sorbent. The experiment was conducted at the initial concentration of studied cations, 200 mg/L, pH 6 and 4 for Sr(II) and Eu(III), respectively, and shaking time=210 min. For the calculating of ∆H° (enthalpy), ∆S° (entropy), and ∆G˚(Gibbs free energy), we used the following Equation^{50–52};

$$
K_d = \frac{q_e}{C_e} \tag{11}
$$

$$
\Delta G^{\rm o} = -RT \ln (K_{\rm d}) \tag{12}
$$

$$
\ln K_d = \frac{\Delta S^o}{R} - \frac{\Delta H^o}{RT}
$$
\n(13)

$$
\Delta G^{\rm o} = \Delta H^{\rm o} - T\Delta S^{\rm o} \tag{14}
$$

 K_d is the distribution coefficient (mL g^{-1}), R is the gas constant, and T is the absolute temperature.

Desorption investigations

The research was done on the desorption of $Sr(II)$ and $Eu(III)$ loaded onto $SnMoT$ sorbent by a batch process with several eluent agents at ambient temperature with a volume-to-sorbent ratio of 100 mL/g. The used eluents are 0.1 M of (HCl, MgCl₂, CaCl₂, AlCl₃, and EDTA). A series of 50 mL bottles, each containing 0.1 g of loaded SnMoT sorbent by Sr(II) and Eu(III) and 10 mL of these eluents was shaken for 24 h, then following the separation of the two phases, the concentrations of Sr(II) and Eu(III) in the solid phase (C_d) and supernatant (C_s) were determined in milligrams per liter. The % of desorption was determined using $(Eq. 15)^{52}$ $(Eq. 15)^{52}$ $(Eq. 15)^{52}$:

$$
\%Desorption = \left(\frac{C_S}{C_d}\right)100\tag{15}
$$

Results and discussion Adsorbent characterization

XRD analysis

The crystalline character of SnMoT sorbent was examined using X-ray diffraction (XRD), as shown in Fig. [1a](#page-4-0). In this Figure, the crystalline structure of SnMoT sorbent is characterized by several sharp peaks at (9.46°, 10.33°, 12.63°, 15.3°, 18.92°, 21.11°, 25.3°, 26.88°, 28.78°, 31.58°, and 45.44°) related to Miller index indications (10-1, 101, 111, 210, 202, 301, 213, 400, 41-1, 124, and 42-5) respectively, with COD 00–406-1583, confrming their crystalline nature with the monoclinic system. Tis result has the same character as the IPT prepared by Mansy et al.³⁵, and P(AA-AN)-talc prepared by Abass et al.^{[28](#page-12-9)}.

FT‑IR analysis

FT-IR spectrum of SnMoT sorbent in Fig. [1b](#page-4-0) exhibits that the metal-O and metal-OH bands are observed at 560 and 683 cm−1 in SnMoT sorbent[53](#page-12-31). Two bands found at 3428 and 1632 cm−1 can be explained by intra-structure water molecules' OH frequencies vibrating in a stretched and bowed manner³⁷ or attributed to Sn-OH groups^{[54](#page-12-32),[55](#page-12-33)}.

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Figure 1. (**a**) XRD analysis (**b**) FT-IR spectrum, and **(c)** TGA and DTA analysis for SnMoT sorbent.

Tree bands observed at (1041, 793, and 471 cm−1) correspond to Sn–O[56](#page-12-34) or due to Si–O, Si–O–Al, and Si–O–Mg bending, respectively^{[57](#page-12-35)}. The Sn–O bond vibrations in the Sn–O–Mo matrix are related to FT-IR bands ranging from 700 to 450 cm^{-1 55}. The bands at 3625 and 943 cm⁻¹ are due to A1–A1–OH (stretching and bending vibra-tion, respectively)^{[57](#page-12-35),[58](#page-13-0)}. The band at 3755 cm⁻¹ is related to Al-OH-Mg bonds in talc powder^{[59](#page-13-1)}.

Termal analysis

Thermogravimetric analyses (TGA) of SnMoT sorbent (Fig. [1c](#page-4-0)), revealed a two-stage process when heated at ten °C/min. The 1st stage (32-201 °C) can be related to the desorption of physically adsorbed water from the surface of the sorbent^{41,58}. The weight loss in this region is 4.05%. The 2nd stage (201-700 °C) may be due to the loss of chemically bonded H_2O^{41} H_2O^{41} H_2O^{41} , the weight loss in this region is 63.39%. Differential thermal (DTA) shows two endothermic peaks at 143 and 265 °C due to free H_2O and chemically bonded H_2O loss. From the TGA data in Fig. [1c](#page-4-0), the weight loss for SnMoT sorbent continued up to 700 °C. The weight loss of SnMoT sorbent with a heating temperature of 10.43% reflects that SnMoT sorbent is more thermally stable than other sorbents²⁸.

SEM analysis

Figure [2](#page-5-0) displays SEM pictures of the SnMoT sorbent material at various magnifcation levels of X500, X1000, and X2000. The findings reveal a varied distribution of tin particles (white) over the molybdate medium (grey); they resemble many tiny islands on the ocean's surface. When the magnifcation power is increased to X1000 and X2000, the surface appears to have very small pores. These particles are sharp and rough, with intermolecular distances that facilitate the physical sorption process on the substance.

XRF analysis

Table [2](#page-6-0) contains the elemental analysis of SnMoT sorbent, which can be determined with XRF. Tese numbers demonstrated that the percentage of metal oxides in the SnMoT sorbent was 34.63, 18.82, 16.93, 16.61, 7.6, and 5.41 for SiO₂, MgO, MoO₃, SnO₂, Fe₂O₃, and Al₂O₃, respectively. These results verified that every component found in SnMoT sorbent is present.

Chemical stability

Table [3](#page-6-1) shows the solubility test of SnMoT sorbent toward various solvents, which refects that the SnMoT sorbent was very steady in common mineral acids and alkalies. These data are useful for the sorption process in

Figure 2. SEM images of SnMoT sorbent at diferent magnifcation powers (**a**) X500 (**b**) X1000 (**c**) X2000.

Table 2. Elemental analysis of SnMoT sorbent using X-ray fuorescence.

Table 3. Chemical stability of SnMoT sorbent in diferent solvents.

diferent media. Table [3](#page-6-1) demonstrates that, in comparison to other sorbents, SnMo sorbent has comparatively high chemical stability $60-62$ $60-62$.

Metal hydrolysis process

The side reaction of metal hydrolysis, which mostly depends on the pH of the solution, primarily affects the separation of the examined elements by the suggested adsorbent. In this context, several tests have been conducted separately. As a result, samples (10 mL each) with 50 mg/L of each element (in distilled water) were made independently at various pH levels ranging from 2 to 9. Samples were shaken for 30 min before being fltered, and every element's concentration at every pH level was tested spectrophotometrically to calculate the precipi-tation %, Table [4.](#page-6-2) Sr(II) and Eu(III) precipitated after pH 8 and 5, respectively. The results of metal hydrolysis show that all subsequent studies were conducted at pH 6 and 4 for Sr(II) and Eu(III), respectively, to avoid the hydrolysis of the metal ions.

Study of sorption

The batch method was used to sorb Sr(II) and Eu(III) from aqueous solutions using the SnMoT sorbent. The diferent parameters infuencing the individual studies of Sr(II) and Eu(III) sorption optimize their sorption on the synthesized SnMoT sorbent. The following sections detail the results that were achieved.

Efect of pH

The % sorption of $Sr(II)$ and Eu(III) from aqueous solutions by the SnMoT sorbent was studied with initial concentration (C_o) 100 mg/L, batch factor (V/m) = 100 mL/g, shaking time (24 h), and pH = (1–8) for Sr(II) and $pH = (1-5)$ for Eu(III) as shown in Fig. [3a](#page-7-0). From this Figure, the % sorption increases with increasing $pH (1-6)$ from 18.0 to 99.4% for Sr(II) and at pH (1–4) from 3.4 to 75.9% for Eu(III). Above this pH value, no change was

Table 4. Sr(II) and Eu(III) hydrolysis (%) at different pHs $(1-9)$ $C_0 = 50$ mg/L and time = 0.5 h].

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Figure 3. Sorption of Sr(II) and Eu(III) onto SnMoT sorbent (**a**) Efect of pH on the % sorption, (**b**) Efect of shaking time on the % sorption, (**c**) Efect of initial concentration on the % sorption, and (**d**) Efect of ionic strength on the % sorption.

observed for the % sorption, and all experimental work was done at pH 6 and 4 for Sr(II) and Eu(III), respectively. Additionally, it was noted that the percentage of Sr(II) and Eu(III) sorption is low at low pH levels. Tis is most likely because the surface active sites are protonated, and the amount of H_3O^+ ions in the aqueous solution increases. Consequently, the competition for the accessible binding surface active site between H_3O^+ and Sr(II) and Eu(III) was brought about by the positively charged surface sites that decreased Sr(II) and Eu(III) uptake. The concentration of OH[−] ions grew, and the concentration of H_3O^+ ions decreased as the original pH values increased, resulting in surface deprotonation of sorbents; these fndings indicate that the SnMoT sorbent's surface typically has a negative charge. As a result, there was more attraction between the sorbent's surface and the solution's positive charge of metal ions.

Infuence of shaking time

The effect of contact time on % sorption of $Sr(II)$ and $Eu(III)$ onto the synthesized SnMoT sorbent was studied, initial concentration (C_0) = 100 mg/L, batch factor (V/m) = 100 mL/g, shaking time (2–270 min), pH = 6 and 4 for Sr(II) and Eu(III), respectively. The obtained data are represented in Fig. [3b](#page-7-0) and show that the % sorption of Sr(II) and Eu(III) onto the synthesized SnMoT increased over time, reaching equilibrium at about 210 min. The rate of Sr(II) and Eu(III) sorption onto SnMoT sorbent rapidly increases from 2 to 180 min and slowly increases from 180 to 210 min, afer which there is no change in the uptake, for additional experimental work, 210 min was utilized as the equilibrium time.

Infuence of concentration

Figure [3c](#page-7-0) reveals the plots between % sorption and amount uptake q_e , (mg/g) of Sr(II) and Eu(III) onto SnMoT sorbent and C_o at the range (50–1000 mg/L) at a fixed temperature (298 ± 1 K), batch factor (V/m) = 100 mL/g, shaking time (210 min), $pH = 6$ and 4 for Sr(II) and Eu(III), respectively. The % sorption of Sr(II) and Eu(III) onto SnMoT sorbent decreases as the initial concentration of Sr(II) and Eu(III) increases. Tese data refect that the % sorption is very high at a small initial concentration due to low competition. Also, the data represented in Fig. [3](#page-7-0)c reflect that q_e of Sr(II) and Eu(III) increases as the initial concentration of Sr(II) and Eu(III) increases and the maximum q_e (33.45 and 24.82 mg/g for Sr(II) and Eu(III), respectively) carried out at initial concentration 1000 mg/L.

Infuence of ionic strength

Plots of the ionic strength of NaCl (0.01–0.5 M) and the percentage of sorption of Sr(II) and Eu(III) onto SnMoT sorbent are displayed in Fig. [3d](#page-7-0). The experiment was carried out at $[C_0=100 \text{ mg/L}, V/m=100 \text{ mL/g},$ agitating time 210 min, $pH = 6$ and 4 for Sr(II) and Eu(III), respectively]. As ionic strength increases, Fig. [3](#page-7-0)d shows a modest decrease in the percentage of sorption of $Sr(II)$ and $Eu(III)$, leading to ionic strength independence. The independence of strong ionic strength is mainly dominated by inner-sphere surface complexation⁶³.

Kinetic study

The adsorption kinetics were examined by applying the PFO and PSO model equations to the experimental data. Two steps are involved in the adsorption of Sr(II) and Eu(III) onto SnMoT sorbent (Fig. [4\)](#page-8-0). For 180 min, the frst step involved rapid adsorption. In the second stage, adsorption was slower and longer, presumably afecting the interior of the adsorbent. The initial phase was swift and dominated in terms of numbers; the second, however, was less rapid and had no quantitative impact. During the frst adsorption phase, the SnMoT surface had several accessible active centers. Following the occupation of these centers, the equilibrium condition was attained, and the second stage, which included the interior regions of the adsorbent, was started. The high concentration of active centers on the surface of SnMoT sorbent causes the fast stage; however, during the slower stage, the adsorption process's efectiveness is decreased as these sites fll more fully. During the initial adsorption stage, several active centers are on the SnMoT sorbent surface. Tese active centers are adsorbed with Sr(II) and Eu(III). As time passes, the number of active centers on the SnMoT sorbent surface grow saturated with Sr(II) and Eu(III),

Figure 4. Kinetic modelling fitting of Sr(II) and Eu(III) onto SnMoT sorbent at $[C_0 = 100 \text{ mg/L}$, $V/m = 100$ mL/g, pH = 6 and 4 for Sr(II) and Eu(III), respectively].

then Sr(II) and Eu(III) gradually difuse through the SnMoT sorbent's pore in the following step. When the PFO and PSO models' R^2 values (Table [5\)](#page-9-0) were contrasted, it was found that the PSO model fit the data better regarding kinetics. Additionally, the proximity of compatibility between the experimental and theoretically derived qe values was demonstrated with the PSO model. These findings showed that the adsorption process followed the PSO rate kinetics. Additionally, the Chi-square (χ^2) is considerably used to determine the differences between values concluded by a model and the values observed experimentally as it has the lowest value of χ^2 . As shown in Table [5,](#page-9-0) the χ^2 of the PSO model was lower than that of the PFO model, indicating the applicability of the PSO.

Sorption isotherms

Various isotherm models were employed to examine the equilibrium data and determine a suitable model for the design procedure. The Langmuir and Freundlich isotherm equations examined Sr(II) and Eu(III) sorption onto SnMoT sorbent. The correlation coefficients (R^2) consistently demonstrate the applicability of isotherm equations. Te interaction mechanism between SnMoT and Sr(II) and Eu(III) at equilibrium was determined using adsorption isotherms. When the R^2 values from the Langmuir and Freundlich isotherm models are compared (Fig. [5,](#page-9-1) and Table [6\)](#page-10-0), the adsorption process of Sr(II) and Eu(III) followed the Langmuir isotherm which ofered a better fit with $R^2 = 0.985$ and 0.994 for Sr(II) and Eu(III), respectively. The results of R_L values were (0.0023 and 0.057), reflecting the favorable sorption isotherms of $Sr(II)$ and $Eu(III)^{64}$ $Eu(III)^{64}$ $Eu(III)^{64}$. The highest amount of sorption that could be achieved was 33.5 and 28.0 mg/g for Sr(II) and Eu(III), respectively. However, the use of R^2 is limited to solving non-linear forms of isotherm equations, but not the errors in isotherm curves. In this concern, it is necessary to analyze the data set using the chi-square test statistic to assess the best-ft isotherm for the sorption

Metal ions	PFO		q_e (exp.)	PSO	
	q_e (cal.)	9.37	9.940	q_e (cal.)	10.45
	K_1	0.060		K_2	0.007
Sr(II)	H_1	0.562		H ₂	0.80
	R ²	0.973		R ²	0.996
	χ^2	0.034		χ^2	0.025
	q_e (cal.)	7.20		q_e (cal.)	7.94
	K_1	0.074		K_{2}	0.012
Eu(III)	H_1	0.530		H ₂	0.76
	R ²	0.972		R^2	0.990
	χ^2	0.026		χ^2	0.0122

Table 5. Kinetic parameters and correlation coefficients (R^2) for the sorption of $Sr(II)$ and Eu(III) onto SnMoT sorbent.

Figure 5. Isothermal modeling ftting of Sr(II) and Eu(III) onto SnMoT sorbent at [Eq. time=210 min, $V/m = 100$ mL/g, pH = 6 and 4 for Sr(II) and Eu(III), respectively].

Table 6. Isotherm parameters for sorption of Sr(II) and Eu(III) onto SnMoT sorbent.

system Eq. ([10\)](#page-3-4). According to the data in Table [6](#page-10-0), by comparing the values of χ^2 for different isotherms, it was found that the lower χ^2 values of Langmuir model pointed to the best fitting isotherm for the sorption of Sr(II) and Eu(III) onto SnMoT sorbent. Therefore, sorption isotherm data are better simulated by the Langmuir model rather than the Freundlich model. Tis reveals that monolayer sorption was the main interaction mechanism of Sr(II) and Eu(III) with SnMoT sorbent used. These findings verified that the Langmuir model is more applicable for the adsorption of Sr(II) and Eu(III) onto SnMoT sorbent.

Termodynamic studies

The influence of temperature on the % sorption of Sr(II) and Eu(III) by SnMoT sorbent was studied at an initial concentration of 200 mg/L, $pH = 6$ and 4 for Sr(II) and Eu(III), respectively, and shaking time = 210 min and the result is represented in Fig. [6](#page-10-1)a. This Figure illustrates how the endothermic nature of the sorption process is refected by an increase in the % sorption of Sr(II) and Eu(III) with increasing reaction temperature. Temperatures of 298, 313, and 338 K were investigated to interpret the thermodynamic behavior of the adsorption process (Fig. [6b](#page-10-1)). The change in ΔH° during the adsorption process was 26.7 and 29.1 kJ/mol for Sr(II) and Eu(III), respectively. Temperature increase showed a positive infuence on Sr(II) and Eu(III) elimination in the endothermic adsorption process. With the rising temperature, the amount adsorbed increased. The entropy change, ΔS°, was 145.8 and 137.6 J/mol.K for Sr(II) and Eu(III), respectively. Tis fnding revealed that the adsorption process was random. A positive entropy could be regarded as an increase in the randomness of the adsorption system as a result of the adsorbent's high affinity⁶⁵. ΔG° s were −16.7,−18.9, and−22.6 kJ/mol for Sr(II), also ΔG°s were−11.9,−13.9, and−17.4 kJ/mol for Eu(III). Te more extensive availability of ΔG˚ at higher temperatures was related to increased mobility of Sr(II) and Eu(III) sorbed onto the SnMoT sorbent surface, increased electrostatic interaction among metal ions sorbed and diferent active groups on the SnMoT sorbent surface.

Desorption investigations

Sr(II) and Eu(III) loaded onto SnMoT sorbent were desorbed using a variety of desorbing agents, and Table [7](#page-11-11) displays the data. The results show that washing with $AICl₃$ hardly desorbed the Sr(II) from the adsorbent surface. While it is relatively desorbed by washing with MgCl₂, CaCl₂, and EDTA solutions and using 0.1 M HCl

Figure 6. (**a**) Efect of reaction temperature on the % sorption of Sr(II) and Eu(III) onto SnMoT sorbent and (**b**) Van't Hoff plot of the adsorption of Sr(II) and Eu(III) onto SnMoT sorbent.

	% Desorption		
Desorbing agents, 0.1 M	Sr(II)	Eu(III)	
HCl	96.8	92.9	
MgCl ₂	34.6	22.5	
CaCl ₂	38.5	18.6	
AICl ₃	8.5	39.6	
EDTA	51.6	29.5	

Table 7. Desorption of Sr(II) and Eu(III) loaded onto SnMoT sorbent using diferent eluents.

as the eluent, high desorption of Sr(II) loaded onto SnMoT sorbent was achieved (96.8%). However, Table [7](#page-11-11) illustrates that the Eu(III) was hardly desorbed from the adsorbent surface by washing with MgCl₂, CaCl₂, and EDTA. While it is relatively desorbed by washing with AlCl₃ great desorption of Eu(III) loaded onto SnMoT sorbent was reached using 0.1 M HCl as eluent (92.9%). As a result, the order of the eluents' efficiency in releasing $Sr(II)$ from the loaded SnMoT sorbent: HCl (96.8%) > > EDTA (51.6%) > CaCl, (38.5%) > MgCl, (34.6%) > AlCl₃ (8.5%). The order is followed by the effectiveness of the eluents employed to release Eu(III) from the loaded SnMoT sorbent: HCl (92.9%) > > AlCl₃ (39.6%) > EDTA (29.5%) > MgCl₂ (22.5%) > CaCl₂ (18.6%). According to the results, it is possible to efectively recover Sr(II) and Eu(III) from loaded SnMoT sorbent with intriguing yields using 0.1 M HCl.

Conclusion

The SnMoT sorbent was prepared using the precipitation process. The SnMo sorbent was described and used to sorb Sr(II) and Eu(III) in batch technique from aqueous solutions. The produced sorbent's equilibrium time (210 min) is confrmed by the sorption data of Sr(II) and Eu(III), obeys the kinetic model of pseudo-2nd order, and is more ftting for the Langmuir isotherm with the highest possible sorption capacity was 33.5 and 28.0 mg/g for Sr(II) and Eu(III), respectively. The thermodynamic parameters displayed that the sorption process was spontaneous and endothermic, suggesting favorable adsorption under the tested conditions. 0.1 M HCl show optimum desorption of Sr(II) and Eu(III). Finally, the obtained results reveal the applicability of the fabricated sorbent as a substance that efectively absorbs Sr(II) and Eu(III) from aqueous solutions and can be used as a promising sorbent.

Data availability

All data generated or analyzed during this study are included in this published article. Datasets are available in the manuscript.

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References

- 1. Nilchi, A., Atashi, H., Javid, A. H. & Saberi, R. Preparations of PAN-based adsorbers for separation of cesium and cobalt from radioactive wastes. *Appl. Radiat. Isot.* **65**, 482–487 (2007).
- 2. Li, W.-A., Peng, Y.-C., Ma, W., Huang, X.-Y. & Feng, M.-L. Rapid and selective removal of Cs⁺ and Sr²⁺ ions by two zeolite-type sulfdes via ion exchange method. *Chem. Eng. J.* **442**, 136377 (2022).
- 3. Shao, Y. *et al.* Efect of environmental conditions on strontium adsorption by red soil colloids in southern China. *Processes* **11**, 379 (2023).
- 4. Hirose, K. 2011 Fukushima Dai-ichi nuclear power plant accident: summary of regional radioactive deposition monitoring results. *J. Environ. Radioact.* **111**, 13–17 (2012).
- 5. Harada, K. H. *et al.* Radiation dose rates now and in the future for residents neighboring restricted areas of the Fukushima Daiichi nuclear power plant. *Proc. Natl. Acad. Sci.* **111**, E914–E923 (2014).
- 6. Nishiyama, Y., Hanafusa, T., Yamashita, J., Yamamoto, Y. & Ono, T. Adsorption and removal of strontium in aqueous solution by synthetic hydroxyapatite. *J. Radioanal. Nucl. Chem.* **307**, 1279–1285 (2016).
- 7. Mikami, S. *et al.* Spatial distributions of radionuclides deposited onto ground soil around the Fukushima Dai-ichi Nuclear Power Plant and their temporal change until December 2012. *J. Environ. Radioact.* **139**, 320–343 (2015).
- 8. Yamashita, J. *et al.* Estimation of soil-to-plant transfer factors of radiocesium in 99 wild plant species grown in arable lands 1 year afer the Fukushima 1 Nuclear Power Plant accident. *J. Plant Res.* **127**, 11–22 (2014).
- 9. Delacroix, D., Guerre, P. J., Leblanc, P. & Hickman, C. Radionuclide and radiation protection data handbook 2002. *Radiat. Prot. Dosim.* **98**, 1–168 (2002).
- 10. Sahoo, S. K. *et al.* Strontium-90 activity concentration in soil samples from the exclusion zone of the Fukushima daiichi nuclear power plant. *Sci. Rep.* **6**, 23925 (2016).
- 11. Yang, H.-M., Jeon, H., Lee, Y. & Choi, M. Sulfur-modifed zeolite A as a low-cost strontium remover with improved selectivity for radioactive strontium. *Chemosphere* **299**, 134309 (2022).
- 12. Yang, J. *et al.* Enhanced removal of radioactive iodine anions from wastewater using modifed bentonite: Experimental and theoretical study. *Chemosphere* **292**, 133401 (2022).
- 13. Youssef, M. A., El-Naggar, M. R., Ahmed, I. M. & Attallah, M. F. Batch kinetics of 134Cs and 152+154Eu radionuclides onto polycondensed feldspar and perlite based sorbents. *J. Hazard. Mater.* **403**, 123945 (2021).
- 14. Attallah, M. F., Borai, E. H. & Shady, S. A. Kinetic investigation for sorption of europium and samarium from aqueous solution using resorcinol-formaldehyde polymeric resin. *J. Radioanal. Nucl. Chem.* **299**, 1927–1933 (2014).
- 15. Mu, W. *et al.* Highly efcient removal of radioactive 90Sr based on sulfonic acid-functionalized α-zirconium phosphate nanosheets. *Chem. Eng. J.* **361**, 538–546 (2019).
- 16. Jin, X., Gu, P., Zhang, G., Shang, X. & Hou, L. Removal of nickel and strontium from simulated radioactive wastewater via a pellet coprecipitation-microfltration process. *J. Radioanal. Nucl. Chem.* **301**, 513–521 (2014).
- 17. Luo, X., Zhang, G., Wang, X. & Gu, P. Research on a pellet co-precipitation micro-fltration process for the treatment of liquid waste containing strontium. *J. Radioanal. Nucl. Chem.* **298**, 931–939 (2013).
- 18. Volkovich, V. A., Grifths, T. R. & Tied, R. C. Treatment of molten salt wastes by phosphate precipitation: removal of fssion product elements afer pyrochemical reprocessing of spent nuclear fuels in chloride melts. *J. Nucl. Mater.* **323**, 49–56 (2003).
- 19. Xu, C., Wang, J. & Chen, J. Solvent extraction of strontium and cesium: a review of recent progress. *Solvent Extr. Ion Exch.* **30**, 623–650 (2012).
- 20. Rawat, N., Mohapatra, P. K., Lakshmi, D. S., Bhattacharyya, A. & Manchanda, V. K. Evaluation of a supported liquid membrane containing a macrocyclic ionophore for selective removal of strontium from nuclear waste solution. *J. Memb. Sci.* **275**, 82–88 (2006).
- 21. Chitra, S. *et al.* Optimization of Nb-substitution and Cs+/Sr+2 ion exchange in crystalline silicotitanates (CST). *J. Radioanal. Nucl. Chem.* **295**, 607–613 (2013).
- 22. Nur, T., Loganathan, P., Kandasamy, J. & Vigneswaran, S. Removal of strontium from aqueous solutions and synthetic seawater using resorcinol formaldehyde polycondensate resin. *Desalination* **420**, 283–291 (2017).
- 23. Hamed, M. M., Holiel, M. & El-Aryan, Y. F. Removal of selenium and iodine radionuclides from waste solutions using synthetic inorganic ion exchanger. *J. Mol. Liq.* **242**, 722–731 (2017).
- 24. Jiao, Z. *et al.* One-pot synthesis of silicon-based zirconium phosphate for the enhanced adsorption of Sr(II) from the contaminated wastewater. *Microporous Mesoporous Mater.* **318**, 111016 (2021).
- 25. Dragan, E. S. & Dinu, M. V. Advances in porous chitosan-based composite hydrogels: Synthesis and applications. *React. Funct. Polym.* **146**, 104372 (2020).
- 26. Kim, G. *et al.* Selective strontium adsorption using synthesized sodium titanate in aqueous solution. *RSC Adv.* **12**, 18936–18944 (2022).
- 27. Zhang, H. *et al.* Bio-Inspired preparation of clay-hexacyanoferrate composite hydrogels as super adsorbents for Cs+. *ACS Appl. Mater. Interfaces* **12**, 33173–33185 (2020).
- 28. Abass, M. R., El-Kenany, W. M. & El-Masry, E. H. High efficient removal of lead(II) and cadmium(II) ions from multi-component aqueous solutions using polyacrylic acid acrylonitrile talc nanocomposite. *Environ. Sci. Pollut. Res.* **29**, 72929–72945 (2022).
- 29. Sani, H. A., Ahmad, M. B. & Saleh, T. A. Synthesis of zinc oxide/talc nanocomposite for enhanced lead adsorption from aqueous solutions. *RSC Adv.* **6**, 108819–108827 (2016).
- 30. Hagag, M. S., Esmaeel, S. M., Salem, F., Zaki, S. A. & Ali, A. H. Uranium sorption from waste solutions by talc phosphogypsum ferri-silicate synthetic new sorbent. *Radiochim. Acta* **110**, 93–106 (2022).
- 31. Basuki, T. & Nakashima, S. Cs adsorption and CsCl particle formation facilitated by amino talc-like clay in aqueous solutions at room temperature. *ACS omega* **6**, 26026–26034 (2021).
- 32. Sprynskyy, M., Kowalkowski, T., Tutu, H., Cukrowska, E. M. & Buszewski, B. Adsorption performance of talc for uranium removal from aqueous solution. *Chem. Eng. J.* **171**, 1185–1193 (2011).
- 33. Wenlei, L. *et al.* Adsorptive characteristics of modifed talcum powder in removing methylene blue from wastewater. *Chem. Speciat. Bioavailab.* **26**, 167–175 (2014).
- 34. Kalantari, K. *et al.* Rapid adsorption of heavy metals by Fe3O4/talc nanocomposite and optimization study using response surface methodology. *Int. J. Mol. Sci.* **15**, 12913–12927 (2014).
- 35. Mansy, M. S., Eid, M. A., Breky, M. M. E. & Abass, M. R. Sorption behavior of 137Cs, 152+154Eu and 131Ba from aqueous solutions using inorganic sorbent loaded on talc. *J. Radioanal. Nucl. Chem.* **332**, 2971–2987 (2023).
- 36. Abass, M. R., Ibrahim, A. B. & Abou-Mesalam, M. M. Retention and selectivity behavior of some lanthanides using bentonite dolomite as a natural material. *Chem. Pap.* **75**, 3751–3759 (2021).
- 37. Abass, M. R., Maree, R. M. & Sami, N. M. Adsorptive features of cesium and strontium ions on zirconium tin(IV) phosphate nanocomposite from aqueous solutions. *Int. J. Environ. Anal. Chem.* **104**, 103–122 (2024).
- 38. Abass, M. R., Breky, M. M. E. & Maree, R. M. Removal of ¹³⁷Cs and ⁹⁰Sr from simulated low-level radioactive waste using tin(IV) vanadate sorbent and its potential hazardous parameters. *Appl. Radiat. Isot.* **189**, 110417 (2022).
- 39. Abass, M. R., Abou-Lilah, R. A. & Abou-Mesalam, M. M. Selective separation of cobalt ions from some fssion products using synthesized inorganic sorbent. *J. Inorg. Organomet. Polym. Mater.* <https://doi.org/10.1007/s10904-023-02957-6>(2024).
- 40. Abou-Mesalam, M. M., Abass, M. R., Abdel-Wahab, M. A., Zakaria, E. S. & Hassan, A. M. Polymeric composite materials based on silicate: II. sorption and distribution studies of some hazardous metals on irradiated doped polyacrylamide acrylic acid. *Desalin. Water Treat.* **109**, 176–187 (2018).
- 41. Abou-Mesalam, M. M., Abass, M. R., Zakaria, E. S. & Hassan, A. M. Metal doping silicates as inorganic ion exchange materials for environmental remediation. *Silicon* **14**, 7961–7969 (2022).
- 42. Metwally, S. S., Hassan, H. S. & Samy, N. M. Impact of environmental conditions on the sorption behavior of ⁶⁰Co and ¹⁵²⁺¹⁵⁴Eu radionuclides onto polyaniline/zirconium aluminate composite. *J. Mol. Liq.* **287**, 110941 (2019).
- 43. Abass, M. R., Ibrahim, A. B., El-Masry, E. H. & Abou-Mesalam, M. M. Optical properties enhancement for polyacrylonitrile-ball clay nanocomposite by heavy metals saturation technique. *J. Radioanal. Nucl. Chem.* **329**, 849–855 (2021).
- 44. Hamed, M. M., Shahr El-Din, A. M. & Abdel-Galil, E. A. Nanocomposite of polyaniline functionalized Tafa: Synthesis, characterization, and application as a novel sorbent for selective removal of Fe(III). *J. Radioanal. Nucl. Chem.* **322**, 663–676 (2019).
- 45. Lagergren, S. Zur theorie der sogenannten adsorption geloster stofe. *K. Sven. vetenskapsakademiens. Handl.* **24**, 1–39 (1898).
- 46. Ho, Y. S. & McKay, G. Te kinetics of sorption of divalent metal ions onto sphagnum moss peat. *Water Res.* **34**, 735–742 (2000).
- 47. Langmuir, I. Te adsorption of gases on plane surfaces of glass, mica and platinum. *J. Am. Chem. Soc.* **40**, 1361–1403 (1918).
- 48. Freundlich, H. Über die adsorption in lösungen. *Zeitschrif für Phys. Chem.* **57**, 385–470 (1907).
- 49. Rizk, H. E. & El-Hefny, N. E. Synthesis and characterization of magnetite nanoparticles from polyol medium for sorption and selective separation of Pd(II) from aqueous solution. *J. Alloys Compd.* **812**, 152041 (2020).
- 50. Abdel-Galil, E. A., Ibrahim, A. B. & Abou-Mesalam, M. M. Sorption behavior of some lanthanides on polyacrylamide stannic molybdophosphate as organic-inorganic composite. *Int. J. Ind. Chem.* **7**, 231–240 (2016).
- 51. Şenol, Z. M. & Şimşek, S. Insights into efective adsorption of lead ions from aqueous solutions by using chitosan-bentonite composite beads. *J. Polym. Environ.* **30**, 3677–3687 (2022).
- 52. Dakroury, G. A., El-Shazly, E. A. A. & Hassan, H. S. Preparation and characterization of ZnO/Chitosan nanocomposite for Cs(I) and Sr(II) sorption from aqueous solutions. *J. Radioanal. Nucl. Chem.* **330**, 159–174 (2021).
- 53. Varshney, K., Jain, V., Agrawal, A. & Mojumdar, S. Pyridine based zirconium(IV) and tin(IV) phosphates as new and novel intercalated ion exchangers: Synthesis, characterization and analytical applications. *J. Term. Anal. Calorim.* **86**, 609–621 (2006).
- 54. Xu, Y. *et al.* Facile assembly of 2D α-zirconium phosphate supported silver nanoparticles: Superior and recyclable catalysis. *New J. Chem.* **44**, 9793–9801 (2020).
- 55. Rao, K. T. V., Souzanchi, S., Yuan, Z., Ray, M. B. & Xu, C. Simple and green route for preparation of tin phosphate catalysts by solid-state grinding for dehydration of glucose to 5-hydroxymethylfurfural (HMF). *RSC Adv.* **7**, 48501–48511 (2017).
- 56. Mekawy, Z. A., El Shazly, E. A. A. & Mahmoud, M. R. Utilization of bentonite as a low-cost adsorbent for removal of ⁹⁵Zr(IV), ¹⁸¹Hf(IV) and ⁹⁵Nb(V) radionuclides from aqueous solutions. *J. Radioanal. Nucl. Chem.*
- 57. Taha, K. K., Suleiman, T. M. & Musa, M. A. Performance of Sudanese activated bentonite in bleaching cottonseed oil. *J. Bangladesh Chem. Soc.* **24**, 191–201 (2011).
- 58. Andrunik, M. & Bajda, T. Modifcation of bentonite with cationic and nonionic surfactants: Structural and textural features. *Materials (Basel).* **12**, 3772 (2019).
- 59. Alabarse, F. G., Conceição, R. V., Balzaretti, N. M., Schenato, F. & Xavier, A. M. In-situ FTIR analyses of bentonite under highpressure. *Appl. Clay Sci.* **51**, 202–208 (2011).
- 60. Abdel-Galil, E. A., Eid, M. A. & Hassan, R. S. Preparation of nanosized stannic silicomolybdate for chromatographic separation of Y(III) from Zr(IV). *Part. Sci. Technol.* **38**, 113–120 (2020).
- 61. Abdel-Galil, E. A., Ibrahim, A. B. & El-Kenany, W. M. Facile fabrication of a novel silico vanadate ion exchanger: Evaluation of its sorption behavior towards europium and terbium ions. *Desalin. Water Treat.* **226**, 303–318 (2021).
- 62. Ibrahim, A. B., Abass, M. R., EL-Masry, E. H. & Abou-Mesalam, M. M. Gamma radiation-induced polymerization of polyacrylic acid-dolomite composite and applications for removal of cesium, cobalt, and zirconium from aqueous solutions. *Appl. Radiat. Isot.* **178**, 109956 (2021).
- 63. Sheng, G. *et al.* Efect of humic acid, fulvic acid, pH, ionic strength and temperature on 63Ni (II) sorption to MnO2. *Radiochim. Acta* **98**, 291–299 (2010).
- 64. Hamed, M. M., Holiel, M. & Ismail, Z. H. Removal of 134Cs and 152+154Eu from liquid radioactive waste using Dowex HCR-S/S. *Radiochim. Acta* **104**, 399–413 (2016).
- 65. Şenol, Z. M. *et al.* Synthesis and characterization of chitosan-vermiculite-lignin ternary composite as an adsorbent for efective removal of uranyl ions from aqueous solution: Experimental and theoretical analyses. *Int. J. Biol. Macromol.* **209**, 1234–1247 (2022).

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Author contributions

M.R.A.: Conceptualization, Visualization, experimental work, writing-reviewing, and editing. R.A.A.-L.: Experimental work, writing—original draft review and editing. L.M.S.H.: Editing, reviewing, and experimental work.

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The authors declare no competing interests.

Additional information

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