



# Bioaugmentation and phytoremediation wastewater treatment process as a viable alternative for pesticides removal: case of pentachlorophenol

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## Abstract

This study focused on the potential for pentachlorophenol removal by a biological process in secondary treated wastewater (STWW). The proposed process is a combined method of phytoremediation using a native plant, *Polypogon maritimus* and *Lemna minor*, and bioaugmentation using a fungus. The bioaugmentation process was performed by a fungal isolate capable of removing PCP, isolated from the compost. The identification of the fungus was performed by morphological, biochemical, and molecular methods. A biological treatment system by bioaugmentation and phytoremediation was set up to estimate the capacity of this process to eliminate a high concentration of PCP. physico-chemical parameters, such as pH, COD, and BOD were tested at experimentation times T<sub>0</sub> (initial) and T<sub>f</sub> (final). The concentration of PCP is controlled by the HPLC method. Thus, the growth of the fungus was determined by spectrophotometry and enumeration on the agar medium. The results obtained show that the isolated and selected fungus is identified by *Penicillium Ilerdanum*. The fungal strain used has a significant capacity for tolerance and elimination of PCP. The results of the physico-chemical parameters showed an improvement in the quality of wastewater after the treatment was carried out. The elimination of PCP came with a release of Common law- and an important decrease in the DOC value in the STWW. The results obtained show that the *Polypogon* treatment shows a significant elimination of PCP by a percentage of the order of 92.01% and 23.58 g. L<sup>-1</sup> chloride concentration. The macrophytes used showed a better ability to tolerate and eliminate PCP with an increase of chlorophyll and its longer sheets.

**Keywords** Wastewater · Bioaugmentation · Phytoremediation · Pentachlorophenol

## Introduction

Environmental pollution because of hazardous waste, organic pollutants, and heavy metals has damaged the natural ecosystem by man [1, 2]. Surface waters are vital for supporting people and ecosystems; however, freshwater availability is under increasing pressure because of a growing human population requiring access to safe water [3, 4]. The surface water pollutants that are of concern and where remediation solutions are being developed. Water pollutants can be broadly categorized as either organic, e.g., hydrocarbons, pesticides, and algal toxins, or inorganic (metals or synthetic and manure-based fertilizers containing excess amounts of N and P, or biological [pathogens and algal toxins] [5]. The current biological treatment plants in the most developed countries are quite effective in removing

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the main pollutants from water. The wastewater treatment process requires the use of biological processes [phytoremediation, bioaugmentation, biostimulation...] which play an important role in the removal of micro-pollutants [6]. Aquatic phytoremediation is a phytotechnology used for the removal of pollutants from surface waters and the restoration of affected water bodies [rivers, streams, lakes, ponds] [7]. In surface waters, plants can be grown to remove pollutants from water and sediment [8, 9] and can be deployed either at the point source or in water bodies where pollution is diffused [10]. Thus, the process of phytoremediation can be used both for sustainable sanitation and as a management strategy for nutrient recovery from water [5]. Emergent and floating macrophytes primarily absorb nutrients and other contaminants through their roots and stem tissue [11]. Thus, Lemna macrophytes have shown high potential for pesticide removal example: *L. minor*, *L. gibba*, and Lemna [12, 13]. Besides, this phytoremediation process depends on microbial activity such as those associated with roots, which represents an important factor in the phytodegradation of pollutants [13–16]. These include bacteria that enhance macrophyte activity through their contribution to nitrification and denitrification and biological mineralization of organic phosphorus [17]. Also, through the process of bioaugmentation, microorganisms are found with the ability to degrade or convert hazardous compounds into less toxic compounds [18]. The bioaugmentation process also allows for the identification of microorganisms capable of degrading or converting hazardous compounds into less toxic ones [18]. According to the literature, most fungi are robust organisms and more tolerant to high levels of pollutants than bacteria, which is justified and supported by their ultra-sophisticated morphological and biological structure [19]. Fungi such as *Phanerochaete chrysosporium*, *Trametes versicolor*, *A. sydowii*, and *Pleurotus* are promising agents for the remediation of PCP-contaminated sites, as reported by Ryu et al. [20] and Zhou et al. [21]. *Lentinus crinitus* CCIBt2611 also exhibited the ability to mineralize hexachlorobenzene and pentachlorophenol and reactive textile dyes [22–25].

This study aimed to examine the removal rates of PCP added in lab microcosm-constructed wetlands composed of fixed and floated plants in secondary wastewater and supplemented with an isolated fungus strain. The novelty is the choice of a macrophyte *Polypogon maritimus* and *Lemna minor* still used for the elimination of heavy metals but it is not used for pesticide and pollutant removal in wastewater. The first part is based on the monitoring of the capacity of the fungal bioaugmentation process in the removal of PCP. The second part is devoted to the analysis of the effect of the combined phytoremediation-bioaugmentation treatment of

PCP on the physico-chemical and microbiological parameters of the wastewater.

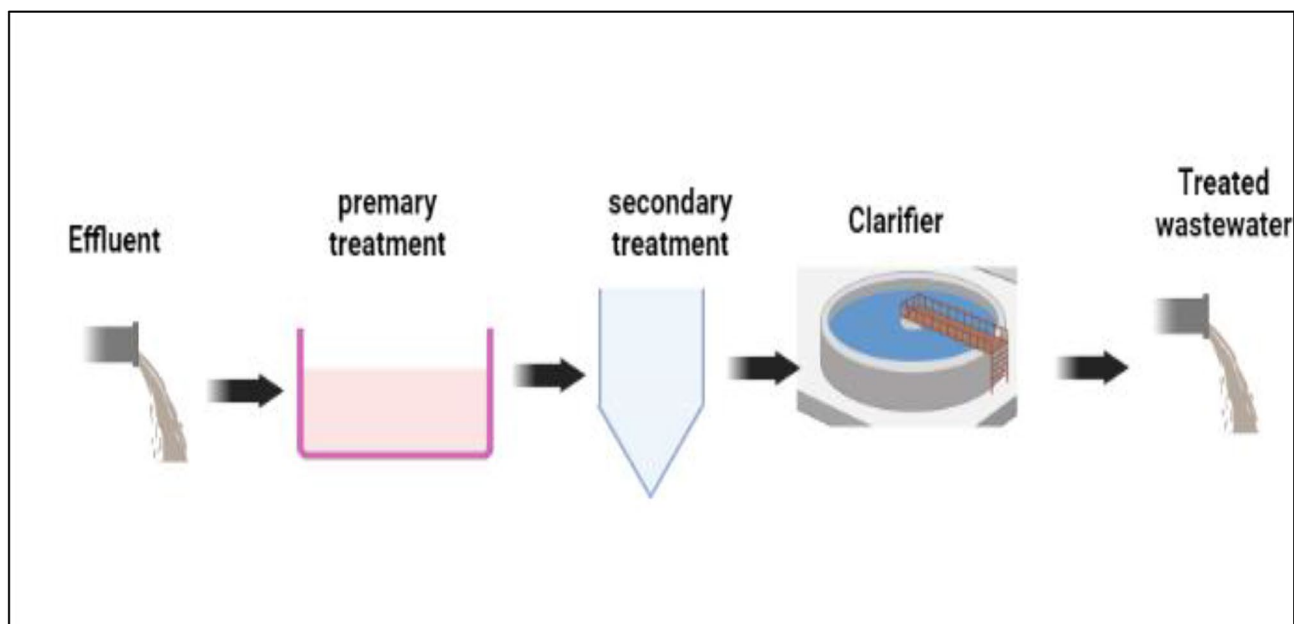
## Materials and methods

### Chemicals and reagents reagent-grade

Samples of wastewater with no detectable PCP were tested at the scale Chargaia I [WWTP] positioned in a residential and business area in Tunis City, northern Tunisia. This mega wastewater plant in Chargaia I of Tunis City treats a wastewater flow of about 60,000 m<sup>3</sup>/day [26]. The WWTP was drained by mixed industrial, urban, domestic, 24 hospitals, and rainwater coming from different zones of the great Tunis. The WWTPs use bio-physical-chemical processes: preliminary treatment, primary settling, secondary treatment by sewage sludge, secondary clarification, and tertiary treatment. A sampling of wastewater is made at the exit of the secondary wastewater clarification system [STWW] [Fig. 1]. The wastewater samples were stored at 4° C for determination of their main physical and chemical characteristics such as cation exchange capacity and pH, total nitrogen [27], and total carbon [28]. In addition, the bacteriological analyzes were realized by NPP standard method as reported in the [www.standardmethods.org/] for an enumeration of the: total coliforms (CT), total streptococci (St), fecal coliforms (FC), Fecal Streptococci (FS), and *Escherichia coli* (EC) for all tested wastewater.

### Selection and identification of fungi strains skilled in degrading PCP

For the selection of fungal isolates skilled in degrading PCP, we have used soil samples beforehand irrigated for 20 years with municipal treated wastewater. Soil suspension dilution in sterile physiological water was prepared, plated in growth potato dextrose Agar (PDA) supplemented by 1000 mg L<sup>-1</sup> of PCP, and cultured for at least three days at 26 °C. The plate showing a predominant one only mycelium type was selected and submitted for further investigation of its main morphological, microbiological characteristics, and molecular examination. Morphological and molecular identification was performed for the selected isolate. Macroscopic characteristics were studied on malt extract agar (MEA) media, according to Samson et al. [29]. The main morphological characteristics considered were colony size, color, and texture, while microscopic observation was carried out on the 7–10-day culture on MEA media using a light microscope (Leica). Genomic DNA was extracted from 7-day-old fungal cultures grown on MEA. For molecular identification, fungal DNA was extracted using the method of Liu



**Fig. 1** Different steps of the wastewater treatment at the scale Chargaia I

et al. [30]. The internal transcript space (ITS) genes used were ITS4, 5'-TCCCGCTTATTGATATGC-3') and (ITS5, 5'-GGAAGTAAAAGTCGTAACAAGG-3') described by White et al. [31]. The fungal isolate sequence was extracted from the NCBI Gen Bank database.

### Bioaugmentation process

The PCP bioaugmentation process was performed using fungi isolate and was selected. The inoculum was prepared in an enriched liquid medium potato dextrose broth for 48 h to obtain around  $10^8$  CFU  $\cdot$  mL<sup>-1</sup>. This inoculum was then transferred to a 250 mL flask containing 100 mL of MSM or sterilized secondary wastewater. The sterilized liquid medium was supplemented with PCP at different concentrations of 50, 100, 500, 1000, and 1500. The composition of MSM earlier reported by Ryu et al. [20] was: 3 g  $\cdot$  L<sup>-1</sup> NaNO<sub>3</sub>, 0.5 g  $\cdot$  L<sup>-1</sup> MgSO<sub>4</sub>, 0.01 g  $\cdot$  L<sup>-1</sup> FeSO<sub>4</sub>, 0.5 g  $\cdot$  L<sup>-1</sup> KCl, 1 g  $\cdot$  L<sup>-1</sup> K<sub>2</sub>HPO<sub>4</sub>, and 5 g  $\cdot$  L<sup>-1</sup> glucose. The MSM-PCP mixture suspension was kept under agitation on an orbital shaker at 150 rpm for 168 h. The pH solution was accustomed to  $7.3 \pm 0.2$ . PCP was mixed with the sterile medium after autoclaving.

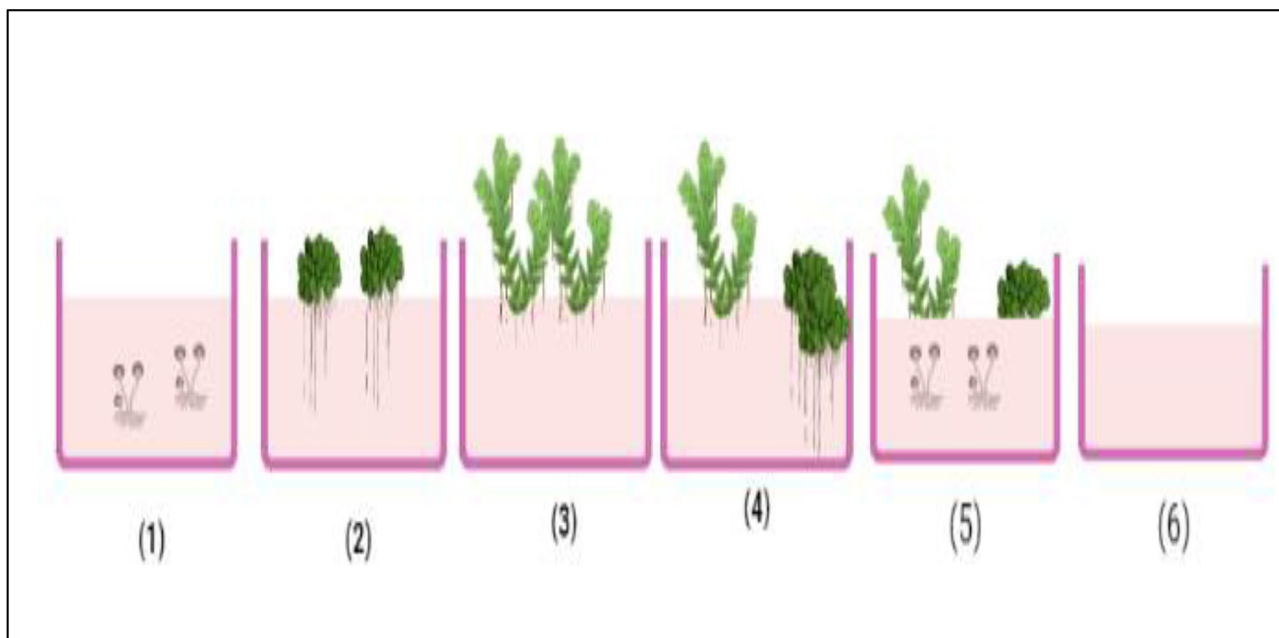
### Phytoremediation–bioaugmentation process

Two phytoremediation processes were realized using two specific varieties of macrophyte plants, *Polypogon maritimus*, and *Lemna minor*, known for their ability to purify PCP-contaminated secondary wastewater (Fig. 2). In the first

step, some small young and healthy seedlings of *Polypogon maritimus* and 20 g of *Lemna minor*, originally collected from a local natural lagoon, were subjected to 6-week growth in the PCP-contaminated secondary wastewater to ensure an adaptation and growth change of these two macrophytes to their new environment. In the second step, the wastewater was contaminated with a known amount of 1000 mg  $\cdot$  L<sup>-1</sup> of PCP as recommended by Manios et al. [32]. PCP with MW 226.34 and superior 99% purity was purchased from Sigma-Aldrich (USA) and prepared in NaOH (0.5 N). Experiments carried out in triplicates (n=3), included two categories of control: control with planting, and control without planting, and were carried out for 20 days.

### PCP content determination

In this study, the PCP removal was controlled for different wastewater treatments of the conducted phytoremediation–bioaugmentation bioremediation process. The PCP removal measure was carried out by high-performance liquid chromatography (HPLC), and extracted by a methanol solution. An aliquot of one milliliter was taken from the different treatments on the first day of the experiment (T0), and after 20 days (TF) of experimentation, was added to 1 mL of methanol. This last suspension of methanol will be vortexed for 5 min and stored at 20 °C for 10 min. This last stored sample will be once again vortexed for 5 min and centrifuged at 8000 rpm for 5 min, and the supernatant was filtered through a 0.22 mm sterile filter. The filtrate was submitted to HPLC analysis (Perkin Elmer Series YL9100



**Fig. 2** Description of treatments process of contaminated Wastewater. (T1) bioaugmentation process, (T2) phytoremediation process (Lemna), (T3) phytoremediation process (polygogon), (T4) phytore-

mediation process (Lemna- polygogon), (T5) phytoremediation- bioaugmentation (Lemna- polygogon- Fungi) and (T6) control SWWT.

system instrument) as reported by Karn et al. [33]. All analyzes were always carried out in triplicates.

### Chloride residual content

In different treatments of bioaugmentation and phytoremediation, the chloride analysis was released by the calorimetric method. One milliliter of  $K_2CrO_4$  (5%), 3 drops of nitric acid (65%,  $d=1.42$ ), and 50 mg pure calcium carbonate were added to a 20 ml sample. The sample was then homogenized by a magnetic stirrer and titrated with 0.0282 N silver nitrate solution until the color changed from yellow to brick red. Blind tests were performed with distilled water. Chloride levels are expressed in  $g L^{-1}$  [1].

### Fungi growth

Indirect estimation of fungal mycelium growth Nephelometric method (OD600 determination) in liquid media. This OD at 600 nm determination of isolated and selected fungi growth is determined by a turbid culture that will increase proportionally as it multiplies. back microbial growth has largely been established in the literature Determine OD by spectrophotometry at 600 nm [28].

### Fungi enumeration

The monitoring of the fungal growth of the selected strain during the treatment period of STWW supplemented with PCP was performed by the culture method on a specific medium. The capability of fungi to tolerate the organic toxicants was revealed by growing strains on solid media Potato dextrose Agar (PDA) at 25 °C for 5 to 7 days. The fungus abundance was measured using the colony-forming units (CFUs) method [36]. All analyzes were carried out in triplicates.

### Plant growth

The plant growth was evaluated by the aerial and roots biomass development based on their dry weight at 80 °C between 24 and 48 h and for the different treatments [28]. All analyzes were carried out in triplicates.

### Photosynthetic pigments

Fresh leaf samples, arbitrarily taken, were cut into 1–2 cm square pieces, frozen, and analyzed for chlorophyll a and b as following described: 0.2 g of prepared fresh leaf samples were submitted to 5 mL of 99.9% acetone extraction and centrifuged at 5000 g for 6 min at 4° C. Thereafter, the absorbance of the supernatant was respectively measured at

470 nm, 646.6 nm, 646.8 nm, 663.2 nm, and 720 nm as recommended by Arnon [34] and Lichtenthaler [35].

Plant photosynthetic pigments were determined using these equations:

$$\text{ChlorophyllA(Clora)} = (12.25 \times \text{Abs663.2}) - (2.79 \times \text{Abs646.8})$$

$$\text{ChlorophyllB(Clora)} = (21.50 \times \text{Abs646.8}) - (5.10 \times \text{Abs663.2})$$

## Statistical analysis

Two-way ANOVA analysis and Principal Component Analysis (PCA) were performed on the physical and chemical parameters of different treatments of phytoremediation-bioaugmentation. Before doing PCA, data were log<sub>10</sub> transformed. Differences between the two parameters were compared by an independent-sample t-test (two-tailed). All data were presented as mean ± SD (n/4 3) and statistical analyzes were performed by using GraphPad Prism 5.0.

## Results

### The characteristics of wastewater

The first important information about the present wastewater treatment system should interest the determination of the main average bio-physical-chemical characteristics of wastewater to be treated in this study. The quality of wastewater is usually expressed in terms of its physical, chemical,

and biological characteristics in Table 1. The secondary treated wastewater (STWW) is characterized by a neutral pH of about 7.06 and a high salinity of about 1827 μs. However, the STWW is characterized by an average COD of the order of 400 mg. L<sup>-1</sup>. According to the obtained results, other parameters are high such as chloride, DOB<sub>5</sub>, SUR, and SS... The characterization of the STWW effluent shows the need for a refining treatment that can be biological such as the phytoremediation process to improve its quality.

### Identification of selected fungal isolates degrading PCP

The main morphological, microbiological, and molecular investigation of the only fungal mycelium grown on a PDA medium allowed us to identify this mycelium as *Penicillium ilderdanum* with an accession number OP862882 Taxonomic description. The photos of colonies and fungal structures of *Penicillium ilderdanum* are presented in Fig. 3. The morphology of this fungal strain confirmed the characteristics of *Penicillium ilderdanum* reported by Alker et al. (2001). The molecular 16 S rRNA gene sequencing followed by NCBI-BLAST analysis confirmed this morphological identification.

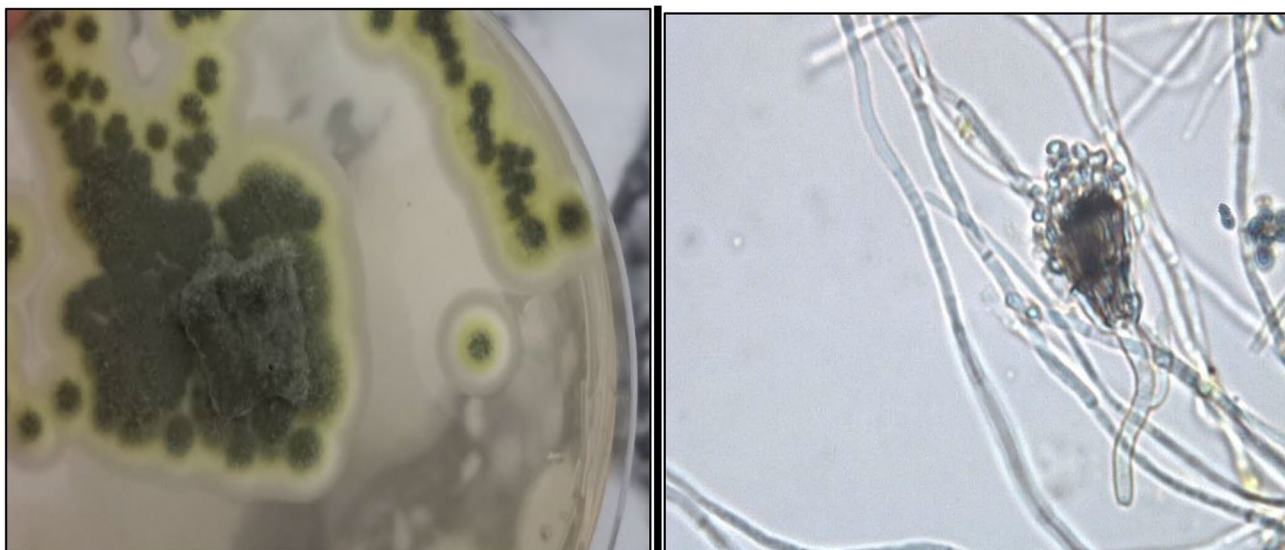
The selected fungus gives a high capacity of resistance and tolerance of PCP at different tested concentrations of 50, 100, 500, 1000, and 1500 mgL<sup>-1</sup> (Fig. 4). According to the results shown in Fig. 4, maximum resistance to PCP was observed at an initially added concentration of about 1000 mg L<sup>-1</sup>. At an initially added concentration of 1000 mg L<sup>-1</sup> of PCP, 57.69% PCP removal, 11.8 g L<sup>-1</sup> release, and growth of 18.10<sup>2</sup> CFU. mL<sup>-1</sup>.

**Table 1** Physical-chemical secondary treated wastewater characteristics

Parameters	Unit	Value
pH <sub>H2O</sub>		7.06 ± 0.13
Conductivity	μs	1827.07 ± 0.13
COD	g. L <sup>-1</sup>	0.4 ± 0.08
Nitrogen		3255 ± 0.12
Total Organic Carbon	g. L <sup>-1</sup>	1.035 ± 0.015
C/N		0.30
Chloride	g Cl <sup>-</sup> . L <sup>-1</sup>	3.125 ± 1.76
Dry matter	g. L <sup>-1</sup>	1.46 ± 0.15
SS	g. L <sup>-1</sup>	0.605 ± 0.01
DOB <sub>5</sub>	g. L <sup>-1</sup>	0.645
NO <sub>3</sub>	mg. L <sup>-1</sup>	14.6 ± 0.28
SUR	mg. L <sup>-1</sup>	64.75 ± 1.77
Total coliform		24.57 · 10 <sup>4</sup> ± 3.25
Total streptococcus	(MPN/100 mL)s	20.24 · 10 <sup>4</sup> ± 2.36
<i>E. coli</i>		952 · 10 <sup>4</sup> ± 2.51

SUR: Surfactant; pH<sub>H2O</sub>: Hydrogen potential (1/5); NO<sub>3</sub>: Nitrate; C: Total organic carbon; N: Nitrogen; BOD<sub>5</sub>: Bio-chemical oxygen demand; SS: Suspended solid; C/N: Carbon/ Nitrogen; COD: Chemical oxygen demand





**Fig. 3** Morphological and microscopic identification of isolated and selected fungi

### Bioaugmentation process by fungi

**Fungi growth:** The selected and identified fungal strain *Penicillium ilderdanum* was subjected to expression in MSM and STWW to verify its efficiency in PCP removal (Fig. 5A). Monitoring of the biomass proliferation of the selected fungus was performed by measuring the OD at 600 nm every 24 h during the incubation period in both MSM and STWW. The tracking of PCP removal is controlled by the variation of PCP content, chloride, and microbial biomass. During the incubation period, we notice three phases: adaptation, exponential and stationary. The bioaugmentation treatment performed in STWW is characterized by a short adaptation phase of 24 h with an OD at 600 nm value of 0.254 and a faster exponential phase during 48 h with a value of 0.784 OD and a decreasing lethal phase after 96 h.

According to the obtained results, the adaptation phase of the fungus *Penicillium ilderdanum* OP862882 in MSM supplemented with PCP for 48 h appeared slow with a growth value about of 0.447. Thus, the exponential phase appeared higher in MSM than in STWW, in the order of 0.98 and 0.78 for 72 h, respectively.

**PCP removal:** Figure 3 A showed the ability of the operational fungi to tolerate and remove PCP in MSM and STWW with an average removal of about 23.32% and 5.36% of PCP, respectively, only during the first day of incubation, and this PCP removal reached 92.33% and 92.36% of PCP after seven days of incubation. According to these last results, the operational fungal isolates selected and used in this investigation showed an important capacity of tolerance and transformation/or degradation or removal of PCP in these two specific media, MSM and STWW.

### Phytoremediation-bioaugmentation process

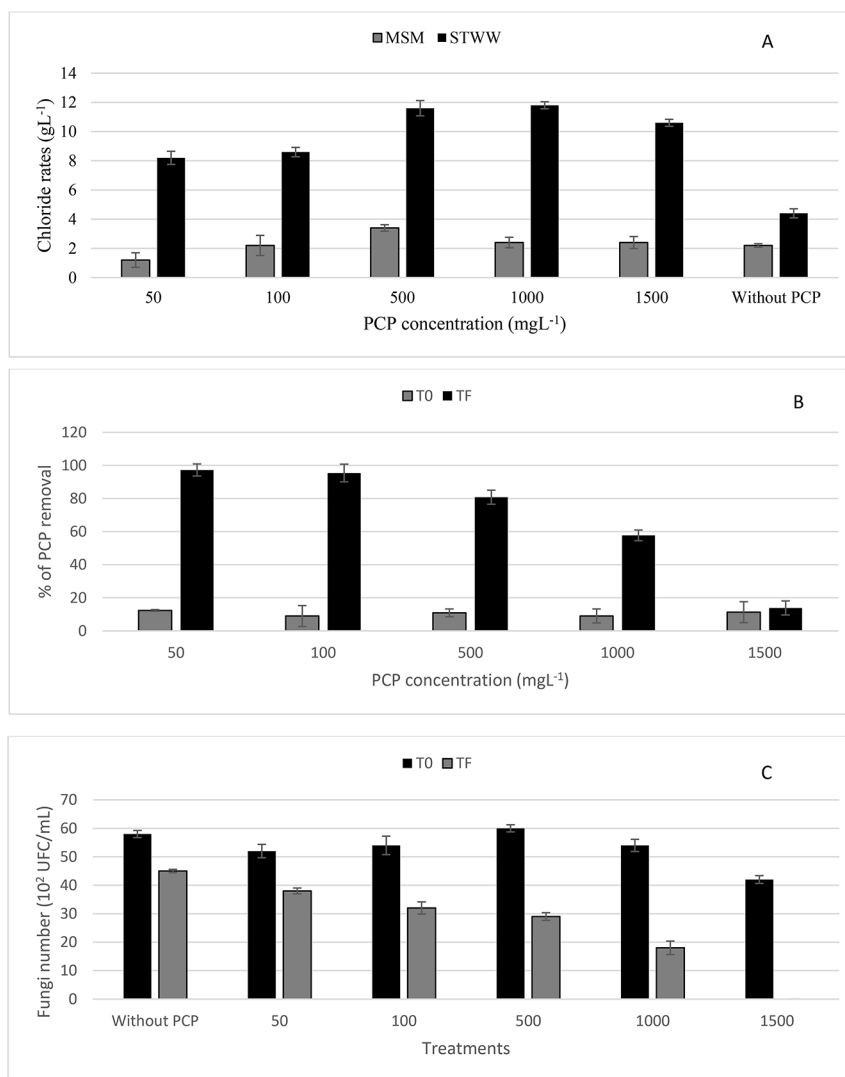
#### PCP content monitoring

The rate of PCP removal registered in the combined phytoremediation-bioaugmentation treatments was determined after HPLC measurement. The results of PCP content are obtained in  $\text{mg. L}^{-1}$  and summarized in Fig. 6. Figure 6 allowed to show that the most important percentage of PCP removal was recorded for Polygogon treatment with a 92.01% removal rate. This last result allowed showing the importance and value of the phytoremediation process regarding PCP bioremediation.

In contrast, the STWW+PCP treatment presented the lowest rate of PCP removal, with only 43.19%. All obtained results allowed us to show the intrinsic value of the microorganisms used in these PCP bioremediation processes. So, the specific fungal genus used in these bioremediation processes contributed intensively and heavily to the efficiency of these processes. The other treatments carried out in this experimentation such as L+PCP, F+L, P+L+F+PCP, L+F+PCP, and P+F+PCP usually provided low PCP removal rates.

#### Chloride content change in wastewater submitted to the bioaugmentation-phytoremediation process

The results of residual chloride content following the different treatments at T0 and TF were summarized in Fig. 6. These results showed that most chloride contents increase in the different treatments of phytoremediation-bioaugmentation either supplemented or not with PCP.



**Fig. 4** Results of chlorides (A), the percent of PCP removal (B) and Fungi growth (C) during bioaugmentation treatment in STWW supplemented with different concentration of PCP (50, 100, 500, 1000 and 1500 mg L<sup>-1</sup>)

The phytoremediation treatment with the *Polypogon* plant (P+PCP) presented the most important value of chloride content of 23.58 g L<sup>-1</sup> when compared to the other investigated treatments such as L+PCP, F+L, P+L+F+PCP, L+F+PCP, and P+F+PCP. The lowest chloride content was recorded in the case of STWW treatment free of PCP with a value of 1.42 g L<sup>-1</sup>. These latter results and the PCP removal ones showed a close relationship between the chloride content in the medium and the PCP removal.

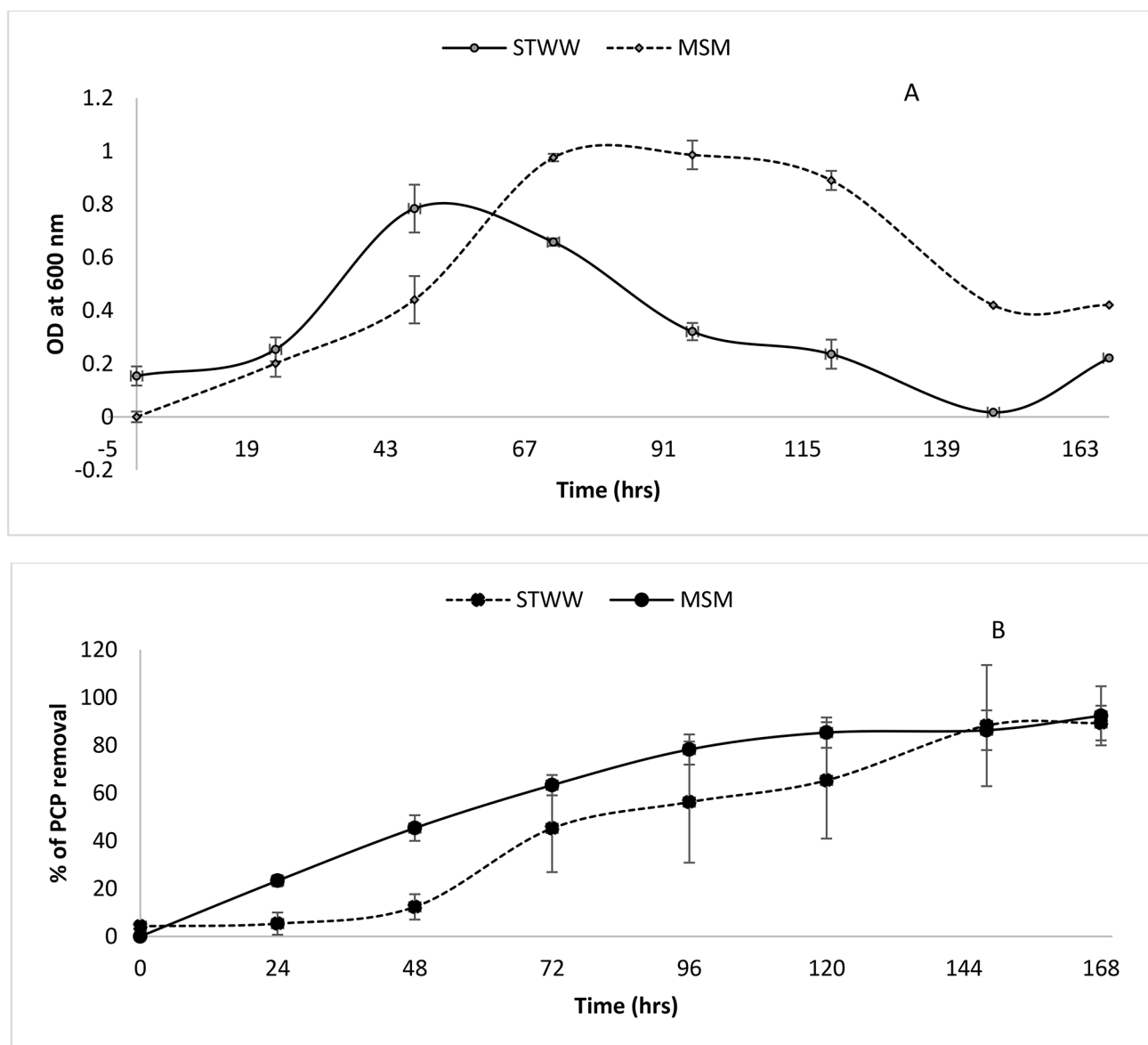
#### Fungal evolution during the varied bioremediation processes

The fungal enumeration in the different treatments investigated is summarized in Fig. 6. The different phytoremediation-bioaugmentation treatments showed fungal mycelium

development with various consistency and intensity in STWW. The L+F+PCP treatment showed the most fungal development, with the highest number of fungi of the order of 3.3 10<sup>3</sup> CFU. mL<sup>-1</sup>. These results allowed us to conclude the absence of antagonism between the fungus and the *Lemna minor* plant. In contrast, the net absence of fungal organisms at the end of the treatment of L+PCP and L+P+F+PCP was noticed. These obtained results prove the fungi *Penicillium ilderdanum* OP862882 viability and the PCP tolerance in STWW during the treatment process.

#### Change of main chemical-physical wetlands characteristics during bioremediation processes

Constructed wetlands have been effective in varying many physical and chemical parameters associated with fungi.



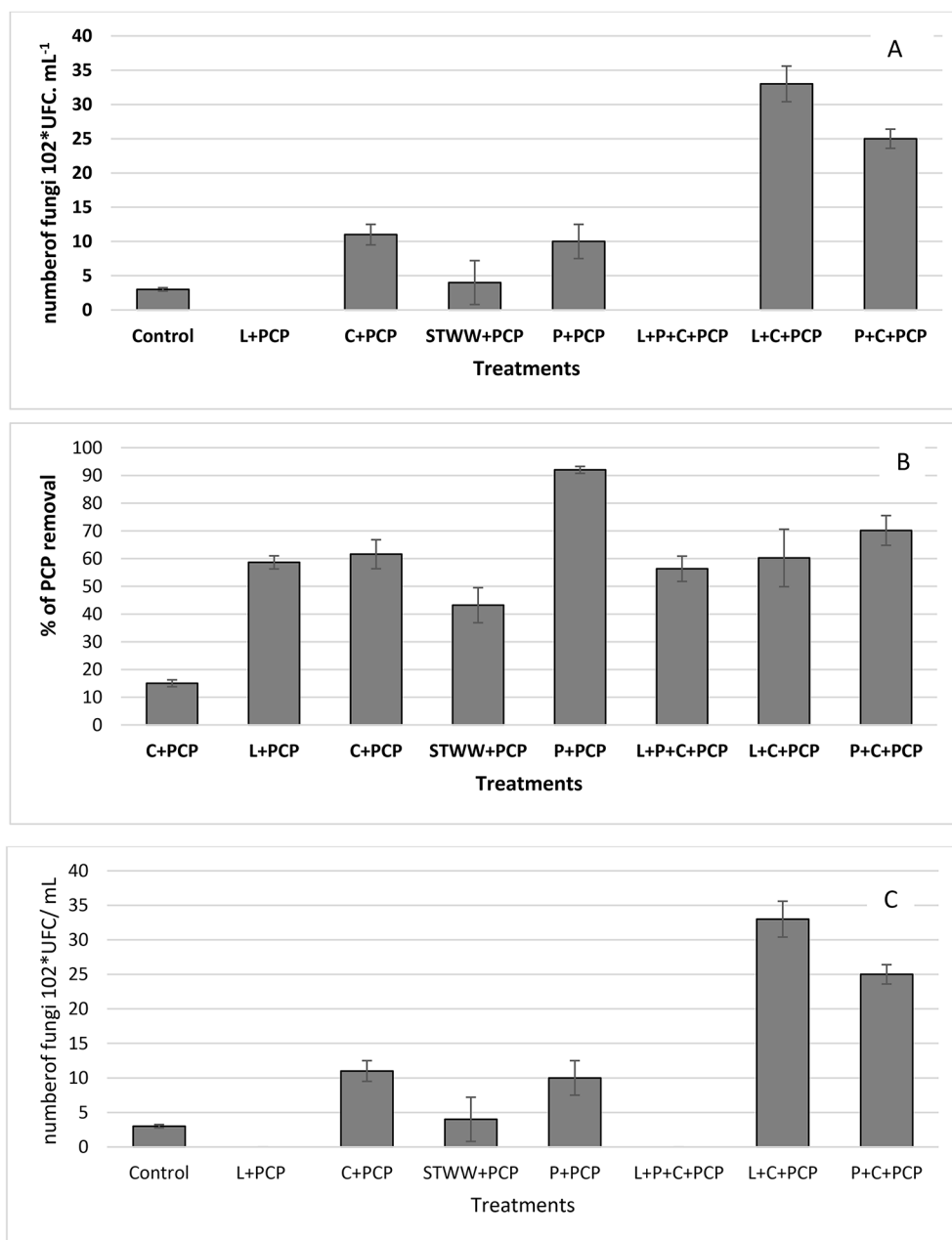
**Fig. 5** PCP percent removal (A) and Bacterial growth in liquid medium 5MSM) and STWW supplemented with PCP ( $1000 \text{ mg L}^{-1}$ ) during 168 h at  $30 \text{ }^\circ\text{C}$

(Table 2). Usually, vegetation has shown an improvement in PCP efficiency removal. The characteristics of the raw effluent were given in Table 1 along with the mixed wastewater effluent from the holding tanks.

The varied phytoremediation-bioaugmentation treatments were submitted to some common physico-chemical parameter analyses, like pH, EC, COD, and OC after 30 days of processing to get the treatment efficiency regarding the treated wastewater quality. The different results obtained from the physico-chemical analysis of STWW are summarized in Table 2. Thus, the STWW submitted to the different PCP bioremediation processes showed a slight

pH increase from 7.06 to around 10. The EC measurement showed a slight increase for all treatments except for the control STWW without PCP. The most important values of EC were recorded in the respective treatments: P+PCP, P+L+F+PCP, and P+F+PCP, with values of 8.44, 8.97, and  $8.35 \mu\text{s. cm}^{-1}$ . Besides, the COD results showed a significant decrease in all STWW treatments. The lowest values of COD were recorded in the treatments P+PCP and L+P+F+PCP with values of the order of 2 and  $3.2 \text{ mg L}^{-1}$ , respectively. The increase in COD may be related to the release of some residual compounds in wastewater likely following the transformation of PCP. As for the OC





**Fig. 6** Chloride contents (A), PCP removal rates (B) and fungal growth (C) change during the wastewater bioremediation process operated at different PCP contents of 50, 100, 500, 1000 and 1500 mg. L<sup>-1</sup>

change in the PCP-STWW treatment process, the highest OC rate was registered in the F + PCP and L + PCP treatment, with respective values of around 54.35% and 53.33%.

### Plant growth analysis

The effectiveness of the biological treatment process of phytoremediation is based on the viability and growth of plants used in wastewater artificially contaminated with PCP. The plant growth assessment summarized in Table 3 was

tracked by counting the number of leaves and measuring their length. According to the obtained results, the *Polypogon maritimus* plant achieved a considerable growth yield in the treatment P + F + PCP. This plant presented a length leaf of about 35.2 cm, with an average number of about 16.

### Chlorophyll determination

Figure 7 showed the chlorophyll a and b evolution during phytoremediation-bioaugmentation treatment in

**Table 2** Physical-chemical characteristics change of contaminated secondary treated wastewater during the processes of phytoremediation

	pH		Salinity (ms. cm <sup>-1</sup> )		COD (mg. L <sup>-1</sup> )		OC (%)		Chloride (mg. L <sup>-1</sup> )	
	T0	Tf	T0	Tf	T0	Tf	T0	Tf	T0	Tf
STWW (control)	10.32 ± 1.12	10.25 ± 1.02	7.88 ± 0.41	7.26 ± 0.14	800 ± 10.2	758 ± 14.3	1.36 ± 0.23	2.36 ± 0.65	1.2 ± 0.12	1.42 ± 0.12
STWW + L + PCP	9.93 ± 0.98	9.59 ± 0.45	7.7 ± 0.31	7.82 ± 0.25	820 ± 7.23	40.00 ± 10.58	14 ± 2.36	54.35 ± 7.21	2.36 ± 0.14	6.38 ± 0.58
STWW + F + PCP	9.6 ± 1.35	9.72 ± 1.56	7.56 ± 0.21	7.72 ± 0.36	852 ± 12.3	66.67 ± 7.23	12.3 ± 2.3	53.33 ± 10.3	4.25 ± 0.13	18.87 ± 2.14
STWW + PCP	9.84 ± 0.95	9.53 ± 2.01	7.55 ± 0.85	7.99 ± 1.3	862 ± 10.25	2.00 ± 8.21	15 ± 3.12	46.67 ± 2.3	3.26 ± 0.45	15.09 ± 1.23
STWW + P + PCP	9.97 ± 1.32	9.68 ± 1.02	7.71 ± 1.47	8.44 ± 0.92	798 ± 12.3	36.67 ± 6.32	14.7 ± 4.21	33.33 ± 1.54	2.56 ± 0.01	23.58 ± 2.14
STWW + L + P + F + PCP	9.83 ± 0.547	9.66 ± 0.85	7.62 ± 2.01	8.97 ± 2.01	995 ± 21.3	3.2 ± 7.21	10.2 ± 2.14	26.67 ± 2.14	3.98 ± 0.23	7.55 ± 1.36
STWW + L + F + PCP	9.76 ± 2.01	9.8 ± 0.12	7.17 ± 1.03	8.07 ± 1.058	889 ± 14.3	93.33 ± 4.25	12.36 ± 1.02	40.00 ± 1.58	3.65 ± 0.15	19.64 ± 1.37
STWW + P + F + PCP	9.86 ± 0.745	9.74 ± 0.36	7.62 ± 0.85	8.35 ± 1.09	795 ± 24.6	99.00 ± 15.2	13.25 ± 1.65	26.67 ± 1.62	4.02 ± 0.41	21.72 ± 2.56

PCP: Pentachlorophenol; STWW: Secondary treated wastewater; L: *Lemna minor*; F: Fungi; P: *Polypogon maritimus*; pH: Hydrogen potential; COD: Chemical organic demand; OC: Organic carbon. ± : Deviation standard; n = 3

**Table 3** Monitoring of *Polypogon maritimus* plant growth during various phytoremediation processes of STWW contaminated with PCP.

	Longer	Number of sheets
P + PCP	27 ± 1.25 <sup>a</sup>	13 ± 2.01 <sup>a</sup>
P + F + PCP	35.2 ± 2.36 <sup>ab</sup>	16 ± 1.3 <sup>b</sup>
L + P + PCP	10 ± 3.01 <sup>a</sup>	3 ± 0.45 <sup>c</sup>
L + P + F + PCP	34 ± 4.25 <sup>b</sup>	14 ± 2.01 <sup>ac</sup>

PCP: Pentachlorophenol; L: *Lemna minor*; C: Fungi; P: *Polypogon maritimus*; F: fungi. a, b, c... : Values affected with the same letter are not significantly according to Student-Newman-Keuls test (IBM SPSS inC, version 26)

PCP-contaminated STWW. Chlorophylls a and b contents in *Polypogon maritimus* and *Lemna minor* plant were elevated in the STWW after 7 days of treatment. The high pigments level registered for these two macrophytes their important viability and tolerance when used for PCP phytoremediation. The *Polypogon maritimus* often showed a higher chlorophyll content and leaves length than *Lemna minor*.

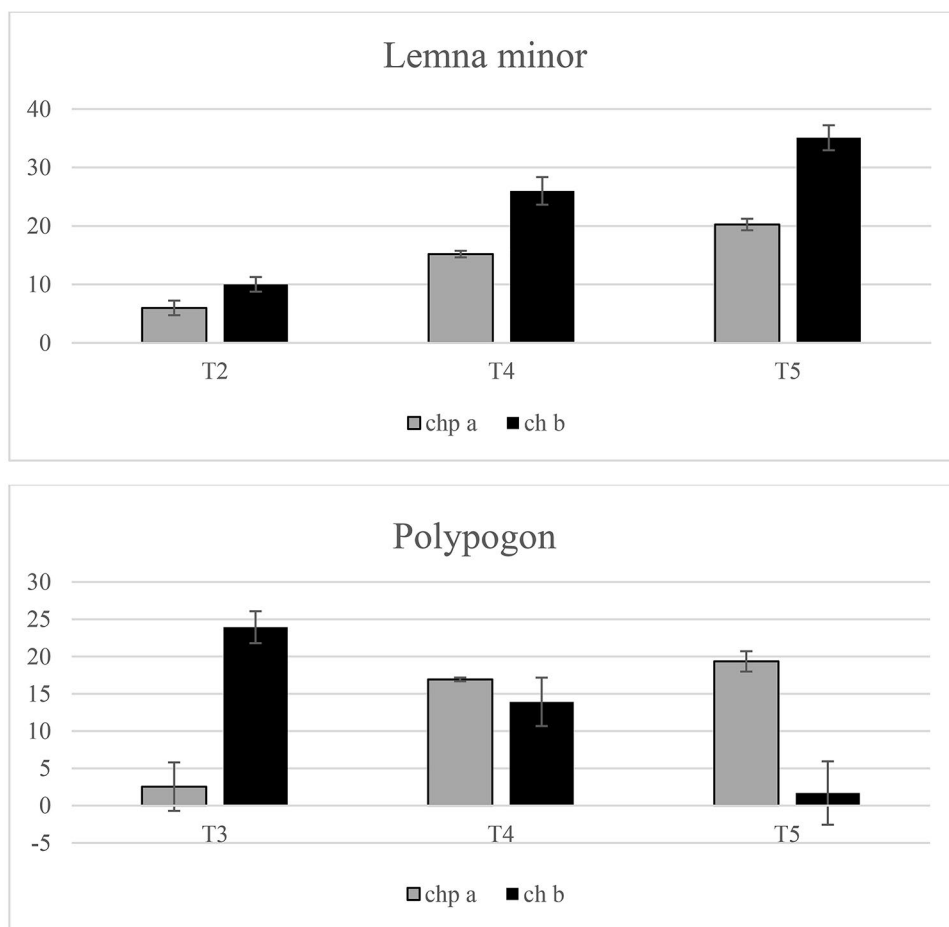
### Principal component analysis (PCA)

To explore the variability observed in the phytoremediation-bioaugmentation treatments of PCP-contaminated wastewater during the incubation times, a discrimination of different treatments by the Principal Component Analysis (PCA) procedure was performed. KMO examination, generous a result of 0.717, and Bartlett sphericity test presenting p-value < 0.0001 showed the suitability of samples to apply the PCA. The procedure allowed the extraction of two principal components reported as principal components and graphically visualized in a scatterplot (Fig. 8). The two main components accounted for 77.3% of the total variance. PC expressed the variability in PCP concentration, fungi number, the organic carbon, and DOC of wastewater (Fig. 8). The separation on PC1 was principally because of the reduction, between the two-sampling time, in wastewater samples of PCP content and the increase in chlorides, bacterial, and fungi number. In T0 time, the different wastewater samples were not discriminated from each other along PC1 and PC2 (Fig. 8).

### Discussion

Raw wastewater treatment processes aimed to change the wastewater quality for safe release in the natural environment, like agricultural soil irrigation, recycling in some important industries requesting large quantities and flows of water, and even why not for noble use as drinking or other suitable worthy purposes. The wastewater treatment took place in wastewater treatment plants, which should be

**Fig. 7** Chlorophyll (leaf)  $\text{g mL}^{-1}$  of the plant *Lemna minor* and polygogon in different wastewater treatments at time 0 (T0) and after 30 days of incubation (TF). (T2) phytoremediation process (*Lemna*), (T3) phytoremediation process (*polygogon*), (T4) phytoremediation process (*Lemna*- *polygogon*), (T5) phytoremediation- bioaugmentation (*Lemna*- *polygogon*- Fungi)

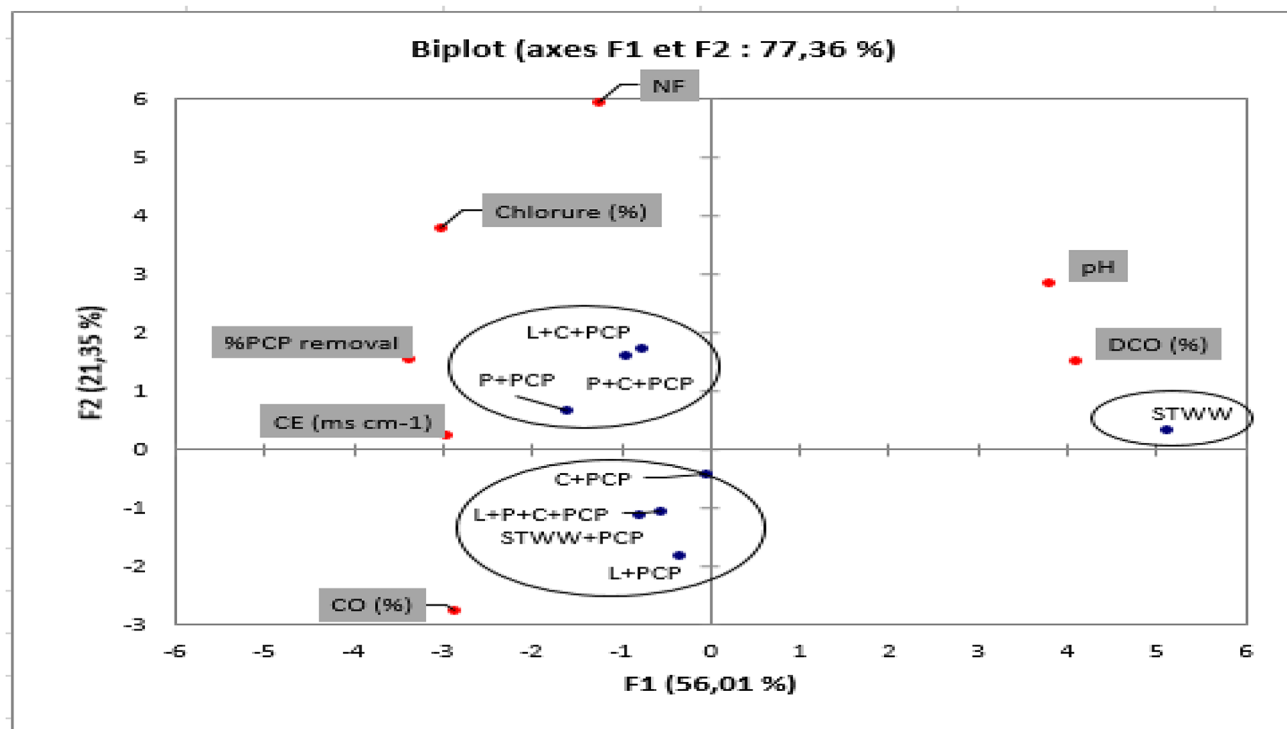


designed and considered under well-defined circumstances and conditions [37]. So, the purpose of wastewater treatment was to remove totally or reduce some important compounds and contaminants from raw water that could cause threats to humans and the natural environment if discharged raw and without proper treatment into receiving water systems like ocean, sea, rivers, wadis, etc [38]. Bioremediation is a common biological technique often used for the removal of main residual compounds from polluted water and makes it safe for discharge in natural ecosystems or recycling usage in many domains as above cited. This technique of bioremediation is essentially based on specific microorganisms or plants for the purification of wastewater under well-defined conditions. The treated wastewater showing relatively good quality could be reused or discharged safely in natural ecosystems [39].

In this study, the bioremediation process investigated a combined phytoremediation-bioaugmentation technique of wastewater artificially contaminated with PCP. This combined process resulted in significant PCP removal rates from wastewater and confirmed the use of microorganisms and specific plants as degrading and removing agents of dyes, cosmetics, detergents, medicines, and pesticides as an

effective method to reduce environmental harm and damage [40–42]. Thus, fungi and bacteria can transform PCP for example by incorporating one or two oxygen from the available  $\text{O}_2$  through the oxygenase process [43, 44]. This process allowed the destruction of the aromatic ring and the subsequent formation of  $\text{CO}_2$  through a slow aerobic transformation; especially for a case of highly chlorinated compounds such as PCP [45].

Microorganisms usually showed reduced growth in the presence of a toxic compound; however, the fungal strain of *Trichoderma harzianum* CBMAI 1677 showed an increased colony mycelium diameter at  $50 \text{ mg L}^{-1}$  of PCP and it is often used in some quantitative biodegradation experiments [46]. Creswell and Curl [47] obtained similar results when *T. harzianum* was experienced with herbicides prometryn, norflurazon, and cyanine. Many fungi have been shown to detoxify PCP by methylation using a specific lignin-degrading system, prevailing to serve other functions such as degradation of wood components such as lignin and cellulose [48–50]. Using reactions catalyzed by pH oxidases such as laccases and peroxidases, fungi can make the primary PCP transformation into pentachloro anisole (PCA) [48–49, 51, 52]. The last process known for PCP removal exploits the



**Fig. 8** Principal Component Analysis on bioremediation (phytoremediation bioaugmentation) of PCP contaminated wastewater at T0 and TF (30 days): projection of the cases and projection of the variables. S.TWW: Secondary wastewater; PCP: pentachlorophenol

adsorption and bio-adsorption processes into organic compounds or the microbial biomass existing in the medium as living or dead organisms. So, it has been found that some microbial cultures showed a particular affinity for binding PCP [51, 53], i.e., the adsorption took place through the charge attraction between PCP and microbial biomass [42, 51, 54–57]. Species of *Lemna minor*, *Azolla filiculoides*, *Typha angustifolia* L., *Chrysopogon zizanioides* (L.) Roberty, *Eichhornia crassipes*, and *Cyperus isocondus* have shown the best efficiency rate with a value  $> 99.9\%$  for the removal of some tested micro-pollutants [58, 59].

The basic principles of bioremediation involved reducing the solubility of these environmental contaminants by changing pH, the redox reactions, and the adsorption of contaminants from a polluted environment [60]. Also, the batch kinetic study of Kobayashi and Kishino [61] showed that around 91% of PCP is removed from municipal wastewater spiked with around  $1 \text{ mg L}^{-1}$  of PCP at an equilibrium time of 5 h. Various reports have been made on enhancing the biosorption of pentachlorophenol (PCP) by altering the pH levels in aqueous solutions. For example, the biosorption abilities of *Aspergillus niger* [62] and *Mycobacterium chlorophenolicum* [63, 64] in the removal of PCP from aqueous solutions were reported to be pH-dependent. Brandt et al. [63] also evaluated the influence of pH on the adsorption and desorption behavior of PCP by *Mycobacterium*

*chlorophenolicum* and reported that pH values were an essential parameter that affected PCP adsorption, and with adsorptive capacity increasing with pH decreasing.

The physical and chemical characteristics of the reactionary environment influenced the absorption and desorption properties of organic as persistent organic pollutants POPs, HAP, PCP, etc [65]; so it is imperative to follow the evolution of a few indicatives and influencing parameters during each bioremediation water treatment. The macrophyte growth and development could be affected because of the well-known PCP phyto-accumulation phenomena within the organ plants as reported by Yang et al. Yang et al. [66]. Other works suggested that the pollutant bioavailability especially depended on the organic matter content of water, the pH, and other various specific properties of the pollutant. Equally severe and prolonged toxicity has shown an increase at low pH. This may be because a lower pH favors a non-ionic form of PCP, which is more easily absorbed and more toxic than ionized PCP [67]. Finally, many works released by our researchers prove that phytoremediation and bioaugmentation can eliminate PCP in wastewater (Table 4).

**Table 4** Description of different results of PCP elimination in secondary treated wastewater (STWW)

Process	Species	Potential	references
Bioaugmentation	Consortium 1: <i>Citrobacter freundii</i> , <i>Klebsiella variicola</i> RCB1013, <i>Staphylococcus equorum</i> YXY-2, <i>Micrococcus lylae</i> ES142con (Bp3), Consortium 2: <i>Leclercia adecarboxylata</i> TPY2, <i>Leclercia adecarboxylata</i> RLF11, <i>Klebsiella</i> sp., <i>Klebsiella oxytoca</i>	after 168 h was 70 and 97.5% in STWW for C1 and C2	Werheni et al. [69]
Phytoremediation + Bioaugmentation	<i>Lemna gibba</i> + <i>Typha angustifolia</i> + <i>Pseudomonas putida</i> HM627618	88% of PCP removal in STWW	Werheni et al. [28]
phytoremediation	<i>Typha angustifolia</i>	96.8% of PCP removal	Werheni et al. [68]
phytoremediation + Bioaugmentation	<i>Typha angustifolia</i> + <i>Aspergillus sydowii</i>	79% of PCP removal in STWW	Werheni et al. [1]
Phytoremediation + Bioaugmentation	<i>Typha angustifolia</i> + <i>Schoenoplectus acutus</i> + <i>Pseudomonas putida</i> HM 627,618	97% of PCP removal in STWW	Werheni et al. [70]
Adsorption	+ date activated carbon		

## Conclusion

Faced with water scarcity caused by global climate change, wastewater reuse remains a promising solution in many industrial and agricultural sectors. Water treatment is the main ecosystem service in the provision of safe and clean water through biological processes such as phytoremediation and bioaugmentation that create new habitats for organisms. In this study, the phytoremediation process improved the sanitary quality of wastewater by removing this pollutant. Thus, based on the results obtained from the physicochemical analyses of the STWW carried out, it proved to be effective in improving the quality of STWW as in the case of DOC, pH... In addition, the combined phytoremediation-bioaugmentation treatment showed the best PCP removal performance with good viability of the two plants, *Polypogon maritimus* and *Lemna minor*, and the selected fungus *Penicillium ilderdanum* OP862882. This biological practice and process, therefore, seems promising for the removal of PCP from PCP-contaminated sites. These processes, proposed in our study, are presented as low-cost and low-maintenance systems to treat domestic wastewater.

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