



OPEN

Levels of potentially toxic and essential elements in Tocantins River sediment: health risks at Brazil's Savanna-Amazon interface

Thiago Machado da Silva Acioly¹, Marcelo Francisco da Silva², José Iannacone³ & Diego Carvalho Viana^{1,4}

The field study aims to address identified research gaps by providing valuable information on the concentration, spatial distribution, pollution levels, and source apportionment of toxic and essential elements in sediment samples from four sampling sites (P1: Beira Rio (urban area), P2: Bananal (rural area), P3: Embiral (rural area), P4: Cidelândia (rural area) distributed along the middle Tocantins River, Brazil. Samples were collected in 2023 from river sections and analyzed using various contamination indices (geoaccumulation index, contamination factor, enrichment factor, pollution load index, sediment pollution index, potential ecological risk coefficients, and integrated risk index). Results indicated that the levels of aluminum, iron, manganese, and selenium exceeded legal standards in that year. Chromium, nickel, copper, zinc, and lead exceeded guidelines, mainly in section P1 for aluminum and section P3 for nickel and lead. Rainy months showed increased presence, indicating seasonal variability. The geoaccumulation index indicated low pollution levels, with lead and nickel notably present near urban and industrial areas. The enrichment factor highlighted elevated concentrations of lead and zinc in industrial areas. Both PLI and SPI indices raise concerns regarding Pb (P4) and Zn (P3) concentrations at specific times of the year. Overall, potential ecological risks were deemed low for most sites. Continuous monitoring and interventions are crucial to preserve water and environmental quality in the region.

Keywords Aquatic toxicology, Biomonitoring, Ecotoxicology, Environmental impact, Heavy metals

The aquatic environment faces constant threats from contamination due to various anthropogenic pollutants, which stem from different industrial, agricultural, and urban activities. These pollutants can have diverse origins and compositions, depending on the specific activities occurring in a given region¹. For instance, mining, agrochemicals, and industrial wastewater contribute to sediment heavy metal contamination^{2,3}. Urban wastewater discharge may contain high levels of Zn, Pb, and Cu⁴ (Ancieta 2012).

Sediments serve a crucial role in preserving the ecological status of water bodies, functioning as essential ecological components in aquatic reservoirs⁵. However, despite their significance, sediments also act as repositories for various pollutants, especially heavy metals, significantly contributing to the mobilization of contaminants within aquatic systems under favorable conditions^{6,7}. Acknowledging their importance as hosts for trace metals, it is imperative to integrate sediments into environmental monitoring programs⁸.

Studies highlight the contamination of potentially toxic elements as a severe ecological crisis in aquatic environments due to their destructive, non-degradable, and bioaccumulative impact on organisms throughout

¹Postgraduate in Animal Science (PPGCA/UEMA), Multi-User Laboratories in Postgraduate Research (LAMP), State University of Maranhão, São Luís 65081-400, Brazil. ²Center for Exact, Natural and Technological Sciences (CCENT), State University of the Tocantina Region of Maranhão (UEMASUL), Imperatriz 65900-000, Brazil. ³Animal Ecology and Biodiversity Laboratory (LEBA), Universidad Nacional Federico Villarreal, 15007 Lima, Peru. ⁴Center of Agrarian Sciences, Center for Advanced Morphophysiological Studies (NEMO), State University of the Tocantina Region of Maranhão (UEMASUL), Imperatriz 65900-000, Brazil. ✉email: tmsacioly@gmail.com; diego_carvalho_@hotmail.com

the food web⁹, Padhye et al.^{10,11}. While essential elements like Na, Ca, Mn, Zn, and Fe are vital for organisms in trace amounts (Amin et al. 2021), non-essential or toxic elements like Cd, Hg, and Pb pose significant risks, even at minimal concentrations^{12,13}. Additionally, these elements strongly adhere to suspended sediments, leading to their rapid deposition in bed sediments and detrimental effects on aquatic diversity¹⁴. In cases of strong sorption, suspended particle settling and sediment formation scavenge contaminants from the water phase, contributing to their accumulation in river and lake beds¹⁵.

The quality of sediments serves as a crucial indicator of water pollution, providing valuable insights into pollutant levels. Sediment contamination, as emphasized by Alsamadany et al.¹⁶, is a significant factor in degrading aquatic environments, leading to the development of various monitoring and management methods. Monitoring sediment quality is essential for assessing water pollution, and it's imperative to share this information with the community and government to develop plans for safeguarding valuable freshwater resources. Only through such measures can we effectively control and prevent damage to the balance of the aquatic ecosystem and the health of the population.

Various indices can be used to assess potential metal pollution levels in sediment. It's important to note that all risk assessment methods provide theoretical, probabilistic results¹⁷. These indices are widely employed for detecting metal contamination and are firmly established in the literature, often mandated by legislation¹⁸. For instance, the Pollution Load Index (PLI) offers a comprehensive assessment of sediment quality regarding metal pollution, aiding in categorizing pollution levels based on established criteria. Spatial analyses of metal levels in sediments, along with differentiation from natural backgrounds, are essential and can help recognize the methods of accumulation and geochemical behavior of metals in aquatic ecosystems.

The Tocantins River, also known as the Araguaia-Tocantins due to its mouth in Pará, flows through the states of Maranhão and Tocantins. This watercourse stands out due to its extensive drainage area of 767,000 km², and hosts the two largest iron ore deposits globally: the Carajás mine in Pará State and the Serra do Carmo iron deposit in Tocantins State¹⁹. Situated in the transitional area between the Amazon and the Cerrado biomes, it plays a crucial ecological role²⁰. The river harbors a highly diverse ichthyofauna, boasting approximately 300 species, 126 genera, and 34 families, predominantly Characiformes, Siluriformes, and Cichlids²¹. Notably, the river is vital for the local population's sustenance, supporting fishing for commerce and subsistence, tourism activities such as river beaches, and traditional crafts. However, over the last 40 years, it has experienced significant degradation, primarily due to the expansion of dams, croplands, irrigation, mining, and aquaculture¹⁹.

The region faces challenges such as deforestation for agriculture, monoculture (e.g., *Eucalyptus*), and sanitation issues driven by urban growth, leading to soil erosion, sedimentation, and the introduction of pollutants²². This degradation affects water and sediment quality, impacting ecological health and community well-being. Several polluted streams flow into the study area, causing unsanitary conditions due to the direct dumping of household waste from bathrooms and kitchens, effectively turning them into open sewers²³. Elevated levels of metals such as Al, Cu, Fe, Mg, and Se exceeding national standards have been observed, indicating these elements are carried into the Tocantins River as suspended particulate matter²⁴. A recent study also found that levels of Al, Cu, Fe, Mg, and Se exceed legal standards. Particularly in urban areas, increased levels of conductivity, total dissolved solids, chlorophyll, NH₄⁺, and NO₂ raise concerns about drinking water quality²². This is critical for residents who rely on untreated river water for recreational and domestic purposes.

The field study aims to address identified research gaps by providing valuable information on the concentration, spatial distribution, pollution levels, and source apportionment of toxic and essential elements in sediment samples from sampling sites distributed along the middle Tocantins River. Additionally, contamination levels have been assessed using various indices such as the Geoaccumulation Index (Igeo), Contamination Factor (CF), Enrichment Factor (EF), Pollution Load Index (PLI), Sediment Pollution Index (SPI), Potential Ecological Risk Coefficients (Er), and their Integrated Risk Index (RI). The lack of studies and the environmental fragility of the middle Tocantins River highlight the urgent need for greater attention to environmental monitoring of this important water resource. Given that anthropogenic activities, particularly urbanization and deforestation, degrade environmental quality, they lead to compromised water and sediment conditions, posing risks to ecosystems and human health.

Material and methods

Study area, sampling, and ethical aspects

The study was carried out in the middle Tocantins River within the Imperatriz area of influence (Maranhão, Brazil). There are four sampling station areas, each equipped with GPS georeferencing using the GARMIN 4215 (Fig. 1). This region is strategically vital as a major transportation hub and gateway to the Amazon, often referred to as the portal to the Amazon. Its strong agribusiness and industrial sectors drive economic growth, generating employment and promoting development. The rainiest/coldest period (January–June) and the hot/dry season (July–December) are highlighted²⁵.

It's worth noting that the region faces several challenges, including deforestation on hillsides, intensive exploitation of fishery resources, sedimentation, sanitation issues, and direct disposal of waste and sewage into the water body, all contributing to contamination of the river waters. Consequently, these activities lead to damage to the health of the Tocantins River ecosystem, affecting both fauna and flora and negatively impacting the quality of life for the population relying on the river.

A recent study provides insights into characterizing the region, where P1 is situated in an urbanized area with significant human activity²². This site is potentially contaminated due to high vessel traffic, recreational beach use, waste disposal, residues from commercial iceboxes, and urban sewage. On the contrary, P2 has a lower potential for contamination, being situated farther away from the city. P3 is located upstream of a paper

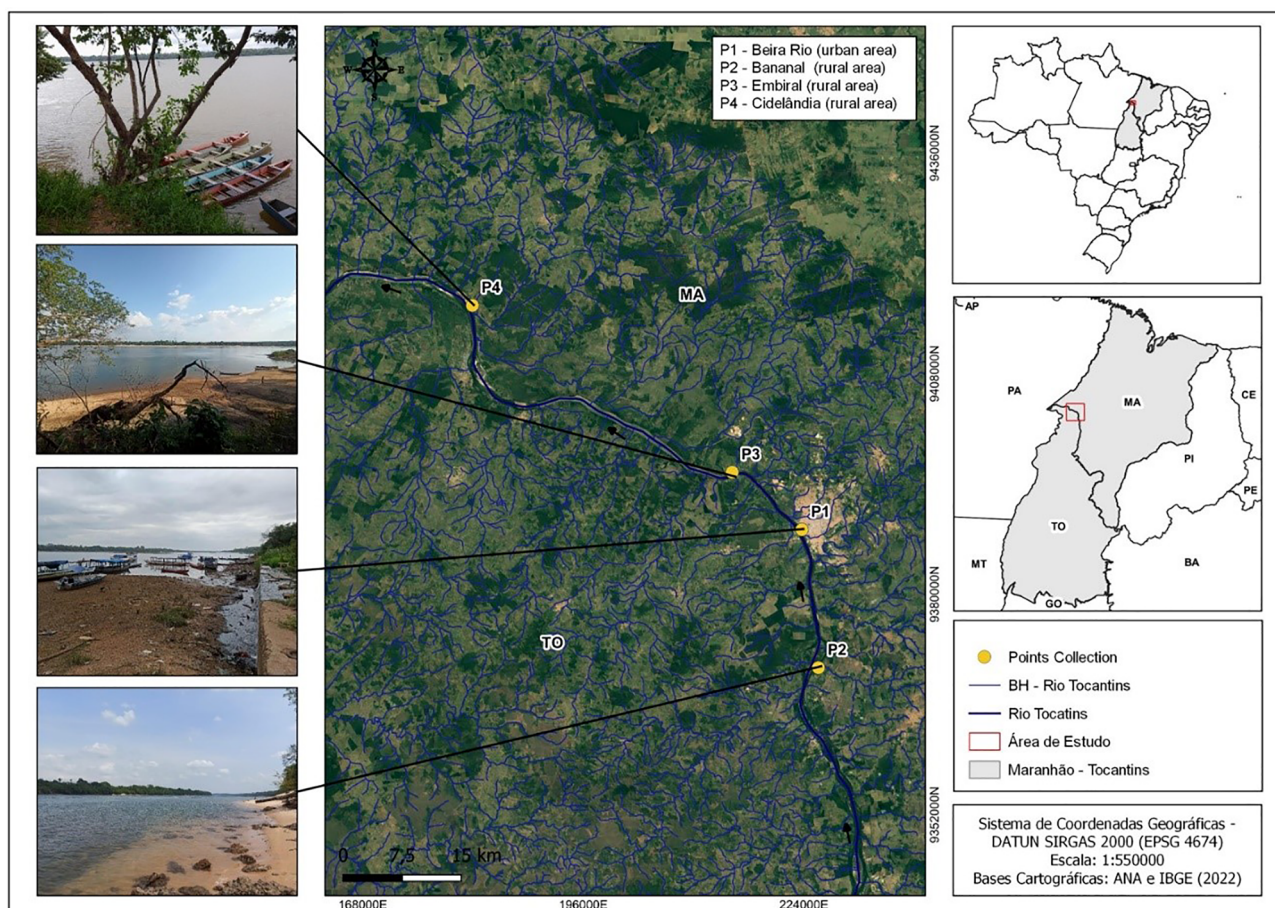


Figure 1. Geographic coordinates of sampling points in the middle Tocantins River, Maranhão, Brazil. P1: Beira Rio (urban area), P2: Bananal (rural area), P3: Embiral (rural area), P4: Cidelândia (rural area).

and cellulose manufacturing and processing industry, while P4 is approximately 100 km downstream from the city, downstream from the paper and cellulose industry.

The degree of inorganic contamination was monitored from January to December 2023, by collecting sediment samples from sections of the middle Tocantins River. Sediment samples were systematically collected monthly, with three replicates at each designated point, resulting in a total of 12 containers. These samples were carefully preserved in pre-cleaned plastic vials, thoroughly rinsed with metal-free water, and maintained within a cold chain (refrigerated at 4 °C) until they reached the laboratory at the State University of the Tocantina Region of Maranhão (UEMASUL).

Both the field collection and subsequent laboratory procedures followed ethical guidelines and were formally reviewed and approved by the Institutional Ethics Committee of the State University of the Tocantina Region of Maranhão (CEUA No: 1025220722) and the State University of Maranhão (CEUA No 040/2023). Furthermore, scientific activities were conducted with explicit authorization (SISBIO Number: 87310–1) by Normative Instruction No. 748/2022 as outlined in the ICMBio Ordinance.

Determination of potentially toxic and essential elements in sediments

Samples of 500 g of surface sediments were collected, with subsamples taken from the central part to avoid contamination. Subsequently, the samples were dried in a drying oven at 70 °C (CienlaB) for three days or until a constant mass was obtained. For the determination of total concentrations of elements in sediments, a 0.5-g sample (dry weight) was taken and 1 mL of HCl (50%), 1 mL of nitric acid (50%), and 8 mL of water were added; following the analytical methodology of the U.S. Environmental Protection Agency (US EPA Method 3050B). The digestion was then accelerated in a digestion block (ECO 16, VELP Scientifica) for 2 h at 120 °C.

The concentrations of both potentially toxic elements (aluminium—Al, antimony—Sb, arsenic—As, barium—Ba, cadmium—Cd, chromium—Cr, mercury—Hg, lead—Pb, lithium—Li, nickel—Ni, strontium—Sr, titanium—Ti, and silver—Ag) and essential elements (beryllium—Be, boron—B, selenium—Se, silica—Si, phosphorus—P, copper—Cu, iron—Fe, calcium—Ca, cerium—Ce, potassium—K, magnesium—Mg, manganese—Mn, molybdenum—Mo, sodium—Na, vanadium—V, cobalt—Co, tin—Sn, and zinc—Zn) were determined using inductively coupled plasma emission spectrometry (ICP-EAS) (SHIMADZU, ICPE-9000, Kyoto, Japan). Mercury (Hg) was determined following the specific methodology EPA 245.7:2005²⁶.

The results were compared with the National Environmental Council resolution no 454/2012²⁷. Additionally, the results were compared with international sediment guidelines, including those from the Canadian Council

of Ministers of the Environment (CCME)²⁸, the National Oceanic and Atmospheric Administration (NOAA)²⁹, and the United States Environmental Protection Agency guidelines for sediments (USEPA)³⁰.

Assessment of anthropogenic pollution in sediments

Geoaccumulation index

The Geo-accumulation Index (Igeo) has been calculated for the six studied metals: Al, Cr, Cu, Ni, Pb, and Zn. Igeo serves as a measure to assess metal contamination in sediments, soils, or other environmental samples, providing insights into the degree of accumulation or contamination relative to natural background levels³¹. The data should be processed using Eq. (1):

$$I_{geo} = \log_2 \left(\frac{C_n}{1.5 \times B_n} \right) \quad (1)$$

In this context, C_n represents the measured concentration of the metal in the sediment samples, while B_n denotes the geochemical background value in the Earth's crust³². To address potential variations in background values resulting from lithogenic variations, a factor of 1.5 is introduced³³.

Muller³¹ categorized the geoaccumulation index into six classes. The classification of contamination levels in sediment, as indicated by the Geo-accumulation Index (Igeo), is determined based on specific ranges^{34,35}. When Igeo is less than or equal to 0, it suggests no contamination. For values between 0 and 1, the sediment is categorized as experiencing no contamination to being moderately contaminated. A range of 1–2 indicates a state of moderate contamination, while 2–3 signifies a condition of moderate to heavy contamination. The levels intensify as Igeo progresses: 3–4 suggests heavy contamination, 4–5 indicates contamination from heavy to extremely contaminated, and an Igeo equal to or greater than 5 denotes an extremely contaminated state.

The logarithm to the base 2 is utilized to express the concentration relationship more clearly. It is important to note that understanding the nature of concentration values, especially about background concentration levels, is crucial for interpreting the index correctly. This classification system offers a comprehensive understanding of the contamination degree in sediment samples, thereby assisting in environmental assessments and the development of management strategies.

Enrichment factor (EF)

In the present study, the concentrations of elements in the Earth's crust were utilized as reference values, sourced from Turekian and Wedepohl³⁶. The Enrichment Factor (EF) assesses the ratio of elemental concentrations in a given sample to those in the Earth's upper continental crust (background). This ratio is calculated using Eq. (2) as proposed by Loring³⁷. The EF serves as a valuable metric for understanding the extent to which elemental concentrations in a sample deviate from the natural background levels, providing insights into potential anthropogenic influences and environmental impacts.

$$EF = \frac{\left(\frac{CM}{CX}\right)_{\text{sample}}}{\left(\frac{CM}{CX}\right)_{\text{earth crust}}} \quad (2)$$

CM represents the concentration of the target element in the sample, while CX denotes the concentration of the same element in the Earth's crust. A value of EF greater than 1 indicates enrichment, signifying an elevated concentration relative to the Earth's crust. Conversely, an EF less than 1 indicates depletion, suggesting a lower concentration compared to the typical Earth's crust. EF is extensively employed in geochemistry and environmental studies to evaluate contamination or the anomalous presence of elements in a sample. Its application provides valuable insights into potential anthropogenic influences and environmental implications.

Pollution load index (PLI)

The Pollution Load Index (PLI), developed by Tomlinson et al.³⁸, serves as a tool to assess the extent of sediment contamination by metals in diverse locations. This index can be calculated using Eq. (3):

$$PLI = \sqrt[n]{(CF_1 \times CF_2 \times \dots \times CF_i)} \quad (3)$$

The PLI indicates how many times the metal content in the sediment exceeds background levels (Soliman et al. 2021), where n represents the number of metals studied. Therefore, interpreting the PLI involves comparing the calculated value with established limits or standards. A PLI equal to 1 suggests that the area is within acceptable limits and is not significantly polluted. Values above 1 indicate an increase in the pollution load, suggesting rising levels of contamination. According to Soliman et al. (2021), a $PLI < 1$ indicates "Unpolluted," $PLI > 1$ "Polluted," and $PLI > 5$ "Very highly polluted".

Sediment pollution index (SPI)

The Sediment Pollution Index (SPI) serves as a comprehensive multi-metals index for assessing sediment quality in terms of both metal content and toxicity. The methodology for calculating SPI, as proposed by Rubio et al.³⁹, is outlined, and it is expressed through Eq. (4):

$$SPI = \frac{\sum (EF_m \times W_m)}{\sum W_m} \quad (4)$$

Toxicity weights (W_m) are assigned to each metal, with values of 2, 5, 5, 5, 1, and 1 for Cr, Cu, Ni, Pb, Al, and Zn, respectively^{40,41}. Additionally, E_{fm} represents the Enrichment Factor for each metal. The SPI index is classified into five categories based on the classification by Soliman et al. (2021): 0–2 (Natural sediment), 2–5 (Low-polluted sediment), 5–10 (Moderately polluted sediment), 10–20 (Highly polluted sediment), and > 20 (Dangerously polluted sediment).

The potential ecological risk coefficients (E_{pr}) and integrated risk index (RI)

To assess potential toxic hazards posed by metals, the potential ecological risk index (RI) derived by Hakanson⁴⁰ was estimated. According to this index, the potential ecological risk coefficient (E_r^i) for a specific metal and the overall potential ecological risk index (RI) induced by the combined impacts of the studied metals were evaluated using the following formulas:

$$C_f^i = \frac{C_s^i}{C_n^i} \quad (5)$$

$$E_r^i = T_r^i \times C_f^i \quad (6)$$

$$RI = \sum_{i=1}^n E_r^i \quad (7)$$

The ecological risk indices (RI) and potential ecological risks (E_r^p) were computed using pollution coefficients (C_f^i), metal content in sediment (C_s^i), baseline values (C_n^i), and toxicity coefficients (T_r^i) (Okbah et al. 2018; Soliman et al. 2019). The pollution coefficients for Cr, Cu, Ni, Pb, Al, and Zn were 2, 5, 5, 5, 1, and 1, respectively. Based on the calculated indices, the ecological risk levels were categorized.

RI values below 150 indicate low risk, while values between 150 and 300 suggest moderate risk⁴². RI values ranging from 300 to 600 indicate considerable risk, and values equal to or exceeding 600 represent significant risk. Similarly, E_{pr} values below 40 signify low risk, while values between 40 and 80 indicate moderate risk. Considerable risk is denoted by E_{pr} values between 80 and 160, while high risk corresponds to values between 160 and 320. E_{pr} values equal to or exceeding 320 indicate very high risk. These categorizations enable a comprehensive assessment of ecological risk levels based on metal concentrations in sediment samples.

Statistical analysis

To evaluate significant differences in the average total concentration of essential and toxic elements across sampling stations or within the same station during different periods of 2023, Analysis of Variance (ANOVA) was employed. Before conducting the analysis, it was essential to perform the Kolmogorov–Smirnov normality test and Levene's test to assess variance homogeneity. The Tukey test ($p < 0.05$) was also employed to compare means. The data were analyzed using the statistical software SPSS version 22.

To identify the association between potentially toxic and essential elements analyses, statistical tools such as correlation matrix and principal component analysis (PCA) were explored on raw data using the free software PAST 4.03 (latest version 2020). The technique reduces data dimensionality, emphasizing principal components explaining most variability, and aiding interpretation of inter-variable relationships for a comprehensive sediment quality analysis.

Results and discussion

Spatial distribution and annual average concentration of potentially toxic and essential elements in sediments from the middle Tocantins River

The sediment monitoring of the middle Tocantins River has revealed significant trends in the concentration of chemical elements throughout the year 2023. Notably, the months of February, April, and May show heightened levels of Al (Fig. 2A), Cr (Fig. 2B), Ni (Fig. 2C), and Pb (Fig. 2D). This period coincides with the rainiest and coldest months (January–June). However, these elements are undetectable (< 0.01 mg/kg) in January and March due to excessive rainfall, which dilutes the metals.

The rains began in January weakly; February saw heavy rain and thunderstorms punctuated by sunny days, while March brought continuous and heavy rainfall⁴³. This seasonal fluctuation suggests a multifaceted interaction influenced by climatic and seasonal factors. At Site P1, there is a higher presence of Al, while Ni exhibits a greater concentration in P3 (Fig. 2). Silva et al.⁴⁴ state that urban activities disrupt natural metal levels in sediments and aquatic ecosystems by altering geochemical processes that mobilize, transport, and disperse hazardous metals. Pb seems to be more prevalent in P3, except in June, where a greater concentration is observed at point P4. These concentration fluctuations underscore the necessity for a more thorough investigation to comprehend the temporal patterns and potential influencing factors linked to the presence of these elements in the waters of the middle Tocantins River. Additionally, spatial analysis of sediments can provide insights into the transport and deposition of contaminants¹⁵.

Senze et al.⁴⁵ also observed that the seasonal variations were reflected in the levels of Al detected in the sediments. The authors noted that across all rivers and their tributaries, autumn values significantly exceeded spring levels, indicating aluminum retention in sediments during the summer months. Conversely, in water, the opposite trend was observed, likely attributed to abundant precipitation and the leaching of aluminum from the catchment. Acioly et al.²² reported that February, part of the rainiest and coldest period (January–July), recorded the highest

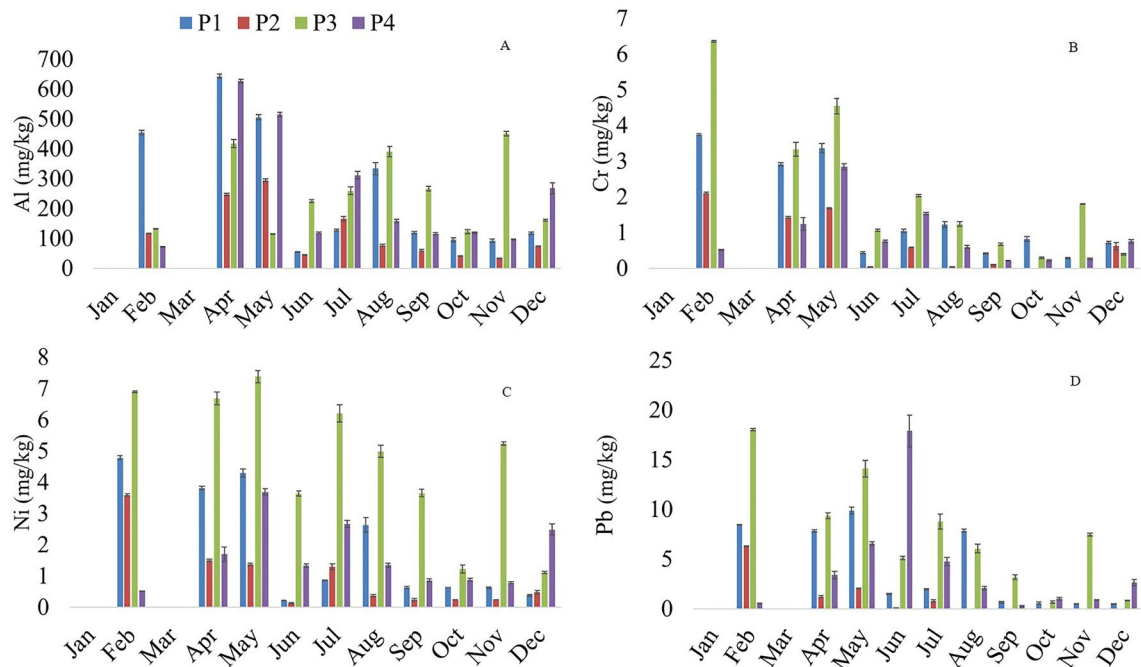


Figure 2. Spatial distribution of potentially toxic elements in the sediments of the middle Tocantins River throughout 2023, Maranhão, Brazil. P1: Beira Rio (urban area), P2: Bananal (rural area), P3: Embiral (rural area), P4: Cidelândia (rural area). Rainiest/coldest period (January to June) and the hot/dry season (July to December) (Koppen and Geiger, 1928). A: Aluminium (Al); B: Chromium (Cr); C: Nickel (Ni); D: Lead (Pb). Values represent mean concentrations (mg/kg) of each metal. Error bars indicate the standard deviation (SD) of the mean concentrations.

values of elements in the water of the middle Tocantins River. In this recent study, annual concentrations of Al exceeded both national and international water quality standards (P1: 0.69 mg/L, P2: 0.71 mg/L, P3: 0.63 mg/L; P4: 0.58 mg/L). Elevated levels of essential elements such as Cu, Fe, Mn, and Se also raise significant concerns for environmental and health quality.

Aluminum demonstrated higher values among the potentially toxic elements; however, it is noteworthy that national and international standards do not provide specific limits for this element (Table 1). Meanwhile, values for Cr, Ni, and Pb exceeded only the sediment guideline set by the National Oceanic and Atmospheric Administration²⁹. The same holds for essential elements, Cu and Zn surpassed the guidelines established by NOAA. No statistical differences were observed in the average concentrations among the studied sections of the middle Tocantins River.

The elements with the highest mean values followed this order: Fe > Al > Si > Mg > K > Na > S. The sediment's mercury content was below the detection limit (Table 1). Manganese and Iron are not explicitly addressed in national and international standards. However, according to Aiman et al.⁴⁶, its safety level is established at 30 mg/kg for sediment quality guidelines (SGV). The same source mentions a safety level of 15 mg/kg for Fe. It is crucial to underscore the lack of specific regulations for aluminum, highlighting a potential gap in standards. Continued monitoring is essential to assess the environmental and health implications of exceeding guidelines for other elements.

Aluminum ranks as the third most abundant element and the predominant metal in the Earth's crust⁴⁷. While commonly present in rocks where it is considered inert and non-bioavailable, weathering of these minerals gradually releases potentially toxic forms of aluminum into the environment, such as Al_3^+ and Al hydroxides^{47,48}. Both natural and human activities contribute to the presence of aluminum in aquatic systems. In sediment, aluminum is typically inert when bound with dissolved organic carbon (DOC) or in the form of silt or clay, serving as a storage reservoir. However, changing conditions, such as a decrease in pH, can mobilize aluminum back into the water column^{49,50}. In such cases, riverbed sediment acts as a pollution store. Health effects of aluminum exposure include impacts on the respiratory, nervous, skeletal, hematological, cardiovascular, hepatic, renal, and endocrine systems, as well as skin, eyes, and body weight^{51–53}.

According to Ohiagu et al.⁵⁴, elements such as Cr, As, Pb, Cu, Fe, Cd, Zn, and Ni are among the metals raising high public health concerns. Several studies indicate that the presence of Cr, Cd, Pb, Mn, and Ni above threshold limits suggests potential carcinogenic and genotoxic effects on humans⁵⁵, Balali-Mood et al.⁵⁶. This situation is alarming and demands continuous monitoring, public awareness, and stringent governmental policy. Notably, Point P3 exhibited slightly elevated values for Cr (2.18 mg/kg), Ni (4.70 mg/kg), and Pb (7.37 mg/kg).

3.2. Assessment of anthropogenic pollution in sediments: Geoaccumulation Index, Enrichment Factor, Contamination Factor, Pollution Load Index, Sediment Pollution Index, and Potential Ecological Risk Index.

For the Igeo, the samples of Al, Cr, and Cu consistently show only negative values, indicating a low level of pollution (Table 2). Among the metals analyzed, Ni and Pb stand out as the highest contaminants. When

| Elements | CONAMA | CCME | USEPA (Not Polluted) | NOAA | | Tocantins River | | | |
|----------------------------|--------|------|----------------------|---------|--------|-----------------|----------------|------------------|----------------|
| | | | | TEL | PEL | P1 | P2 | P3 | P4 |
| Potentially toxic elements | | | | | | | | | |
| Al | | | | | | 254.97 ± 43.13 | 116.04 ± 16.94 | 254.51 ± 29.74 | 240.32 ± 39.91 |
| As | 5.9 | 5.9 | 3 | | | 0.57 ± 0.13 | 0.15 ± 0.06 | 2.38 ± 0.34 | 0.63 ± 0.13 |
| Au | | | | | | 3.08 ± 0.34 | 1.52 ± 0.20 | 5.29 ± 0.60 | 2.75 ± 0.27 |
| Ba | | | | | | 5.10 ± 0.83 | 2.66 ± 0.31 | 9.64 ± 1.30 | 4.35 ± 0.48 |
| Cr | 37.3 | 37.3 | 25 | 0.0523 | 0.16 | 1.50 ± 0.25* | 0.66 ± 0.15* | 2.18 ± 0.39* | 0.90 ± 0.18* |
| Hg | 0.17 | 0.17 | | 0.13 | 0.7 | < 0.06 | < 0.06 | < 0.06 | < 0.06 |
| In | | | | | | 1.86 ± 0.31 | 1.08 ± 0.26 | 3.51 ± 0.54 | 1.81 ± 0.25 |
| Ni | 18 | | 20 | 0.0159 | 0.0428 | 1.78 ± 0.36* | 0.90 ± 0.21* | 4.70 ± 0.54* | 1.63 ± 0.25* |
| Pb | 35 | 35 | 40 | 0.03024 | 0.112 | 3.98 ± 0.73* | 1.05 ± 0.35* | 7.37 ± 1.19* | 4.02 ± 1.60* |
| Sb | | | | | | 1.41 ± 0.28 | 0.62 ± 0.17 | 4.04 ± 0.51 | 1.55 ± 0.22 |
| Essential elements | | | | | | | | | |
| B | | | | | | 2.05 ± 0.43 | 1.14 ± 0.24 | 2.25 ± 0.34 | 1.07 ± 0.18 |
| Ca | | | | | | 291.65 ± 26.34 | 133.19 ± 17.19 | 399.93 ± 48.36 | 145.47 ± 18.29 |
| Cu | 35.8 | 35.7 | 25 | 0.0187 | 0.39 | 1.32 ± 0.13* | 0.60 ± 0.05* | 2.11 ± 0.23* | 0.94 ± 0.10* |
| Co | | | | | | 2.98 ± 0.43 | 1.18 ± 0.24 | 5.35 ± 0.66 | 2.58 ± 0.34 |
| Fe | | | | | | 774.50 ± 139.39 | 378.81 ± 57.50 | 1105.75 ± 163.97 | 549.99 ± 83.79 |
| K | | | | | | 16.72 ± 3.82 | 11.75 ± 2.76 | 31.92 ± 6.47 | 14.87 ± 2.33 |
| P | | | | | | 2.72 ± 0.43 | 0.79 ± 0.20 | 5.29 ± 0.72 | 1.85 ± 0.32 |
| Mg | | | | | | 47.19 ± 8.02 | 30.00 ± 4.50 | 136.02 ± 17.60 | 50.35 ± 6.77 |
| Mn | | | | | | 13.63 ± 1.42 | 7.95 ± 0.96 | 29.14 ± 2.69 | 14.04 ± 1.49 |
| Na | | | | | | 15.88 ± 3.09 | 11.69 ± 1.48 | 15.02 ± 2.38 | 11.38 ± 1.65 |
| S | | | | | | 9.24 ± 2.77 | 12.09 ± 7.34 | 24.74 ± 4.70 | 4.32 ± 0.66 |
| Se | | | | | | 7.50 ± 0.93 | 3.55 ± 0.67 | 13.37 ± 1.55 | 6.16 ± 0.70 |
| Si | | | | | | 255.84 ± 31.71 | 186.83 ± 20.04 | 341.25 ± 46.73 | 218.79 ± 22.74 |
| Sn | | | | | | 3.85 ± 0.54 | 2.27 ± 0.44 | 8.55 ± 0.99 | 3.52 ± 0.41 |
| V | | | | | | 1.57 ± 0.24 | 1.01 ± 0.26 | 2.12 ± 0.32 | 1.11 ± 0.18 |
| Zn | 123 | 123 | | 0.124 | 0.271 | 4.15 ± 1.11* | 2.11 ± 1.04* | 9.03 ± 4.35* | 3.08 ± 1.58* |

Table 1. Comparison of the concentration of potentially toxic and essential elements (annual average, mg/kg) in sediment from the middle Tocantins River with national and international regulations. P1: Beira Rio (urban area), P2: Bananal (rural area), P3: Embiral (rural area), P4: Cidelândia (rural area). CONAMA: National Environmental Council resolution no 454/2012²⁷. USEPA: United States Environmental Protection Agency guidelines for sediments³⁰. CCME: Canadian Council of Ministers of the Environment²⁸. NOAA: National Oceanic and Atmospheric Administration²⁹, Sediment Guidelines. TEL: Threshold Effects Level. PEL: Probable Effects Level. These are metrics similar to the effects range metrics (i.e., Effects Range Low and Effects Range Median) used to assess the potential effects of contaminants on sediments. The Threshold Effects Level is the average of the 50th percentile and the 15th percentile of a dataset and the Probable Effects Level is the average of the 50th percentile and the 85th percentile of a dataset. *Parameter concentrations above the quality standards of some regulations in comparison. The Tukey test was performed ($p < 0.05$) for mean comparison, where “a” and “b” alone denote significant differences about other groups. The detection limit was 0.01 mg/kg for Al, Au, Hg, Ca, Pb, S, and Si, 0.02 mg/kg for Ar, B, Ba, Cu, Cr, Fe, K, Mg, Mn, Na, Ni, Se, Sn, and Zn, and 0.03 mg/kg for In, Cd, Co, P, Sb, and V.

examining the average data for 2023, only Pb recorded a positive value for P3 (0.54); nevertheless, it still falls within the range of no contamination to moderately contaminated. However, considering the maximum values obtained for metal concentration in the Igeo equation, both Ni and Pb recorded values higher than 1 for points P1 (1.50), P3 (1.92), and P4 (3.00).

The calculations indicate a minimal presence of zinc pollution in the samples. The Igeo serves as a valuable tool for evaluating the prevailing contamination levels in specific sedimentary areas^{57,58}. For the majority of points and situations, the Igeo values were predominantly negative, indicating a lack of contamination. These values offer a descriptive investigation of sediment contamination status and the accumulation profile of heavy metals^{59,60}. The average EF for the studied metals exhibited a decrease in the following order: Pb > Ni > Zn > Cu > Cr > Al. Moreover, when evaluating the maximum metal concentration values in the EF equation, Pb was the sole element surpassing a value of 1 (P4 = 2.41), whereas Zn recorded a value above 1 for P3 (P3 = 1.347) (Table 2). The EF stands out as a crucial index for identifying the sources of metals in both soil and sediment⁶¹. A value of EF greater than 1 indicates enrichment, signifying an elevated concentration relative to the Earth's crust^{62,63}. Conversely, an EF less than 1 indicates depletion, suggesting a lower concentration compared to the typical Earth's crust.

| Elements | Index value | Igeo | | | | Enrichment factor | | | |
|----------|-------------|-------|-------|-------|-------|-------------------|-----------|---------|-----------|
| | | P1 | P2 | P3 | P4 | P1 | P2 | P3 | P4 |
| Al | Min | -3.96 | -4.83 | -3.52 | -3.47 | 0.00062 | 0.00028 | 0.00083 | 0.00088 |
| | Average | -2.39 | -3.26 | -2.40 | -2.45 | 0.00319 | 0.00145 | 0.00318 | 0.00300 |
| | Máx | -2.82 | -2.82 | -1.61 | -1.39 | 0.01097 | 0.00433 | 0.00858 | 0.01053 |
| Cr | Min | -4.29 | -4.80 | -3.91 | -4.48 | 0.00078 | 0.00056 | 0.00189 | 0.00067 |
| | Average | -2.00 | -2.88 | -1.72 | -2.42 | 0.01667 | 0.00733 | 0.02422 | 0.01000 |
| | Máx | -1.20 | -2.11 | -0.95 | -1.47 | 0.05278 | 0.02667 | 0.07844 | 0.04078 |
| Cu | Min | -2.18 | -3.04 | -2.08 | -2.63 | 0.01133 | 0.00489 | 0.01333 | 0.00667 |
| | Average | -1.49 | -2.35 | -1.08 | -1.93 | 0.02933 | 0.01333 | 0.04711 | 0.02089 |
| | Máx | -0.84 | -1.72 | -0.37 | -1.14 | 0.08356 | 0.02800 | 0.13644 | 0.06467 |
| Ni | Min | -5.15 | -7.14 | -2.20 | -4.05 | 0.00138 | 0.00016 | 0.01926 | 0.00441 |
| | Average | -1.16 | -2.10 | -0.43 | -1.32 | 0.02618 | 0.01324 | 0.06912 | 0.02397 |
| | Máx | 0.61 | -0.10 | 1.08 | 0.02 | 0.08588 | 0.05882 | 0.17471 | 0.06941 |
| Pb | Min | -2.39 | -2.85 | -1.79 | -1.73 | 0.035 | 0.022 | 0.0545 | 0.0595 |
| | Average | -0.22 | -2.12 | 0.54 | -0.03 | 0.199 | 0.0525 | 0.369 | 0.201 |
| | Máx | 1.50 | 0.85 | 1.92 | 3.00 | 0.695 | 0.33 | 0.93 | 2.41 |
| Zn | Min | -6.01 | -7.62 | -3.34 | -6.01 | 0.00105 | 0.0002526 | 0.01036 | 0.0003368 |
| | Average | -1.38 | -2.53 | -0.65 | -2.06 | 0.04368 | 0.02221 | 0.09474 | 0.03242 |
| | Máx | -0.54 | -0.41 | 0.92 | -0.14 | 0.2653 | 0.3021 | 1.347 | 0.4811 |

Table 2. Geoaccumulation index and enrichment factor for investigated metals in the sediment from the Middle Tocantins River, Maranhão, Brazil. Min: minimum value observed for the element; Average: average value observed for the element; Máx: maximum value observed for the element. P1: Beira Rio (urban area), P2: Bananal (rural area), P3: Embiral (rural area), P4: Cidelândia (rural area). Igeo values: ≤ 0 (no contamination), 0–1 (none to moderate contamination), 1–2 (moderate contamination), 2–3 (moderate to heavy contamination), 3–4 (heavy contamination), 4–5 (heavy to extreme contamination), ≥ 5 (extreme contamination). An EF > 1 indicates enrichment (elevated concentration relative to Earth's crust), while an EF < 1 indicates depletion (lower concentration than typical Earth's crust).

The mean values for the studied metals revealed a higher PLI for Ni (0.235) and Pb (0.184), both observed at P3 (Table 3). On the other hand, considering the observed maximum values obtained through ICPE analysis, the PLI hierarchy is as follows: Pb (1.205 at P4) > Zn (1.041 at P3) > Ni (0.595 at P3) > Cr (0.282 at P3) > Cu (0.246 at P3). These results suggest that, based on the mean values, there is no imminent concern. However, at certain moments or periods of the year, there is an escalation in pollution levels (PLI > 1 “Polluted” according to Soliman et al. 2021), particularly for lead and zinc. This emphasizes the importance of monitoring and implementing environmental management strategies to mitigate further contamination risks.

A variety of pollution indices are at our disposal for evaluating the quality of soil and sediment. The selection of these indices is intricately tied to specific objectives, taking into account factors such as contamination levels, metal origins, and potential ecological risks^{64–66}. Furthermore, the mathematical and statistical methodology behind these indices makes them practical and highly efficient tools for environmental studies⁶¹. Each index specializes in investigating contamination, emphasizing the need to tailor the choice based on the distinct goals of the assessment.

Similar to the PLI index, the SPI provides a composite measure of sediment contamination by various pollutants, allowing for a comprehensive assessment of sediment quality concerning pollution^{67,68}. In this study, all investigated metals showed low pollution levels (Natural sediment), with Pb being the most concerning, having a global SPI value of 0.444 (Table 3). The overall PLI value was also higher for Pb. However, when considering the maximum concentration of metals found, Pb stands out with a SPI of 2.410 (P4), classified as “Low-polluted” sediment. Despite Al showing among the highest mean values, its accumulation index values were low, likely due to its high reference value (a relatively abundant element in the Earth's crust).

The estimation results of potential ecological risk coefficients (Epr) and their integrated risk index (RI) for sediment samples are presented in Table 4. The order of Epr for the studied metals is Pb > Ni > Zn > Cu > Cr > Al. Consequently, the estimated Er values for these elements are all lower than 40, indicating a low level of ecological risk⁴². As for the RI values, the combined potential ecological risks are assessed to be at a low degree of contamination for the majority of the studied sites. The highest RI value was observed for Pb (15.749) (Table 4). However, even with this value, the overall risk level for the middle Tocantins River remains classified as “Low-grade” in terms of potential ecological risk.

All estimations were conducted across the minimum, average, and maximum values observed in sediment samples collected in 2023. This approach allows for a comprehensive understanding of the temporal variability of the risks involved. By considering the full range of values within the sediment samples, we can better grasp how ecological risk levels fluctuate over time. This comprehensive analysis enhances our ability to assess and manage potential environmental risks associated with metal contamination in sediment ecosystems. Comprehensive

| Elements | Index value | Pollution load index (PLI) | | | | Global | Sediment pollution index (SPI) | | | | Global |
|----------|-------------|----------------------------|---------|-------|---------|--------|--------------------------------|---------|---------|----------|--------|
| | | P1 | P2 | P3 | P4 | | P1 | P2 | P3 | P4 | |
| Al | Min | 0.089 | 0.074 | 0.112 | 0.097 | 0.135 | 0.00062 | 0.00028 | 0.00083 | 0.00088 | 0.011 |
| | Average | 0.153 | 0.114 | 0.147 | 0.145 | | 0.00319 | 0.00145 | 0.00318 | 0.00300 | |
| | Máx | 0.200 | 0.151 | 0.182 | 0.196 | | 0.01097 | 0.00433 | 0.00858 | 0.01053 | |
| Cr | Min | 0.003 | 0.002 | 0.007 | 0.002 | 0 | 0.00156 | 0.00112 | 0.00378 | 0.00134 | 0.044 |
| | Average | 0.060 | 0.026 | 0.087 | 0.036 | | 0.03334 | 0.01466 | 0.04844 | 0.020 | |
| | Máx | 0.190 | 0.096 | 0.282 | 0.147 | | 0.10556 | 0.05334 | 0.15688 | 0.08156 | |
| Cu | Min | 0.020 | 0.009 | 0.024 | 0.012 | 0.066 | 0.01133 | 0.00489 | 0.01333 | 0.00667 | 0.043 |
| | Average | 0.053 | 0.024 | 0.084 | 0.038 | | 0.02933 | 0.01333 | 0.04689 | 0.02089 | |
| | Máx | 0.150 | 0.050 | 0.246 | 0.116 | | 0.08356 | 0.028 | 0.13644 | 0.06467 | |
| Ni | Min | 0.005 | 0.0005 | 0.065 | 0.015 | 0.128 | 0.00138 | 0.00016 | 0.01926 | 0.00441 | 0.024 |
| | Average | 0.089 | 0.045 | 0.235 | 0.081 | | 0.01309 | 0.00647 | 0.03397 | 0.01176 | |
| | Máx | 0.292 | 0.200 | 0.595 | 0.236 | | 0.04294 | 0.02941 | 0.06985 | 0.02765 | |
| Pb | Min | 0.017 | 0.011 | 0.027 | 0.030 | 0.305 | 0.035 | 0.022 | 0.0545 | 0.0595 | 0.444 |
| | Average | 0.099 | 0.026 | 0.184 | 0.100 | | 0.199 | 0.052 | 0.368 | 0.201 | |
| | Máx | 0.347 | 0.165 | 0.465 | 1.205 | | 0.695 | 0.330 | 0.930 | 2.410 | |
| Zn | Min | 0.0008 | 0.00019 | 0.008 | 0.00026 | 0.112 | 0.00105 | 0.00025 | 0.01037 | 0.000337 | 0.090 |
| | Average | 0.034 | 0.017 | 0.073 | 0.025 | | 0.01379 | 0.00705 | 0.03063 | 0.01047 | |
| | Máx | 0.205 | 0.234 | 1.041 | 0.372 | | 0.08526 | 0.09726 | 0.43158 | 0.15410 | |

Table 3. Pollution load index and sediment pollution index for investigated metals in the sediment from the middle Tocantins River, Maranhão, Brazil. Min: minimum value observed for the element; Average: average value observed for the element; Máx: maximum value observed for the element. P1: Beira Rio (urban area), P2: Bananal (rural area), P3: Embiral (rural area), P4: Cidelândia (rural area). A PLI < 1 indicates “Unpolluted,” PLI > 1 “Polluted,” and PLI > 5 “Very highly polluted.” SPI index: 0–2 (Natural), 2–5 (Low), 5–10 (Moderate), 10–20 (High), > 20 (Dangerous).

| Sampling site | Index value | Potential ecological risk coefficients (Epr) | | | | | |
|----------------------------|-------------|--|-------|-------|-------|--------|-------|
| | | Al | Cr | Cu | Ni | Pb | Zn |
| P1 | Min | 0.001 | 0.016 | 0.056 | 0.070 | 0.175 | 0.001 |
| | Average | 0.003 | 0.033 | 0.147 | 0.131 | 0.995 | 0.044 |
| | Máx | 0.008 | 0.106 | 0.419 | 0.416 | 1.724 | 0.265 |
| P2 | Min | 0.001 | 0.011 | 0.024 | 0.009 | 0.110 | 0.000 |
| | Average | 0.002 | 0.015 | 0.067 | 0.059 | 0.262 | 0.022 |
| | Máx | 0.004 | 0.053 | 0.140 | 0.294 | 0.825 | 0.303 |
| P3 | Min | 0.001 | 0.038 | 0.067 | 0.101 | 0.272 | 0.010 |
| | Average | 0.003 | 0.052 | 0.233 | 0.410 | 1.842 | 0.096 |
| | Máx | 0.006 | 0.157 | 0.687 | 0.882 | 2.790 | 1.347 |
| P4 | Min | 0.001 | 0.013 | 0.033 | 0.022 | 0.297 | 0.000 |
| | Average | 0.003 | 0.020 | 0.104 | 0.114 | 0.402 | 0.033 |
| | Máx | 0.008 | 0.082 | 0.305 | 0.319 | 6.055 | 0.482 |
| Integrated risk index (RI) | | 0.041 | 0.596 | 2.282 | 2.827 | 15.749 | 2.603 |

Table 4. The potential ecological risk index for investigated metals in the sediment from the middle Tocantins River, Maranhão, Brazil. P1: Beira Rio (urban area), P2: Bananal (rural area), P3: Embiral (rural area), P4: Cidelândia (rural area). Min: minimum value observed for the element; Average: average value observed for the element; Máx: maximum value observed for the element. Epr values: < 40 (Low risk), 40–80 (Moderate risk), 80–160 (Considerable risk), 160–320 (High risk), ≥ 320 (Very high risk). RI values: < 150 (Low risk), 150–300 (Moderate risk), 300–600 (Considerable risk), ≥ 600 (Significant risk).

ecological risk assessment is of great significance for the restoration of watershed ecosystem health, and the appropriate and effective assessment method is the premise of ecological risk assessment⁶⁶.

Correlation coefficients and principal component analysis (PCA)

The strongest positive correlations are observed within specific pairs of data: Cu/Si ($r = 0.999$), Ni/Sn ($r = 0.999$), Au/Co ($r = 0.999$), As/Mg ($r = 0.999$), and Cr/Fe ($r = 0.996$) (Table S1). A negative correlation was not observed. Correlation coefficients offer valuable insights into the strength and direction of relationships among different

parameters related to water quality. They are commonly used to analyze associations between persistent contaminants in aquatic ecosystems^{69–71}. A coefficient closer to 1 or –1 indicates a stronger correlation, while values near 0 suggest a weaker or nonexistent linear relationship.

Correlation coefficients provide valuable insights into the strength and direction of relationships among analyzed parameters, enhancing our understanding of sediment quality dynamics and potential aquatic contaminant sources^{72,73}. Consequently, numerous studies utilize both univariate and multivariate statistical analyses to identify highly correlated sediment pollutants and relevant industries^{74,75}. This enables researchers to effectively trace these pollutants and industries, leading to the development of more efficient pollution management strategies and mitigation efforts.

The PCA results indicate that the first two components explain 99.38% of the data variability (Fig. 3). The first component accounts for 97.5% of the variance and is positively loaded with As, B, Ba, Ca, Co, Cr, Cu, and Fe. Conversely, the second component accounts for 1.92% of the variance and is positively loaded with Al and Pb, while exhibiting negative loadings for Ca, Mg, S, and Zn. In this study, the correlation matrix was used to better capture the relationships between metals, minimizing the impact of different variance scales among them. The loadings, scores, and eigenvalues of the principal component analysis are available in the supplementary material (Tables S2, S3, and S4).

Highlighting the considerable role of aluminum in sediment contamination, PCA offered valuable insights into the underlying dynamics. These sediments can serve as significant sources of metal contamination in aquatic environments, as deposited or stored contaminants may re-suspend or dissolve back into the water. Hence, employing this technique was crucial for pinpointing pollution sources and extracting reliable information to elucidate clearer relationships^{76,77}.

The PCA analysis unveiled three distinct groups based on sediment quality. The first group comprised stations P1 (located in an urban area) and P3 (close to industrial waste discharge). P1 showed a positive association with the first and second components but a negative association with the third, whereas P3 exhibited a positive association with the first and third components. The second group solely included station P2, situated near the village of “Bananal” in a rural area, which displayed a negative association with all PCA components. The third group consisted solely of station P4, also located in rural areas but in proximity to industrial waste discharge (from pulp and paper production). P4 showed a positive association with both the second and third components but a negative association with the first. This underscores the effectiveness of PCA in identifying pollution sources and providing valuable insights for conservation and management efforts.

This research is one of the few studies on the quality of sediments in the middle Tocantins River, highlighting high concentrations of metals such as Al, Cr, Ni, Pb, Cu, and Zn in an urbanized region of the river. Unregulated urbanization has led to environmental changes, adding high loads of untreated domestic and industrial wastewater, and altering biological, physical, and chemical processes in natural systems, directly or indirectly affecting the population's quality of life^{78–80}. The river's maximum flow is approximately 11,000 m³/s, contributing 8,388 t/day of suspended solid discharge to the ocean⁸¹, supporting marine biodiversity at its mouth. Additionally, the Tocantins River basin is among the most degraded in the Amazon^{19,82}. According to Schmitz⁸³, tropical deforestation and unsustainable policies pose current challenges in this basin, which serves as the ecotone between the Amazon and Cerrado biomes. This region is a primary target for agricultural expansion in Brazil, as indicated in Presidential Decree 8447 of 2015^{19,84}. Furthermore, Brazil stands as the world's second-largest agricultural producer and is forecasted to undergo substantial growth in agricultural output over the next four decades^{85,86}. Despite these challenges, freshwater ecosystems in this region play vital roles in biodiversity preservation, water purification, and flood control.

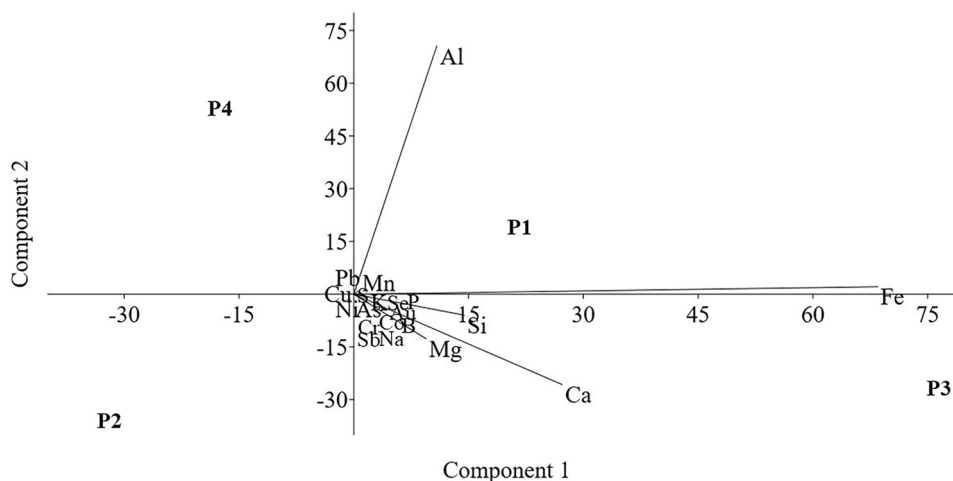


Figure 3. Principal component analysis (PCA) of sediment quality monitoring data of the middle Tocantins River, Maranhão, Brazil. P1: Beira Rio (urban area), P2: Bananal (rural area), P3: Embiral (rural area), P4: Cidelândia (rural area).

In this scenario, this region emerges as a pivotal case study, and measures might be implemented to decrease the levels of these potentially toxic elements. Effective measures to address these issues include the treatment of industrial and urban effluents and the reforestation and conservation of riparian forests^{87,88}. Industries and urban areas should be required to properly treat their effluents before discharging them into the river. Water treatment technologies such as chemical precipitation, adsorption, and filtration can effectively remove heavy metals^{89–91}. Additionally, vegetation along riverbanks (riparian forests) acts as a natural filter, reducing the amount of sediment and contaminants reaching the river^{92,93}. Reforestation projects can help restore these areas, enhancing their capacity to mitigate pollution and protect aquatic ecosystems.

Often, environmental degradation stems from a lack of environmental responsibility and ineffective public policies. Thus, the ongoing deforestation, even in conservation-oriented scenarios, raises questions about the effectiveness of current protection measures, highlighting the need for reassessment and strengthening of conservation strategies. Therefore, ongoing monitoring and enforcement, along with education and awareness efforts, are essential^{94–96}. Establishing a continuous monitoring system for pollution levels and rigorous enforcement to ensure adherence to environmental laws is crucial. Furthermore, educational campaigns for local communities and industries on the impacts of pollution and the importance of sustainable practices can foster changes in behavior.

Conclusion

The levels of aluminum, iron, manganese, and selenium exceeded legal standards through 2023. Although aluminum showed the highest values among the potentially toxic elements, there are no specific limits in national and international standards for this element. Chromium, nickel, copper, zinc, and lead only surpassed the guidelines established by NOAA (National Oceanic and Atmospheric Administration). The seasonal fluctuation indicates a complex dynamic influenced by climatic or seasonal factors. The rainy months of February, April, and May showed heightened levels of aluminum, chromium, nickel, and lead. Particularly, in urban area P1 had a higher presence of Al, while Ni and Pb were more concentrated in P3.

Igeo indicates low pollution levels, with lead and nickel standing out in urban areas (P1) and near the pulp and paper industry (P3 and P4). The EF highlights high concentrations of Pb (P4) and Zn (P3). Both PLI and SPI indices raise concerns regarding Pb (P4) and Zn (P3) concentrations at specific times of the year. The overall risk level for the middle Tocantins River remains classified as “Low-grade” concerning potential ecological risk. These findings stress the importance of continuous monitoring and interventions to uphold water and environmental quality in the region.

Data availability

The datasets used and/or analyzed during the current study are available from the corresponding author on reasonable request.

Received: 19 March 2024; Accepted: 2 July 2024

Published online: 04 August 2024

References

- Guabloche, A., Alvarino, L., Acioly, T. M. D. S., Viana, D. C. & Iannaccone, J. Assessment of essential and potentially toxic elements in water and sediment and the tissues of *Sciaena deliciosa* (Tschudi, 1846) from the Coast of Callao Bay, Peru. *Toxics* **12**(1), 68. <https://doi.org/10.3390/toxics12010068> (2024).
- Omwene, P. I., Öncel, M. S., Çelen, M. & Kobya, M. Heavy metal pollution and spatial distribution in surface sediments of Mustafakemalpaşa stream located in the world's largest borate basin (Turkey). *Chemosphere* **208**, 782–792. <https://doi.org/10.1016/j.chemosphere.2018.06.031> (2018).
- Ullah, R., Muhammad, S. & Jadoon, I. A. Potentially harmful elements contamination in water and sediment: Evaluation for risk assessment and provenance in the northern Sulaiman fold belt, Baluchistan, Pakistan. *Microchem. J.* **147**, 1155–1162. <https://doi.org/10.1016/j.microc.2019.04.053> (2019).
- Guadie, A. *et al.* Effluent quality and reuse potential of urban wastewater treated with aerobic-anoxic system: A practical illustration for environmental contamination and human health risk assessment. *J. Water Process Eng.* **40**, 101891. <https://doi.org/10.1016/j.jwpe.2020.101891> (2021).
- Li, K. The application of reweighted markov chains in water-sediment prediction in inland river basins: A case study of the shiyang river basin. *Sustain. Cities Soc.* **73**, 103061. <https://doi.org/10.1016/j.scs.2021.103061> (2021).
- Liao, J. *et al.* Heavy metals in river surface sediments affected with multiple pollution sources, South China: Distribution, enrichment and source apportionment. *J. Geochem. Explor.* **176**, 9–19. <https://doi.org/10.1016/j.gexplo.2016.08.013> (2017).
- Peng, W., Li, X., Xiao, S. & Fan, W. Review of remediation technologies for sediments contaminated by heavy metals. *J. Soils Sediments* **18**, 1701–1719. <https://doi.org/10.1007/s11368-018-1921-7> (2018).
- Khaled, A., Abdel-Halim, A., El-Sherif, Z. & Mohamed, L. A. Health risk assessment of some heavy metals in water and sediment at Marsa-Matrouh, Mediterranean Sea, Egypt. *J. Environ. Protect.* **8**(01), 74. <https://doi.org/10.4236/jep.2017.81007> (2017).
- Niede, R. & Benbi, D. K. Integrated review of the nexus between toxic elements in the environment and human health. *AIMS Public Health* **9**(4), 758. <https://doi.org/10.3934/publichealth.2022052> (2022).
- Marengo, M. *et al.* Ecological and human health risk assessment of potentially toxic element contamination in waters of a former asbestos mine (Canari, Mediterranean Sea): Implications for management. *Environ. Monitor. Assess.* **195**(1), 150. <https://doi.org/10.1007/s10661-022-10737-x> (2023).
- Padhye, L. P. *et al.* Contaminant containment for sustainable remediation of persistent contaminants in soil and groundwater. *J. Hazard. Mater.* <https://doi.org/10.1016/j.jhazmat.2023.131575> (2023).
- Demir, C., Öner, M., Çetin, G. & Bakırder, S. A simple and effective graphene oxide-zinc oxide nanocomposite based solid phase extraction method for the determination of cadmium in tea matrix at trace levels by flame atomic absorption spectrophotometry. *J. Food Compos. Anal.* <https://doi.org/10.1016/j.jfca.2024.105968> (2024).
- Kanici Tarhane, A., Aluc, Y., Kiziltepe, S. & Ekici, H. An Investigation of heavy metal concentrations in the sera of cattle grazed in different locations in the Kars Province of Türkiye. *Bull. Environ. Contam. Toxicol.* **112**(1), 1. <https://doi.org/10.1007/s00128-023-03821-6> (2024).

14. Haghazari, H. *et al.* Potentially toxic elements contamination in surface sediment and indigenous aquatic macrophytes of the Bahmanshir River, Iran: Appraisal of phytoremediation capability. *Chemosphere* **285**, 131446. <https://doi.org/10.1016/j.chemosphere.2021.131446> (2021).
15. Chiaia-Hernández, A. C. *et al.* Sediments: Sink, archive, and source of contaminants. *Environ. Sci. Pollut. Res.* **29**(57), 85761–85765. <https://doi.org/10.1007/s11356-022-24041-1> (2022).
16. Alsamadany, H., Al-Zahrani, H., Selim, E. & El-Sherbiny, M. Spatial distribution and potential ecological risk assessment of some trace elements in sediments and grey mangrove (*Avicennia marina*) along the Arabian Gulf coast, Saudi Arabia. *Open Chem.* **18**(1), 77–96. <https://doi.org/10.1515/chem-2020-0010> (2020).
17. Satapathy, S., Panda, C. R. & Jena, B. S. Risk-based prediction of metal toxicity in sediment and impact on human health due to consumption of seafood (*Saccostrea cucullata*) found in two highly industrialised coastal estuarine regions of Eastern India: A food safety issue. *Environ. Geochem. Health* **41**, 1967–1985. <https://doi.org/10.1007/s10653-019-00251-4> (2019).
18. Soliman, N., Thabet, W. M., El-Sadaawy, M. M. & Morsy, F. A. Contamination status and potential risk of metals in marine sediments of Shalateen coast, the Red Sea. *Soil Sediment Contam: Int. J.* **31**(1), 40–56. <https://doi.org/10.1080/15320383.2021.1903832> (2022).
19. Pelicice, F. M. *et al.* Large-scale degradation of the Tocantins-Araguaia River basin. *Environ. Manag.* **68**, 445–452. <https://doi.org/10.1007/s00267-021-01513-7> (2021).
20. Serrão, E.A.O. Impactos das mudanças climáticas e do uso da terra na produção hidrelétrica na Amazônia Oriental. 2022. <http://dspace.sti.ufcg.edu.br:8080/jspui/handle/riufcg/25111>. Accessed 22 Jan 2024.
21. Santos, G.M. Catálogo dos Peixes Comerciais do Baixo Rio Tocantins. Eletronorte. CNPq. INPA (1984).
22. Acioly, T. M. S., Silva, M. F., Barbosa, L. A., Iannacone, J. & Viana, D. C. Levels of potentially toxic and essential elements in water and estimation of human health risks in a river located at the interface of Brazilian Savanna and amazon biomes (Tocantins River). *Toxics* **12**(7), 444. <https://doi.org/10.3390/toxics12070444> (2024).
23. Targa, M. S., Oliveira, F. S. L., Balduino, A. R., Souza Catelani, C. & Castro, M. P. Análise das ações antrópicas na bacia hidrográfica do Riacho Bacuri no município de Imperatriz-MA. *Rev. Técnica Ciências Ambient.* **5**, 1–9 (2021).
24. Nascimento, B. L. M. *et al.* Comportamento e avaliação de metais potencialmente tóxicos (Cu (II), Cr (III), Pb (II) e Fe (III)) em águas superficiais dos Riachos Capivara e Bacuri Imperatriz-MA, Brasil. *Eng. Sanitária e Ambiental* **20**, 369–378. <https://doi.org/10.1590/S1413-41522015020000113620> (2015).
25. Köppen, W. & Geiger, R. *Klimate der Erde* (Verlag Justus Perthes, 1928).
26. USEPA. United States Environmental Protection Agency. 2005. Mercury in Water by Cold Vapor Atomic Fluorescence Spectrometry, Method 245.7. https://www.epa.gov/sites/default/files/2015-08/documents/method_245-7_rev_2_2005.pdf. Accessed 26 Feb 2024.
27. CONAMA. Conselho Nacional do Meio Ambiente. Resolução No 454 - 01 de novembro de 2012. Conselho Nacional do Meio Ambiente. Brasília, DF. 2012. https://www.icmbio.gov.br/cepsul/images/stories/legislacao/Resolucao/2012/res_conama_454_2012_materiaisdragadoemaguasjurisdicionaisbrasileiras.pdf. Accessed 26 Jan 2024.
28. Canadian Council of Ministers of the Environment (CCME). Canadian environmental quality guidelines, Winnipeg. 1999. <https://ccme.ca/en/resources/sediment#>. Accessed 26 Jan 2024.
29. NOAA (National Oceanic and Atmospheric Administration). Screening quick reference tables (SQUIRTs). 1999. <https://response.restoration.noaa.gov/sites/default/files/SQUIRTs.pdf>. Accessed 26 Jan 2024.
30. USEPA. United States Environmental Protection Agency. Guidelines for Sediments. <https://www.epa.gov/caddis-vol2/sediments>. Accessed 26 Jan 2024.
31. Müller, G. Die Schwermetallbelastung der Sedimente des Neckars und Seiner Nebenflüsse. *Chemiker-Zeitung* **6**, 157–164 (1981).
32. Taylor, S. R. & McLennan, S. M. The geochemical evolution of the continental crust. *Rev. Geophys.* **33**, 241–265 (1995).
33. Chakravarty, M. & Patgiri, A. D. Metal pollution assessment in sediments of the Dikrong River, N.E. India. *J. Human Ecol.* **27**(1), 63–67. <https://doi.org/10.1080/09709274.2009.11906193> (2009).
34. Fodoué, Y., Ismaila, A., Yannah, M., Wirmvem, M. J. & Mana, C. B. Heavy metal contamination and ecological risk assessment in soils of the pawara gold mining area, Eastern Cameroon. *Earth* **3**(3), 907–924. <https://doi.org/10.3390/earth3030053> (2022).
35. Hu, B. *et al.* Application of portable XRF and VNIR sensors for rapid assessment of soil heavy metal pollution. *PLoS ONE* **12**, e0172438. <https://doi.org/10.1371/journal.pone.0172438> (2017).
36. Turekian, K. K. & Wedepohl, K. H. Distribution of the elements in some major units of the earth's crust. *Geol. Soc. Am. Bull.* **72**(2), 175–192. [https://doi.org/10.1130/0016-7606\(1961\)72\[175:DOTEIS\]2.0.CO;2](https://doi.org/10.1130/0016-7606(1961)72[175:DOTEIS]2.0.CO;2) (1961).
37. Loring, D. H. Normalization of heavy-metal data from estuarine and coastal sediments. *ICES J. Mar. Sci.* **48**(1), 101–115. <https://doi.org/10.1093/icesjms/48.1.101> (1991).
38. Tomlinson, D. L., Wilson, J. G., Harris, C. R. & Jeffrey, D. W. Problems in assessment of heavy metals in estuaries and the formation of pollution index. *Helgol. Mar. Res.* **33**(1), 1566–1575. <https://doi.org/10.1007/BF02414780> (1980).
39. Rubio, B., Nombela, M. A. & Vilas, F. J. M. P. B. Geochemistry of major and trace elements in sediments of the Ria de Vigo (NW Spain): An assessment of metal pollution. *Mar. Pollut. Bull.* **40**(11), 968–980. [https://doi.org/10.1016/S0025-326X\(00\)00039-4](https://doi.org/10.1016/S0025-326X(00)00039-4) (2000).
40. Hakanson, L. An ecological risk index for aquatic pollution control. A sedimentological approach. *Water Res.* **14**(8), 975–1001. [https://doi.org/10.1016/0043-1354\(80\)90143-8](https://doi.org/10.1016/0043-1354(80)90143-8) (1980).
41. Xu, Z. Q., Ni, S. J., Tuo, X. G. & Zhang, C. J. Calculation of heavy metals' toxicity coefficient in the evaluation of potential ecological risk index. *Environ. Sci. Technol.* **31**(2), 112–115 (2008).
42. Dash, S., Borah, S. S. & Kalamdhad, A. S. Heavy metal pollution and potential ecological risk assessment for surficial sediments of Deepor Beel, India. *Ecol. Indic.* **122**, 107265. <https://doi.org/10.1016/j.ecolind.2020.107265> (2021).
43. Weather Spark. Condições meteorológicas médias de Apucarana. 2024. <https://pt.weatherspark.com/h/m/30244/2023/3/Condi%C3%A7%C3%B5es-meteorol%C3%B3gicas-hist%C3%B3ricas-em-mar%C3%A7o-de-2023-em-Imperatriz-Brasil#Figures-Temperature>. Accessed 28 Feb 2024.
44. Silva, C. P., Silveira, E. L., Santos, A. M. V. & Campos, S. X. Metal cross-contamination relationships between sediments and loricariidae species (siluriform) in a neotropical riverine system. *Environ. Res.* **258**, 119412. <https://doi.org/10.1016/j.envres.2024.119412> (2024).
45. Senze, M., Kowalska-Górska, M., Czyż, K., Wondolowska-Grabowska, A. & Łuczynska, J. Aluminum in bottom sediments of the lower Silesian Rivers supplying dam reservoirs versus selected chemical parameters. *Int. J. Environ. Res. Public Health* **18**(24), 13170. <https://doi.org/10.3390/ijerph182413170> (2021).
46. Aiman, U., Mahmood, A., Waheed, S. & Malik, R. N. Enrichment, geo-accumulation and risk surveillance of toxic metals for different environmental compartments from Mehmood Booti dumping site, Lahore city, Pakistan. *Chemosphere* **144**, 2229–2237. <https://doi.org/10.1016/j.chemosphere.2015.10.077> (2016).
47. USEPA. United States Environmental Protection Agency. Final Aquatic Life Criteria for Aluminum in Freshwater. EPA-822-R-18-001. 2018. <https://www.epa.gov/wqc/2018-final-aquatic-life-criteria-aluminum-freshwater>. Accessed 26 Feb 2024.
48. GC. Priority Substance List Assessment Report for Aluminum Salts (Environment Canada and Health Canada, 2010).
49. British Columbia. Ministry of Water, Land, and Resource Stewardship. Aluminum Water Quality Guidelines: Freshwater Aquatic Life. Water Quality Guideline Series, WQG-09-1. Prov. B.C., Victoria B.C. 2023. <http://www2.gov.bc.ca/gov/content/environment/air-land-water/water/waterquality/water-quality-guidelines/approved-water-quality-guidelines>. Accessed 02 Feb 2024.

50. Yan, J. *et al.* Effects of switching redox conditions on sediment phosphorus immobilization by calcium/aluminum composite capping: Performance, ecological safety and mechanisms. *Chemosphere* **343**, 140294. <https://doi.org/10.1016/j.chemosphere.2023.140294> (2023).
51. Buranatrevedh, S. Health risk assessment of workers exposed to metals from an aluminium production plant. *J. Med. Assoc. Thai.* **93**(12), 136 (2011).
52. Liaquat, L., Batool, Z. & Haider, S. Developmental toxicity of aluminium and other metals: Areas unexplored. *Biochem. Mech. Alum. Induced Neurol. Disord.* <https://doi.org/10.2174/9781681088839121010008> (2021).
53. Ungureanu, E. L., & Mustatea, G. Toxicity of heavy metals. in *Environmental Impact and Remediation of Heavy Metals*. (IntechOpen, 2022). <https://doi.org/10.5772/intechopen.102441>
54. Ohiagu, F. O., Chikezie, P. C., Ahaneku, C. C. & Chikezie, C. M. Human exposure to heavy metals: Toxicity mechanisms and health implications. *Mater. Sci. Eng. Int. J.* **6**(2), 78–87 (2022).
55. Fuentes-Gandara, F., Pinedo-Hernandez, J., Marrugo-Negrete, J. & Diez, S. Human health impacts of exposure to metals through extreme consumption of fish from Colombian Caribbean Sea. *Environ. Geochem. Health* **40**, 229–242. <https://doi.org/10.1007/s10653-016-9896-z> (2016).
56. Balali-Mood, M. *et al.* Toxic mechanisms of five heavy metals: Mercury, lead, chromium, cadmium, and arsenic. *Front. Pharmacol.* <https://doi.org/10.3389/fphar.2021.643972> (2021).
57. Kimijima, S., Sakakibara, M., Pateda, S. M. & Sera, K. Contamination level in geo-accumulation index of river sediments at artisanal and small-scale gold mining area in Gorontalo Province, Indonesia. *Int. J. Environ. Res. Public Health* **19**, 6094. <https://doi.org/10.3390/ijerph19106094> (2022).
58. Senoro, D. B. *et al.* Quantitative assessment and spatial analysis of metals and metalloids in soil using the geo-accumulation index in the capital town of Romblon Province, Philippines. *Toxics* **10**(11), 633. <https://doi.org/10.3390/toxics10110633> (2022).
59. Karkain, R. M., Karkain, S. M. & Sadeq, M. A. Assessment of heavy metal contamination in the sediments and soils in the City of Dubai based on the geoaccumulation index. *Emirati J. Civil Eng. Appl.* **2**(1), 78–84. <https://doi.org/10.54878/fk289z29> (2024).
60. Haris, H. *et al.* Geo-accumulation index and contamination factors of heavy metals (Zn and Pb) in urban river sediment. *Environ. Geochem. Health* **39**, 1259–1271. <https://doi.org/10.1007/s10653-017-9971-0> (2017).
61. Ahirvar, B. P., Das, P., Srivastava, V. & Kumar, M. Perspectives of heavy metal pollution indices for soil, sediment, and water pollution evaluation: An insight. *Total Environ. Res. Themes* **6**, 100039. <https://doi.org/10.1016/j.totert.2023.100039> (2023).
62. Al Hejazi, A., Davis, M., Prete, D., Lu, J. & Wang, S. Heavy metals in indoor settled dusts in Toronto, Canada. *Sci. Total Environ.* **703**, 134895. <https://doi.org/10.1016/j.scitotenv.2019.134895> (2020).
63. Huang, C. C. *et al.* A comprehensive exploration on the health risk quantification assessment of soil potentially toxic elements from different sources around large-scale smelting area. *Environ. Monitor. Assess.* **194**(3), 206. <https://doi.org/10.1007/s10661-022-09804-0> (2022).
64. Kowalska, J. B., Mazurek, R., Gasiorek, M. & Zaleski, T. Pollution indices as useful tools for the comprehensive evaluation of the degree of soil contamination: A review. *Environ. Geochem. Health* <https://doi.org/10.1007/s10653-018-0106-z> (2018).
65. Pejman, A., Bidhendi, G. N., Ardestani, M., Saeedi, M. & Baghvand, A. A new index for assessing heavy metals contamination in sediments: A case study. *Ecol. Indic.* **58**, 365–373. <https://doi.org/10.1016/j.ecolind.2015.06.012> (2015).
66. Yang, T. *et al.* Comprehensive ecological risk assessment for semi-arid basin based on conceptual model of risk response and improved TOPSIS model—a case study of Wei River Basin, China. *Sci. Total Environ.* **719**, 137502. <https://doi.org/10.1016/j.scitotenv.2020.137502> (2020).
67. Badawy, W. *et al.* Characterization of major and trace elements in coastal sediments along the Egyptian Mediterranean Sea. *Mar. Pollut. Bull.* **177**, 113526. <https://doi.org/10.1016/j.marpolbul.2022.113526> (2022).
68. Zhang, J. *et al.* Assessment of heavy metal pollution and water quality characteristics of the reservoir control reaches in the middle Han River, China. *Sci. Total Environ.* **799**, 149472. <https://doi.org/10.1016/j.scitotenv.2021.149472> (2021).
69. Mahler, B. J. *et al.* Biofilms provide new insight into pesticide occurrence in streams and links to aquatic ecological communities. *Environ. Sci. Technol.* **54**(9), 5509–5519. <https://doi.org/10.1021/acs.est.9b07430> (2020).
70. Xiang, M. *et al.* Heavy metal contamination risk assessment and correlation analysis of heavy metal contents in soil and crops. *Environ. Pollut.* **278**, 116911. <https://doi.org/10.1016/j.envpol.2021.116911> (2021).
71. Yüksel, B., Ustaoglu, F., Tokatli, C. & Islam, M. S. Ecotoxicological risk assessment for sediments of Çavuşlu stream in Giresun, Turkey: association between garbage disposal facility and metallic accumulation. *Environ. Sci. Pollut. Res.* **29**(12), 17223–17240. <https://doi.org/10.1007/s11356-021-17023-2> (2022).
72. Geng, M. *et al.* Assessing the impact of water-sediment factors on water quality to guide river-connected lake water environment improvement. *Sci. Total Environ.* **912**, 168866. <https://doi.org/10.1016/j.scitotenv.2023.168866> (2024).
73. Olomukoro, J. O. & Enabulele, C. O. Assessment of heavy metal contamination and sediment characteristics in Ozomu lake, southern Nigeria: Implications for environmental health. *Kuwait J. Sci.* **51**(2), 100192. <https://doi.org/10.1016/j.kjs.2024.100192> (2024).
74. Capparelli, M. V. *et al.* Use of an integrated geochemical and ecotoxicological approach to evaluate sediment metal contamination in three protected estuarine areas along the Coast of São Paulo State, Brazil. *Bull. Environ. Contam. Toxicol.* **106**, 355–362. <https://doi.org/10.1007/s00128-020-03076-5> (2021).
75. Vezzone, M. *et al.* Metal pollution in surface sediments from Rodrigo de Freitas Lagoon (Rio de Janeiro, Brazil): Toxic effects on marine organisms. *Environ. Pollut.* **252**, 270–280. <https://doi.org/10.1016/j.envpol.2019.05.094> (2019).
76. Vitkauskaitė, A. & Salytė, M. Air quality investigation using functional data analysis methods Lietuvos matematikos rinkinys Lietuvos matematikų draugijos darbai. *Ser. B* **64**, 36–53. <https://doi.org/10.15388/LMR.2023.33658> (2023).
77. Wu, H. *et al.* Health risk assessment based on source identification of heavy metals: A case study of Beiyun River, China. *Ecotoxicol. Environm. Saf.* **213**, 112046. <https://doi.org/10.1016/j.ecoenv.2021.112046> (2021).
78. Tariq, A. & Mushtaq, A. Untreated wastewater reasons and causes: A review of most affected areas and cities. *Int. J. Chem. Biochem. Sci.* **23**, 121–143 (2023).
79. Wear, S. L., Acuña, V., McDonald, R. & Font, C. Sewage pollution, declining ecosystem health, and cross-sector collaboration. *Biolog. Conserv.* **255**, 109010. <https://doi.org/10.1016/j.biocon.2021.109010> (2021).
80. Akhtar, N., Syakir Ishak, M. I., Bhawani, S. A. & Umar, K. Various natural and anthropogenic factors responsible for water quality degradation: A review. *Water* **13**(19), 2660. <https://doi.org/10.3390/w13192660> (2021).
81. Lima, J.E.F.W. Carvalho, N.D.O., & Silva, E.M. Diagnóstico do fluxo de sedimentos em suspensão na Bacia Araguaia-Tocantins. Planaltina, DF: Embrapa Cerrados; Brasília, DF: ANEEL: ANA (2004). <https://ainfo.cnptia.embrapa.br/digital/bitstream/CPAC-2010/24058/1/lima-01.pdf>. Accessed 25 June 2024.
82. Swanson, A. C., Kaplan, D., Toh, K. B., Marques, E. E. & Bohlman, S. A. Changes in floodplain hydrology following serial damming of the Tocantins River in the eastern Amazon. *Sci. Total Environ.* **800**, 149494. <https://doi.org/10.1016/j.scitotenv.2021.149494> (2021).
83. Schmitz, M. H. Landscape of the Tocantins-Araguaia Basin: challenges, dynamics, and future prospect for conservation (Doctoral dissertation, Universidade Estadual de Maringá. Departamento de Biologia. Programa de Pós-Graduação em Ecologia de Ambientes Aquáticos Continentais) (2024). <https://aquadocs.org/handle/1834/43029>
84. Polizel, S. P. *et al.* Analysing the dynamics of land use in the context of current conservation policies and land tenure in the Cerrado: MATOPIBA region (Brazil). *Land Use Policy* **109**, 105713. <https://doi.org/10.1016/j.landusepol.2021.105713> (2021).

85. Camargo, F. A. O. *et al.* Brazilian agriculture in perspective: great expectations versus reality. *Adv. Agron.* **141**, 53–114. <https://doi.org/10.1016/bs.agron.2016.10.003> (2017).
86. Escobar, N. *et al.* Spatially-explicit footprints of agricultural commodities: Mapping carbon emissions embodied in Brazil's soy exports. *Glob. Environ. Change* **62**, 102067. <https://doi.org/10.1016/j.gloenvcha.2020.102067> (2020).
87. Melo, M. P. D. *et al.* Effects of local land use on riparian vegetation, water quality, and in situ toxicity. *Rev. Ambient. Água* **17**, e2856. <https://doi.org/10.4136/ambi-agua.2856> (2022).
88. Qasem, N. A., Mohammed, R. H. & Lawal, D. U. Removal of heavy metal ions from wastewater: A comprehensive and critical review. *Npj Clean Water* **4**(1), 1–15. <https://doi.org/10.1038/s41545-021-00127-0> (2021).
89. Saleh, T. A., Mustaqeem, M. & Khaled, M. Water treatment technologies in removing heavy metal ions from wastewater: A review. *Environ. Nanotechnol. Monitor. Manag.* **17**, 100617. <https://doi.org/10.1016/j.enmm.2021.100617> (2022).
90. Dhenkula, S. P., Shende, A. D., Deshpande, L. & Pophali, G. R. An Overview of Heavy metals treatment & management for laboratory waste liquid (LWL). *J. Environ. Chem. Eng.* <https://doi.org/10.1002/cben.201600010> (2024).
91. Singh, V. *et al.* Toxic heavy metal ions contamination in water and their sustainable reduction by eco-friendly methods: Isotherms, thermodynamics and kinetics study. *Sci. Rep.* **14**(1), 7595. <https://doi.org/10.1038/s41598-024-58061-3> (2024).
92. Deng, J. *et al.* Wide riparian zones inhibited trace element loss in mining wastelands by reducing surface runoff and trace elements in sediment. *Toxics* **12**(4), 279. <https://doi.org/10.3390/toxics12040279> (2024).
93. Pandey, S., Kumari, T., Verma, P., Singh, R. & Raghubanshi, A. S. Impact of anthropogenic stresses on riparian ecosystem and their management perspectives. *Ecol. Signif. River Ecosyst.* <https://doi.org/10.1016/B978-0-323-85045-2.00004-2> (2022).
94. Awewomom, J. *et al.* Addressing global environmental pollution using environmental control techniques: A focus on environmental policy and preventive environmental management. *Discov. Environ.* **2**(1), 8. <https://doi.org/10.1007/s44274-024-00033-5> (2024).
95. Guevara-Herrero, I., Bravo-Torija, B. & Pérez-Martín, J. M. Educational practice in education for environmental justice: A systematic review of the literature. *Sustainability* **16**(7), 2805. <https://doi.org/10.3390/su16072805> (2024).
96. Li, S., Sun, H., Sharif, A., Bashir, M. & Bashir, M. F. Economic complexity, natural resource abundance and education: Implications for sustainable development in BRICST economies. *Resour. Policy* **89**, 104572. <https://doi.org/10.1016/j.resourpol.2023.104572> (2024).

Acknowledgements

We express our heartfelt gratitude to CNPq (403097/2022-3), CAPES, and FAPEMA (193956/2022) for their invaluable support, laying the foundation for our scientific research.

Author contributions

T.M.S.A.—conceptualization, methodology, data curation, formal analysis, investigation, validation, writing—original draft, writing—review, and editing; M.F.S.—visualization and methodology; J.I.—methodology, validation, and writing—review; D.C.V.—resources, supervision, project administration, and validation.

Competing interests

The authors declare no competing interests.

Additional information

Supplementary Information The online version contains supplementary material available at <https://doi.org/10.1038/s41598-024-66570-4>.

Correspondence and requests for materials should be addressed to T.M.d. or D.C.V.

Reprints and permissions information is available at www.nature.com/reprints.

Publisher's note Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.



Open Access This article is licensed under a Creative Commons Attribution 4.0 International License, which permits use, sharing, adaptation, distribution and reproduction in any medium or format, as long as you give appropriate credit to the original author(s) and the source, provide a link to the Creative Commons licence, and indicate if changes were made. The images or other third party material in this article are included in the article's Creative Commons licence, unless indicated otherwise in a credit line to the material. If material is not included in the article's Creative Commons licence and your intended use is not permitted by statutory regulation or exceeds the permitted use, you will need to obtain permission directly from the copyright holder. To view a copy of this licence, visit <http://creativecommons.org/licenses/by/4.0/>.

© The Author(s) 2024