ORIGINAL ARTICLE

Zinc doping of different nanoparticles of TiO₂ Sachtopore for improved **elimination of the methyl orange by photocatalysis**

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Received: 10 May 2022 / Accepted: 31 July 2022 / Published online: 12 August 2022 © Qatar University and Springer Nature Switzerland AG 2022

Abstract

Heterogeneous photocatalysis is an increasingly popular method used to decontaminate polluted water. Dyes have become one of the major sources of water pollution and methyl orange is one of the most commonly used dyes in the textile industries. The objective of this study was to remove MO and compare the efficiency of the synthesized photocatalysts. In the current work, photodegradation of methyl orange dye, selected as a pollutant model is carried out at room temperature in aqueous solution of Zn -doped TiO₂. The photocatalysts tested in this study were prepared by coprecipitation method. The catalysts were characterized by XRD, TEM, SEM, EDS, and DRS, before proceeding to photodegradation. X-ray difraction showed that samples are only composed of $TiO₂$ anatase, except $TiO₂$ P25, which contains some crystallites in the rutile form. According to TEM, all nanoparticles of the diferent photocatalysts have a spherical, clearly dispersed and fne morphology. A strong agglomeration is observed for TiO₂ P25. Energy dispersive X-ray spectroscopy confirmed Zn doping by the presence of peaks attributed to zinc. The UV–Vis difuse refectance spectroscopy (UV–Vis DRS) shows a shift of absorption from the UV region to the visible light region and shows that all nanoparticles have very similar absorption thresholds, thus the doped catalysts are promising materials for the use of visible light. The obtained results during the study of the adsorption showed that the order of adsorbability varies in the following decreasing direction: TiO₂ Sachtopore (ZnCl₂/ $TiO_2 = 0.2\%$ > TiO₂ Sachtopore (ZnCl₂/TiO₂= 0.1%) > pure TiO₂ Sachtopore > TiO₂ P25. The photodegradation kinetics was studied as a function of several parameters. A comparison of the degradation kinetics of methyl orange has shown that TiO₂ P25 provides comparable performances to doped TiO₂ Sachtopore at 0.2% of Zn. 1 g/L is the optimum concentration for all catalysts used. The synthesized nanomaterials showed good photodegradation ability in MO decomposition, as well as the results reveal that photocatalysis is a very suitable technique for the purifcation of water containing methyl orange.

Highlights

- Compared to other photocatalysts, doped $TiO₂$ Sachtopore (ZnCl₂/TiO₂ = 0.2%) has a higher adsorption capacity.
- **1 g/L is the optimum concentration for all samples used during this study.**
- **0.2% of zinc reveals signifcant results than other percentages.**

Keywords Adsorption · Photodegradation · Methyl orange · $ZnCl₂/TiO₂$ · Water treatment

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1 Introduction

Dyes are organic compounds used in many industries: stationery, cosmetics, food, textiles, pharmaceuticals, and in medical diagnostics $[1-3]$ $[1-3]$. Water contamination could be due to the color effluents coming from different industries, which is a current problem all over the world [\[4\]](#page-11-2). Their removal/pollutant remediation is a major problem because they are highly toxic and non-biodegradable [\[5](#page-11-3), [6\]](#page-11-4). Advanced Oxidation Processes (POA) for the degradation of organic recalcitrant has been

rapidly developed in the recent years. In addition, the POA driven by sunlight are renewable and ecological in restoration of the environment [\[7](#page-11-5)]. Among these POA, photocatalysis has gradually emerged as a promising alternative solution for the removal of soluble organic compounds and it has attracted great interest [\[8](#page-11-6), [9\]](#page-11-7). It is a fast growing green technology used for the efficient removal of wastewater effluents and the production of ideal hydrogen from water and sunlight [[10](#page-11-8)]. Thus, it offers a great advantage over the other treatment techniques used because it allows the transformation, minimization, and deactivation of persistent compounds in water, as well as the mineralization of all pollutants (bacteria, dyes, organic soils, air pollution …) while respecting the integrity of the environment and human health [\[11](#page-12-0)]. It has extended its application in water treatment due to its potential characteristics such as: feasible operational pressure and temperature, economic reliability and complete degradation of pollutants including parent and intermediate compounds without generating secondary pollutants [[12](#page-12-1)]. Heterogeneous photocatalysis, in the presence of $TiO₂$, is one of the effective photoactive semiconductors that gave better results used in wastewater photodegradation than in the gas phase and also by its potential for application at treatment levels polluted water [\[8](#page-11-6)]. The mechanism of this technique is widely studied [[13](#page-12-2)[–17\]](#page-12-3), based on the irradiation of the semiconductor by photons of energy greater than or equal to that of its band gap. This absorption results in an activation of the semiconductor. The latter results in the passage of an electron from the valence band to the conduction band, giving rise to an electron/hole pair. These two charge carriers can either recombine by releasing heat or separate to react with species adsorbed on the surface of the semiconductor. Holes in the valence band can oxidize donor molecules (reducing) and react with water molecules to give hydroxyl radicals and electrons can reduce electron acceptors (oxidizing) and react with oxygen species to give superoxide radicals. The reactive oxygen species produced (ROS) interact with contaminants and break them down into less harmful materials [[18\]](#page-12-4). The photocatalytic mechanism is explained by the following equations (Eqs. (1) (1) – (9) (9) (9)).

$$
TiO_2 + ZnCl_2 + hv \to e^-{}_{(CB)} + h^+{}_{(VB)} \tag{1}
$$

$$
O_2 + e^-_{(CB)} \rightarrow O_2 \tag{2}
$$

$$
H_2O + h^+_{(VB)} \rightarrow OH^{\bullet} + H^+ \tag{3}
$$

$$
H_2O_2 + e^-{}_{(CB)} \rightarrow (OH)^{\bullet} + OH^-
$$
 (4)

$$
OH^- + h^+_{(VB)} \to OH^* \tag{5}
$$

$$
OH^{\bullet} + e^{-}(CB) \rightarrow OH^{-} \tag{6}
$$

$$
MO + OH^{\bullet} \rightarrow MO + H_2O \rightarrow MO degradation products + CO_2 + H_2O
$$
 (7)

(8) $MO + h^{+}_{V\)({VB})} \rightarrow MO^{+} \rightarrow MO degradation products + CO_{2} + H_{2}O$

(9) $MO + e^{-}$ _(CB) → MO^{-} → MOdegradationproducts + CO₂ + H₂O

The use of this technique results in the complete mineralization of these pollutants at room temperature and under atmospheric pressure, while maintaining the integrity of the environment and human health [[19,](#page-12-5) [20](#page-12-6)].

The use of titanium dioxide for advanced wastewater treatment and water purifcation processes has attracted growing attention since 1972 [[21](#page-12-7), [22\]](#page-12-8). To extend the photoresponse of $TiO₂$ to the visible region, many modification methods, such as metal ion doping [[23\]](#page-12-9), non-metal doping [[24\]](#page-12-10), noble metal deposition [[25](#page-12-11), [26\]](#page-12-12), composite semiconductors [[27](#page-12-13), [28\]](#page-12-14), surface dye sensitization [[29](#page-12-15)], and photosensitive material modifcations [\[30\]](#page-12-16), heterojunction with other semiconductors, and copolymerization have been reported. Among these strategies, doping is a suitable and efective technique to modulate the band gap, reduce the high interlayer resistance, diferentiate the reaction sites for reduction and oxidation reactions, and improve the photocatalytic activity of $TiO₂$ for water purification. Doping has been a "rising star" technique to overcome the limitations ofered by pristine photocatalysts [[31\]](#page-12-17).

Furthermore, elemental doping is considered a versatile strategy and capable of tailoring the band gap, promoting a targeted application via incorporation of cations, anions or codoping. Due to doping, the absorbed light can be easily converted into charge carriers and ultimately into reactive species [\[31](#page-12-17)].

The aim of this work is to study the adsorption and photodegradation of methyl orange dye in an aqueous medium in the presence of pure and doped $TiO₂$ Sachtopore nanoparticles, using different percentages of zinc $(ZnCl₂)$ synthesized by the coprecipitation method, and $TiO₂$ Degussa P25 was used as reference, because of its high purity and wide use in industrial applications and in scientifc research [[32](#page-12-18)].

2 Materials and methods

2.1 Reactants

Methyl orange is an anionic dye and classifed among the azo dyes which are the most frequently used by the textile industries, because of their high dyeing power on several fbers on the one hand and their thermodynamic stability

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on the other (Fig. [1\)](#page-2-0). However, the choice of MO dye is their hydrophobic property and the solubility in effluents, which makes methyl orange one of the major pollutants of liquid discharges of textiles dyeing [\[33\]](#page-12-19). In addition, this dye, which has been widely used to simulate organic pollutants in wastewater in laboratories [[34–](#page-12-20)[36\]](#page-12-21), has been chosen as a pollutant in our work.

The photocatalytic degradation was carried out using TiO₂ photocatalysts for comparison: TiO₂ P25, TiO₂ Sachtopore, TiO₂ Sachtopore (ZnCl₂/TiO₂=0.1%), TiO₂ Sachtopore $(ZnCl₂/TiO₂=0.2%).$

2.2 Synthesis of Zn-doped TiO₂

 $TiO₂$ Sachtopore was purchased without further purification. This source was evaluated without further processing. The co-precipitation process was used to prepare powders of different weight percent $ZnCl₂/TiO₂$ ratio = 0.1% and 0.2% based on $TiO₂$ Sachtopore semiconductor nanoparticles. $TiO₂$ and $ZnCl₂$ were weighed in an appropriate amount and dissolved in a solution of hydrochloric acid (HCl). The mixture was put into a sonicator for 2 h to ensure dissolution. To adjust the pH, NaOH was added to the solution. NaOH acts as a co-precipitant. Afterwards, the solution was fltered and the precipitate was obtained, washed and dried at 50 °C in an oven. At the end, grinding photocatalysts synthesized with a mortar and then calcined at 500° C.

2.3 Photoreactor and light source

The photocatalytic reactor consists of a pyrex beaker, of a capacity equal to 250 mL. The beaker containing the aqueous solution of the dye and the semiconductor was placed in the reactor away from daylight. UV irradiations are provided from four lamps of 15 W.

2.4 Procedure

In the last few decades, several photocatalysts have been developed to address the challenges of water pollution, such

as $TiO₂$, $CuS₂$, ZnO , and $MoS₂$. Among these photocatalysts, $TiO₂$ is of particular interest because of its crystallographic, electronic and optical properties [\[37](#page-12-22)] intro as well as it is a material that simultaneously combines adsorption and photocatalysis. TiO₂ plays the role of adsorbent in the first place, by fxing on its surface ions or molecules (gases, metals, organic molecules…) in a more or less reversible way. There is a transfer of matter from the aqueous or gaseous phase to the solid surface. Once the $TiO₂$ is saturated, it starts to play the role of a photocatalyst after irradiation, where it degrades the pollutants. This represents a practical and low cost advantage.

A volume of known aqueous methyl orange solution (5 mg/L or 10 mg/L) was introduced into the beaker and then into the photoreactor at $pH = 5.8$ in the presence of (0.2 g/L or 1 g/L) TiO₂ semiconductor (as a suspended powder). Then, the solution was stirred in the dark for 40 min (until equilibrium adsorption of methyl orange on the catalyst surface was reached). Adsorption is a very important step for our experiment. Then, the lamps were turned on and photocatalytic degradation began for 120 min. Regular samples every 10 min were taken, then fltered through the micro-flter (Millipore type) (0.2 μm porosity) to remove any catalyst particles in suspension, and fnally the samples are analyzed through a UV–visible spectrophotometer, which allows direct measurement of the optical densities corresponding to the maximum absorption wavelength of the dye (464 nm).

2.5 Characterization

The micrographs of the prepared powders were obtained by transmission electron microscopy (TEM) using Tecnai G2 120 kV and the Bruker AXS D-8 difractometer was used to identify the morphology of $TiO₂$ samples. The Lambda 950 UV–vis-NIR spectrophotometer (Perkin-Elmer) allowed us to record the transmission spectra of $TiO₂$ suspensions. For the morphological analysis of TiO₂ powders, it was performed by scanning electron microscopy (SEM) JEOL JFC-2300HR JEOL JSM-IT 100 with an accelerating voltage of 20 kV. The samples were fltered through a 0.2 μm Millipore membrane. Absorbance measurements were performed using the Mapada 1600 UV/Visible spectrophotometer. The maximum absorption wavelength of methyl orange (λmax) is 464 nm [\[38](#page-12-23)].

3 Results and discussion

3.1 Analysis of nanoparticles with transmission electron microscopy (TEM)

Information such as morphology, size distribution, and the chemical contrast in the nanoparticles could be deduced **Fig. 1** The orange methyl structure from micrographs of electron microscopy [[39](#page-12-24)]. Figure [2](#page-3-0)

Fig. 2 TEM pictures of samples **A** TiO₂ P25, **B** pure TiO₂ Sachtopore, **C** doped TiO₂ Sachtopore (ZnCl₂/TiO₂=0.1%), **D** doped TiO₂ Sachtopore ($ZnCl₂/TiO₂=0.2%$)

illustrates the micrograph of TiO₂ P25 (A), which shows a mixture of dark and gray nanoparticles with a uniform distribution with a spherical morphology and a heavy agglomeration. Pure $TiO₂$ Sachtopore (B) shows clearly regular nanoparticles and spherical dark. The micrographs of doped TiO₂ with Zn (C) and (D) show nanoparticles that are fine, dispersed, and of diferent sizes.

3.2 X‑ray difraction nanoparticles characterization (XRD)

The crystalline structure of samples was studied using X-ray difraction. Figure [3](#page-4-0) shows difractograms of pure and doped $TiO₂$ powders. By comparing the diffraction peaks of the samples to $TiO₂$ diffraction references in the anatase and/or rutile phase, we found that the samples are composed only by crystallites of TiO₂. The crystalline anatase structure is known for its activity in the oxidation of organic pollutants [\[40](#page-12-25)], probably due to the large production of hydroxyl radicals of catalysts comparing to rutile [[41\]](#page-12-26). In the case of $TiO₂$ P25, the crystallites of $TiO₂$ are in anatase and rutile phases. However, in the case of pure and doped $TiO₂$ Sachtopore with zinc, the samples showed exclusively the anatase phase. It has been observed that the intensity of the difraction peaks of the anatase phase gradually decreased with the increasing of Zn content. This could be attributed to the deformation of the crystal lattice. One can also observe for all the samples: the presence of an intense peak can be attributed to crystallites of $TiO₂$ anatase. Low intensity peaks could be found in the internal structure of $TiO₂$ or just on the surface. In addition, the maximum intensities become wider; this is probably due to the formation of small TiO₂ crystallites $[42]$ $[42]$ $[42]$.

The absence of peaks characteristic of zinc oxide phases may indicate the high metal dispersion or be due to the low concentration of Zn doping. This could be attributed to the fine dispersion of zinc ions in small amounts in most $TiO₂$ nanopowders. We found that there is no appearance of crystallinity difference between doped and undoped $TiO₂$. In addition, no signifcant characteristic zinc peak was found in the doped $TiO₂$, also due to the lower amount of zinc.

All Sachtopore samples show similar patterns well attributed to the anatase crystal phase of $TiO₂$ and there is no other phase in the Sachtopore samples. In each pattern, the six peaks located at angles 2θ: 25.18, 37.84, 48.37, 53.91, 55.02, and 62.65 were observed. These peaks correspond to the (101), (004), (200), (105), (211), and (204) difraction planes of the anatase phase. The index was made using (JCPCDS card 21–1272).

(D)

1000 n

Fig. 3 Diffractograms of TiO₂ samples

Table 1 The crystallite size values calculated on the basis of the Debye–Scherrer equation

From the experiments carried out in XRD, we calculated the crystallite's size of the samples using the Debye–Scherrer relation. The results of the average sizes of the various samples are put together in Table [1](#page-4-1). We analyzed the evolution of the crystallite's sizes as a function of the zinc content. It is noted that the size of the nanoparticles of the doped $TiO₂$ is bigger compared to the TiO₂ P25. Thus, doping has a great infuence on the size of the nanoparticles of pure Sachtopore $TiO₂$. The higher the Zinc content is, the greater the size of the crystallites become.

3.3 Difuse refectance spectroscopy study (DRS)

The transmittance of light of different $TiO₂$ nanoparticles was measured by difuse refectance spectroscopy (DRS) (Fig. [4](#page-5-0)). From Fig. [4,](#page-5-0) the samples exhibit the same behavior. The spectra of these samples show an absorption edge around 800 nm, meaning that photons with wavelengths below 800 nm (more energy) will be absorbed by these

samples. The threshold for absorption of the samples is related to the presence of $TiO₂$ and Zn- $TiO₂$.

The absorbance curves clearly show the infuence of the crystalline phase on the position threshold of the absorption of TiO₂. In particular, we find that the results of the literature put the optical gap of anatase at 3.2 eV and the rutile at a lower energy, at 3.0 eV, but both in the UV region [\[43](#page-13-0)]. In order to determine the optical gap of these samples, the graphical method from the Tauc equations has been used [[43,](#page-13-0) [44\]](#page-13-1). On the other hand, the band gap energies can also be determined based on the Kubelka–Munk theory, where F(R) = (1-R)²/2R = α [\[45](#page-13-2)]. Plotting the (F(R∞).hν)^{1/2} plot as a function of $(h\nu)$ allows the value of the band gap energy to be determined by extrapolating the linear portion of the curve to the x-axis. For more clarity, the plots are shown in Fig. [5](#page-5-1) and the bandgap values are summarized in Table [2.](#page-5-2) The band gap values of all photocatalysts were between the values found in the literature for the anatase and rutile phases $[46]$ $[46]$ $[46]$.

The absorption data obtained by DRS allow the nanoparticles to divide into two types. TiO₂ P25 has an absorption threshold in the UV which corresponds to the absorption threshold of $TiO₂$ in rutile and anatase phases. While the pure $TiO₂$ Sachtopore, Zn- $TiO₂$ absorb the energies corresponding to the absorption threshold of $TiO₂$ in the anatase phase. The absorption in this area is related to the superposition of the energy levels of dopants, which are located between the valence band and the conduction one for $TiO₂$. Namely that the dopants could create a level between the

 $TiO₂$ samples

Fig. 4 Transmittance spectra of

Fig. 5 Modified curves $(h\alpha v)^{1/2}$ of α absorbance of diferent titanium dioxide powders, according to the energy of the incident photon hν

Table 2 The optical gaps of pure and doped $TiO₂$ Sachtopore and $TiO₂$ P25

	Catalysts TiO ₂ Sachto- pore pure	$TiO2$ Sachto- pore doped (ZnCl ₂ / $TiO_2 = 0.1\%$	$TiO2$ Sachto- pore doped (ZnCl ₂ / $TiO_2 = 0.2\%$	TiO ₂ P25
Gap/ev	3.18	3.14	3.10	3.3

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valence and conduction bands and decrease the forbidden band. This allows a small amount of dopants to be inserted into the $TiO₂$ network and the level of dopants has appeared between the valence band and the $TiO₂$ conduction band [[47\]](#page-13-4).

From the results, we can say that there is a small decrease in the energy gap values for all the catalysts.

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3.4 SEM images and EDS analysis

SEM images of pure and doped $TiO₂$ sachtopore nanoparticles show almost the same, all grains are spherical, dark and light. The particles are assembled and agglomerated and they are homogeneous. So, we can say that the morphology of $TiO₂$ did not change after doping. EDS spectroscopy was used to show the presence of zinc. From the EDS

 (b)

Fig. 6 SEM images of powders (**a**) TiO₂ Sachtopore, (**b**) TiO₂ Sachtopore (ZnCl₂/TiO₂=0.1%), EDS spectra for (**c**) TiO₂ Sachtopore, (**d**) TiO₂ Sachtopore $(ZnCl_2/TiO_2=0.1\%)$

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spectrometer of pure Sachtopore $TiO₂$, titanium and oxygen atoms are observed, which shows that the powder contains a quantity of titanium and oxygen (Fig. 6). For zinc-doped $TiO₂$, the same thing is observed but with Zn peaks, so the powder contains titanium and oxygen as well as zinc; these peaks show that the Zn dopant is homogeneously distributed in the TiO₂ matrix and indicate the presence of the $ZnCl₂$ component. The EDS results confrm the doping of sachtopore $TiO₂$ with zinc.

3.5 Adsorption of methyl orange on different TiO₂ catalysts

The adsorption on the catalyst surface is an initial and essential step for the photocatalytic reaction [[48](#page-13-5), [49](#page-13-6)]. In order to identify the optimal conditions for dye degradation, we frst started by studying the adsorption of MO on different $TiO₂$ catalysts.

3.6 Kinetics of methyl orange adsorption on different TiO₂ catalysts

This study aims to verify the adsorption of methyl orange on TiO₂ catalysts and to compare their efficiencies and to determine the necessary time to achieve the adsorption equilibrium state [\[50\]](#page-13-7). Monitoring of adsorption kinetics was performed in darkness under continuous stirring for 2 h (120 min) and the results obtained are grouped in Fig. [7.](#page-7-0) Regardless of the catalyst, the curves show that the amount adsorbed increases with increasing agitation time and tends toward a limit value corresponding to the maximum amount after about 40 min. This time indicates that the adsorption equilibrium is reached. It is therefore imperative to let the mixture agitate for at least this amount of time prior to irradiation of the mixture. From a kinetics point of view, the adsorption occurs in two steps: the frst is quick and the second is slower. It has been observed that over 95% of the

Fig. 7 Kinetic adsorption of methyl orange (0.5 g/L) on diferent studied catalysts (10 ppm)

limiting adsorption capacities of catalysts were reached after 30 min of agitation. Beyond this time, the adsorption rate becomes slower. The curves in this fgure show that under the same conditions, the dye does not have the same affinity for the solid. Dye adsorbs well on doped $TiO₂$ nanoparticles than pure $TiO₂$. In addition, the adsorption capacity depends on the amount of dopant. The amounts of the dye adsorbed on the catalysts equilibrium varies in the following order: doped TiO₂ Sachtopore (ZnCl₂ /TiO₂=0.2%) > doped $TiO₂$ Sachtopore $(ZnCl₂/TiO₂=0.1%)$ > Pure TiO₂ Sachto $pore > TiO₂ P25$. This improved result can be justified by the increase of the adsorption sites of the catalyst or it could be due to the structure of each catalyst [\[50](#page-13-7)].

3.7 Methyl orange photodegradation in the presence of the various catalysts

The objective of this part of the work is to study the infuence of several parameters such as the concentration of methyl orange, the concentration and the nature of the catalyst, as well as the doping of the catalyst on the degradation kinetics of the methyl orange dye in the presence of diferent pure and doped $TiO₂$ nanoparticles, in aqueous medium. Different $TiO₂$ catalysts are used to compare their photocatalytic efficiencies. This study is very important in the design and optimization of industrial systems**.**

3.7.1 The efect of initial concentration of methyl orange on the kinetics of degradation

The kinetic study of the photocatalytic reaction of methyl orange depending on initial concentration of MO was conducted for two diferent concentrations of 5 mg/L and

Fig. 8 Curves of $\ln(C/C_0)$ in function of irradiation time of MO degradation (10 ppm) in the presence of $TiO₂$ (1 g/L)

Fig. 9 Curves of $\ln(C/C_0)$ in function of irradiation time of MO degradation (5 ppm) in the presence of TiO₂ (1 g/L)

10 mg/L. Figure [8](#page-7-1) and Fig. [9](#page-8-0) show the photodegradation kinetics of methyl orange and $TiO₂ (1 g/L)$. The kinetics of degradation depends on the methyl orange concentration and the nature of the catalyst. The decrease of the degradation of the dye during a high concentration of the latter is explained by the increase of the organic molecules of this pollutant with a number of oxidants generated which is fxed in the solution, all this leads to a low percentage of degradation [\[51](#page-13-8)]. On the other hand, perhaps, this is due to the coverage of the photocatalytic sites by the molecules of the dye and decrease in the generation of the oxidizing radicals on the surface of the catalyst [[52\]](#page-13-9).

So, the lower the initial concentration of methyl orange is, the higher and faster the degradation will be.

On the other hand, the higher the initial concentration of the dye, the lower the trajectory of the photons entering the solution, which influences the degradation efficiency $[53]$ $[53]$.

The linearity of the curves of ln (C/C_0) as a function of time indicates that the degradation follows an apparent frstorder (Ln (C/C_0) = -K.t). The value of the kinetics constant and the half-life time are shown in Table [3](#page-8-1). It was clearly observed that the lower the initial concentration of MO

3.7.2 The effect of initial concentration of TiO₂ **on the degradation kinetics**

In heterogeneous catalysis, the reaction rate depends generally on the quantity of the catalyst. In contrast, in the photocatalytic performance, only the $TiO₂$ grains that are directly exposed to the radiation are active. It is obvious that the mass needed depends strongly not only on the geometry of the reactor volume, but also on the form in which the photocatalyst is used [[54\]](#page-13-11). The experiments were carried out using two different concentrations 0.2 g/L and 1 g/L of TiO₂ and the initial concentration of methyl orange used was 10 ppm.

The obtained results (Fig. 10 and Fig. 11) show that when the concentration of $TiO₂$ increases, the degradation kinetics increase and the time to remove MO decreases. Indeed, for TiO₂ concentration of 1 g/L, the photo-disappearance is much more important than for a $TiO₂$ concentration of 0.2 g/L. This study showed that a concentration of 1 g/L $TiO₂$ for all samples is the optimum concentration. The rate constant varies similarly with the catalyst concentration between 0.2 g/L and 1 g/L (Table [4](#page-9-1)), which allow us to

Fig. 10 Plotting of ln (C/C_0) in function of irradiation time of MO degradation (10 ppm) in the presence of TiO₂ (0.2 g/L)

Table 3 Kinetic parameters of TiO₂ (1 g/L) and 10 ppm, 5 ppm of MO and the correlation coefficient R^2

R^2	R^2
	MO(5 ppm)
0.9857	0.9953
0.9924	0.9912
0.9802	0.9913
	MO(10 ppm)

Fig. 11 Plotting of $\ln(C/C_0)$ in function of irradiation time of MO degradation (10 ppm) in the presence of TiO₂ (1 g/L)

predict that all grains are photoactive. The degradation rate is therefore proportional to the mass of $TiO₂$ and therefore to the concentration of the adsorbed molecule.

In general, increasing the catalyst concentration promotes the adsorption of the contaminant on the surface. However, for larger quantities, we have noted that degradation kinetics decreases and the reaction rate decreases due to the saturation of the absorption of photons or the loss of light scattering [\[54\]](#page-13-11). The linearity of the curves of $\ln(C/C_0)$ in function of the time indicates that this degradation follows an apparent frst-order kinetics. The value of the kinetics constant and the half-life time are summarized in (Table [4](#page-9-1)). In addition, an increasing on the catalyst concentration promotes the degradation of MO dye. Indeed the half-life time decreased for various catalysts.

3.7.3 The efect of catalyst nature on the degradation kinetics

The purpose of this study is to make a comparison between two types of TiO₂ catalysts (pure TiO₂ P25 and pure TiO₂ Sachtopore) in order to assess the efectiveness of the catalyst on the methyl orange degradation kinetics. Methyl orange solution of 5 ppm was introduced into a reactor in

Fig. 12 Plotting of $\ln(C/C_0)$ in function of irradiation time of MO degradation (5 ppm) in the presence of TiO₂ P25, pure TiO₂ Sachtopore (1 g/L)

the presence of 1 g/L of TiO₂ (Fig. [12](#page-9-2)). The photocatalytic efficiency of $TiO₂$ depends on the nature of the catalysts used. The linearity ln (C/C_0) shows that the degradation kinetics does not change and keeps following the frst order (Fig. [12](#page-9-2)). The values of the rate constant and the half-life time are listed in (Table [5\)](#page-10-0). The photocatalyst $TiO₂$ P25 shows greater photocatalytic activity than pure Sachtopore $TiO₂$. This difference in the efficiency may be associated with the difference between textural and structural characteristics of each catalyst. The high performance of $TiO₂$ P25 photocatalyst can also be explained by these features (80% anatase and 20% rutile 317 m²/g surface size, nonporous) [\[54\]](#page-13-11). Also, TiO₂ P25 contains anatase crystallites centered by a small rutile core. According to this form, the strong photocatalytic efficiency has been attributed to catalytic "hot spots" at the intersection of the two phases due to the unique surface chemistry [\[55](#page-13-12)].

It was clearly observed that the kinetic constant obtained with pure $TiO₂$ Sachtopore is smaller than that obtained by TiO₂ P25, while the half-life time in the presence of TiO₂ P25 is lower than the one obtained in the presence of pure $TiO₂$ sachtopore. This result shows that methyl orange degrades faster in the presence of $TiO₂$ P25.

Table 4 Kinetic parameters of various TiO₂ catalysts at a concentration of 0.2 g/L and 1 g/L at 10 ppm of MeO and the correlation coefficient R^2

K_{app} (min ⁻¹) TiO ₂ (0.2 g/L)	K_{app} (min ⁻¹) TiO ₂ (1 g/L)	$t_{1/2}$ (min) TiO ₂ (0.2 g/L)	$t_{1/2}$ (min) TiO ₂ (1 g/L)	R^2 TiO ₂ (0.2 g/L)	R^2 TiO ₂ (1 g/L)	
0.0020	0.027	347	257	0.9889	0.9857	
0.0011	0.0024	630	288	0.9788	0.9924	
0.0010	0.0019	693	365	0.9747	0.9802	

Table 5 Kinetic parameters of $TiO₂$ and the coefficient of determination R^2

	TiO ₂ P25	Pure $TiO2$ Sachto- pore	
K_{app} (min ⁻¹)	0.0105	0.0043	
$t_{1/2}$ (min)	66	161	
R^2	0.9953	0.9912	

Fig. 13 Plot of \ln (C/C₀) versus irradiation time of MO degradation (5 ppm) in the presence of pure and doped $TiO₂$ Sachtopore (1 g/L)

3.7.4 The efect of dopant on the kinetics of degradation

Several important studies have been done on the modifcation of materials in order to obtain stronger photocatalysts, such as modification or surface formation [\[56](#page-13-13)[–58](#page-13-14)] of heterostructures with other substances [\[59](#page-13-15), [60](#page-13-16)]. Element doping (N, C, Cu, Ag, Nb…) is an available and practical method [\[61](#page-13-17), [62](#page-13-18)] and effective in improving the photocatalytic yield.

Moreover, doping with alkali metals (Na, K), nonmetals (B, C, N, O, P, S, Cl, and Br, etc.), transition elements (Fe, Cd, Co, Mo, V, etc.), and rare earth metals (Ce, Eu, Se, Y) induces a reduction of the band gap and a shift of the photocatalytic response to the visible region of the solar spectrum [\[31](#page-12-17)].

The purpose of this study is to make a comparison between pure $TiO₂$ Sachtopore and doped with different percentages of zinc chloride $(ZnCl₂)$ to evaluate the efficiency

of the catalyst on the degradation of methyl orange. Methyl orange solution of 5 ppm was introduced into a reactor in the presence of 1 g/L of TiO₂ (Fig. [13](#page-10-1)).

The curves ln $(C/C_0) = f(t)$ show that under the same conditions, MO dye does not exhibit the same affinity for the solid. The photocatalytic degradation efficiency of the catalysts varies in the following descending order: doped TiO₂ Sachtopore (ZnCl₂/TiO₂=0.2%) > Pure TiO₂ Sachtopore > doped TiO₂ Sachtopore (ZnCl₂ /TiO₂ = 0.1%).

Table [6](#page-10-2) groups the values of the half-life time and the speed constant of each catalyst.

Doping improves the photo-reactivity of $TiO₂$ Sachtopore. Zn^2 ⁺ can play both as an electron and a trap hole, which does not leave the recombination of photogenerated e^- / h ⁺ pairs on the surface of the semiconductors [[66](#page-13-19)].

In addition, zinc reveals better results with a mass percentage of 0.2%. However, 0.1% of zinc does not improve the photocatalytic performance of $TiO₂$ Sachtopore.

3.8 Recent work on photocatalytic degradation using TiO₂ or Zn doped composites as catalysts.

Several works are interested in the study of photocatalytic degradation using $TiO₂$ or zinc-doped composites as catalysts in order to remove diferent pollutants; in this paragraph, we will quote some studies found in the literature in order to have a recent report and also to make a comparison.

We found a study on the degradation of bisphenol S (BPS) by persulfate activation under UV-LED using mixed catalysts: Synergistic effect of $Cu-TiO₂$ and $Zn-TiO₂$ for catalysis. They used the sol–gel method to prepare $Zn-TiO₂$ and $Cu-TiO₂$. The two catalysts showed a clear synergistic efect in the photocatalytic degradation process. When the persulfate is 5 mM and the catalyst is 0.3 g/L, the removal rate is 100% in 18 min, in the case of the mixed catalyst (0.2 g/L Zn-TiO₂ and 0.1 g/L Cu-TiO₂), while in the case of 0.3 g/L Zn-TiO₂ the degradation rate is 58% and 90% in the case of 0.3 g/L Cu-TiO₂. Several factors were investigated: the Cu-TiO₂/Zn-TiO₂ ratio, the catalyst dosage, the persulfate dosage, the initial BPS concentration and the initial pH of the solution and the optimal conditions were determined. In summary, in this study, the degradation efficiency of BPS is improved if single metal-doped $TiO₂$ is replaced by mixed $TiO₂$. This technique makes the catalyst more efficient [\[67](#page-13-20)].

In another work, the authors developed a solar-derived $TiO₂/g-C₃N₄$ heterostructure photocatalyst without noble

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Table 6 Kinetic parameters obtained

metal by the microwave-assisted hydrothermal route for high photocatalytic activity. The prepared photocatalysts were investigated for photocatalytic hydrogen production and methylene blue dye degradation. 7.5 mg is the weight of the TiO₂/g-C₃N₄ photocatalyst that showed the highest dye degradation removal rate (84.61%) and also a photocatalytic hydrogen productivity of 110 μ molg⁻¹ h⁻¹, which is 2.0 times higher than that of pristine $TiO₂$. This significant increase in hydrogen productivity results from the extended UV–visible light absorption capacity, which increases the charge carrier separation very efficiently. Moreover, they also showed that the enhanced photocatalytic activity is due to the interfacial contact between $TiO₂$ and $g-C₃N₄$, which effectively reduces electron–hole recombination. This technique provides highly efficient photocatalysts for a wide range of applications such as organic dye degradation, hydrogen evolution, and water treatment [[68\]](#page-13-21).

There is also another work on photo-induced charge separation at the interface of $Zn-Pd/TiO₂$ hybrids for enhanced photocatalytic and electrochemical activity. In this work, the authors used diferent hybrids to change the role between Ohmic and Schottky barriers to improve the photocatalytic activity. To form hybrids, metal nanoparticles (Pd and Zn) and bimetallic nanoparticles (ZnPd) are opted to study the interface modifcation by UPS and XPS. The study showed that $TiO₂$ nanorods fabricated with Pd and ZnPd (TiO₂ NR) show the constitution of a Schottky barrier with upward band bending at the interface. Thus, $TiO₂$ nanorods (TiO₂ NR) fabricated with Zn show an ohmic barrier with a downward band bend at the interface. At the electrochemical analysis, they analyzed that in $ZnPd/TiO₂ NR$, the electrical conductivity is higher, which facilitates the charge carrier transport. Better photocatalytic activity with ZnPd/TiO_2 NR than with Pd/TiO₂ NR and Zn/TiO_2 NR, was due to the improved charge separation at the interface of $ZnPd/TiO₂ NR [69]$ $ZnPd/TiO₂ NR [69]$.

4 Conclusion

Nowadays, the protection of the environment is a major problem for mankind. The treatment and reuse of industrial water have become an urgent necessity. Methyl orange is one of the most widely used dyes in industries; for this reason, we have been interested in the photocatalytic removal of this dye and have been working on the development of the photocatalytic efficiency of our $TiO₂$ Sachtopore catalyst by doping, which is very important in order to obtain a better photo-active solid during degradation.

The photocatalytic degradation of methyl orange was studied using $TiO₂ P25$, pure $TiO₂$ Sachtopore, and doped TiO₂ Sachtopore at two different percentages 0.1% and 0.2% . Upon adsorption, the doped $TiO₂$ Sachtopore shows a strong adsorption capacity higher than that of $TiO₂$ P25. The initial

concentration of methyl orange plays an important role in the photocatalytic degradation kinetics. By decreasing the concentration of methyl orange, the degradation rate increases. Doping also plays an important role in the photocatalytic process and improves the photoreactivity of 0.2% TiO₂ Sachtopore. The photocatalytic degradation efficiency of the catalysts varies in a decreasing order: for $TiO₂$ Sachtopore ($ZnCl₂/TiO₂=0.2%$) is 51% > pure TiO₂ Sachtopore is 36% > TiO₂ Sachtopore (ZnCl₂/TiO₂ = 0.1%) is 32% at a concentration of 5 ppm in the presence of 1 g/L of $TiO₂$.

Furthermore, the photocatalytic degradation depends on the nature and concentration of the catalyst. The degradation of methyl orange follows apparent frst order kinetics for all catalysts.

Declarations

Conflict of interest The authors declare no competing interests.

Life science reporting No life science threat was practiced in this research.

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