Research Article

Photocatalytic degradation of chlorpyrifos and methylene blue using α‑Bi2O3 nanoparticles fabricated by sol–gel method

Subhash Dharmraj Khairnar1 · Vinod Shankar Shrivastava[1](http://orcid.org/0000-0002-2065-1068)

© Springer Nature Switzerland AG 2019

Abstract

Organic dyes and pesticides, i.e. methylene blue and chlorpyrifos are the most frequently used in the various felds in everyday life for the diferent purpose. The presence of organic dyes and pesticides within the aquatic environment causes serious problems to the both types ecosystems. The aim of this study, to evaluate the photocatalytic degradation of methylene blue and chlorpyrifos in aqueous solution using α-Bi₂O₃ nanoparticles (NPs). The α-Bi₂O₃ NPs were successfully synthesized by sol–gel method and characterized by XRD, SEM, TEM, EDX and FTIR techniques. Moreover, the photocatalytic degradation of methylene blue and chlorpyrifos was investigated by optimizing the various parameters like pH, an initial concentration of the solution, irradiation time and photocatalyst dose, etc. The photocatalytic efficiency of α-Bi₂O₃ NPs was measured using a UV–visible double beam spectrophotometer. The results show that the α-Bi₂O₃ NPs have an average crystallite size of 32.47 nm with a highly rough and porous surface area. The optimized condition for the photocatalytic degradation was obtained at $pH=2$ and $pH=5$, the initial concentration of solution 10 mg/L and irradiation time 90 min for methylene blue and chlorpyrifos respectively. Besides this, the kinetics of photocatalytic degradation were investigated and the results indicate that the photocatalytic degradation of methylene blue and chlorpyrifos follows the pseudo-frst order kinetics.

 \boxtimes Vinod Shankar Shrivastava, drvinodshrivastava1962@gmail.com; Subhash Dharmraj Khairnar, subhashkhairnar20@gmail.com | ¹Nano-Chemistry Research Laboratory, G. T. Patil College, Nandurbar, M.S 425412, India.

SN Applied Sciences (2019) 1:762 | https://doi.org/10.1007/s42452-019-0761-4

Received: 12 April 2019 / Accepted: 11 June 2019 / Published online: 21 June 2019

Graphic abstract

Keywords Chlorpyrifos · Organic pesticides · α-Bi₂O₃ · Pseudo first order · Methylene blue

1 Introduction

In India the majority of peoples totally rely on agricultural production because it is the main occupation of most of the Indian villagers. Agriculture is a standout amongst the most essential factors in the Indian economy and contributes 18% to the GDP. Which ensuring the nourishment, security for more than 1.37 billions of Indian populations with the diminishing cultivable land resource is a Herculean task. India is the second largest producer of agricultural products. India accounts for 7.39% of the total global agricultural output. In the process of accomplishing the target pesticides play an important role in Indian farming. Pesticides are the agrochemicals, one of the invaluable inputs in sustaining agricultural production as substantial food production is lost due to insect pests, plant pathogens, weeds, etc. However, since the green revolution (1966) has been begun in India, the application of these chemicals expanded in excess of multiple times and causing tremendous loss to the environment and human health. Internationally big effort is made to safe use of chemicals. In India nearly 65% of the workforce derives a livelihood from agriculture and are therefore exposed to chemical pesticides. The rampant use of pesticides has played havoc with living beings and the environment as these chemicals persists and seep into the environment for

SN Applied Sciences A SPRINGER NATURE journal a long time because of more water solubility, tendency to adsorb to the soil (soil adsorption) and more half-life that tends to persist in the environment.

Chlorpyrifos, O,O-diethyl O-(3,5,6-trichloro-2-pyridyl) phosphorothioate $[C_9H_{11}Cl_3NO_3PS]$ (CPS), is a well-known organophosphorothioate pesticide that is widely used in agricultural and nonagricultural felds [[1\]](#page-8-0). The Indian government has restricted various chlorinated hydrocarbons pesticides such as Aldrin, DDT and Chlordane etc., and the use of CPS has increased as an alternate option [[2\]](#page-8-1). Exposure to CPS and its metabolites has been related to a variety of nerve disorders in humans. CPS shows a wide spectrum of biological activity and is used to control a range of insects, pests as well as soil-dwelling grubs, rootworms, borers and subterranean termites. The contamination has been found up to about 24 km from the site of application. Symptoms of acute poisoning include headache, nausea, muscle twitching and convulsions and in some extreme cases even death. A human birth defects have also been associated with exposure to CPS and its products. It also afects the male reproductive system. CPS is toxic to a variety of benefcial arthropods, including bees, ladybird beetles and parasitic wasps. It kills fsh at concentrations as low as a few parts per trillion $[3-5]$ $[3-5]$ $[3-5]$. Therefore, the degradation of CPS present in the environment has become a public concern.

Fig. 1 3D structures of **a** methylene blue, **b** chlorpyrifos

Nanotechnology is imperative with regards to environmental remediation [\[6](#page-8-4), [7\]](#page-8-5). Photocatalytic degradation has been proved to be a promising method for the treatment of wastewater contaminated with organic and inorganic pollutants. In recent years, advanced oxidation processes (AOPs) have been proposed as innovative water treatment technologies. The prominent feature of these AOPs is based on the in situ generations of active species (i.e., H_2O_2 , HO, O2⁻⁻, and O₃) for the mineralization of refractory organic compounds, water pathogens and disinfection by products [[8](#page-8-6)[–14](#page-8-7)]. Heterogeneous photocatalysis, which accelerating the photoreaction by a catalyst, an attractive and efficient method for the degradation of environmental pollutants or non-biodegradable toxics present in domes-tic sewage, industrial or agricultural effluents [\[15](#page-8-8)]. Among various semiconductors a p-type $Bi₂O₃$ heterogeneous semiconductor was considered as one of the most profcient photocatalysts and important in modern solid state because of its unique structures and physical characteristics like high refractive index, high oxygen-ion conductivity, dielectric permittivity and thermal stability, $Bi₂O₃$ is inert towards neutral water and possess band gap energy in the visible region $($ \sim 2.9 eV) can oxidise water and produce highly reactive species for initiating oxidation reaction for degradation of pesticides and dyes [\[15,](#page-8-8) [16\]](#page-8-9). Pure Bi₂O₃ exists five polymorphisms: α-Bi₂O₃, β-Bi₂O₃, γ-Bi₂O₃, δ-Bi₂O₃, ε-Bi₂O₃ [[17](#page-8-10)]. α-Bi₂O₃ transforms into δ-Bi₂O₃at 729 °C, upon cooling δ-Bi₂O₃ transfer into β-Bi₂O₃ at 650 °C [[17](#page-8-10), [18\]](#page-8-11). Among them, high-temperature δ-phase and the low-temperature α-phase are stable, whereas β -Bi₂O₃, γ-Bi₂O₃ and ε-Bi₂O₃ forms are metastable [[17\]](#page-8-10). Among all phases the band gap of the low-temperature a phase is 2.8 eV and therefore found to be active in the visible region [\[19,](#page-8-12) [20](#page-8-13)].

Methylene blue (MB) is a severe pollutant since it is widely used as a colorant in various industries for the different purpose. It is harmful and carcinogenic to human beings [[21](#page-8-14), [22](#page-8-15)]. Therefore, it is necessary to remove such hazardous dyes and pesticides from natural water bodies. In this study, we frst time report a study on the

photocatalytic degradation of CPS and MB independently using α -Bi₂O₃ NPs fabricated by the sol-gel method.

2 Experimental

2.1 Materials and methods

All the chemicals used in the present study are of A.R.grade. Bismuth nitrate pentahydrated $[Bi(NO₃)₃·5H₂O]$, Citric acid, Polyethylene glycol 600 (PEG600), Sodium hydroxide (NaOH), Methylene blue and nitric acid were purchased from S. D. Fine Chemicals, Mumbai, India and Chlorpyrifos were purchased from local markets and used without any purifcation. The 3D structures of Methylene blue and chlorpyrifos are as shown in Fig. [1](#page-2-0)a, b.

2.2 Preparation of dye and pesticide solutions

The stock solution (100 mg/L) of MB and CPS was prepared using deionized water. The experimental solution of the desired concentration of MB and CPS (10, 15, and 20 mg/L) was prepared by further dilution of the stock solution with deionized water.

2.3 Synthesis of α-Bi₂O₃ nanoparticles

0.1 M of Bi $(NO_3)_3.5H_2O$ was dissolved in 100 ml of nitric acid solution (1:5, HNO₃: H₂O) and mixed with 2 gm of citric acid. In order to prevent agglomeration, a small amount of PEG600 was added as a surfactant. The pH of the solution was adjusting 3 by adding NaOH and the above solution was stirred for 2 h, then a sol formed. The sol solution was heated to 80 °C for 3 h to form a yellowish gel. This gel was decomposed at 140 °C in oven. The gel initially started to swell and flled the beaker producing a foamy precursor. This foam consists of homogeneous fakes of very small particle size.

Fig. 2 a The XRD patterns of α-Bi₂O₃ NPs, **b** particle size distribution from SEM images of α-Bi₂O₃ NPs

2.4 Photocatalytic degradation experiments

The photocatalytic degradation of MB and CPS were carried out in a photocatalytic reactor (Photocatalytic reactor, Lelesil Innovative System, Mumbai) under the UV visible light irradiation with the help of 160 W mercury lamp (wavelength range 200–690 nm). The synthesized α -Bi₂O₃ NPs was added in 50 ml of an aqueous solution of MB and CPS. Then the solution is kept in the photocatalytic reactor with continuous stirring to ensure that the suspensions of the catalyst were uniform over the span of reaction. At regular intervals, adequate amount of sample solution was removed and centrifuged for 10 min to separate α -Bi₂O₃ NPs. After that the supernatant solution was analyzed with the help of UV–visible double beam spectrophotometer at the wavelength of absorbance maximum (for MB λ_{max} = 663 nm and CPS λ_{max} = 308 nm) to obtain the concentration of MB and CPS in solution. The percentage of degradation of dye has been calculated by using Eq. [1](#page-3-0).

$$
\%Degradation = \frac{(C_0 - C_t)}{C_0} \times 100\tag{1}
$$

where $\mathsf{C}_{\textup{o}}$ is the initial concentration and $\mathsf{C}_{\textup{t}}$ is the concentration after time t.

2.5 Characterizations

The synthesized α -Bi₂O₃ NPs was characterized by scanning electron microscopy (SEM Hitachi S-4800 Japan), Transmission electron microscopy (TEM), X-ray difraction (XRD-Bruker D 8 Advance X-ray difractometer Germany), electron dispersive X-ray spectroscopy (EDX-Bruker X Flash 5030), Fourier-transform infrared spectroscopy (Shimadzu FTIR-8400) techniques and the photocatalytic activity was performed by a UV–Vis spectrophotometer (Systronics 2203 India).

3 Results and discussion

3.1 X‑ray difraction analysis

The X-ray diffraction analysis of synthesized α -Bi₂O₃ NPs are as shows in Fig. [2](#page-3-1)a. The diffraction patterns of α-Bi₂O₃ were (121), (202), (212), (113), (041) and (241) refection planes at 28.55, 30.92, 34.55, 38.15, 47.06, 54.77 and 55.64, 2θ position enlightening the monoclinic crystal structure of α -Bi₂O₃ in accordance with JCPDS card No. 00-041-1449, Confirming the phase purity of synthesized α -Bi₂O₃ NPS. The high intensity peak of (121), (113) and (041) peaks were suggested the better growth of α -Bi₂O₃ NPs. The inset of Fig. [2a](#page-3-1) shows the magnifed image of a plane (121) at 28.55 2θ position. Figure [2](#page-3-1)b shows the particle size distribution obtained from SEM images. The results clearly indicated the synthesized α -Bi₂O₃ NPs have a particle size under the 20 nm and the maximum particle has a diameter in the range of 2–4 nm.

Table [1](#page-4-0) shows the lattice parameters and the average crystallite size of α -Bi₂O₃ NPs, which were calculated from FWHM of XRD peaks at 28.55, 30.92, 34.55, 38.15, 47.06, 54.77 and 55.64, 2θ position using Scherer formula given in Eq. [2](#page-3-2) [[23](#page-8-16)].

$$
D = \frac{0.94\lambda}{\beta \text{Cos}\theta} \tag{2}
$$

where D is the average crystalline size, λ is a wavelength in A^0 , β is the FWHM in radian and θ is diffraction angle in degree. The average crystallite size of the α -Bi₂O₃ NPS were 32.47 nm.

3.2 UV–visible absorption spectra

The UV–visible absorption spectra of α -Bi₂O₃ NPs is as shown in Fig. [3.](#page-4-1) The absorption data further utilized to

Fig. 3 UV–visible absorption spectra of α -Bi₂O₃ NPs and inset image is the Tauc's plot

calculate the band gap energy, obtained by plotting the graph between $(ahv)^2$ Vs hv as shown in inset of Fig. [3](#page-4-1). Where h, α and ν represent the Planck constant, absorption

coefficient, and the incident light frequency respectively. Based on the Tauc's plot the direct band gap (E_a) of α -Bi₂O₃ Nps are 2.91 eV which shows the good agreement with previous literature [\[24](#page-9-0)] and indicating the high absorption of α -Bi₂O₃ Nps in the visible as well as in UV region.

3.3 Morphology

To achieve the information regarding the particle morphology, the synthesized α -Bi₂O₃ NPs were examined by scanning electron microscopy (SEM). The SEM results of α-Bi₂O₃NPs revealed a large number of small, interconnected spherical grains with a high rough and porous surface Fig. [4](#page-4-2)a, b. The nanoporous surface of α -Bi₂O₃ photocatalyst enables the electrolyte to permeate into the surface, react with photogenerated holes directly on the particle surface. So, the porous surface of the photocatalyst could potentially promote the charge transfer process [[24,](#page-9-0) [25\]](#page-9-1) and eventually enhance the photocatalytic efficiency. Also, a confrmation of morphology and crystallite size was carried out by Transmission electron microscopy (TEM) analysis. The TEM images of α -Bi₂O₃ NPs shows that

SN Applied Sciences A SPRINGER NATURE journal

Fig. 6 FTIR spectra of α-Bi₂O₃ NPs

the α -Bi₂O₃ NPs were spherical grains with porous morphology and crystallite size in the range of 1–100 nm as shown in Fig. [4](#page-4-2)c, d.

3.4 Energy‑depressive X‑ray spectroscopy (EDS)

The formation of α -Bi₂O₃ NPs is also confirmed by EDX analysis and the results are shown in Fig. [5](#page-5-0). The energy dispersive X-ray spectroscopy of a α -Bi₂O₃ indicate the sharp peaks of Bismuth and oxygen at 10.8 keV and 0.5 keV respectively, whereas the α -Bi₂O₃ consisted of 55% bismuth and 45% oxygen. No other impurity was detected in sample spectrum, which confirms the purity of the α -Bi₂O₃ NPs.

3.5 Functional group analysis (FTIR)

For the determination of functional group present in the synthesized NPS, Fourier-transform infrared spectroscopy (FTIR) analysis was carried out. Figure [6](#page-5-1) shows the FT-IR spectra of α -Bi₂O₃ NPs. The absorption band at 3427 and 1577 cm⁻¹ is due to the –OH stretching and bending vibrations [26]. The band at 1377 cm⁻¹ is the characteristic peak

SN Applied Sciences A SPRINGER NATURE journal of the interlayer nitrate. The absorption peak at 1106 cm^{-1} is due to the –OH group stretching vibration of the H_2O molecule in α -Bi₂O₃ lattice [\[27\]](#page-9-3). In general, metal oxide shows the absorption band below 1000 cm^{-1} therefore, the absorption band at 889 and 665 cm⁻¹ is typical bismuth oxygen bond stretching and bending vibrations.

3.6 Photocatalytic studies

3.6.1 Efect of pH of solution

In photocatalytic degradation, the pH of the sample solution is very important because it influences the degradation efficiency and states about the surface charge properties of the photocatalyst. To study the effect of pH solution was carried out in the range of pH 2–10 at a constant initial concentration of MB and CPS (10 mg/L) with catalyst dose 0.1 gm/L. The zero point of charge (pH_{ZPC}) of α -Bi₂O₃ is 6.3 which indicates that at pH < pH _{7PC} catalyst surface is positively charged while at pH > pH_{ZPC} negatively charged respectively. In the case of MB dye, it shows the maximum degradation at pH 2 although it is a cationic dye. At pH 2 catalyst surface is positively charged, and the redox potential of the photogenerated valence band holes is sufficiently positive to generate hydroxyl radicals. Also, the positive hole plays vital role to trap the MB molecule and leads to the direct oxidation of MB dye and degrade up to 72% in 60 min Fig. [7.](#page-6-0) [\[28](#page-9-4)].

For the CPS, the Pk_a value is 4.6 therefore, at pH the $pH < Pk_a$ value positively charged protonated CPS predominates, while at $pH > Pk_a$ value negatively charged CPS anions predominates. Hence shows the maximum degradation at pH 5 i.e. greater than the Pk_a of CPS. At this pH the positively charged surface of photocatalyst electrostatically attracts the negatively charge CPS anions and leads to the maximum degradation up to 67% in 60 min Fig. [7](#page-6-0). [\[29](#page-9-5)].

Fig. 7 Efect of pH on photocatalytic degradation of MB and CPS, at, conditions: pH=2–10, dye concentration 10 mg/L and catalyst dose 0.1 gm/L for the irradiation time 60 min

3.6.2 Efect of initial concentration

The infuence of initial concentration of MB and CPS on degradation efficiency of α -Bi₂O₃ NPs were studied by varying the initial concentration of MB and CPS from 10 to 20 mg/L at optimized pH of both and the catalyst doses 0.1 gm/L for 60 min. The results are as shown in Fig. [8](#page-6-1)a, b, the photocatalytic degradation efficiency of α -Bi₂O₃ NPS are decreasing with an increase in the initial concentration of MB and CPS solution from 10 mg/L to 20 mg/L. The photocatalytic degradation is completely depends on the generation of hydroxyl radicals and positive holes. But as the initial concentration of MB and CPS increases, the generated of hydroxyl radicals and positive holes will be reduced. Hence the degradation efficiency of α -Bi₂O₃ NPs is decreasing as the initial concentration of MB and CPS increases [[30](#page-9-6)].

3.6.3 Efect of irradiation time

The effect of irradiation time on the photocatalytic efficiency of the catalyst was examined by keeping the other parameter constant. The results are as shown in Fig. [9.](#page-6-2) It

Fig. 8 Efect of initial dye concentration on the photocatalytic degradation **a** MB, conditions pH=2, dye concentration 10, 15, 20 mg/L and catalyst dose 0.1 gm/L for irradiation time 60 min. **b** CPS, conditions: pH=5, CPS concentration 10, 15, 20 mg/L and catalyst dose 0.1 gm/L for irradiation time 60 min

Fig. 9 Efect of irradiation time on photocatalytic degradation of MB, conditions pH=10, dye concentration 10 mg/L and catalyst dose 0.1 gm/L irradiation time 120 min. and for CPS conditions: pH=5 CPS concentration 10 mg/L and catalyst dose 0.1 gm/L irradiation time 120 min

indicates that the photocatalytic efficiency of the catalyst increase with irradiation time up to 90 min. As seen in Fig. [8](#page-6-1). When the irradiation time was longer than 90 min, the photocatalytic efficiency is not increased. Because large numbers of small molecules are generated by the photocatalytic degradation process as the irradiation time too long and these small molecules were adsorbed on the catalyst surface, results in the decreased in the formation of hydroxyl radical that degrade the MB and CPS. Hence the increase in the irradiation time does not lead to the maximum degradation [[31\]](#page-9-7).

3.6.4 Efect of catalyst dose

The effect of photocatalyst dose was studied at the optimized condition of MB and a CPS by varying the catalyst dose from 0.2 to 1.0 gm/L. The results are as shown in Fig. [10](#page-7-0). It shows that the photocatalytic efficiency were increased with catalyst dose for both MB and CPS, this is

Fig. 10 Efect of catalyst doset on the photocatalytic degradation of MB, conditions pH=2, dye concentration 10 mg/L and catalyst dose 0.2–1 gm/L irradiation time 90 min. CPS, conditions: pH=5 CPS concentration 10 mg/L and catalyst dose 0.2–1 gm/L irradiation time 90 min

Scheme 1 Mechanism of photocatalytic degradation of MB and CPS

 $a_{1.2}$

 1.0

 0.8

 0.6

 0.4

n Co/C

because of as we know that the photocatalytic efficiency completely based on the hydroxyl radical and positive hole generation. Therefore, as the catalyst dose increases the surface area also increases, which leads to the greater number of hydroxyl radical and positive holes were generated. Hence the photocatalytic efficiency, enhanced as the catalyst dose increased [[32](#page-9-8)]. Based on the above study the possible mechanism for the photocatalytic degradation of MB and CPS can be explained by the Scheme [1.](#page-7-1)

3.6.5 Kinetics study

3.6.5.1 Pseudo frst order kinetics To study the kinetics of photocatalytic degradation of the MB and CPS in aqueous suspension, the experiment was performed at the optimized condition of both MB and CPS. The photocatalytic degradation of both follows the pseudo frst order kinetics Fig. [11](#page-7-2)a, b. The pseudo frst order kinetics can be repre-sented as shown in Eq. [3](#page-7-3). [\[33\]](#page-9-9).

$$
\ln \frac{C_0}{C_t} = kt \tag{3}
$$

where k is rate constant, C_0 is the initial concentration and C_t is the concentration at time t. From Fig. [11](#page-7-2)a, b, it is clearly seen that the $\ln C_0/C_t$ is directly proportional to the irradiation time. The rate constant k was determined from the slope of the straight line. The rate constant for MB is 0.01842 and 0.02993 min⁻¹ respectively, and having a good correlation with linear regression coefficient (R^2) .

4 Conclusion

In the present study, α -Bi₂O₃ NPs were successfully synthesized by sol–gel method and characterized by XRD, SEM, TEM, EDX and FTIR techniques. The results of XRD, SEM and TEM confirmed that the α -Bi₂O₃ NPs are present in

Fig. 11 Pseudo frst order kinetics for photocatalytic degradation of **a** MB, conditions pH=2, dye concentration 10 mg/L and catalyst dose 1 gm/L irradiation time 90 min **b** CPS, conditions: pH=5 CPS concentration 10 mg/L and catalyst dose 1 gm/L irradiation time 90 min

SN Applied Sciences A SPRINGER NATURE journal

monoclinic phase with crystalline size below the 1–100 nm and having spherical grains with porous surface morphology. The synthesized α -Bi₂O₃ NPs were utilized as a photocatalyst for the degradation of MB and CPS. To the best of author knowledge, this is the frst time report on the photocatalytic degradation of dye and pesticide. The photocatalytic degradation efficiency of photocatalyst increases with the increase in the irradiation time up to 90 min after that it attains the equilibrium. The maximum degradation was obtained with photocatalyst dose 1 mg/L at pH-2 for MB and pH-5 for CPS with 10 mg/L initial concentration of both MB and CPS solution respectively. The photocatalytic degradation of MB and CPS follows the pseudo frst kinetics with rate constant is 0.01842 and 0.02993 min−1. Thus the synthesized α -Bi₂O₃ NPs shows the excellent photocatalytic efficiency against the MB dye as compared to the CPS pesticide under UV–visible irradiation.

Acknowledgements The authors gratefully acknowledge to Kaviyitri Bahinabai Chaudhari North Maharashtra University, Jalgaon for providing fnancial support under the scheme of Shree G. H. Raisoni fellowship for doctoral research and XRD. SEM, EDS and FTIR analysis. Authors are thankful to DST-FIST providing the necessary Grant for instruments. Also, authors are thankful to UGC-DAE-Consortium for Scientifc Research, Indore for the TEM analysis.

Compliance with ethical standards

Conflict of interest On behalf of all authors, the corresponding author states that there is no confict of interest.

References

- 1. Kamel A, Byrne C, Vigo C, Ferrario J, Stafford C, Verdin G, Siegelman F, Knizner S, Hetrick J (2009) Oxidation of selected organophosphate pesticides during chlorination of simulated drinking water. Water Res 43(2):522–534. [https://doi.](https://doi.org/10.1016/j.watres.2008.10.038) [org/10.1016/j.watres.2008.10.038](https://doi.org/10.1016/j.watres.2008.10.038)
- 2. Bootharaju MS, Pradeep T (2012) Understanding the degradation pathway of the pesticide, chlorpyrifos by noble metal nanoparticles. Langmuir 28(5):2671–2679. [https://doi.](https://doi.org/10.1021/la2050515) [org/10.1021/la2050515](https://doi.org/10.1021/la2050515)
- 3. Racke KD (1993) Environmental fate of chlorpyrifos. In: Ware GW (ed) Reviews of environmental contamination and toxicology. Springer, New York, pp 1–150
- 4. Devi LG, Murthy BN, Kumar SG (2009) Photocatalytic activity of V^{5+} , Mo⁶⁺ and Th⁴⁺ doped polycrystalline TiO₂ for the degradation of chlorpyrifos under UV/solar light. J Mol Catal A Chem 308(1):174–181. [https://doi.org/10.1016/j.molca](https://doi.org/10.1016/j.molcata.2009.04.007) [ta.2009.04.007](https://doi.org/10.1016/j.molcata.2009.04.007)
- 5. Hossain MS, Fakhruddin ANM, Chowdhury MAZ, Alam MK (2013) Degradation of chlorpyrifos, an organophosphorus insecticide in aqueous solution with gamma irradiation and natural sunlight. J Environ Chem Eng 1(3):270–274. [https://doi.](https://doi.org/10.1016/j.jece.2013.05.006) [org/10.1016/j.jece.2013.05.006](https://doi.org/10.1016/j.jece.2013.05.006)
- 6. Warner CL, Addleman RS, Cinson AD, Droubay TC, Engelhard MH, Nash MA, Yantasee W, Warner MG (2010) High-performance, superparamagnetic, nanoparticle-based heavy metal sorbents

for removal of contaminants from natural waters. Chem Sus Chem 3(6):749–757. <https://doi.org/10.1002/cssc.201000027>

- 7. Campelo JM, Luna D, Luque R, Marinas JM, Romero AA (2009) Sustainable preparation of supported metal nanoparticles and their applications in catalysis. Chem Sus Chem 2(1):18-45. [https](https://doi.org/10.1002/cssc.200800227) [://doi.org/10.1002/cssc.200800227](https://doi.org/10.1002/cssc.200800227)
- 8. Chong MN, Jin B, Chow CWK, Saint C (2010) Recent developments in photocatalytic water treatment technology: a review. Water Res 44(10):2997–3027. [https://doi.org/10.1016/j.watre](https://doi.org/10.1016/j.watres.2010.02.039) [s.2010.02.039](https://doi.org/10.1016/j.watres.2010.02.039)
- 9. Gomez S, Marchena CL, Pizzio L, Pierella L (2013) Preparation and characterization of $TiO₂/HZSM-11$ zeolite for photodegradation of dichlorvos in aqueous solution. J Hazard Mater 258– 259:19–26. <https://doi.org/10.1016/j.jhazmat.2013.04.030>
- 10. Gaya UI, Abdullah AH (2008) Heterogeneous photocatalytic degradation of organic contaminants over titanium dioxide: a review of fundamentals, progress and problems. J Photochem Photobiol C 9(1):1–12. [https://doi.org/10.1016/j.jphotochem](https://doi.org/10.1016/j.jphotochemrev.2007.12.003) [rev.2007.12.003](https://doi.org/10.1016/j.jphotochemrev.2007.12.003)
- 11. Shifu C, Gengyu C (2005) Photocatalytic degradation of organophosphorus pesticides using floating photocatalyst TiO₂·SiO₂/beads by sunlight. Sol Energy 79(1):1-9. [https://doi.](https://doi.org/10.1016/j.solener.2004.10.006) [org/10.1016/j.solener.2004.10.006](https://doi.org/10.1016/j.solener.2004.10.006)
- 12. Wei L, Shifu C, Wei Z, Sujuan Z (2009) Titanium dioxide mediated photocatalytic degradation of methamidophos in aqueous phase. J Hazard Mater 164(1):154–160. [https://doi.org/10.1016/j.](https://doi.org/10.1016/j.jhazmat.2008.07.140) [jhazmat.2008.07.140](https://doi.org/10.1016/j.jhazmat.2008.07.140)
- 13. Avasarala BK, Tirukkovalluri SR, Bojja S (2011) Photocatalytic degradation of monocrotophos pesticide—an endocrine disruptor by magnesium doped titania. J Hazard Mater 186(2):1234–1240. <https://doi.org/10.1016/j.jhazmat.2010.11.132>
- 14. Chen S, Liu Y (2007) Study on the photocatalytic degradation of glyphosate by $TiO₂$ photocatalyst. Chemosphere 67(5):1010-1017.<https://doi.org/10.1016/j.chemosphere.2006.10.054>
- 15. Leontie L, Caraman M, Alexe M, Harnagea C (2002) Structural and optical characteristics of bismuth oxide thin flms. Surf Sci 507–510:480–485. [https://doi.org/10.1016/S0039](https://doi.org/10.1016/S0039-6028(02)01289-X) [-6028\(02\)01289-X](https://doi.org/10.1016/S0039-6028(02)01289-X)
- 16. Fan HT, Teng XM, Pan SS, Ye C, Li GH, Zhang LD (2005) Optical properties of δ -Bi₂O₃ thin films grown by reactive sputtering. Appl Phys Lett 87(23):231916. [https://doi.org/10.1063/1.21363](https://doi.org/10.1063/1.2136351) [51](https://doi.org/10.1063/1.2136351)
- 17. Deng H-Y, Hao W-C, Xu H-Z (2011) A transition phase in the transformation from α-, β- and ∊- to δ-bismuth oxide. Chin Phys Lett 28(5):056101. <https://doi.org/10.1088/0256-307x/28/5/056101>
- 18. Salazar-Pérez A, Camacho-López M, Morales-Luckie R, Sánchez-Mendieta V, Ureña-Núñez F, Arenas-Alatorre J (2005) Structural evolution of Bi_2O_3 prepared by thermal oxidation of bismuth nano-particles. Superf Vacío 18(3):4–8
- 19. Fan H, Pan S, Teng X, Ye C, Li G, Zhang L (2006) δ-Bi₂O₃ thin films prepared by reactive sputtering: fabrication and characterization. Thin Solid Films 513(1–2):142–147
- 20. Bian Z, Zhu J, Wang S, Cao Y, Qian X, Li H (2008) Self-assembly of active Bi_2O_3/TiO_2 visible photocatalyst with ordered mesoporous structure and highly crystallized anatase. J Phys Chem C 112(16):6258–6262
- 21. Nezamzadeh-Ejhieh A, Hushmandrad S (2010) Solar photodecolorization of methylene blue by CuO/X zeolite as a heterogeneous catalyst. Appl Catal A 388(1):149–159. [https://doi.](https://doi.org/10.1016/j.apcata.2010.08.042) [org/10.1016/j.apcata.2010.08.042](https://doi.org/10.1016/j.apcata.2010.08.042)
- 22. Buthiyappan A, Aziz ARA, Daud WMAW (2016) Recent advances and prospects of catalytic advanced oxidation process in treating textile effluents. Rev Chem Eng 32(1):1-47
- 23. Khairnar SD, Patil MR, Shrivastava VS (2018) Hydrothermally synthesized nanocrystalline Nb_2O_5 and its visible light

photocatalytic activity for the degradation of Congo-red and methylene blue. Iran J Catal 8(2):143–150

- 24. Yang X, Lian X, Liu S, Wang G, Jiang C, Tian J, Chen J, Wang R (2012) Enhanced photocatalytic performance: a $β$ -Bi₂O₃ thin film by nanoporous surface. J Phys D Appl Phys 46(3):035103
- 25. Aroutiounian V, Arakelyan V, Shahnazaryan G (2005) Metal oxide photoelectrodes for hydrogen generation using solar radiationdriven water splitting. Sol Energy 78(5):581–592
- 26. Viruthagiri G, Kannan P, Shanmugam N (2018) Photocatalytic rendition of Zn^{2+} -doped Bi_2O_3 nanoparticles. Photon Nanostruct 32:35–41
- 27. Jha RK, Pasricha R, Ravi V (2005) Synthesis of bismuth oxide nanoparticles using bismuth nitrate and urea. Ceram Int 31(3):495– 497. <https://doi.org/10.1016/j.ceramint.2004.06.013>
- 28. Amalraj A, Pius A (2015) Photocatalytic degradation of monocrotophos and chlorpyrifos in aqueous solution using $TiO₂$ under UV radiation. J Water Process Eng 7:94–101. [https://doi.](https://doi.org/10.1016/j.jwpe.2015.06.002) [org/10.1016/j.jwpe.2015.06.002](https://doi.org/10.1016/j.jwpe.2015.06.002)
- 29. Verma A, Dixit PD (2012) Photocatalytic degradability of insecticide Chlorpyrifos over UV irradiated titanium dioxide in aqueous phase. Int J Environ Sci 3:743–755
- 30. Shankar MV, Cheralathan KK, Arabindoo B, Palanichamy M, Murugesan V (2004) Enhanced photocatalytic activity for the destruction of monocrotophos pesticide by TiO₂/Hβ. J Mol Catal A Chem 223(1):195–200. [https://doi.org/10.1016/j.molca](https://doi.org/10.1016/j.molcata.2004.03.059) [ta.2004.03.059](https://doi.org/10.1016/j.molcata.2004.03.059)
- 31. Wei L, Shifu C, Wei Z, Sujuan Z (2009) Titanium dioxide mediated photocatalytic degradation of methamidophos in aqueous phase. J Hazard Mater 164(1):154–160. [https://doi.org/10.1016/j.](https://doi.org/10.1016/j.jhazmat.2008.07.140) [jhazmat.2008.07.140](https://doi.org/10.1016/j.jhazmat.2008.07.140)
- 32. Muhamad SG (2010) Kinetic studies of catalytic photodegradation of chlorpyrifos insecticide in various natural waters. Arab J Chem 3(2):127–133. [https://doi.org/10.1016/j.arabj](https://doi.org/10.1016/j.arabjc.2010.02.009) [c.2010.02.009](https://doi.org/10.1016/j.arabjc.2010.02.009)
- 33. Patil MR, Khairnar SD, Shrivastava VS (2016) Synthesis, characterisation of polyaniline–Fe₃O₄ magnetic nanocomposite and its application for removal of an acid violet 19 dye. Appl Nanosci 6(4):495–502. <https://doi.org/10.1007/s13204-015-0465-z>

Publisher's Note Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.