

Life Cycle Assessment of a Plastic Part Injected with Recycled Polypropylene: A Comparison with Alternative Virgin Materials

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Abstract

Plastics recycling is becoming a common action to reduce our products and processes' environmental impact, and it is of the utmost importance to introduce circular economy strategies. However, for most of the diferent types of thermoplastics, recycling is not currently its usual end of life due to the technical difficulties in the sorting and recycling processes. This paper presents the complete life cycle assessment of an industrial component made with three diferent thermoplastics; two virgin thermoplastics typically used for similar parts in the market as Polyamide 6 and Polypropylene, and an alternative source of 100% recycled Polypropylene. All life cycle stages are included in the study. After carrying out the life cycle inventory, calculations of the environmental impact of each life cycle steps have been performed with ReCiPe 2016 EndPoint (H/A) v1.03/World and with IPCC 2013 GWP 100a v1.03 methodologies, comparing all three materials under the same conditions. A sensibility assessment has also been performed, calculating a worst-case scenario of the recycled material, and considering higher material acquisition distances. This study shows that recycled Polypropylene contributes to reducing the overall environmental impact of the component life cycle by 29.8% under ReCiPe, and by 42.8% under Carbon Footprint when compared to virgin Polypropylene. For the worst-case scenario, these reductions in the environmental impact of the component life cycle are also signifcant: 23.2% and 36.4%, respectively, showing that the use of recycled polymers is a key approach to reduce the environmental impact of plastic components.

Keywords Life cycle analysis · LCA · Recycling · Polypropylene · Injection · Post-industrial · Induction hobs

Abbreviations

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1 Introduction

The consumption of plastic materials, especially polypropylene, has increased exponentially in recent decades, and it is now stabilized in its historic maximum [[1\]](#page-9-0). This growth is due to the increase in technological possibilities for this type of material, the rise of consumption in society, and the new manufacturing techniques as additive manufacturing $[2-4]$ $[2-4]$. This culture of consumism makes it necessary to establish sustainable environmental management that optimizes the collection, separation, and handling of waste [\[5](#page-10-0)], but it also minimizes its use by modifying the social perception of

when it should be used [[6](#page-10-1)] and the affections it causes [\[7](#page-10-2)]. Otherwise, some studies foresee an unsustainable situation in the medium term $[8]$ $[8]$, both because of the accumulation of waste [[9\]](#page-10-4), and because of the signifcant environmental impact of certain specifc materials [[10\]](#page-10-5).

Consumer electronics, household appliances, and other large electrical devices account for much of the waste [\[11](#page-10-6)]. These appliances are often challenging to treat and require specialized treatment capable of separating materials that is not always carried out [\[12\]](#page-10-7). This is why the European Union published the Waste Electrical and Electronic Equipment (WEEE) Directive [[13\]](#page-10-8), which promotes the recycling, recovery, and even the reuse of materials from these appliances; and the WEEE2 [[14\]](#page-10-9), which classifes and diferentiates between appliances and components. It takes into account both internal components and packaging waste.

WEEE and WEEE2 are complemented by the European Directives on Registration, Evaluation, Authorization, and Restriction of Chemicals (REACH) [[15\]](#page-10-10) and the one addressing Restriction of Hazardous Substances (RoHS) [\[16](#page-10-11)], as well as the RoHS2 [[17\]](#page-10-12), which control the use of hazardous raw materials in electrical and electronic equipment.

Since 2019, the European Union has taken a more active role in defning a framework for the member countries' environmental activities [\[18](#page-10-13)]. This framework is to be transformed into an action plan to make the Union economies part of the environmental transformation [[19](#page-10-14)].

All these restrictions applied by the European Union afect directly to the product design process. These limitations are mainly focused on energy consumption and the availability of the required documentation to stablish the compliance with each regulation [[20](#page-10-15)]. The trend of product design considering environmental impact, also known as ecodesign [[21\]](#page-10-16), includes the environmental impact as another factor when making design decisions [\[22,](#page-10-17) [23](#page-10-18)], afecting the choice of raw materials, manufacturing processes [\[24\]](#page-10-19), assembly and disassembly methods [[25](#page-10-20), [26](#page-10-21)]. The choice of less polluting materials means less impact in terms of raw materials, but it must be accompanied by the objective of reducing the overall impact in all activities of the design [[27\]](#page-10-22). One of the options proposed by ecodesign to reduce the environmental impact is facilitating the separation and identifcation of materials and its recyclability. The ecodesign criteria recommend the use of dismountable joints whenever possible, as Maitre-Ekern and Dalhammar [\[28\]](#page-10-23), proposed as prevention against planned obsolescence. This also benefts the working time in the treatment centers and the their cost.

These environmental design criteria, put into practice at the end of the appliance's life, favors the trend towards "0 landfll" [[29\]](#page-10-24), which implies that all the materials that make up a product should be reusable or recoverable and recyclable [[30](#page-10-25)]. The European Union has also promoted strategies to push circular economy, and the valorization of plastic components [\[31](#page-10-26), [32\]](#page-10-27), and, therefore, companies are starting to defne strategies to introduce circular economy in their business [[33\]](#page-10-28).

In the European Union there are laws such as the Energyusing Products (EuP) [\[34](#page-10-29)] and the Energy-related Products (ErP) [[35\]](#page-10-30) that establish ecodesign requirements for products, especially regarding energy-efficiency. More ambitious approaches have been analyzed by some authors, such as Kiling et al. [[36](#page-10-31)], which proposes a method to integrate functional, economic, and environmental assessments to comply with legislation and achieve even more environmentally conscious designs.

Ecodesign is also directly related to the circular economy that studies the use and valuation of waste [[37\]](#page-10-32) and its effect on both the environmental and economic impact [[38,](#page-11-0) [39\]](#page-11-1). Other authors, such as Ungerman, study circular economy models applied to business models [\[40](#page-11-2)]. The materials recovered and reused, either in their original form or treated later, can reduce the problem of the amount of waste originated from consumption. It has been shown that energy intake is also lower in the recycling of certain elements concerning the energy required for their primary extraction [[41](#page-11-3)].

EEA producers and recyclers can share information on materials used due to the methodology proposed in the technical report published by the International Electrotechnical Commission, IEC/TR 62635:2012 [[42\]](#page-11-4). Thanks to this information exchange, more appropriate end-of-life (EOL) treatments can be used [\[43](#page-11-5), [44](#page-11-6)]. It also allows the calculation of a recyclability index and recovery rate of raw materials [\[45](#page-11-7)]. After that, the user can select the most appropriate working methods and the most signifcant components to recover. On the other hand, the REAPro method was proposed by the European Union to evaluate resource efficiency, identify critical points, and improve potentials [\[46](#page-11-8)].

Thermoplastic materials with fllers have diferent characteristics that make them difficult to reuse and use them again as raw materials [[47\]](#page-11-9), and are specifed in the design and manufacturing processes, either by their aesthetic fnish or by their characteristics and physical properties [[48](#page-11-10)].

Plastic materials can be recycled through two processes that can be compared in terms of results and their environmental impact: mechanical recycling and chemical recycling [[49\]](#page-11-11).

When recycling plastic by mechanical methods, the most critical step is the sorting process, as studied by Peeters with the case of diferent technologies of a single product [\[50](#page-11-12)]. Diferent polymers cannot be mixed if the objective is to obtain a recycled polymer with similar characteristics to the virgin one [[51](#page-11-13)]. The same applies to fllers, binders, and additives, which must be taken into account when classifying and separating materials [\[52](#page-11-14)]. To start with, this material has to be characterized in order to know the starting properties to be, then, crushed into small pieces and fed into the extruder. The size and shape of the raw material shaped in the form of chippings are critical for the recovery process and the industrialization of the procedure as described in the recycling process and installation patented by Rialti SpA [\[53](#page-11-15)]. The main sources of recycled material are textile fiber scraps and industrial waste. This procedure also includes mixing it with the necessary additives and fllers in the fnal product. The step in which these fllers are added to the raw material determines the result and afects water and energy consumptions. These materials or compounds are necessary to stabilize properties such as fuidity, fame resistance, or mechanical strength, among others. The mixture is made by adjusting the starting material so that the resultant one ofers the properties stipulated for the fnal product.

Chemical recycling does not require the proper cleaning of the components to be processed, not even the total separation of the plastics, which reduces the pre-treatment separation costs. This process is much newer and less technologically and industrially developed [\[54](#page-11-16)]. It decomposes the materials into monomers or shorter chains than the starting polymers [\[55\]](#page-11-17), which can be used for the subsequent polymerization into new thermoplastics.

The result of the recycled material will be a new compound, recycled by mechanical means, with physical properties (mechanical, rheological), that allows the material to be a substitute of the materials recycled to obtain it [[56\]](#page-11-18).

Depending on the origin of the thermoplastic wastes, two main groups can be described. Post-consumer and postindustrial materials.

Post-consumer materials are those extracted from waste separation, that is, materials from cities or companies that are, subsequently, separated in specifc waste treatment facilities [\[57](#page-11-19)]. The separation is usually done manually, as the diferent thermoplastics on the market may be similar in density or vary in their properties depending on the additives. Some materials, such as polyethylene terephthalate (PET) used in commercial bottles or containers, are easily identifable and recyclable since their use is widespread for certain products [[58](#page-11-20), [59](#page-11-21)].

Another source of post-consumer materials is the one of product recovery. Appliances such as household ones, air conditioners, telephones, etc., are managed mainly in two ways. Either they are taken to specialized treatment centers [\[60](#page-11-22)] where the diferent components are separated according to their material [\[61\]](#page-11-23) and their recovering feasibility; or they are crushed and separated later, looking for the most valuable materials and separating the rest. The materials obtained by separation in specialized centers are considered to be of better quality and purity than those extracted in the separation of domestic waste [[62\]](#page-11-24). But even so, they are unreliable as they can be subject to human error or cross-contamination [[63\]](#page-11-25). The

common agreement is that this is still a feld with signifcant room for improvement [[64\]](#page-11-26).

Post-industrial materials are those that come from industrial processes, so their traceability is easier [\[65](#page-11-27)], and are derived from excess production, product non-conformities, or other causes of waste in a product's manufacturing process which are not reusable in its primary process [[66\]](#page-11-28).

ISO 14040 [\[67](#page-11-29)] defnes the principles and framework of the LCA for a product. These principles include the defnition of the scope, the limitations of the assessment, and the interpretation phase. In addition, the requirements and guidelines for the study are listed under the standard ISO 14044 [[68](#page-11-30)].

The LCA methodology has been successfully applied to assess the impact of PP manufacturing [[69\]](#page-11-31) or other materials [\[70\]](#page-11-32). The methodology can also reveal the signifcant phases in which the impact can be identifed and reduced, such as the study carried by Zhang et al. on the minimization of the environmental impact by optimizing the cutting parameters [[71\]](#page-11-33). For example, this approach has also been implemented to reveal the most critical areas where not all inventory information is available thanks to sensitivity analyses [\[72](#page-11-34)]. Alternative use of the LCA aims to compare design solutions and the impact within the whole life cycle, such as the definition of the manufacturing material [[73,](#page-12-0) [74](#page-12-1)].

There are many examples of current LCA applied to assess the impact of products [[75](#page-12-2), [76\]](#page-12-3) or processes [\[77](#page-12-4)[–79](#page-12-5)]. Industry and its processes can also be assessed in order to analyze the impact on resources [\[80](#page-12-6), [81\]](#page-12-7) or industrial scenarios [\[82](#page-12-8), [83\]](#page-12-9), and are primarilyrelated to energy production $[84, 85]$ $[84, 85]$ $[84, 85]$ $[84, 85]$ and consumption $[86, 87]$ $[86, 87]$ $[86, 87]$ $[86, 87]$ $[86, 87]$. LCA can also be mixed with economic factors to include in the development decisions [[36\]](#page-10-31).

This paper will analyze the environmental impact of a functional unit depending on the material used. One of the proposed alternatives studied is 100% recycled, while the other two are diferent virgin thermoplastic alternatives commonly used for the same purpose. PA6 and PA66 have been widely applied to electronic housings. However, they are being replaced by other thermoplastics like PP due to economic reasons. Product designers modifed the frst concepts used in induction hobs to reduce the mechanical requirements of the materials. In the case of BSH cooktops, introducing a metallic chassis allowed the use of materials with lower mechanical properties, so PP replaced PA6 in most induction cooktops a few years ago after a comprehensive study to prove the feasibility of the material.

2 Materials and Methods

To validate the suitability of using recycled materials, two counter-type materials are studied. In this case, polypropylene with 40% talc and HB fre resistance characteristics,

which is the least fame retardant rating defned by a slow burning on a horizontal part. The comparison begins with rheological tests, and after confrming that both materials are comparable, samples for testing are injected [[88](#page-12-14)]. Lifetime and release tests are carried with the test samples. This comparison aims to establish that the properties of the recycled material are similar to those of the virgin material in order to study the environmental impact of the part manufactured with both raw materials.

After a brief comparative study, it was found that virgin and recycled material shared the range of mechanical and rheological characteristics required so that both were interchangeable (Table [1](#page-3-0)).

In the comparative environmental impact study, the functional unit is also analyzed with unflled Polyamide 6.

2.1 LCA Methodology

The comparison and analysis of the environmental impact were carried out by means of a LCA based on ISO 14040 [[67](#page-11-29)] and 14044 [\[68](#page-11-30)], and following the ReCiPe 2016 EndPoint (H/A) v1.03/World (2010) methodology. This approach allows the comparison of the environmental impact on diferent areas, and the diferent impact categories are weighted and unifed into a single value. This study's result offers a comparable result among different materials, products or services [\[89](#page-12-15)].

The study also calculates the Carbon Footprint following the IPCC 2013 GWP 100a v1.03 methodology. This method is also commonly used because of its social relevance, as countries should try to reduce their greenhouse gas (GHG) emissions in order to limit global warming and minimize its negative impact on ecosystems, food security, cities, and so forth [[90,](#page-12-16) [91\]](#page-12-17).

2.1.1 Goal and Scope Defnition

This study aims to compare the environmental impact associated with the manufacture of a plastic component present in the induction cooktops of a home appliance manufacturer. The study includes an analysis of three diferent raw materials and will also analyze the infuence of the raw materials on the functional unit's impact, the manufacturing processes, distribution, and its end of life (EOL).

2.1.2 Functional Unit

The environmental impact study of the recycled polypropylene was carried out on an injected component used in the $6th$ generation of induction cooktops manufactured by the home appliance producer. This component is defned in polypropylene with 40% talc, and the impact of its variant on unflled PA6 is also analyzed.

The part is then injected with three raw materials. Two polypropylene references from diferent suppliers and one PA6. Its dimensions are $534 \times 460 \times 35$ mm, with a thickness of 1.8 mm (Fig. [1\)](#page-3-1). In order to focus on the impact of the material and the processing, the functional unit is defned as one injected part, delivered to the fnal consumer, and including the EOL.

This part houses diferent power electronic boards and manages the electrical routing required for the correct performance of the appliance.

Due to the assembly's design, the efforts of the part and, therefore, its mechanical requirements are minimized. Its working position is fxed on a metal component that supports

Fig. 1 Plastic component injected in virgin PP

Table 1 Material properties

the mechanical stresses made by the components' weight and the compression of the internal springs. This makes it possible to use a thermoplastic with inferior mechanical characteristics and lower price compared to the PA6, which was traditionally used for this type of parts. The maximum temperature measured in the part in cases of extreme use is 73 ºC, so even thermal stresses do not require the use of special technical materials. In fact, those with higher usetemperature specifcations are needed in other parts of the hob due to the higher temperatures reached.

2.1.3 System Boundaries

Fig. 2 System boundaries

inputs, and outputs

The LCA limits are established from obtaining the raw material and its manufacture to its transport to the induction cooktop factory in Zaragoza (Spain) where it is assembled. After the complete appliance assembly, it is considered a road distribution by full truck according to the percentages of distribution by countries given by the home appliance producer. The main destination countries are, among others: Germany, France, Spain, and The United Kingdom. In the calculation of the environmental impact, the average route followed by an appliance is calculated (Fig. [2\)](#page-4-0).

2.1.4 Inventory Data, Cut‑Of Criteria, and Assumptions

Both polypropylene materials are injected under similar conditions, the same injection machine and the same mold. The injection parameters used are the same. The variant manufactured in PA6 uses the same tools, but it is adapted to the raw material requirements such as drying and injection parameters.

Manufacturing data for virgin PA and PP are extracted from the Plastics Europe publications on plastics production in Europe [[92](#page-12-18)]. Both materials are manufactured in Europe and, therefore, this data is representative for calculations and comparisons with recycled PP, whose calculation is the purpose of the paper.

The environmental impact value of virgin polypropylene is combined with the one of talc, which supposes up to 40% in terms of weight; being both values obtained from the Ecoinvent datasets.

To know exactly the energy consumption of the process, the machine's consumption during the automatic injection process has been measured (Fig. [3\)](#page-5-0) with a Circutor C80 power analyzer [[93\]](#page-12-19).

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All calculations have been made with the SimaPro 9 software [[94\]](#page-12-20) and the EcoInvent database v3.5 [[95\]](#page-12-21). The impact of the energy mix used is also calculated according to the country where the process takes place in.

3 Life Cycle Inventory

All the processes involved in the life-cycle of the functional units assessed are broken down into single processes that can be quantifed and compared. The impact of these simple processes can be quantifed thanks to material or process datasets, and will be later used for calculating the real impact of the functional unit.

Information for material or thermoplastic recycling is not included in such datasets and, therefore, it has been calculated in detail as the required data is available from the recycled material producer. This manufacturer of recycled raw materials has consumption measurements at all points of its manufacturing process to be accurate.

3.1 Polymer Manufacturing and Transportation to the Injection Plant

Values of the environmental impact of the virgin raw materials included in the assessment are obtained from an EcoInvent dataset based on Plastics Europe data [\[92\]](#page-12-18). However, transport values and distances are defned according to the real situation of induction cooktop components. Distances are selected from current suppliers of both raw materials and injected parts.

The manufacture of recycled material begins with the identifcation and transport of post-industrial raw material sources. These materials are transported to the producer by freight trains, freight ships and ">32 metric ton" trucks; depending on the origin of the post-industrial material. In this case, the manufacturing process of recycled polypropylene with 40% talc and HB fame resistance has been studied (E-RIALFILL H 07 40 T).

Firstly, it is established that the manufacturing process is the result of mixing three sources:

- Ground material, crushed and stored on the premises until use.
- Scrap from internal manufacturing processes that are not compliant but recoverable.
- Masterbatch, fllers, stabilizers, and other components needed to obtain the specifed product.

Materials with these three diferent origins are mixed in adequate proportions according to the manufacturer's expertise.

Under the description given in the patented process [\[53](#page-11-15)], the material is treated to allow its further treatment and to increase the efficiency of processing. Nevertheless, the complete diagram is not included due to the confdentiality agreement reached with the company.

The thermoplastic sources from the first points are homogenized and, if necessary, crushed to obtain sizes and shapes admissible by subsequent processes.

The quantities defined for each type of material are loaded into the production line, where they are extruded, cooled down, and cut. According to the technical data sheet, the result is 100% recycled material with mainly mechanical and other physical characteristics (Fig. [4](#page-6-0)). These properties are adjusted for each production batch according to the results of the previous analysis of the raw material in the input warehouse.

SCADA systems control each point in the manufacturing process so that there is total traceability both in terms of quantities of material and the energy which contributed to the system; and the result of the line in the fnal product, line stoppages, and inefficiencies. These systems are used to control processes parameters at a distance.

Fig. 3 Power consumption measurement in injection facilities. Detail of the power analyzer connection

Fig. 4 Post-industrial waste converted to recycled raw material compound

The quantities of material supplied by each of the dosing units of the extrusion line analyzed and, therefore, the composition of the recycled material, can be broken down as follows:

- 55% PP.
- \bullet 40% Talc.
- 2% Carbonates.
- 3% Masterbatch: additives, lubricant, stabiliser, etc.

For confidentiality reasons, we cannot specify the materials supplied as masterbatch in full detail, which are within the manufacturer's knowledge. These materials will be taken into account for the calculation of the environmental impact, but the manufacturing process will not be fully detailed.

The average production range of recycled raw materials with the defned manufacturing process is between 3000 and 4000 kg/h, while the average energy consumption is between 0.2 and 0.3 kWh/Kg. As said, although the calculations are made with exact values, they are not shown to protect the confdentiality of the process.

These values are obtained as the average calculated from the supplier's data during average production.

Once the material is manufactured, it travels from the supplier by truck to the injection plant. There it is stored in silos for later use. In all cases, it is considered a full " >32 metric ton" truck transport. The transport distances considered for each of the assessed materials are 1295 km for PA6, 1477 km for virgin PP and 1272 km for recycled PP.

To calculate the value of the recycled PP, it has been considered that the primary raw material has neither a positive, nor a negative impact in itself since it is a material that would have ended up in a landfll or incinerated. However, the transport to the recycling plant does generate an impact, just as the electrical energy or other sources do, such as transport and auxiliary components necessary to calculate the total impact of the material produced in the plant.

3.2 Injection Molding Process

Once the impact of the raw materials is known, it is introduced into the analysis of the whole piece defned in the scope of the study.

The injection process uses the same moud and injection machine Engel Duo 550 [\[96](#page-12-22)] for all three materials. The differences in the process will be only those related to the specifc raw material. These diferences are analyzed, including the consumption of the necessary instrumentation.

In order to evaluate this process, electricity consumption measurements were collected during the injection of the various materials.

To avoid consumption peaks that would distort the measurement obtained, they have been carried out with the previously detailed equipment for periods of several hours in automatic mode [[97\]](#page-12-23).

The results attained are compared per injected part, as the density of both materials is slightly diferent, which means a variation in the fnal weight of the piece (Table [2](#page-7-0)). Data of the energy consumption of the PA6 processing includes the consumption during the drying of the material necessary before the injection.

3.3 Transportation to the Assembly Plant and the Customers

After the injection process, the part is manually handled for packaging and subsequent shipment. The supplier that injects the part is located at a distance of 9 km from the manufacturer's induction cooktop assembly plant in Montañana, Zaragoza. A ">32 metric ton" milk-run truck, shared with other components, does this transportation to the assembly plant.

After the assembly of the appliances, the distribution scenario of the fnal product is considered in all cases, but the diferent weight of each part is taken into account. Freight trains, freight ships, and ">32 metric ton" trucks,

are used depending on the country of destination. The distances are obtained as the average of the values given by the logistics department of the home appliance manufacturer, and represent the average, in km, carried out by an average worktop. This value has been weighted according to the volume of worktops sent to each country.

3.4 End of Life

The EOL of the device and its components are considered and, for this purpose, the average values for recovery and disposal according to IEC-TR 62635 [\[42\]](#page-11-4) are also taken into account. These values state diferent percentages of recycling, landfll disposal, or incineration for each raw material type.

As the part understudy will be treated with the EOL of the product (an induction hob), and the most typical EoL treatment is shredding, it has been considered that the part will not be recycled. The IEC-TR 62635 considers nonrecyclable plastic parts with flling, as these materials are required to be manually separated, a scenario that is not always fulflled. Therefore, the most restrictive scenario has been considered, with 5% incineration and 95% landflling, as proposed by IEC-TR 62635.

4 Results

Results compare the impact under both methodologies. The diferent materials are presented broken into the different processes studied. Firstly, a base case is assessed comparing the results of the functional unit (one injected part, delivered to the fnal consumer, and including the EOL) under the three diferent raw materials at each step of the complete process.

Once diferences are analyzed, a worst-case is defned and assessed in order to evaluate the variation of the results under diferent conditions.

Table 3 ReCiPe impact of one injected part delivered to the fnal consumer and including the EOL

4.1 Base Scenario Results

After assessing the functional unit, results show a reduction in the overall values of the ReCiPe environmental impact of 29.8% compared to virgin material and 74.1% compared to PA6. This aggregated reduction will be separated into diferent processes to obtain a better insight into the reduction potential in each situation.

The same process is held with IPCC or Carbon Footprint. The reduction of the impact when using recycled materials represents 42.8% compared to virgin materials and 81.2% with PA6.

After the required calculation of the single material impact, the analysis compares the impact of the functional units' complete life cycles in the three diferent cases.

Using recycled material reduces the environmental impact in comparison with virgin alternatives. Breaking the result down on the diferent stages of the process included in the inventory is possible to know each process's contribution to the complete impact (Table [3](#page-7-1)).

The main impact relates to the production of the raw material. The production of PA6 is signifcantly higher compared to the production or recycling of the PP but, in all cases, it represents the highest percentage of the result. For virgin PP material production represents 48% of the impact, 81% for PA6, and only 27% for recycled PP.

The relevance of the transport processes is remarkable. Adding all afected transportation, the accumulated portion of PA6 transportation is 6% lower due to the higher overall impact. For the virgin ones, PP represents 20%, and for recycled PP 28%. This explains the signifcant impact of transport-related processes [\[98\]](#page-12-24). However, this impact can be reduced by either selecting closer sources of materials and suppliers or improving transport selection [[99](#page-12-25)]. Transportation impact is also afected by the weight of the functional unit. PA6, with its lower density, refects a lower impact in the LCA phase.

The analyzed material for recycling comes from an extensive provider network, as it arrives from several countries. Achieving a local supplier or, at least, increasing their contribution to the mix of processes material would reduce the average transport and, therefore, the associated impact.

The diferent kinds of thermoplastic afect the impact of the processing and injection process, as it can be seen in the case of PA6 given that the electric consumption required for the injection is higher, and requires a previous drying stage which is not needed in the other materials.

Regarding the carbon footprint, the comparative impact shows a signifcant reduction when using any of the studied PP (Table [4](#page-8-0)).

The use of recycled material reduces the result of $CO₂$ eq in 81.2% compared to the use of PA6. The impact is reduced in 42.8% when compared to virgin PP.

The most damaging processes remain the same as in the case of ReCiPe methodology. Material production accounts for 3.824 kg $CO₂$ eq. of the PA6 functional unit material production, representing 86.7% of the total. For virgin PP the material production casts 0.868 kg CO₂ eq., 60% of the total impact, and for the recycled PP, 0.25 kg CO₂ eq., only 30% of the total impact.

Values are similar in all three cases due to transport. The diferences appear because of the diferent distances traveled in each case, and the diferent materials' density. However, they account for a diferent percentage due to the diferent overall impact. In the case of PA6, transport-related emissions weight 6.4%, and the value increases for virgin PP (22%) and recycled PP (37.6%).

4.2 Worst Case Scenario Analysis

The result obtained from the study is calculated based on characteristics and data obtained in an average real case.

In the study carried out, the existing variation in the recycled polypropylene's origin has been verifed. The study was carried out with average values, but given the wide variety of suppliers, it is possible to anticipate diferent environmental impact values depending on the batch used to manufacture the material.

It was decided to evaluate the piece's environmental impact by modifying the scenario for obtaining the raw material by the recycled material producer (Worst Case or

Table 4 Global warming potential IPCC [Kg $CO₂$ eq.] of one injected part delivered to the fnal consumer and including the EOL

IPCC [$KgCO$, eq.]	PA6	Virgin PP	Recycled PP
Material production	3.824	0.868	0.250
Material transport	0.063	0.078	0.068
Injection molding process	0.180	0.131	0.135
Transport to final client	0.223	0.240	0.243
EOL.	0.119	0.128	0.130
Total	4.410	1.444	0.827

WC). In the scenario proposed, the material is obtained from a more distant source, such as textile production plants in Turkey. The transport proposed is a combination of road and sea transport to the production plant in Italy.

Of all the sources of impact analyzed in the study, the only one afected in this new scenario is the manufacturing of the recycled polypropylene (Table [5\)](#page-8-1), since the rest of the values remain constant, and it is considered that the modifcation of the origin of the material does not afect the mechanical properties or the electrical energy required for its processing.

The only value modifed is the impact of the material production, which includes the impacts of transporting postindustrial waste to the recycling plant.

The Carbon Footprint results (Table [6](#page-8-2)) show a rise in the overall impact again. This increase focuses on the production of the recycled material, which includes the necessary transportation of the selected waste to the recycling plant.

The analysis result shows an increase of 9.5% in ReCiPe impact and 11.2% in Carbon Footprint.

Taking into account these calculations, the maximum distances of PP material acquisition that make the recycled results equal to the virgin PP alternative can also be obtained. Following ReCiPe, these theoretical maximum distances would be around 9400 km if all the transportation was carried out by $a > 32$ metric ton truck, and more than 50,000 km if it was performed by freight ship. With IPCC methodology, these theoretical maximum distances are even higher (around 12,000 km and more than 140,000 km, respectively). All these distances are theoretical and clearly not realistic. Local sources will always be preferred due to their lower environmental impact. However, these high theoretical distance values show that, for the studied process, PP recycling is an interesting option from an environmental point of view, almost regardless of the origin of the materials (as there are relatively low diferences between the base scenario and the worst-case scenario already showed).

Table 5 ReCiPe impact of one injected part Base Case vs WC

ReCiPe [mPt]	Recycled PP	Recycled PP WC
Material production	11.29	15.23
Total	41.63	45.57

Table 6 Global warming potential IPCC [Kg $CO₂$ eq.] of one injected part Base Case vs WC

Obviously, this statement would not apply from an economic analysis, which is out of this paper's scope.

5 Conclusions

Diferences in results demonstrate the importance of materials selection in the design process. The environmental impact of a single part can vary up to 30% only depending on the chosen supplier of material and up to 75% depending on the type of thermoplastic selected.

The recycled material can be an option to replace common virgin materials reducing by 30% the environmental impact of the material. This replacement can be executed directly when the mechanical requirements comply with the datasheet of the substituting material.

Using recycled material reduces the impact of raw material manufacturing by 61%. Even when the worst-case scenario is assessed, the reduction is a signifcant 48%. This focuses on the impact reduction in the recycling process, as there are no signifcant diferences in other processes, but only those related to the distance travelled or the diferent density among materials.

Transportation of the materials to the injection plants brings signifcant damage to the environment, as seen in the sensibility assessment. These afections are reduced by 9.5% by selecting local suppliers. When there is no possibility of reducing transport, the impact can be optimized either by modifying the transport processes (train or boat instead of truck for long distances); or by distributing the impact in a more signifcant share of material (transporting raw material in full containers or at full capacity instead of moving fnal parts which are more delicate and in need of more space). This mainly afects the Carbon Footprint results (11.2%), that emphasizes the use of fossil fuels in processes and transport.

The impact of the injection molding process remains at similar values among all materials, as electric consumption is the primary source of impact. However, PA6 requires a previous stage of drying which implies an additional step and increases the process impact by 33% compared to the recycled material.

Regarding Carbon Footprint, which is more afected by the electrical consumptions, the injection is the most variable stage of the process after material production, which varies by 37.4% comparing the use of PA6 to virgin PP. The slight diference between virgin and recycled PP is due to the diferent melting temperatures and mold conditioning in the process.

The EOL stage causes the same impact regardless of the chosen thermoplastic as the most conservative approach, landflling, has been selected for all of them. Most plastic materials, especially if they included fllers, are not usually recycled and can only be landflled or incinerated, as the IEC/TR 62635:2012 shows [[42\]](#page-11-4).

The standardization of the use of recycled thermoplastics is the best way to develop a material waste market with the potential to be recycled. The European Union is creating plans to promote circular economy and waste valorization [[19\]](#page-10-14) as one of the main areas of the European Green Deal [[18\]](#page-10-13).

The design of the complete system must take into account legal requirements that can afect the thermoplastic selection. For the same functional unit, and depending on the raw material, the impact reduction can vary by 50% comparing two raw materials, and by 74%, when the selected thermoplastic is exchanged and when using recycled material. As the presence of renewable energies is increasing in the electricity mix, the environmental impacts of recycled polymers will decrease in the future, as the electrical consumption is relevant for the recycling processes.

After the assessment of post-industrial waste, future research lines would include the evaluation of post-consumption thermoplastic waste. It also opens the analysis of a closed-loop cycle, which would involve re-using the waste generated by the same products in a circular economy aligned scenario.

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Declarations

Conflict of Interest On behalf of all authors, the corresponding author states that there is no confict of interest.

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