

Hydrochemical investigations of groundwater quality for drinking and irrigational purposes: two case studies of Koprivnica-Križevci County (Croatia) and district Allahabad (India)

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Received: 4 April 2017 / Accepted: 25 September 2017 / Published online: 7 October 2017
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Abstract The hydrochemical study was conducted to outline the suitability of groundwater for drinking and irrigational purposes of two distinctly located geographical areas, Koprivnica-Križevci County, Croatia and district Allahabad, India. A total 160 groundwater samples were collected from sources (in hot and cold seasons) of both countries during 2 years (2011–2012) to evaluate the recent hydrochemical status of groundwater. Samples were analyzed for 12 various physico-chemical parameters. The Koprivnica-Križevci County has Ca–HCO₃ water type, whereas in Allahabad it prevails as Mg–HCO₃. All the

parameters were below the maximum acceptable value (MAV) of standards (WHO, US EPA, Indian and Croatian Standards). The usage and consumption of groundwater should be of no concern with regard to inorganic pollutants. Irrigational indices (residual sodium carbonate content, magnesium hazard, permeability index, exchangeable sodium percentage, and salinity hazard) showed that groundwater is suitable for irrigational purposes. Hierarchical clustering analysis (HCA) has resulted into four groups due to the processes on natural and anthropogenic factors. The principal component analysis (PCA) has shown that first factor (all physico-chemical analyzed variables) explains 27.75% of the total variability and another factor (geographical distribution) accounts only for 10.46%. We found that local environmental conditions are more important predictor than geographical distribution. The present study may be helpful in dealing for further studies concerning groundwater quality issues in the distinctly located geographical areas. The groundwater quality is essential for management and sustainability of water resources, economic development, and human health.

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Keywords Groundwater quality · Drinking water · Irrigation · Croatia · India

Introduction

Groundwater is a regional resource, but water scarcity and pollution are becoming a global issue due to increasing population, economic growth, and climate change. Development of new sources of water beside its efficient use, together with conservation measures, should be an important component of any (developed and developing) country's national water plan. In many parts of the world

watersheds, river, ponds, aquifers and the associated ecosystems have undergone significant modifications; hence the vitality, quality, and availability of the water resources have been facing further human threats (Singh et al. 2015a). The causes of groundwater pollution are natural and human-induced. In recent years, human-induced causes have increased and could be easily linked to unprecedented population growth; a changing climate, rapid urbanization, expansion of infrastructure, migration; land conversion; extensive agricultural activities and over exploitation (Singh et al. 2010, 2013a, b, 2015a, b; Srivastava et al. 2013; Nemčić-Jurec et al. 2013; Nemčić-Jurec and Jazbec 2017; Jacintha et al. 2016; Rawat et al. 2017a, b; Singh et al. 2017). To understand the impact of human activities, scientists have studied chemical speciation, saturation, transport of pollutant and decay, dissolution, precipitation, ion-exchange, sorption, and desorption, together with the residence time occurring along the flow path and change in land use control the variation in chemical of groundwater (Belkhiri et al. 2011; Singh et al. 2012; Thakur et al. 2015). These studies help in developing and defining the feasible strategies for the protection of groundwater resources from pollution (Chidambaram et al. 2013a, b).

The hydrogeological setting is important, the alluvial aquifers constitute a hydrological unit formed by the alluvial deposits and are characterized by a linear and shallow feature as they are spread along the fluvial valleys of the basin (Helena et al. 1999). The shallow aquifers have high permeability, hence are vulnerable to pollution (Ravikumar and Somashekar 2012). Hence, it is very important to monitor the quality of groundwater to suppress the contamination and to minimize the pollution. According to literature (Singh et al. 2009, 2015; Tziritis et al. 2016), groundwater chemistry is generally influenced by the characteristics of catchment area, such as natural (geogenic) factors like climatic conditions, bedrock geology, soil properties, and transport or flow of water through environment. Hydrogeochemical characteristics are dependent on the chemistry of rock-forming minerals and on the physical process of erosion, respectively, mineral dissolution. As a result, groundwater is enriched in ions and other elements (Robinson and Ayotte 2006). Anthropogenic impact is mainly related with extensive agricultural practices which favor nitrate enrichment and salinization (Nemčić-Jurec and Jazbec 2017) or related to industry and waste water sources (Gautam et al. 2013).

Groundwater quality has been investigated by many researchers (Mikac et al. 1998; Ahel et al. 1998; Stambuk-Giljanovic 2005; Yidana et al. 2008; Ravikumar and Somashekar 2012; Singh et al. 2012, 2013a, b; Tiwari and Singh 2014; Chandra et al. 2015; Thakur et al. 2015; Gautam et al. 2016). Multivariate statistical techniques (MST) help in effective management of large and complex groundwater data. Principal component analysis (PCA) and

cluster analysis (CA) are widely used statistical techniques for the characterization and evaluation of groundwater quality (Singh et al. 2004, 2009). Researchers have used MST to characterize and evaluate surface and groundwater quality and have found them very useful for studying the variations caused by different factors (Vega et al. 1998; Singh et al. 2005, 2009, 2012, 2013a, b, 2015a, b; Shrestha and Kazama 2007; Belkhiri et al. 2011; Tziritis et al. 2016). Astel et al. (2006) have applied MST for the determination of chloro/bromo disinfection by-products in drinking water at 12 locations in the Gdańsk area (Poland). Ujevic et al. (2012) have used cluster, factor, and discriminant analyses to study the groundwater chemistry of Eastern Croatia and to identify the main geochemical processes responsible for high arsenic concentrations in analyzed groundwater.

Despite the importance of groundwater in Allahabad (India) and Koprivnica-Križevci County (Croatia), it is known about natural and anthropogenic influences but only in some areas of countries. The importance of the groundwater resources in the area should not be underestimated. Groundwater quality indices provide a simple and easily understandable tool for managers on the water quality and possible uses for irrigation water. An individual quality factor alone is not enough to evaluate the drinking or irrigational water quality because it could be restrictive and can give unfavorable qualification. The groundwater hydrochemistry of Koprivnica-Križevci County and Allahabad district, especially for irrigation purposes, is poorly understood. Increased knowledge of hydrochemical water quality of Allahabad, India and Koprivnica-Križevci County of Croatia will be leading to sustainable development of water resources. In some parts of India, irrigation is applied without control and supervision, and in Croatia irrigation is rarely for now. During the past two decades, irrigational uses in India have been rapidly increased and started in Croatia too but often without adequate planning. The aim of this paper was: (1) to study groundwater quality on intense agricultural area of Koprivnica-Križevci County (Croatia) and on agricultural and urban area of Allahabad (India) to assess the influences on groundwater quality and in the same way estimate the suitability of groundwater for drinking and irrigational uses, and (2) to assess whether local conditions are more important predictor for groundwater quality than geographical distribution. Assessing of groundwater quality could be applied in such a way in other unexplored areas of the two countries and provide better water management and human health.

Study area

In this paper, there are two study areas that were targeted to assess the groundwater quality. The first is Koprivnica-Križevci County located between 46°10'N and 16°50'E and

covers an area of 1748 km² (Fig. 1a). The study area belongs to the catchment area of the River Drava where prevailing quaternary gravel-sand aquifer (Mayer et al. 1996). On the north-western hydrogeological area, dominate carcasses of mountain Kalnik. In the deeper layers laying reservoir of groundwater, lime dolomite aquifer. Some parts on north-west hydrogeological area stretch on the Pleistocene terrace prevailing gravel and sand aquifer.

The climate is characterized by hot and dry season which includes spring and summer and cold season which includes autumn and winter. Average annual amount of rainfall on study area is 803 mm. The amount of rainfall is similar in the whole area. July and September are the 2 months of maximum rainfall. Average annual

temperature is 10.9–11.2 °C and the climate is moderately warm; July usually being the hottest month year with mean daily maximum temperature at 38.2 °C and minimum temperature 7 °C; January being the coldest month of the year with mean daily maximum temperature 5 °C and minimum – 22.6 °C. The relative humidity in study area was 80–96% during the rainy and cold season and about 15–20% during the dry and hot season. Average annual wind speed for the study area is 2.5–3 m/s and maximum wind speed is about 14 m/s.

The research area (Fig. 1a) is located on three different hydrogeological units (Fig. 2a). The first area of sources C1, C4, and C5 is located in the water area of central Podravina. This water area extends from line-Legrad Koprivnica, in the east, to the line of Podravska Slatina-Sopje, in the west. From the north area, Podravina is bounded by the state border with Hungary, and in the south area, it is bounded by northern slopes of Bilogora and Papuk. The area of the old alluvial terraces composed of layers of the Drava River and its tributaries, and terraced sediments. Hydrogeological characteristics were caused by geological structure of the area. To the hydrogeological features, the most important is Quaternary gravel-sand aquifer, with a thickness of 30 m to the south (C4 and C5 locations) and in some parts of the river basin district reaches up to 50 m (C1 location). Top part—the roof of the aquifer is composed of dust, sand, and clay with a significant occurrence of live sands and marsh loess mainly in southern and eastern part of the river basin district. The thickness of this layer and its texture strongly influence on the permeability of the substances from the surface layers to the aquifer. On the west side from Virovitica, the thickness of the roof is about 10 m. In the vicinity of the Virovitica, the roof is thicker and further to the east regularly it is thicker than 20 m. There is no unique hydrogeological regime on the water area. The impact of the Drava River on the direction of groundwater flow is clearly visible in the zone width of 2–3 km along the Drava River and the area from Legrad to Pitomača (C1, C4, and C5). In this zone, groundwater imitates the regime of the Drava River during a year. In the rest part of the area, the influence of the Drava is not clear and unambiguous. The aquifer is filled by the infiltration of rainwater through the low permeable cover and by seepage from the bed of the Drava River in the upstream part of the territory. Natural water quality varies from site to site (Mayer et al. 1996). The second and third hydrogeological unit covers an area of Prigorje region (C2 and C3 locations). The C2 landscape area dominate the carcasses of mountain Kalnik. Morphologically most prominent parts were built from the Paleogene and Neogene rocks while the northern area of the Kalnik was built from the Mesozoic deposits. The massif of mountain Kalnik is built from porous lime

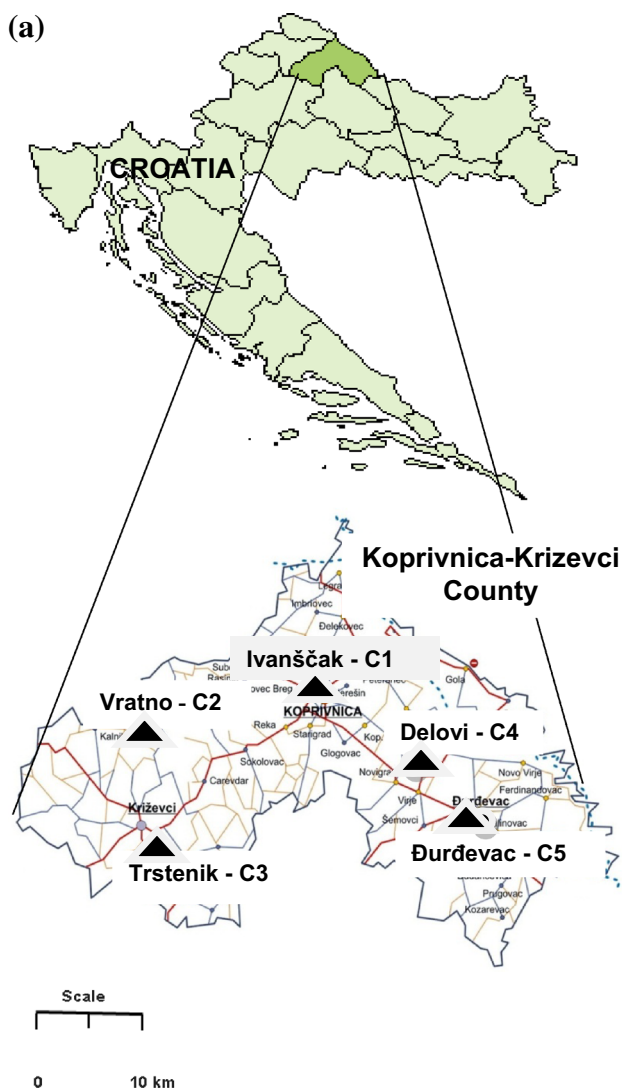


Fig. 1 Map showing the water sampling locations of the two study area. **a** Koprivnica-Križevci County of country Croatia (C1–C5); **b** Allahabad district, state Uttar Pradesh of country India (I1–I30) [see “Appendix 1” Station code and sampling locations in Croatia ($n = 5$) and India ($n = 30$)]

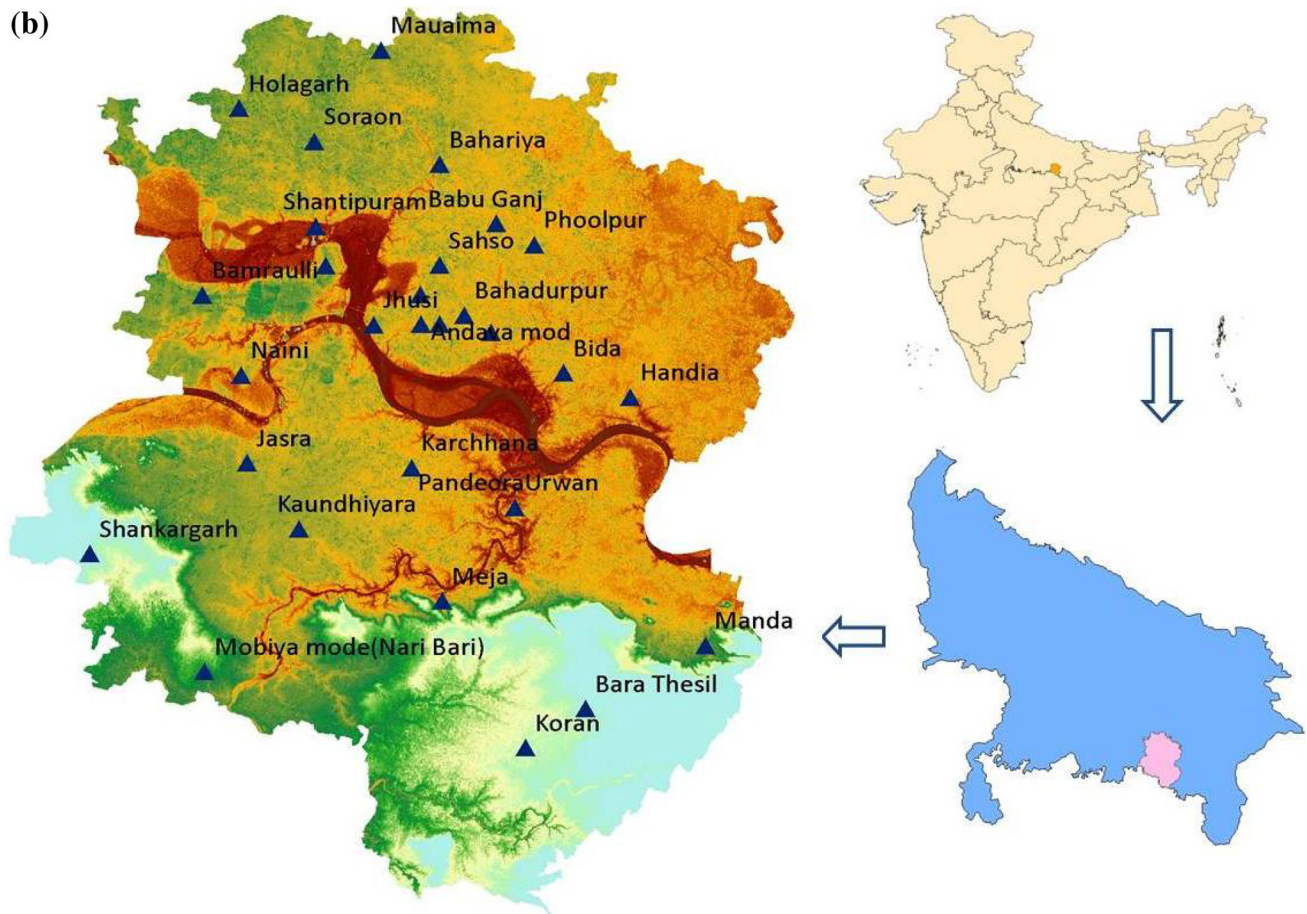


Fig. 1 continued

dolomite breccia and permeable to poor permeable Tortonian deposits. Deposits in the surface area represent collector of rainwater and in the deeper layers is reservoir of underground water, lime dolomite aquifer (Šimunic 1984). Area of C3 location, which also belongs to the Prigorje, stretches on the Pleistocene terrace. After 7–9 m of cover clay, lies aquifer to 9 m thick build from various grit of sand and gravel. Beneath aquifer, lies a thin layer of coal, clay and marl clay. The roof layers build from soil and Pleistocene clay are poorly permeable materials that protect the aquifer (Sebastian and Britvic 1989).

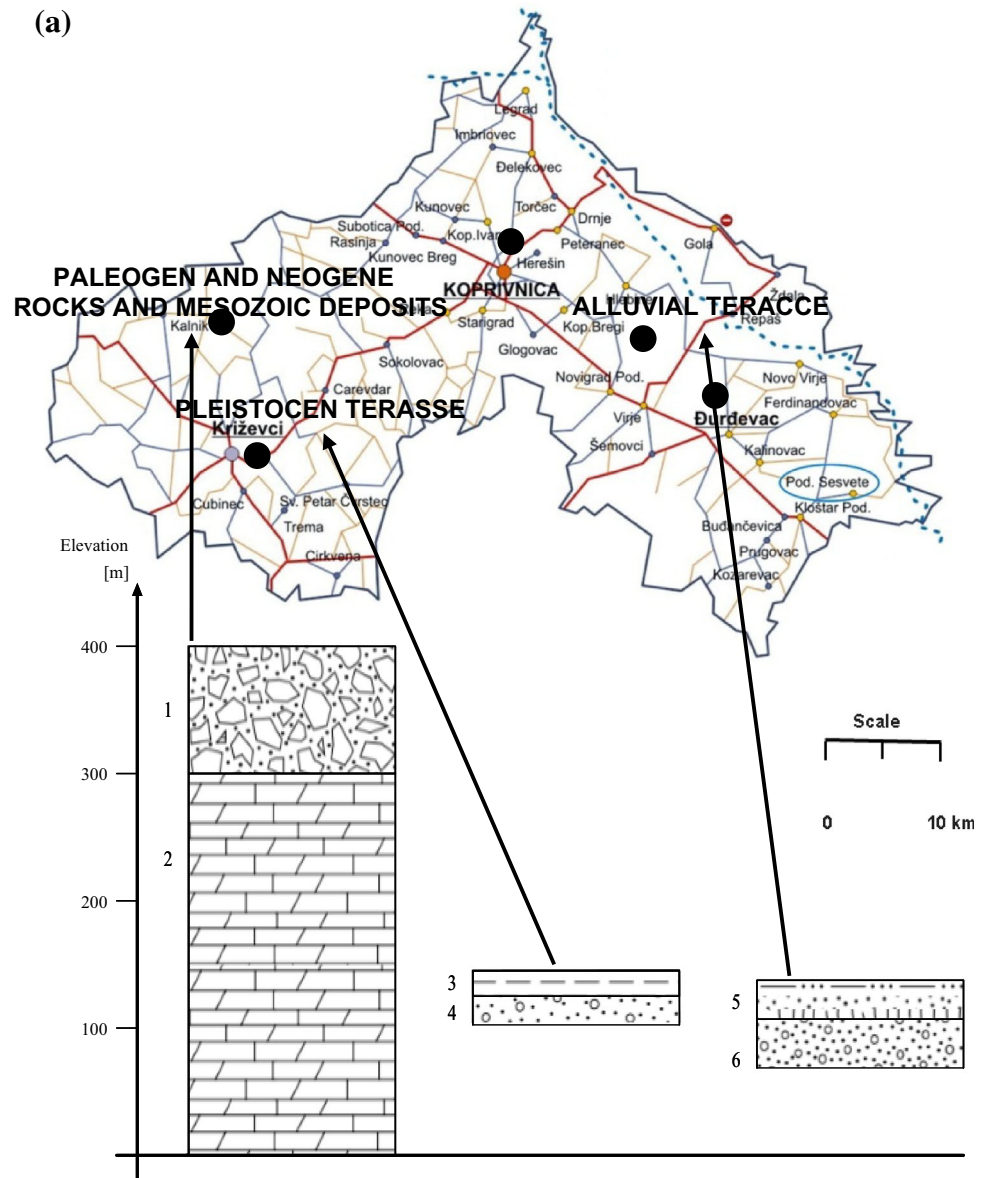
The second study area is Allahabad district located ($24^{\circ}47'N$ – $25^{\circ}47'N$ latitudes and $81^{\circ}19'E$ – $82^{\circ}21'E$ longitudes) by covering an area of 5246 km^2 in India. Geologically and hydrogeologically, it is an interesting area; the two major rivers Ganga and Yamuna flow through the district (Fig. 1b). The whole Trans-Ganga tracts, the greater portion of doab (interfluvial region), are composed of gangetic alluvium (Fig. 2b). The alluvial detritus of the Vindhyan is found in the southern part of the doab. The Trans-Yamuna tract and the Vindhyan detritus merge in the gangetic sand and silt. The gangetic alluvium consists of

alterations of fluvial deposition of sand, silt, and clay. The thickness of alluvium decreases from north to south. Glass sand deposits are found in the neighborhood of Shankargarh.

The climate of region is characterized by a long and hot summer (March–middle of June), monsoon (June–September), and cold seasons (mid-November–February). The rainfall generally decreases from the south-east to the north-west. About 88% of the annual rainfall is received during the monsoon season. The average annual rainfall is 975.4 mm with 48 rainy days in a year. From about the middle of November, the temperature begins to fall rapidly and in January (the coldest month) the mean daily maximum is 23.7°C . The heat in the summer season, particularly, in May and the early part of June is intense; May usually being the hottest month of the year with the mean daily maximum temperature at 41.8°C and the mean daily minimum at 26.8°C .

The study area of Allahabad district is a part of Quaternary alluvium and Vindhyan Plateau and ranges from Proterozoic to recent geological origin. The basement in the area is formed by Quartzite of Kaimur group and is

Fig. 2 Depicting **a** hydrogeological setting of Koprivnica-Križevci County (Sources: Mayer et al. 1996; Sebastian and Britvic 1989), and **b** hydrogeological setting of Allahabad district (Sources: Central Groundwater Board, India, Report, 2008–2009)



Legend:

- 1 porous limy dolomite and permeable deposits (100 m)
- 2 limy dolomite (300 m) – aquifer
- 3 clay (7 – 9 m)
- 4 sand – gravel (9 m) – aquifer
- 5 silt, sand, clay and marsh loess (10 – 20 m)
- 6 gravel – sand (30 – 50 m) – aquifer

unconformably overlain by Quaternary alluvium. The distribution of groundwater is dependent on the type of formation; in study area, it ranges from unconsolidated to alluvial formation and groundwater occurs under unconfined to confined conditions. The groundwater is shallow (northern part of the district) to deep aquifers (southern part

of the district) and depth ranges from 2 to 20 m during pre-monsoon period, and in post-monsoon period the groundwater recharge and hence the shallow water tables recorded as 1–18 m. In the Vindhyan formation (consolidated), the water tables are much dynamic in nature and it ranges from 3 to 10 m below ground level during pre-monsoon period

(b)

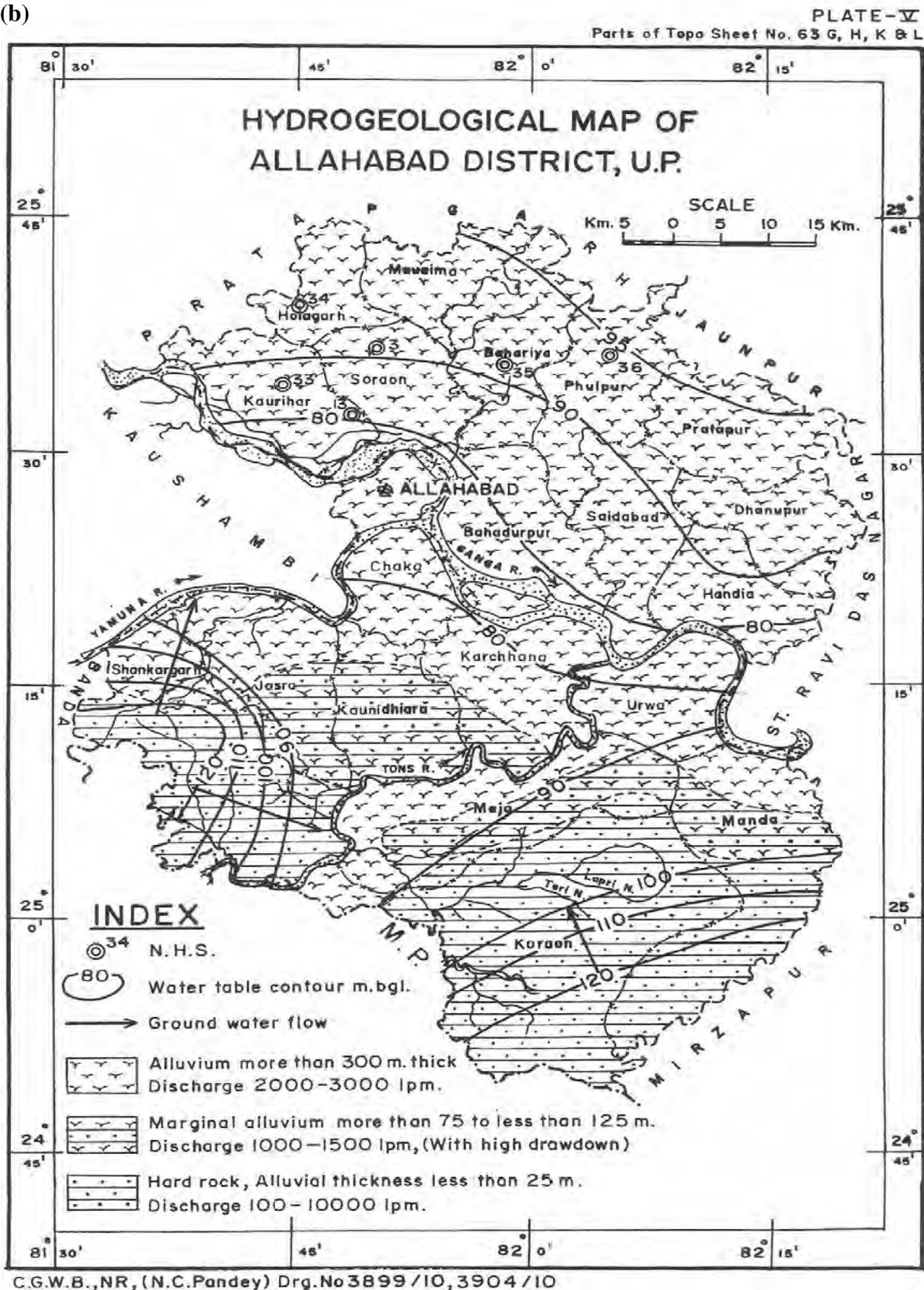


Fig. 2 continued

and 2–8 mg/l in post-monsoon period. The seasonal fluctuations were ranging from 1 to 4 m. The lithologies in Trans-Ganga and Tran-Yamuna area suggest that these areas have unique geological and hydrogeological characteristics. Further, the alluvial plain is also classified as younger and older alluvial plain; in the study area, older plain have two subdivisions (Fig. 2b) as Banda older alluvium and Varanasi alluvium (Singh et al. 2015b). The groundwater exploration studies have revealed a three-tier aquifer systems in the alluvial area that have unique granular zones as follows: (1) I aquifer group (shallow) ranging from 0.0 to 110 mbgl; (2) II aquifer group (middle aquifer) ranging from 120 to 250 mbgl; and (3) III aquifer group (deeper aquifer) lying below 260 down to depth 400 mbgl. The aquifer material is medium-to coarse-grained sand admixed with gravel at places. The tube well which is located in the alluvium plains have high yield 2000–3000 lpm (liter per minute), whereas the of Vindhya regions have limited fractured zones, and sustain less water and exists down to 125 m only and yield is low 500–1000 lpm; within this sandstone domain, a silica sand horizon exists having a thickness of 5–40 m which also contains groundwater (Singh et al. 2015b).

Materials and methods

Water sampling and analysis

During 2011–2012, a total of 40 samples of groundwater were taken during hot and dry season which includes spring and summer and cold season which includes autumn and winter (4 times per year). The groundwater quality has been assessed systematically for all samples and from all the five locations of sources of public water supply (Fig. 1a). Samples were collected (according to HRN ISO 56667-5 2000) in clean polyethylene bottles (1 l volume) for laboratory analysis. The water sampling from wells of sources was done after releasing the water for 2–3 min on the outlet of a water supply system. The samples were stored in the portable fridge, in the dark at 1–5 °C for transportation. After delivery to the laboratory, the samples were stored at the same temperature until the analysis (within 24 h).

pH, electrical conductivity (EC), and total dissolved solids (TDS) were measured using pH meter/conductometer system (Mettler Toledo) employing standard methods (Clesceri et al. 1998; Rice et al. 2012). Anions (Cl^- , SO_4^{2-} , F^- , and NO_3^-) were measured by ion chromatography (Dionex-ICS 3000, Ion pack AS15, using a $\text{Na}_2\text{CO}_3/\text{NaHCO}_3$ -Eluent with a flow rate of 1 ml/min). ASRS ULTRA II-4 mm performed chemical suppression. The analysis was done according to HRN EN ISO 10304-1

(2009)/Ispr.1:2012. Cations (Na^+ , K^+ , Mg^{2+} , and Ca^{2+}) were measured by ion chromatography (Dionex-ICS 3000, Ion pack CS16, using a 30 mM MSA-Eluent with a flow rate of 1 ml/min). CSRS 300 4-mm has been used to perform chemical suppression. The analysis of cations was done according to HRN EN ISO 14911 (2001). Total alkalinity (as HCO_3^-) were measured by a method prescribed in standard methods (Clesceri et al. 1998; Rice et al. 2012), by titration with HCl. Precision judged (% RSD) from the duplicates is on an average < 10% for all parameters.

Collection and analysis of Indian samples ($N = 120$) during hot (pre-monsoon) and cold season (post-monsoon) have been performed as specified in the standard international methods (APHA 1996). Total alkalinity (as HCO_3^-) and physical parameters, such as EC, total dissolved solids (TDS), and pH, were measured in situ using digital probe. Alkalinity is measured using a Hach field titration kit (through titration with 0.1 M HCl). The major cations (Mg^{2+} , Ca^{2+} , Na^+ , and K^+) were analyzed using an atomic absorption spectrometer (Varian, 280 FS), and major anions (F^- , Cl^- , SO_4^{2-} and NO_3^-) were analyzed using an ion chromatography (Dionex Dx-120) using anion (AS12A/AG12) columns coupled to an anion self-regenerating suppressor (ASRS) in recycle mode. Bicarbonate (HCO_3^-) was determined by titration method as described in the standard methods for the examination of water and wastewater (APHA 1996). The precision and accuracy of the analyses are within 5% (evaluated through repeated analyses of standards and samples).

Data analysis

Descriptive statistic of physicochemical parameters of groundwater samples ($N = 160$) collected from studied areas. Afterwards, the normalized concentrations anions and cations were plotted as Piper plot, it comprises three pieces: a ternary diagram (in the lower left representing the cations, in the lower right representing the anions, and a diamond plot in the middle representing a combination of the two). The intersected third point on diamond plot represents a sample.

Groundwater quality datasets were further subjected to two MST: HCA and PCA. HCA was designed to perform classification by assigning observations to groups in more-or-less homogeneous and distinct forms. Its method of analysis was Euclidean distance as distance of measure and complete linkage as linkage algorithm; however, helps in creation and identification of natural groupings for samples. Moreover, PCA was carried out to decipher the sources of ions in the groundwater samples and hydrogeochemical interpretation (Singh et al. 2009, 2015a, b). PCA we calculated with respect to their annual average

concentration of parameters which has given the eigen value and percentage of the variance explained by each factor. It also gives a communality of unity for each component. Communality is the variance in observed variables accounted for by common factors. The results were compared to the maximum acceptable level (MAV) of the Regulation (2013) and to other standards (WHO 2011; Indian Standard 2012) for drinking water purposes. The suitability of water for irrigation purposes was identified by the following irrigation indices: residual sodium carbonate content (RSC), magnesium hazard (MH), permeability index (PI), salinity hazard (SH), and exchangeable sodium percentage (ESP). The statistical analysis was carried out through statistical software packages as: descriptive statistic (SAS 9.2 2008), HCA and PCA (STATISTICA 8.0 2011), and graphical presentations through both statistical packages. The Piper was plotted with the help of AqQA (version 1.1.1) software.

Results and discussions

Hydrochemistry

The analytical results are shown in Table 1. Hydrogeochemistry of groundwater is characterized by the prevalence of calcium and bicarbonate as dominant cation and anion in Koprivnica-Križevci County of samples. The descending order of cations is as follows: ($\text{Ca}^{2+} > \text{Mg}^{2+} > \text{Na}^+ > \text{K}^+$) at sampling locations. The Na/Cl molar ratio greater than 1 typically suggests that the Na^+ ion was released in water from silicate by weathering and dissolution of silicate minerals, and it contributes to Ca^{2+} and Mg^{2+} ions in the groundwater (Gautam et al. 2015). With respect to anion descending order, it is as follows: $\text{HCO}_3^- > \text{NO}_3^- > \text{SO}_4^{2-} > \text{Cl}^- > \text{F}^-$ or $\text{HCO}_3^- > \text{SO}_4^{2-} > \text{NO}_3^- > \text{Cl}^- > \text{F}^-$. The HCO_3^- may be derived from soil zone carbon dioxide, parent rock materials (carbonates and silicates) through weathering, and dissolution (Singh et al. 2012, 2013a, b, Gautam et al. 2015). NO_3^- is normally considered as a minor anion in non-polluted waters hence their elevated concentration are indicator for the ongoing agricultural impact from the excessive use of manure or mineral fertilizers (Nemčić-Jurec et al. 2013; Nemčić-Jurec and Jazbec 2017). The maps for the spatial distributions of the hydro-chemical parameters are shown on Fig. 3a. Interactions between groundwater and surrounding rocks are mostly the main processes responsible for the observed characteristics of groundwaters on the locations (Belfar et al. 2017). Figure 3a shows an anthropogenic factor (higher concentrations of nitrate and chloride) controlling the groundwater composition on C1 location (Nemčić-Jurec et al. 2013; Nemčić-Jurec and Jazbec 2017).

Mg^{2+} and HCO_3^- are dominant ions at most of the sampling locations at district Allahabad. Cations are in the descending order of concentrations as follows: $\text{Mg}^{2+} > \text{Ca}^{2+} > \text{Na}^+ > \text{K}^+$, while anions are highly variable and the descending order depends on location. At most sites, anions have been following the descending order of concentrations as $\text{HCO}_3^- > \text{Cl}^- > \text{SO}_4^{2-} > \text{NO}_3^- > \text{F}^-$. Nitrate concentration was found low in groundwater samples at many locations of Allahabad which suggests low agricultural activities compared to Koprivnica-Križevci County of samples. The areas have more trees combined cultivating crops which may decrease the probability of groundwater contamination by nitrate. The studies suggest that even in areas with high nitrogen input despite the extensive use of fertilizers, and even well-drained soils (Nolan et al. 1998) have devoid of nitrate in the groundwater. Hubbard and Sheridan (1989) have suggested explanations for low nitrate in shallow groundwater that include dilution from precipitation, denitrification, and uptake by plants. Lasagna and De Luca (2016) and Eiche et al. (2016) have suggested the influence of oxido-reductive process on groundwater which results in low nitrate concentration. In Table 1, it was shown that standard deviation of nitrate is higher in India in comparison with Koprivnica-Križevci County. Generally, our results indicate higher variability of ion concentrations in Allahabad samples. The water chemistry of Allahabad water samples are controlled primarily by rock–water interaction along distinct subsurface flow paths. The Allahabad water samples are located lying in the thick alluvial plain, thin alluvial plain, and marginalized alluvial plain (hard rock region in southern part). The temporal variation is controlled by mixing-primarily dilution during rainfall periods that replenishes the groundwater of the area. It may be possible that seasonal variations of pCO_2 have played a role in temporal variation. Calcium, bicarbonate, and magnesium are derived from carbonate dissolution. The relatively high concentrations of magnesium indicate dissolution of very magnesium-rich calcite mineral (Meyer 1999). The contour maps of the spatial distributions of the hydro-chemical parameters at district Allahabad are shown in Fig. 3b. The interaction of soil–rock–water with human (weathering, geochemical reactions, alkaline nature of water, ion-exchange, and high variability in temperature range) mostly controls processes of elevated concentration of ions in the groundwaters of the study area (Singh et al. 2013a, b). The anthropogenic contribution of nitrate and chloride is high at few locations. The average elevated concentration of F^- was found at Koraon (or Koran), Bahariya and Karchhana. Weathering of limestone, dolomite, gypsum and anhydrite, and cation-exchange processes release Ca^{2+} and Mg^{2+} ions into groundwater (Garrels 1976). The possible sources of Na^+ in groundwater are the dissolution of rock salts and

Table 1 Descriptive statistics of physicochemical properties of groundwater samples collected during the 2011 and 2012 years (hot and cold season); concentration of ions in mg/l and EC in $\mu\text{S}/\text{cm}$; min—minimum; max—maximum; avg—average; std—standard deviation

Parameters	Year 2011								Year 2012							
	Hot season/Summer				Cold season/Winter				Hot season/Summer				Cold season/Winter			
	Min	Max	Avg	Std	Min	Max	Avg	Std	Min	Max	Avg	Std	Min	Max	Avg	Std
Croatia																
pH	7.0	8.3	7.8	0.52	7.3	7.9	7.6	0.21	6.5	8.3	7.4	0.81	6.9	7.4	7.1	0.22
EC	449	545	481	38	410	609	487	77	452	541	487	47	418	569	499	76
TDS	228	325	269	36	215	292	266	31	223	308	268	37	200	304	251	44
F ⁻	0.1	0.2	0.2	0.04	0.1	0.2	0.2	0.03	0.1	0.2	0.2	0.05	0.1	0.2	0.2	0.02
Cl ⁻	2.5	19.7	9.5	7.60	3.3	22.0	10.1	8.39	3.2	16.0	8.9	6.10	2.2	20.0	10.1	7.88
HCO ₃ ⁻	105	148	126	18	107	152	125	17	114	151	124	15	106	143	118	15
SO ₄ ²⁻	7.2	34.4	21.2	12.32	6.2	38.0	21.4	13.50	11.0	37.0	22.2	11.19	9.0	34.0	21.4	11.63
NO ₃ ⁻	1.5	29.9	13.6	11.85	2.6	22.0	12.2	9.01	1.7	25.0	12.5	10.29	2.1	30.0	14.5	11.87
Ca ²⁺	56.2	87.0	66.9	11.73	53.0	94.0	69.8	15.7	56.0	91.0	72.4	12.8	51.0	78.0	60.0	11.1
Mg ²⁺	15.0	26.3	22.3	4.33	14.0	22.0	19.2	3.11	17.0	25.0	19.2	3.35	14.0	22.0	17.4	3.58
Na ⁺	2.4	18.1	9.0	6.13	2.5	12.0	7.3	4.04	3.1	13.0	8.0	4.30	3.4	18.0	8.7	6.18
K ⁺	0.3	1.0	0.7	0.25	0.4	1.1	0.7	0.25	0.4	1.1	0.7	0.25	0.5	1.1	0.8	0.28
ESP	2.0	13.3	7.4	4.44	2.1	9.8	6.4	3.16	2.6	11.1	6.7	3.37	3.3	14.2	8.0	4.53
SAR	0.1	0.5	0.2	0.16	0.1	0.3	0.2	0.11	0.1	0.4	0.2	0.12	0.1	0.5	0.3	0.17
Hardness	221	314	258	36	190	317	253	48	214	305	259	38	185	264	221	32
RSC	- 1.0	- 0.2	- 0.5	0.31	- 1.0	- 0.2	- 0.5	0.38	- 0.9	- 0.3	- 0.6	0.21	- 0.4	- 0.1	- 0.3	0.11
PI	24.4	37.8	33.1	5.1	24.3	38.6	33.2	5.5	27.4	39.4	32.3	4.39	30.9	39.7	36.9	3.6
MH	27.9	41.0	35.5	5.7	26.0	34.7	31.5	3.47	25.6	35.2	30.6	4.18	26.4	39.3	32.4	5.48
India																
pH	6.8	7.9	7.3	0.22	7.0	8.0	7.4	0.22	7.1	8.0	7.5	0.29	7.1	8.7	7.5	0.42
EC	545	1289	829	156	476	1805	891	353	329	1270	634	214	326	1501	760	340
TDS	376	700	523	95	315	736	469	121	305	611	430	68	299	661	453	89
F ⁻	0.1	1.3	0.6	0.34	0.2	1.7	0.7	0.39	0.0	1.7	0.5	0.31	0.0	1.7	0.6	0.47
Cl ⁻	8.1	156.0	24.0	28.90	7.4	130.0	27.5	27.10	1.1	60.2	12.1	14.56	0.9	58.1	16.5	17.38
HCO ₃ ⁻	203	596	365	76	157	534	340	120	244	456	325	47	213	464	342	66
SO ₄ ²⁻	1.6	71.0	12.6	16.73	1.4	68.0	11.7	15.87	0.6	99.7	12.7	18.73	0.2	57.5	11.4	15.22
NO ₃ ⁻	0.3	1.1	0.8	0.25	0.1	1.0	0.6	0.25	0.1	2.0	0.8	0.51	0.1	9.8	0.9	1.73
Ca ²⁺	26.0	69.0	39.5	9.68	8.0	51.0	19.6	10.47	10.0	72.0	29.2	16.10	10.0	92.0	27.5	28.77
Mg ²⁺	12.7	119.0	51.1	20.42	10.6	69.0	31.5	14.67	16.9	49.2	30.2	7.38	17.9	65.3	33.1	11.32
Na ⁺	5.7	82.0	27.2	24.41	5.6	89.0	35.0	25.75	5.1	71.4	17.3	13.43	2.9	76.7	18.5	18.26
K ⁺	1.1	5.6	2.3	1.39	1.2	3.9	2.0	0.90	0.9	3.2	2.0	0.53	1.1	3.6	1.9	0.55
ESP	3.1	45.6	15.8	11.30	5.7	61.4	29.3	15.27	7.0	45.9	16.4	8.95	4.2	48.2	17.1	11.47
SAR	0.1	2.0	0.7	0.60	0.2	2.9	1.2	0.86	0.2	2.3	0.5	0.44	0.1	2.4	0.6	0.59
Hardness	143	574	308	94	81	311	178	63	124	324	197	44	103	447	204	92
RSC	- 0.4	8.2	2.9	1.45	0.9	7.8	3.8	2.07	- 1.8	5.3	2.8	1.66	0.5	6.3	3.6	1.39
PI	28.1	99.6	51.6	14.7	48.0	137.8	81.4	26.2	34.2	95.9	62.8	14.7	32.8	110.1	72.2	23.3
MH	33.3	85.2	66.4	8.84	43.2	91.2	70.8	13.7	40.5	83.0	64.2	13.5	32.1	87.4	71.1	16.4

weathering of sodium-bearing minerals. The HCO₃ concentration in groundwater is derived from carbonate weathering as well as dissolution of carbonic acid in the aquifers (Jeevanandam et al. 2007; Singh et al. 2012, 2013a, b).

Hydrochemical facies

The Piper's tri-linear diagram (Piper 1944) includes three triangles. Back and Hanshaw (1965) have defined the subdivisions of the diamond field that have represented

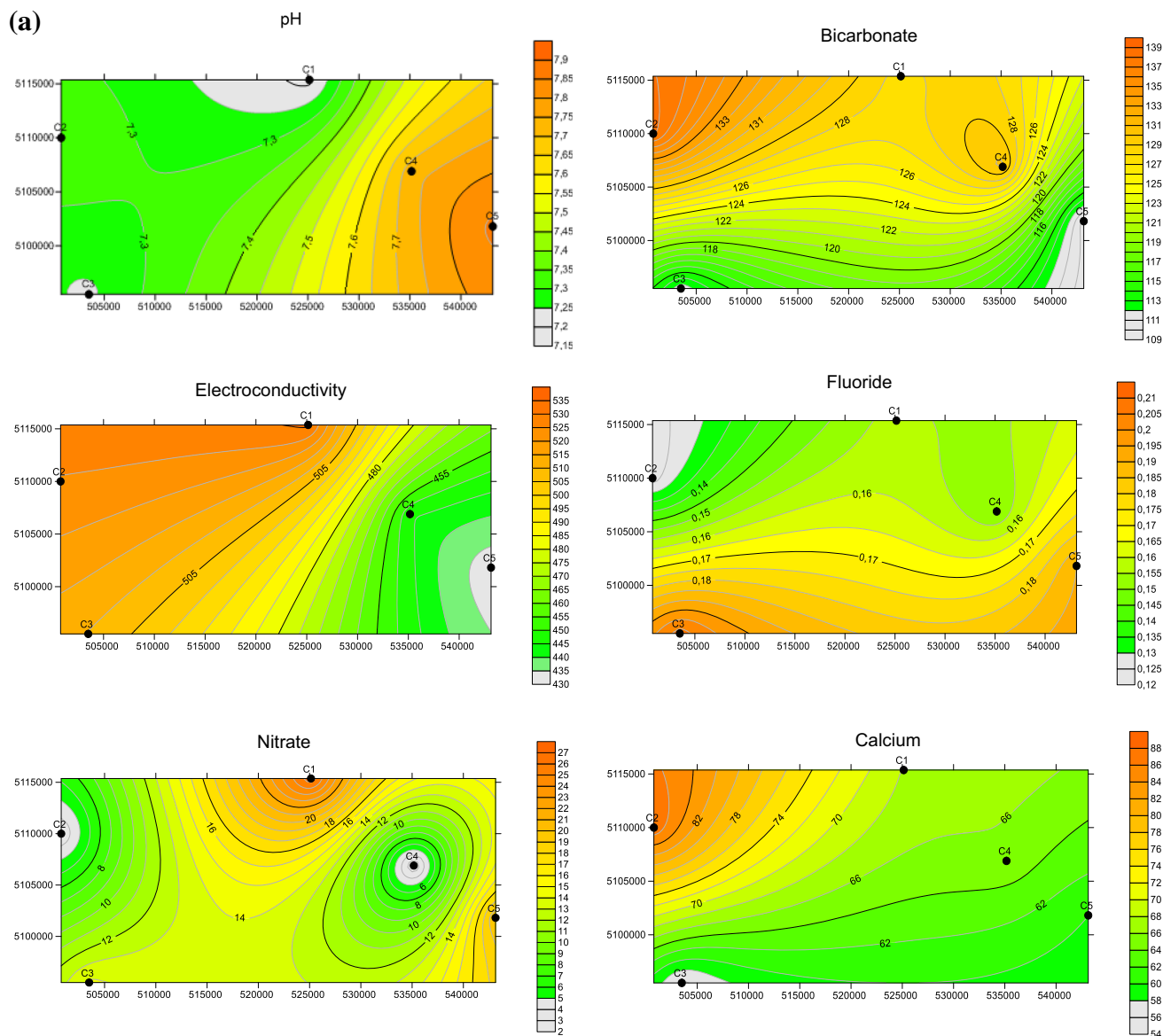


Fig. 3 Contour plots of hydro-chemical parameters of **a** Koprivnica-Križevci County and **b** Allahabad district

water-type or categories that form the basis for one common classification scheme for natural waters. Lithology, solution kinetics, and flow patterns of the aquifer control hydrochemistry of any facies. Based on the Fig. 4a, b, it is evident that the dominant water type of Indian sample is $Mg-HCO_3$ and at few location water type is $Ca-HCO_3$ or shallow fresh groundwater and $Na-HCO_3$ or deep groundwater influenced by ion exchange, whereas of Koprivnica-Križevci County it is $Ca-HCO_3$ or shallow fresh groundwater in all the season. The seasonal variations of few types are visible in Allahabad samples, whereas it is absent in Koprivnica-Križevci County samples. During summer and winter season of 2012, the sample of I1c and I8d was $Ca-HCO_3$, whereas the samples of I9a (summer) and I9b and I10b (winter season) of 2011 was $Ca-HCO_3$

and I14c, I26c, I27c, I28c, I29c (summer), I14d, I21d, I23d, I24d (winter) of 2012 was reported as $Ca-HCO_3$. The $Na-HCO_3$ water type was reported from I19a (summer 2011), and I21c (summer 2012), it is attributed to drop of water table, whereas I11b, I12b, I19b, I20b, I22b, I27b, and I28b in winter season of 2011 and I19d, I28d in winter season of 2012, this may be due to less infiltration of rain water. The chemical composition of $Ca-HCO_3$ water type may be evolved from silicate hydrolysis, cation exchange or a combination of both (Toran and Saunders 1999; Singh et al. 2013a, b). Few authors have also reported the involvement of irrigation return flow and anthropogenic activities (Jeevanandam et al. 2007; Gautam et al. 2015; Derby and Casey 2009).

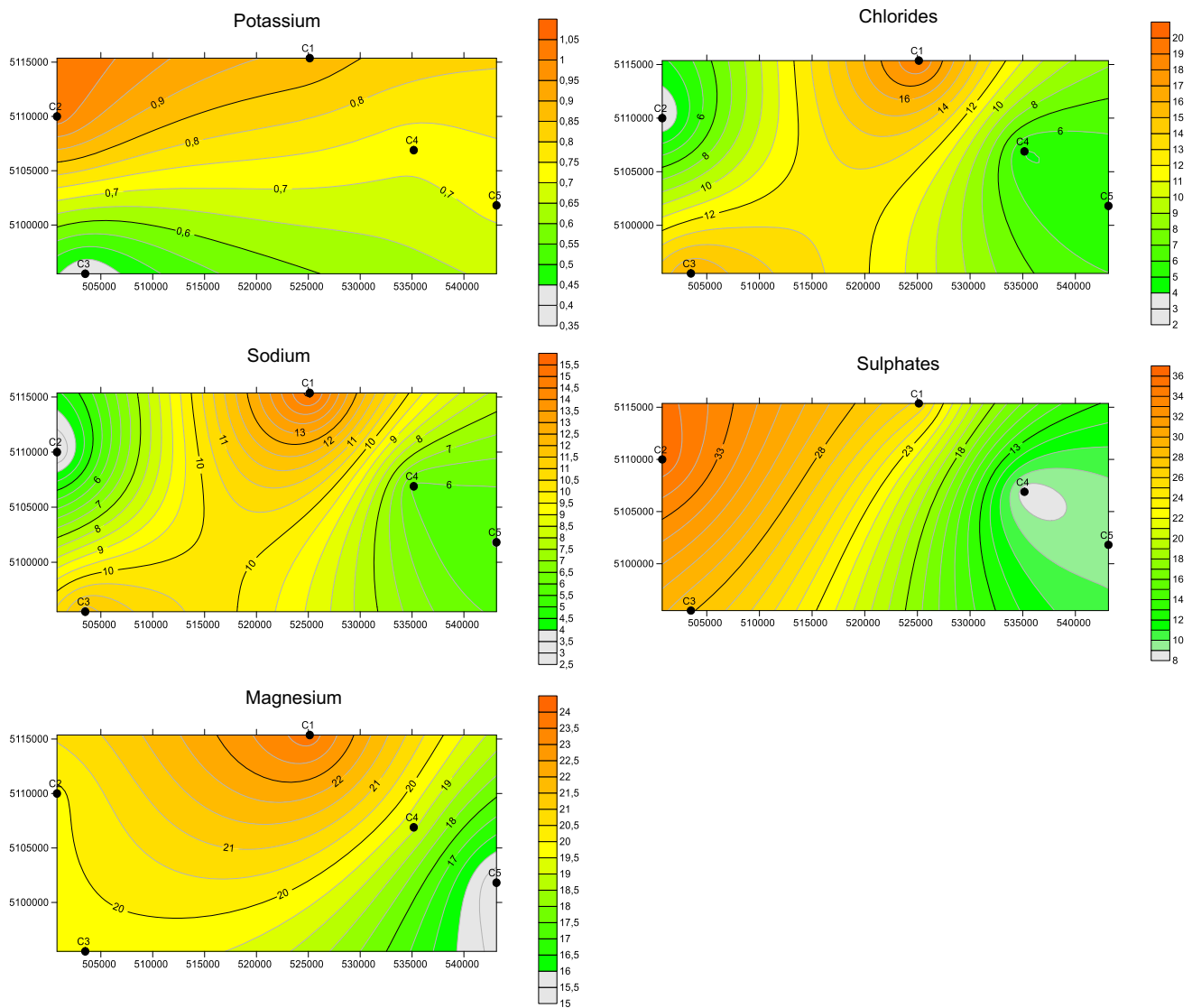


Fig. 3 continued

Gibbs (1970, 1972) has suggested a graphical representation based on these illustrations; there are three main mechanisms, which govern the chemical composition or chemistry of any waters. These are as follows: (1) evaporation dominance, (2) rock–water interaction or weathering dominance, and (3) precipitation dominance that control the chemistry of waters. The distribution of Koprivnica-Križevci County and Allahabad water sample points demonstrated that the major ion chemistry of the groundwater of two regions seems to be largely controlled through chemical weathering of rock-forming minerals and anthropogenic activities.

Gibbs diagram has illustrated (Fig. 5) that the samples of both the locations in different seasons of year 2011 and 2012 are weathering dominance. The water–rock/soil interaction is responsible for the chemical composition of

the groundwater in both the region. The chemical composition of samples does not originate from evaporation–crystalline and precipitation dominance.

Water quality for drinking water purposes

The chemical parameters of the groundwater were compared to the standard guideline values as recommended by the Regulation (2013) for the drinking and public health purposes (Table 2). Concentrations of all the measured parameters ($N = 480$) in the research area of Koprivnica-Križevci County in groundwater from wells of public water supply were below the maximum acceptable value (MAV) of Regulation (2013) and other standards (WHO 2011), indicating that the usage and consumption of groundwater should be of no concern with regard to inorganic pollutants.

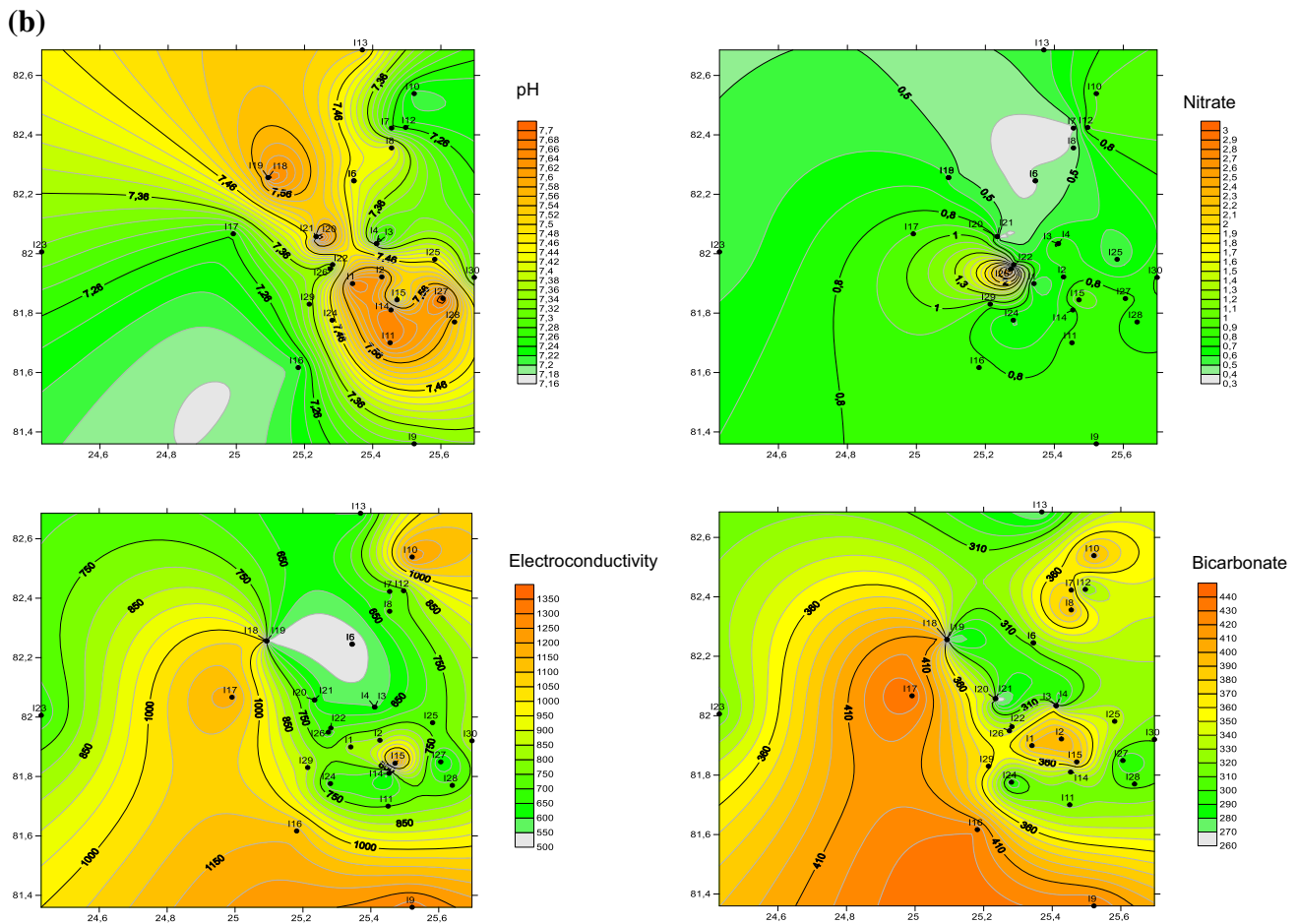


Fig. 3 continued

Water quality is safe for drinking water purposes. However, nitrate concentration in some wells was at or above the recommended level value (RLV) of Nitrate Directive (Council Directive 1991). It was shown (Nemčić-Jurec et al. 2013, Nemčić-Jurec and Jazbec 2017) that nitrate pollution is present in research area particularly in groundwater from shallow wells. Scientists have different opinions about the concentrations considered as pollution (Hallberg 1989; Jalali 2005; Kazemi 2004; Amano et al. 2016). We assume that the contamination is present in the research area and after an increase in the intensity of use of agricultural land, an increase in nitrate concentration above MAV followed. Eiche et al. (2016) have shown the influence of these processes. Nitrate concentration was higher in some of the investigated wells which are mainly located on the densely populated and agriculturally dominated areas compared to most other water sources.

Groundwater as a source of water supply has been utilized in India from ancient time, mainly for domestic needs and as well as for irrigation. Exploitation of groundwater has exceeded 80% in Gujarat, Haryana, and Punjab; 70% in Rajasthan, Tamil Nadu, and Uttar Pradesh; while in

Andhra Pradesh and Maharashtra it is 49 and 55%, respectively, and in the remaining states, it is below 40%, touching a low mark of 2% in Manipur (ARDC 1979). Over the few decades, shallow hand pumps are installed in large numbers in the rural and suburban region of the Gangetic Plains to over withdrawal and improperly used for the disposal of domestic and industrial wastes which adversely affected the quality of groundwater. In India, Allahabad, concentrations of all the measured parameters ($N = 1440$) from all groundwater samples were below MAV of standards WHO (2011) and Indian Standard (2012) (Table 2). In researched part of India in this paper, groundwater is safe for drinking water purposes with regard to inorganic parameters. Since water is not hygienic in all areas of Allahabad, the deterioration of groundwater quality is due to overexploitation (Kamra et al. 2002; Singh et al. 2002; Négrel et al. 2007) resulting in the higher salinity, fluoride (Vaishya et al. 2014), nitrate, iron, and other heavy metals in groundwater. Many researchers have assessed the groundwater quality in various parts of India (Prasad 1998; Kaul et al. 1999; Abbasi et al. 2002; Jagdap et al. 2002; Gupta and Deshpande 2004; Khaiwal and Garg

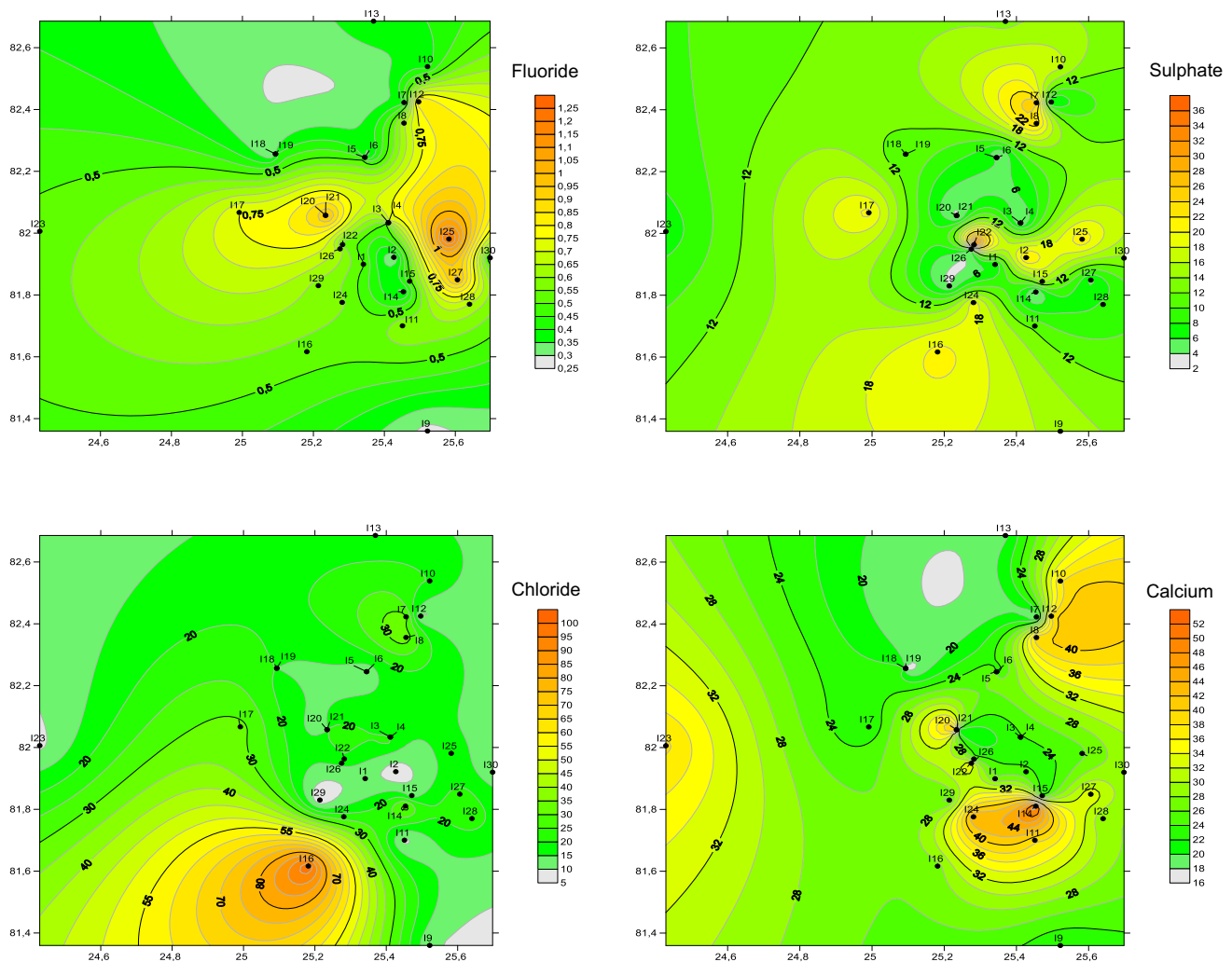


Fig. 3 continued

2006; Prakash and Somashekar 2006; Singh and Chandel 2006; Shivran et al. 2006; Bishnoi and Arora 2007; Gupta et al. 2008; Srinivasamoorthy et al. 2011; Singh et al. 2009, 2012, 2013a, b, 2015b; Kumar et al. 2014). It is important to examine similar researches in other areas of India. We consider that it is important to examine other parameters and indicators of water safety prescribed by the laws and regulations related to drinking water that are not examined in this paper.

Water quality for irrigation purposes

Residual sodium carbonate content (RSC)

The suitability of water for irrigation is often affected by the available quantity of bicarbonate and carbonate in the excess of alkaline earths ($\text{Ca}^{2+} + \text{Mg}^{2+}$) (Singh et al. 2013a). When the total carbonates and bicarbonates are in excess of total of Ca^{2+} and Mg^{2+} , there may be possibility

of complete precipitation of these ions in water. RSC index has been used to quantify the effects of carbonate and bicarbonate by the following Eq. (1).

$$\text{RSC} = (\text{CO}_3^- + \text{HCO}_3^-) - (\text{Ca}^{2+} + \text{Mg}^{2+}). \quad (1)$$

Higher value of RSC in water has led to fast adsorption of sodium in soil (Singh et al. 2013a). If the RSC values are > 2.5 meq/l, these water sources are considered to be harmful for the growth of plants, and < 1.25 are safe, whereas with RSC values > 1.5 – 2.5 meq/l water is not suitable for irrigation purposes (Table 3). All the water samples of Koprivnica-Križevci County have shown that these are suitable for irrigation, but for Indian samples very few are suitable for irrigation uses.

Magnesium hazard (MH)

The presence of high concentrations of Mg^{2+} in water has affected the soil quality by converting it into alkaline soil,

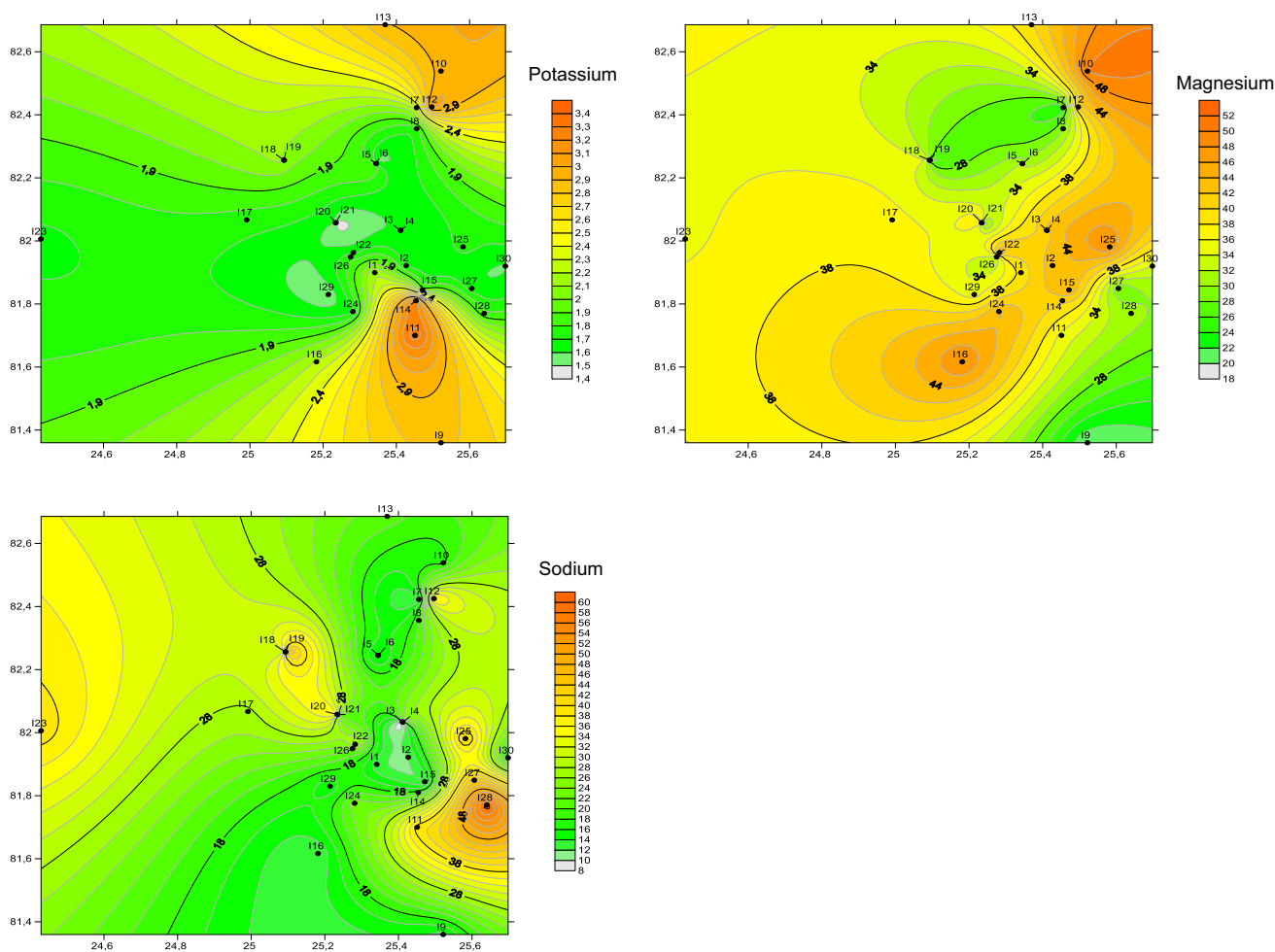


Fig. 3 continued

which means low crop yield (Gowd 2005). Szabolcs and Darab (1964) have proposed MH index to know identify whether the water sample is suitable for irrigation applications as defined by the following Eq. (2)

$$MH = \frac{Mg^{2+}}{Ca^{2+} + Mg^{2+}} \times 100. \tag{2}$$

The $MH > 50$ means that water is unsafe for irrigation use, whereas < 50 suggests that water is safe for irrigation uses (Table 3). As per this index, all the water samples of Koprivnica-Križevci County samples fall under safe category, whereas for Indian samples the majority of sample does not fall under the safe category. Only few samples fall in the safe category.

Permeability index (PI)

Porosity and permeability are the important physical properties of soils; the permeability is defined as the ability of the soil to transmit water and air. An impermeable soil is good for aquaculture as the water loss through seepage or

infiltration is low. The soil permeability is generally affected by the long-term applications of irrigation water as it is influenced by the groundwater which contains ions, such as Na^+ , Ca^{2+} , Mg^{2+} and HCO_3^- content of the soil (Singh et al. 2013a, b). Excess concentration of Na^+ leads to the development of an alkaline soil that can change the physical properties of soil by reducing the permeability (Raju 2007). We applied Doneen (1964) method of classification of irrigation water based on the PI. The expression of PI is given by following Eq. (3)

$$PI = \frac{Na^+ + \sqrt{HCO_3^-}}{Ca^{2+} + Mg^{2+} + Na^+} \times 100, \tag{3}$$

where, concentrations are given in meq/l. PI values > 75 , 25–75, and < 25 which fall in class I (safe), class II (marginally safe), and class III (unsafe), respectively, are shown in Table 3.

Based on Fig. 6, it is evident that all the Koprivnica-Križevci County samples fall under the category of marginally safe, whereas the majority of Indian samples fall under marginally and safe category in the Doneen’s chart;

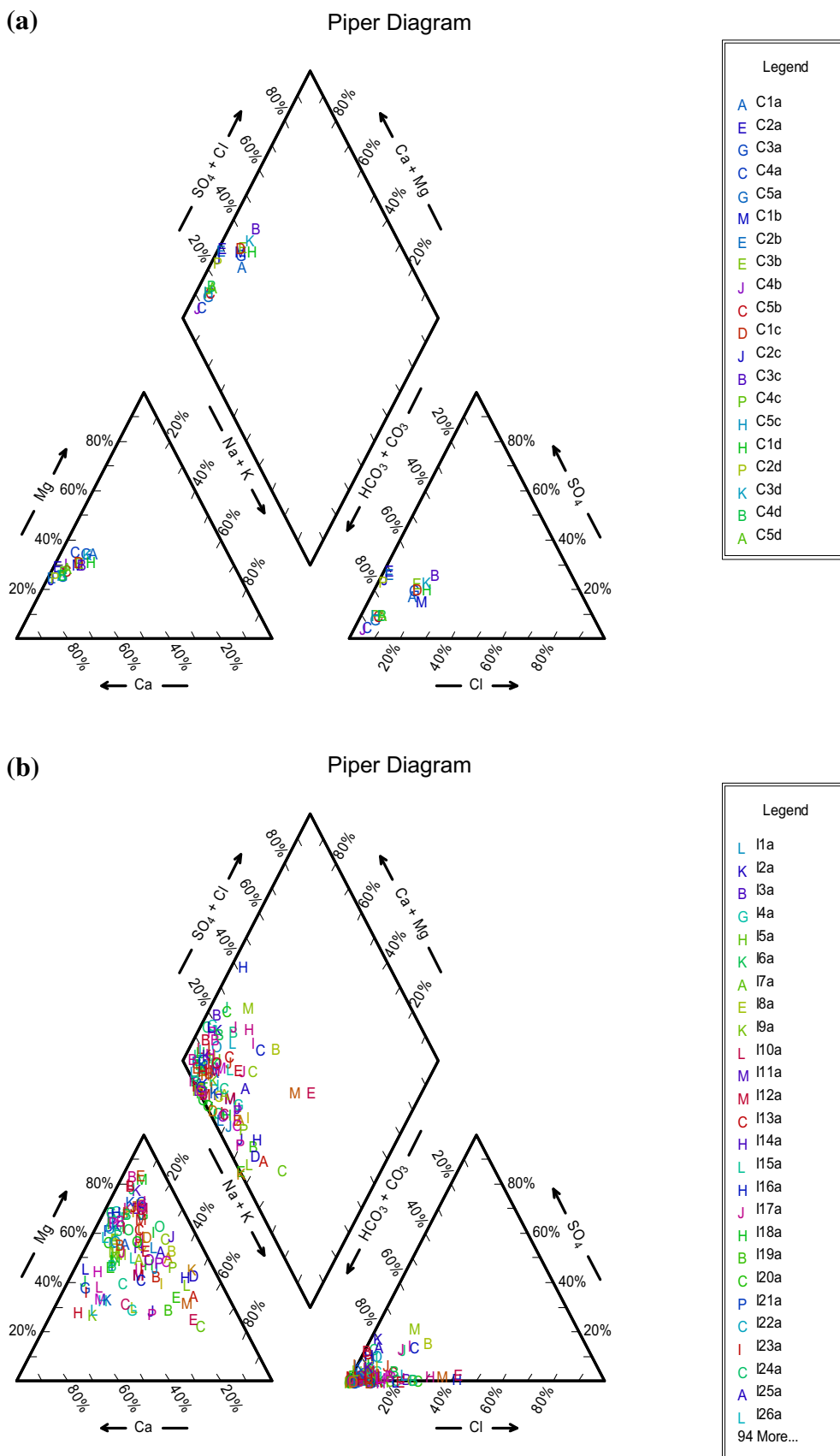


Fig. 4 Piper plots of groundwater chemistry of **a** Koprivnica-Križevci County and **b** Allahabad district

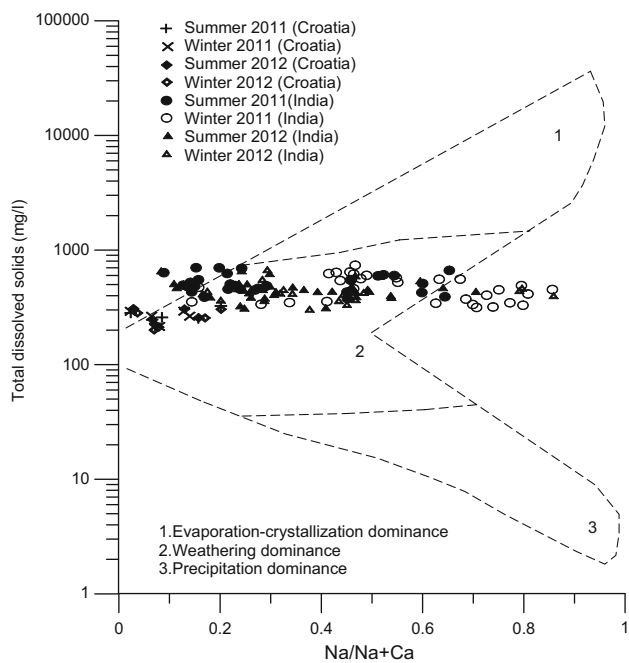


Fig. 5 Gibbs plot of groundwater samples indicating the origin of water types of the studied area

it implies that the water which has higher value of PI is safe for irrigation purposes.

Exchangeable sodium percentage (ESP)

The percentage of Na has been calculated using the Eq. (4)

$$ESP = \frac{Na^+}{Ca^{2+} + Mg^{2+} + Na^+ + K^+} \times 100, \quad (4)$$

where all the concentration units are in meq/l. The sodium reacts with soil to reduce its permeability; higher amount of sodium in water has reduced crop yield. Wilcox (1955) has used ESP and EC (Fig. 7) to classify groundwater and

divided into five categories. It is recommended that the ESP should not exceed 60% in water which is used for irrigation purposes. The ESP was lowest (1.95%) and highest (14.13%) in Koprivnica-Križevci County water sample. All the Koprivnica-Križevci County water samples have low concentration of Na% and fall in excellent to good category. The Indian water samples have lowest (3.14%) and highest (61.45%) ESP. Majority of Indian samples falls under the excellent to good category and few samples lie in the category of good to permissible. Hence, generally water is suitable for agriculture application.

Salinity hazard (SH)

Total concentration of soluble salts or salinity is an important parameter. The higher concentration of salts in groundwater will greatly affect the growth of plants by changing the physical properties of soil (soil structure, affecting soil permeability, aeration, texture, and makes soil hard) (Trivedy and Goel 1984; Ramesh and Elango 2012). The United States Salinity Laboratory (USSL) (Richards 1954) diagram demonstrates the combined effect of EC and SH, and categorizes the water samples for irrigation uses. According to Singh et al. (2015b), total concentration of soluble salts in irrigation water can be expressed for the purpose of classification of irrigation water as low ($EC \leq 250 \mu S/cm$), medium ($250-750 \mu S/cm$), high ($750-2250 \mu S/cm$), and very high ($2.250-5.000 \mu S/cm$) in salinity zones (Richards 1954).

The calculated value of SH in the groundwater ranges from 0.06 to 0.5 in case of Koprivnica-Križevci County water samples, whereas in Allahabad water samples it ranges from 4.19 to 61.45. The resultant plot of groundwater sample of Koprivnica-Križevci County and Allahabad on US salinity diagram shows that the groundwater majority of samples fall in the category C2S1 (medium

Table 2 Summary table of limits for drinking water parameters defined by WHO (2011), Croatian Regulation (2013) and Indian Standards (2012) BIS—10500: 2012; concentration of ions in mg/l and EC in $\mu S/cm$

Parameters	WHO (2011)	Indian Standard (2012) IS:10500		Croatian Regulation (2013)
	Health based guidelines	Acceptable limit	Permissible limit	Maximum available value
pH	6.5–8.5	6.5–8.5	6.5–8.5	6.5–9.5
EC	1400	–	–	2500
TDS	1000	500	2000	–
HCO ₃ ⁻	1000	200	600	–
SO ₄ ²⁻	400	200	400	250
Cl ⁻	250	250	1000	250
NO ₃ ⁻	50	45	100	50
F ⁻	1.5	1.0	1.5	1.5
Ca ²⁺	75	75	200	–
Mg ²⁺	50	30	100	–
Na ⁺	200	–	–	200
K ⁺	55	–	–	12

Table 3 Threshold limits for some irrigation water quality parameters (Ayers and Westcot 1994); concentration of ions in mg/l and EC in $\mu\text{S}/\text{cm}$ or dS/m

Irrigation problem	Units	Degree of restriction on use		
		Safe/suitable/none	Marginally safe/slight to moderate	Unsafe/harmful/severe
Salinity (affects crop water suitability)				
EC_w	dS/m	< 0.7	0.7–3.0	> 3.0
(or)				
TDS	mg/l	< 450	450–2000	> 2000
Infiltration (affects infiltration rate of water into the soil)				
SAR = 0–3	and $\text{EC}_w =$	> 0.7	0.7–0.2	< 0.2
= 3–6	=	> 1.2	1.2–0.3	< 0.3
= 6–12	=	> 1.9	1.9–0.5	< 0.5
= 12–20	=	> 2.9	2.9–1.3	< 1.3
= 20–40	=	> 5.0	5.0–2.9	< 2.9
Specific ion toxicity				
Na^+	SAR	< 3.0	3.0–9.0	> 9.0
Cl^-	meq/l	< 4.0	4.0–10.0	> 10.0
Miscellaneous effects				
Nitrogen (NO_3^-)	mg/l	< 5.0	5–30	> 30
HCO_3^-	meq/l	< 1.5	1.5–8.5	> 8.5
pH	–		Normal range 6.5–8.4	
RSC	meq/l	< 1.25	1.25–2.5	> 2.5
PI	%	> 75	25–75	< 25
MH	%	< 50	–	> 50

EC electrical conductivity, *SAR* sodium absorption ratio, *TDS* total dissolved solids, *RSC* residual sodium carbonate content, *PI* permeability index, *MH* magnesium hazard

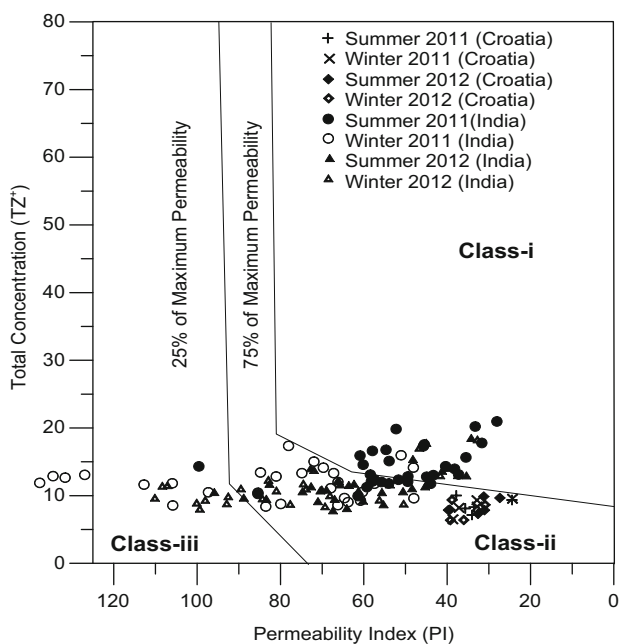


Fig. 6 Doneen plot of groundwater samples for hot and cold season in 2011 and 2012 of the studied areas

salinity hazard and low sodium hazard), and rest samples fall C3S1 (high salinity hazard and low sodium hazard) region (Fig. 8). The groundwater in the world is very often contaminated with one or more pollutants (Derby and Casey 2009; Amano and Nakagawa 2016; Liu et al. 2003; Sharma et al. 2016). Some pollutants degree with time and some remain unchanged. Researchers usually investigate within a country or within the neighboring countries (Ujevic et al. 2010, 2013; Noshadi and Ghafourian 2000), but we examined whether there is a similarity in the groundwater quality in some locations within two geographically different and distant countries (different geographical distribution). We estimate whether influences of local environment are more important predictor for groundwater quality than geographical distribution.

Cluster analysis

The groundwater in the world is very often contaminated with one or more pollutants (Derby and Casey 2009; Amano and Nakagawa 2016; Liu et al. 2003; Sharma et al.

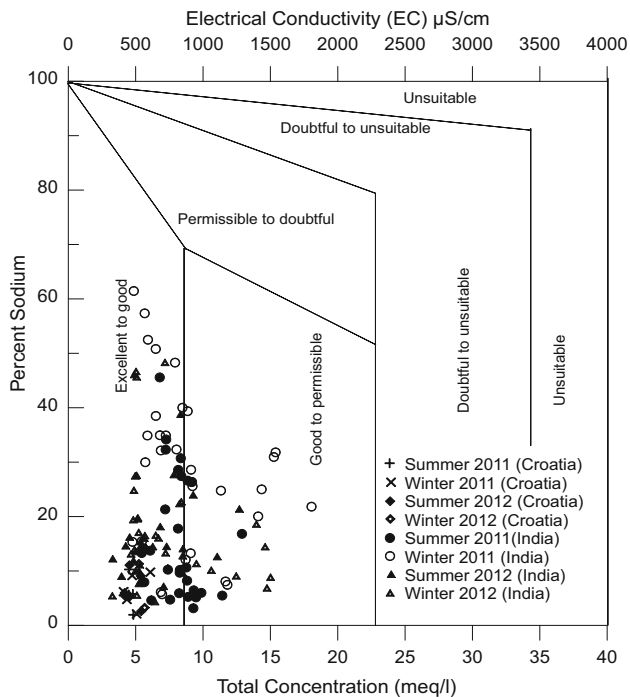


Fig. 7 Illustration of ESP of groundwater samples of the two study area

2016). Some pollutants degree with time and some remain unchanged. Researchers usually investigate within a country or within the neighboring countries (Ujevic et al. 2010; Ujević et al. 2013; Noshadi and Ghafourian 2000), but we examined whether there is a similarity in the groundwater quality in some locations within two geographically different and distant countries (different geographical distribution). We estimate whether influences of local environment are more important predictor for groundwater quality than geographical distribution.

Based on Fig. 9, four clusters (groups) with different hydrogeochemical characteristic can be distinguished. The first cluster includes the locations of Allahabad (India) (I9, I10, I15, I16, I17 and I18). In this cluster, groundwater conductivity ranges from 1007 to 1805 $\mu\text{S}/\text{cm}$. Bicarbonates dominate but as anions chlorides and sulfates are also present. As cations dominate magnesium but natrium and potassium are also present. Groundwater from these locations is rich in salts and the highest conductivity was reported. Kura et al. (2013) found that some ions HCO_3^- , K^+ , and Mg^{2+} and other major ions strongly influence the TDS and EC of groundwater. These locations (I9 and I10) come under the agricultural influences, whereas locations (I15, I16, I17 and I18) are mainly influenced by geological setting and hence anthropogenic and geogenic influences are contributing the elevated concentrations of these ions (Singh et al. 2013a; Nakagawa et al. 2016; Amano et al. 2016).

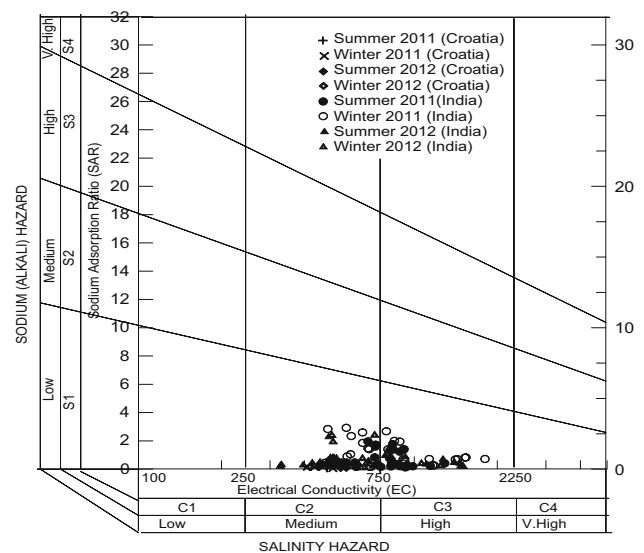


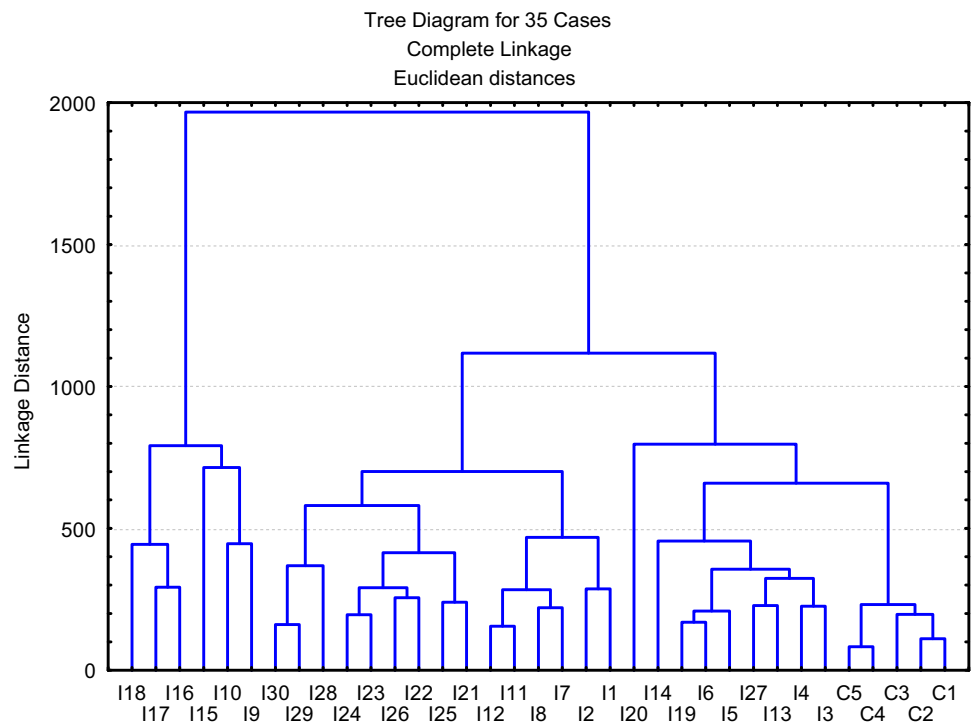
Fig. 8 Classification of groundwater samples of the two study area based on Wilcox

The second cluster consists of Allahabad locations (I1, I2, I7, I8, I11, I12, I21, I22, I23, I24, I25, I26, I28, and I29). Groundwater conductivity ranged from 484 to 1167 $\mu\text{S}/\text{cm}$. This monitoring groundwater sampling sites have alkaline pH, higher values of TDS, EC, NO_3^- , F^- , Cl^- , Ca^{2+} , Mg^{2+} , Na^+ , and low value of K^+ and SO_4^{2-} . The sites are at urban–rural fringe (I1, I2), under agricultural activities (I7, I8), urban residential area (I11, I12), area under agriculture and industrial dominance (I21, I22, I23, I24), and area under extensive agriculture (I25, I26, I28, and I29) (Singh et al. 2013a). Anthropogenic activities and geological settings have major influences on the groundwater quality but minor than in the first cluster (Hingston and Gailitis 1976).

Isolated third cluster of I20 forms due to EC, and it ranges from 330 to 951 $\mu\text{S}/\text{cm}$. The pH and F^- at this location is higher and the concentration of sulfates, calcium, and magnesium is low. This may be attributed to local geological settings and may be due to regional differences in rainwater salt composition (Hingston and Gailitis 1976).

The fourth cluster consists of both Indian and Koprivnica-Križevci County locations (C1, C2, C3, C4, C5, I3, I4, I5, I6, I13, I14, I19, and I27). Groundwater conductivity in this cluster ranged from 327 to 814 $\mu\text{S}/\text{cm}$. This group explains weathering or leaching of ions from sewage (Chidambaram et al. 2013a) and agricultural areas (Aris et al. 2007), anthropogenic pollution (Sundaray 2010), respectively. In this cluster, groundwater sample has similar mineralization on an average over all the seasons for both Indian and Koprivnica-Križevci County locations. Some other researchers (Eiche et al. 2016) have found that

Fig. 9 Dendrogram of cluster analysis of the concentration of all analyzed variables on locations for hot and cold season in 2011 and 2012



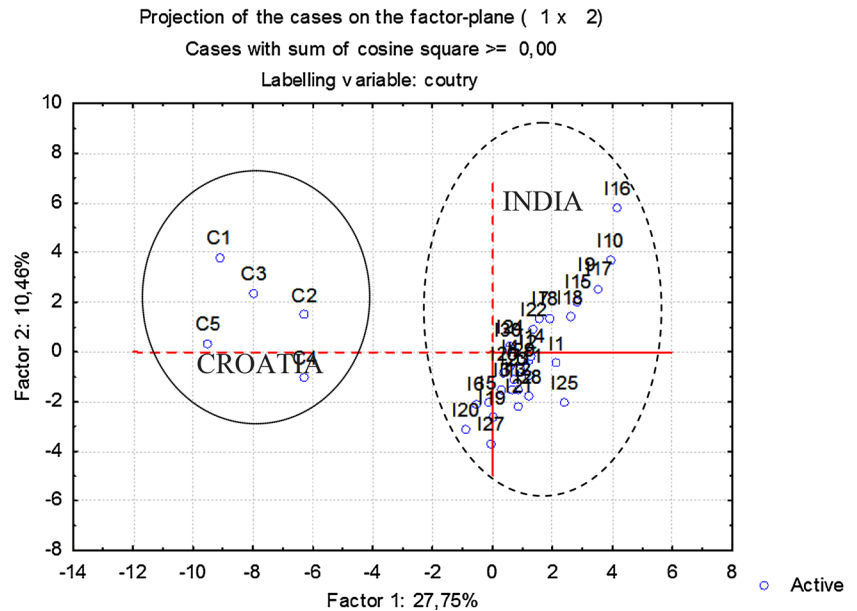
the mineralization as well as ion concentrations in some locations have shown a dependence on the season with decreasing values during rainy season. In comparison with Allahabad locations, nitrate concentrations are considerably higher in Koprivnica-Križevci County and ranges from 1.97 to 26.7 mg/l (Nemčić-Jurec et al. 2013; Nemčić-Jurec and Jazbec 2017). Allahabad samples of groundwater have shown a specific hydrochemistry with low nitrate of < 1 mg/l. In Koprivnica-Križevci County locations, C2 (hard rock) and C4 (alluvial) has shown low nitrate in groundwater. On C2 location, source of groundwater is 400 m deep (Mayer et al. 1996) and C4 location lies on the area rich in iron (Ujević et al. 2010, 2013). Since the groundwater in some parts of Koprivnica-Križevci County are generally anoxic and under reducing conditions (Ujević et al. 2010, 2013), reductive dissolution of iron (hydr) oxides can result in the release of iron or arsenic. Nitrate concentration of the groundwater gets also clearly influenced by the oxido-reductive process which is indicated by the low nitrate concentration (Lasagna and De Luca 2016; Eiche et al. 2016; McMahon and Chapelle 2008). In Indian case, sampling sites are located in the thick alluvial plain and few on marginally alluvial terrain. In study area, direct dumping of organic waste like untreated sewage is applied. Further, continuous use of groundwater causes soil salinization in parts of the aquifer(s) which is caused by the irrigation water return flow due to the mismanagement of the irrigation scheme and the overexploitation of groundwater resources. Over-pumping introduces excess dissolved oxygen that oxidizes immobile minerals, releases

some metals by reductive dissolution of (hydr) oxides, and increases metals in water. We assume that by oxido-reductive processes nitrate is reduced indicating low nitrate concentration on agricultural areas of Allahabad (Liu et al. 2003; Lasagna and De Luca 2016; Eiche et al. 2016; McMahon and Chapelle 2008). Dilution from precipitation, denitrification, and uptake by plants contributes to low nitrate in groundwater in Allahabad (Amano et al. 2016; Nakagawa et al. 2016). We presented all the processes on research areas but we have not revealed which processes prevail.

Principal component analysis (PCA)

To compare and confirm similarity or differences of groundwater quality from both countries with respect to average annual concentration of all analyzed variables (2011–2012), we used PCA (Fig. 10). PCA includes correlated variables with the purpose of reducing the numbers of variables and explaining the same amount of variance with fewer variables (principal components). The new variables created, the principal components scores (PCS), are orthogonal and uncorrelated to each other, being linear combinations of the original variables. They are obtained in such a way that the first PC explains the largest fraction of the original data variability, the second PC explains a smaller fraction of the data variance than the first one and so forth (Ielpo et al. 2012; Singh et al. 2015a, 2017). Our results show that first factor (all

Fig. 10 Projections of the groundwater sources from Croatia (C) and India (I) in respect to the first two factors (PCA), and this in respect to the concentrations of all analyzed variables on locations for period in 2011 and 2012



analyzed variables) explains 27.75% of the total variability and another factor (geographical distribution) accounts only for 10.46%. It means that the location inside country and environment nearby source of groundwater is more important predictor of groundwater quality than the geographical distribution. Love et al. (2004) in the two case studies determined that the difference in groundwater quality is due to all analyzed variables too. Except difference in groundwater quality, they showed contaminated and uncontaminated areas that are in accordance with our results in case of Koprivnica-Križevci County and India. The hydrogeochemical conditions of aquifer units are mainly controlled by the agricultural practices in Koprivnica-Križevci County case indicating high nitrate concentration in some locations. In Indian case, anthropogenic influences in some areas and influence on variability of groundwater quality and high mineralization and conductivity are determined.

Fig. 8 also shows how locations are grouped based on the average annual concentration of all analyzed variables indicating two groups: Koprivnica-Križevci County (C1–C5) and India (I1–I30) suggested that geographical distribution does not reveal the sources of the constituents and pollution (Liu et al. 2003; Kim et al. 2005) but contribute to the total variability of groundwater quality (Sharma et al. 2016).

Conclusion

Based on groundwater analysis in case of Koprivnica-Križevci County (Croatia), the water facies was Ca–HCO₃ type while in Allahabad (India) case mostly Mg–HCO₃ and

few waters type have reported Ca–HCO₃ and Na–HCO₃. The results suggests decreasing values of cation as follows Ca²⁺ > Mg²⁺ > Na⁺ > K⁺ in Koprivnica-Križevci County and Mg²⁺ > Ca²⁺ > Na⁺ > K⁺ in Allahabad. Anions follow decreasing order HCO₃⁻ > NO₃⁻ > SO₄²⁻ > Cl⁻ > F⁻ or HCO₃⁻ > SO₄²⁻ > NO₃⁻ > Cl⁻ > F⁻ in Koprivnica-Križevci County and HCO₃⁻ > Cl⁻ > SO₄²⁻ > NO₃⁻ > F⁻ in the most of locations in Allahabad. Groundwater resources of both the regions are controlled by different physical, chemical, and biological factors, whose contribution is oriented in the form of land use activities, bedrock geology and hydrogeological setting, climate of the area, and depth of the water tables. The suitability of groundwater quality for drinking water purposes is evaluated based on comparison with the standard guideline values as recommended by the Regulation (2013) and other standards (WHO 2011; Indian Standard 2012) in Koprivnica-Križevci County case. In Allahabad case, values were evaluated based on comparison with WHO 2011. Concentrations of all the measured parameters are indicating that the usage and consumption of groundwater should be of no concern with regard to inorganic pollutants. Water quality is safe for drinking water purposes in both countries.

The suitability of groundwater for irrigation is evaluated based on residual sodium carbonate, magnesium hazard, permeability index, salinity hazard, and sodium absorption ratio. Based on RSC, MH, PI, and SH, all the water samples of Koprivnica-Križevci County have shown suitability for irrigation. In Indian case, most of the sample fall in the suitable range for irrigation purpose but few samples are exceeding the permissible limits. Based on sodium absorption ratio, all the samples in Koprivnica-Križevci County and

Allahabad are falling under the excellent to good category and few samples in India lies in the category of good to permissible. Generally, water is suitable for agriculture application in the research area of Koprivnica-Križevci County and in the most locations of the research area in India. Hierarchical clustering analysis (HCA) resulted into four groups which difference is based on natural and anthropogenic factors. It has revealed that Allahabad samples are more influenced by weathering, rainfall, and anthropogenic impact generally that contribute to higher variability of groundwater. Elevated NO_3^- concentrations at few locations are reported from Koprivnica-Križevci County samples due to impact of ongoing agricultural practices. Although direct dumping of organic waste like untreated sewage is determined in Indian case, the values of nitrate are low. Dilution from precipitation, denitrification, oxido-reductive processes and uptake by plants contribute to low nitrate. Principal component analysis (PCA) showed that all analyzed variables have explained 27.75% of the total variability. Another factor (geographical distribution) accounts only for 10.46%. In case of comparison with Koprivnica-Križevci County and Allahabad in groundwater quality, we confirmed that the local environmental conditions are more important factor for groundwater quality than geographical distribution. In general, the application of rational irrigation practices and the control of aquifer overexploitation in India and good agricultural practice in Koprivnica-Križevci County are significant for a strategic management plan of sustainable groundwater resources management. The present study may be helpful for further similar studies concerning groundwater quality issues in these countries and in other distinctly located geographical areas. This is a very important prerequisite for the reduced risk for human health.

Acknowledgements Author SKS express his thanks to University Grants Commission, New Delhi, India for providing the financial support Project Grant number (F. no. 42-74/2013(SR)). All the authors express their gratitude to the Editor in Chief of Journal.

Appendix 1

See Table 4.

Table 4 Station code and sampling locations in Croatia ($N = 5$) and India ($N = 30$)

Sr. no.	Station code	Sampling locations
Croatia		
1	C1	Ivanščak
2	C2	Vratno
3	C3	Trstenik
4	C4	Delovi
5	C5	Đurđevac

Table 4 continued

Sr. no.	Station code	Sampling locations
India (all locations in Allahabad district)		
1	I1	Jhusi
2	I2	Saraye Inayat
3	I3	Hanuman Ganj
4	I4	Bida
5	I5	Handia
6	I6	Andava mod
7	I7	Habibpur
8	I8	Sahson
9	I9	Babu Ganj
10	I10	Phulpur
11	I11	Bamraulli
12	I12	Teliar Ganj
13	I13	Shantipuram
14	I14	Bara Thesil
15	I15	Shankargarh
16	I16	Mobiya mode (Nari Bari)
17	I17	Koraon
18	I18	Manda
19	I19	Meja
20	I20	Urwa
21	I21	Pandeora
22	I22	Naini
23	I23	Bahadurpur
24	I24	Jasra
25	I25	Bahariya
26	I26	Karchhana
27	I27	Soraon
28	I28	Holagarh
29	I29	Kaundhiyara
30	I30	Mau aima

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