ORIGINAL RESEARCH



Synthesis of conductive polymer-coated mesoporous MCM-41 for textile dye removal from aqueous media

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Abstract In this paper, we aimed to evaluate Acid Blue 62 removal from aqueous media, using mesoporous silicate MCM-41, loaded with polypyrrole (PPy) and polyaniline (PAni) composites. PPy/MCM-41 nanocomposite showed higher performance than PAni/MCM-41 due to its smaller molecule size. For characterizing the synthesized composites, different methods were applied. The Langmuir model showed the greatest agreement with the experimental findings $(q_{\rm m}, 55.55~{\rm mg~g^{-1}})$. The kinetic study also confirmed the compatibility between the pseudo-second-order model and adsorption. Moreover, we measured Gibbs free energy changes $(\Delta G^{\rm o})$ and enthalpy changes $(\Delta H^{\rm o})$. Considering the negative $\Delta G^{\rm o}$ and positive $\Delta H^{\rm o}$, AB62 adsorption on PPy/MCM-41 nanocomposite can be considered a spontaneous, endothermic reaction.

Keywords Adsorption · Acid Blue 62 · Polypyrrole · Polyaniline · MCM-41

Introduction

The presence of dyes in industrial wastewaters is a major issue in different countries. Different industries, including textile printing, plastic, pharmaceutical, and food industries, use dyes in their procedures. About 20% of dye production in the world is lost during the dyeing process

 [1]; therefore, the presence of dyes in wastewater is inevitable.

Dye removal from wastewater is significant, as the quality of water greatly depends on color; even a very low dosage of dye (<1 ppm) can be easily seen in water and is found to be unfavorable. Besides, most dyes result in the occurrence of diseases, such as skin inflammation, skin irritation, carcinomas, and mutations in humans [1]. Removal of colored contaminants from wastewater is carried out using various techniques, including biological and physicochemical technologies.

Oxidation [2], coagulation–flocculation [3], separation of membranes [4], and adsorption [5, 6] are among the major processes involved in treatment. Among the mentioned processes, adsorption is the most frequently used technique, showing feasibility, high yield rate, and less expenditure [5]. Species in the adsorption process are moved to the solid phase and can minimize the effluent volume [1].

Researchers have recently introduced favorable adsorbents [5–7]. Highly porous materials such as mobile crystalline material-41 (MCM-41) [8, 9], modified hexagonal mesoporous silica (HMS) [5, 10], and nanoporous silica (SBA-15) [11] seem to be appropriate for removing different dyes. Different types of surfactants are used for the preparation of these materials. Surfactants act as templates throughout sol–gel and hydrothermal processes. The main characteristics of mesoporous materials include great specific surface area, uniform and limited pore size distribution, and great thermal stability [12, 13].

Recently, MCM-41 has been used for the adsorption of dyes from wastewater. Considering the presence of SiO and SiOH (known to adsorb cationic dyes and inhibit anionic dye adsorption), the structure of MCM-41 is negatively charged [14]. The importance of interaction





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between MCM-41 and safranin (a large positively charged dye) was examined in a previous study [15]. To enhance the MCM-41 capacity in adsorbing specific substances, it is necessary to make surface modifications. Considering the adsorbent–adsorbate interactions, surface modifications also improve the selectivity of MCM-41 [16].

Amines, as well-known functional groups, have been used in many studies for surface modification of different types of adsorbents. To eliminate mercury ions from water, MCM-41 was functionalized with diethylenetriamine (DETA) [9]. In a previous study, aminopropyltrimethoxysilane (APTS) was applied for MCM-41 (NH₃ + -MCM-41) modification to remove four types of anionic dyes from the aqueous media [16].

Transition metals including nickel are extensively applied for the modification of MCM-41 structure [17]. In a previous study, to remove methyl blue from aqueous solutions, highly ordered nickel-supplemented MCM-41 adsorbents were synthesized (with varying nickel contents) [14]. Overall, according to our literature research, few studies have been conducted on the use of polypyrrole (PPy)/MCM-41 and polyaniline (PAni)/MCM-41 in anionic dye adsorption.

With this background in mind, PPy/MCM-41 and PAni/MCM-41 nanocomposites were synthesized in the current study. Nanocomposites were described using different methods, including transmission electron microscopy (TEM), scanning electron microscopy (SEM), Fourier transform infrared spectroscopy (FTIR), X-ray diffraction (XRD) analysis, and BET method. Through performing batch experiments, the efficiency of PPy/MCM-41 and PAni/MCM-41 in AB62 elimination from aqueous media was also examined. Moreover, the significance of solution pH, contact time, temperature, and adsorbent dose was evaluated. Finally, thermodynamic and kinetic evaluations were performed to measure the parameters.

Materials and methods

Substances

The reagents used for the preparation of samples and experimental tests included cetyl-trimethyl-ammonium bromide (CTMABr), ammonium hydroxide, tetraethyl ortho silicate (TEOS, SiC₈H₂₀O₄), aniline, pyrrole, potassium iodate (KIO₃), ferric chloride, sulfuric acid, acetone, hydrochloric acid (HCl), NaOH, Na₂HPO₄, NaH₂PO₄, and deionized water (Merck Co., Germany). Moreover, Dystar Co. (Germany) provided anionic dye (Acid Blue 62 or AB62, $\lambda_{\rm max} = 595$ nm). Figure 1 illustrates the chemical composition of AB62.

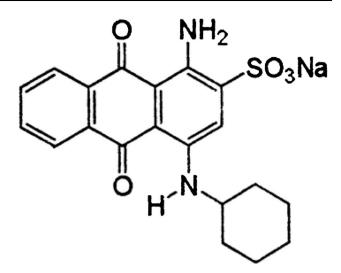


Fig. 1 Molecular structure of dye Acid Blue 62

MCM-41 synthesis

MCM-41 synthesis was performed in line with the technique proposed by Kamarudin et al. [18]. In brief, 2.4 g of cetyltrimethylammonium bromide (CTMABr) as the template was added to deionized water (120 g) and stirred to form a uniform solution. Afterward, ammonium hydroxide (8 mL) was added and stirred over 5 min. Following that, 10 mL of tetraethyl orthosilicate (TEOS), as the silicon source, was added and stirred for 24 h.

The solution was moved to a steel autoclave and stored at 145 °C for 2 days. The pH was adjusted typically during 48–72 h to attain stability. The obtained product was filtered, washed, and stored at 100 °C for 1 day. In the final step, calcination was performed at 600 °C for 5 h.

PAni/MCM-41 nanocomposite preparation

For the preparation of PAni/MCM-41, potassium iodate (KIO₃, 1 g) and 1 M sulfuric acid (100 mL) were mixed and stirred during 10 min. Afterward, 0.2 g of cetyl trimethylammonium bromide surfactant, along with 1 g of MCM-41, was added. Following a 20-min interval, distilled aniline monomer (1 mL) was mixed in the solution. The color of the solution changed from light to dark violet, indicating the occurrence of polymeric reaction and formation of PAni.

The reaction was performed at room temperature during 5 h. Finally, the composite was filtered and impurities were eliminated. PAni/MCM-41 composite was rinsed several times using distilled water and acetone and placed in an oven during 1 day at a temperature of 70 °C. Finally, PAni/MCM-41 composite was stored in a desiccator for subsequent experiments [6, 7].





PPy/MCM-41 preparation

For preparing PPy/MCM-41 nanocomposite, 5.4 g of ferric chloride was added to distilled water and stirred for 20 min. Afterward, MCM-41 (1 g) and pyrrole (1 mL) were mixed in the filtered solution. For fulfilling the polymerization reaction, the final mixture was stirred for 5 h. The synthetic composite was filtered and dried for 24 h [11].

Characterization

XRD patterns were obtained by a refractometer (35 kV, 28.5 mA, and 298 K; Philips Instruments, Australia). The BET surface area was estimated with respect to nitrogen adsorption at −196 °C. The BJH method (Quantachrome NovaWin2, USA) was utilized to measure the size and volume distribution of pores, based on the N2 adsorptiondesorption isotherm curve. The existence of functional groups on MCM-41 surface was determined with FTIR (8400S, Shimadzu, Japan) at wavenumbers 400-4000 cm⁻¹ based on the KBr method. We used UV-Vis spectrophotometry for determining dye concentration before and after adsorption. TEM images of the adsorbents were obtained by a Philips camera (CM120, the Netherlands) operating at 150 kV. Also, SEM analysis (HITACHI S-4160) was used for identifying the sample morphology.

Batch experiments

Batch experiments were performed with 100 mL of various concentrations of dye solution (20, 40, 60, 80, and 100 mg L^{-1}). The experiments were performed in an incubator at 200 rpm. The pH of dye solutions was adjusted to 2–10 with HCl and NaOH solutions (0.1 M). Also, the impact of adsorbent dosage, primary dye dosage, duration of contact, and temperature was examined.

To obtain the equilibrium concentrations, a UV–Vis spectrophotometer was used (6310, JENWAY, UK). After contact durations of 15, 30, 45, 60, 90, 120, and 180 min, sampling was performed; the samples were centrifuged at a speed of 6000 rpm during 30 min. Equation (1) was used to measure the adsorption potential of the adsorbent:

$$q_{\rm t} = (C_{\rm o} - C_{\rm t})V/M. \tag{1}$$

In this equation, q_t (mg g⁻¹) represents the adsorption efficiency at time t until equilibrium, C_o and C_t (mg L⁻¹) denote dye dosage at baseline and time t, respectively, V denotes dye volume (L), and M represents the concentration of the adsorbent (g). In addition, Eq. (2) was used to calculate the dye removal efficiency:

Removal efficiency =
$$(C_o - C_t)/C_o \times 100$$
. (2)

Results and discussion

Characterization

Figure 2 demonstrates the N_2 adsorption/desorption isotherms related to MCM-41, PAni/MCM-41, and PPy/MCM-41 nanocomposites. Isotherm type IV was recognizable for MCM-41 with hysteresis. An obvious rise in the adsorbed N_2 was reported at P/P_0 of 0.6–0.8, which characterizes mesoporous materials [19–21]. It was concluded from the BJH method that MCM-41 has a limited pore size distribution (average, 6.7 nm). Table 1 demonstrates parameters including the average pore diameter, as well as pore volume in MCM-41 and synthesized nanocomposites.

After pyrrole and aniline polymerization, the inflection point of the nanocomposite isotherm changed to a lower P/P_0 ; based on this finding, PPy and PAni were in the MCM-41 channels [21, 22]. Moreover, the reduction in BET surface area, as well as pore volume and size, in PPy/MCM-41 and PAni/MCM-41 nanocomposites clearly shows that the polymers penetrated into the MCM-41 channels [19, 21, 22].

According to the findings, the MCM-41 BET surface area reduced from 1003.61 to 434.89 and 535.34 $\rm m^2~g^{-1}$ after loading with PPy and PAni, respectively. PPy with smaller molecules could rapidly fill the MCM-41 pores in comparison with PAni; therefore, it caused pore obstruction and decreased the BET surface area.

Figure 3 indicates the XRD analysis of synthesized MCM-41, PAni/MCM-41, and PPy/MCM-41 in the range of $0 < 2\theta < 10$. There was a sharp peak at $2 < 2\theta < 3$, as well as two weaker peaks at $3 < 2\theta < 5$, which are

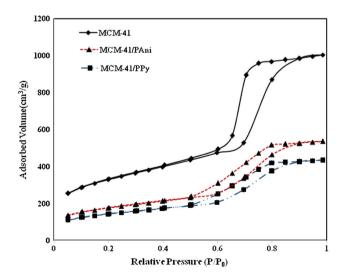


Fig. 2 N_2 adsorption–desorption isotherms of MCM-41, PAni/MCM-41, and PPy/MCM-41



Table 1 Characterization results of MCM-41, PAni/MCM-41, and PPy/MCM-41							
Sample	Surface area (BET) (m ² g ⁻¹)	Pore diameter (BJH) (nm)					

Sample	Surface area (BET) (m ² g ⁻¹)	Pore diameter (BJH) (nm)	Pore volume (BJH) (cm ³ g ⁻¹)
MCM-41	1003.61	6.7	0.78
MCM-41/PAni	535.34	4.8	0.47
MCM-41/PPy	434.89	3.6	0.32

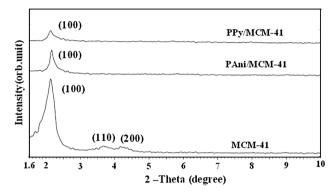
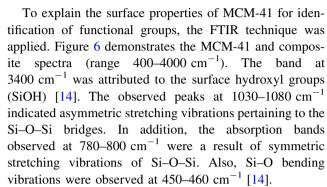


Fig. 3 XRD patterns of MCM-41, PAni/MCM-41, and PPv/MCM-41

attributed to (100), (110), and (200) planes of the MCM-41 structure [23]. According to Fig. 3, following PPy and PAni loading, the XRD pattern of MCM-41 was comparable with pure MCM-41. Nevertheless, the diffraction intensity of PPy/MCM-41 was below that of MCM-41 and PAni/MCM-41. MCM-41 pore filling with PPy might be the probable reason for the reduction in peak intensity.

Coating of MCM-41 outer surface by amine groups did not majorly affect MCM-41 diffraction intensity [24]. According to this finding, it can be concluded that quick pore filling in the presence of PPy is attributed to the smaller molecular size of PPy, compared to PAni. On the other hand, as for the small size of PPy particles, they can penetrate into the MCM-41 structure, generating more positions to absorb a constant volume. Since a larger number of molecules can penetrate into the pore throughout MCM-41 loading, the structure rapidly collapses and leads to the reduction of diffraction intensity [19, 22, 25].

In Fig. 4, SEM results are presented. The MCM-41 and composites are formed with small particle sizes (almost uniform). The increase in MCM-41 particle size after loading indicates the contribution of PPy and PAni to the formation of composites. In addition, TEM findings are presented in Fig. 5. It can be observed that both MCM-41 and composites had hexagonal well-ordered mesoporous structures, where organic compounds could easily access the active sites. Therefore, MCM-41 pore structure after PPy and PAni loading remained undamaged, and PPy and PAni were evenly scattered on the MCM-41 surface [14]. Also, the light and black points on the surface of the composite are representative of MCM-41 nanopowder and PAni-PPy, respectively.



The C=C bonds at 1584 and 1492 cm⁻¹ were related to quinonoid and benzenoid rings, respectively, in the PAni/ MCM-41 composite; the peaks at 1300 cm⁻¹ were related to C-N [26]. The peak at 810 cm⁻¹ indicated out-of-plane C–H deformation in the Π -disubstituted benzene ring. The strong broad band at 3427 cm⁻¹ for PPy/MCM-41 was attributed to PPy stretching vibrations. Also, C-H vibrations accounted for the bands at 2918 cm⁻¹, whereas bands at 2356 cm⁻¹ were related to stretching vibrations of C-N.

C=C ring stretching of pyrrole accounted for the absorption band at 1635 cm⁻¹. The band at 1305 cm⁻¹ was attributed to C-H vibrations. Also, the peak at 1091 cm⁻¹ was related to C–O symmetric stretching and in-plane O–H deformation. C-H deformation vibrations in the CH=CH group could account for the peak at 910 cm⁻¹ [21, 25, 27].

Comparison of MCM-41 and PPv/MCM-41 FTIR patterns indicated that some bands had disappeared and, therefore, PPy was incorporated in MCM-41 particles. In addition, the distinguished absorption peak of PPy/MCM-41 near 3427 cm⁻¹ decreased following AB62 adsorption. This finding showed that PPy/MCM-41 pore structure changed with dye adsorption probably because of inherent disorder, and not the MCM-41 structure collapse.

Adsorption assessments

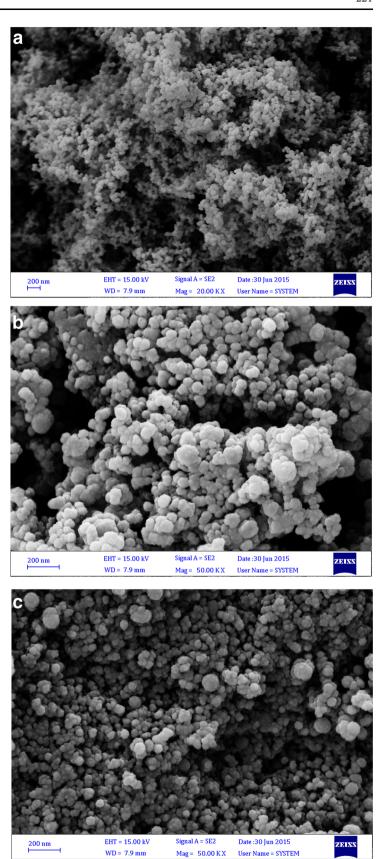
pH

Overall, pH is the most effective parameter during adsorption, which suppresses the quantity of dye ions on adsorbent active sites [28]. The removal efficiency of AB62 was examined in different pH ranges (2-10), whereas other variables including the adsorbent quantity, dye dosage, and temperature remained constant. The

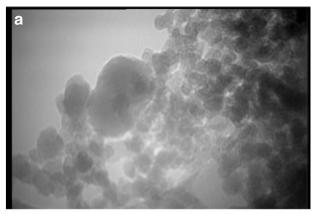


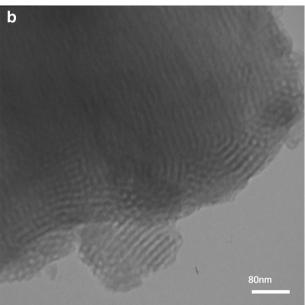


Fig. 4 SEM images of MCM-41, PAni/MCM-41, and PPy/MCM-41









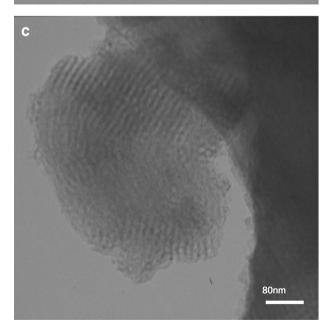


Fig. 5 TEM images of MCM-41, PAni/MCM-41 and PPy/MCM-41

importance of pH in AB62 adsorption on the origin and modified MCM-41 is presented in Fig. 7.

The primary doses of dye and composites were 40 mg L⁻¹ and 0.02 g, respectively. As can be seen, a rise in pH from 2 to 10 led to a decline in dye adsorption. The composite surface could attract positive charges at a lower pH [29], and the powerful electrostatic bond between the positive and negative charges of composite and dye molecules, respectively, could enhance dye adsorption. In addition, lower AB62 adsorption, observed at alkaline pH, might be associated with electrostatic repulsion among anionic dye molecules and OH ions [5].

In comparison with PPy, PAni has a larger molecular size and a lower dye adsorption considering the steric hindrance. In fact, steric hindrance enhances, as the molecular size of the polymer increases [30]. In addition, unmodified MCM-41 had a lower dye removal efficiency than MCM-41 modified with PPy; also, PAni/MCM-41 showed lower efficiency in acidic conditions.

Adsorbent concentration

The significance of adsorbent concentration was examined, using 0.02–0.2 g of both composites and 40 mg L⁻¹ of dye solution in 100 mL solution (pH 2). As presented in Fig. 8, by increasing the concentration of PPy/MCM-41 and PAni/MCM-41 to 0.1 and 0.15 g, respectively, we could enhance the removal efficiency (up to 81 and 38%, respectively). However, further increase of composites had no effects on removal efficiency due to aggregation of composites [31, 32]. Therefore, 0.1 g of PPy/MCM-41 and 0.15 g of PAni/MCM-41 were considered as the optimal adsorbent dosages for AB62 removal.

Temperature

For evaluating the effect of temperature on PPy/MCM-41 and PAni/MCM-41 removal, assessments were performed at various temperatures with primary dye doses of 40–100 mg L⁻¹, 0.1 g of PPy/MCM-41, and 0.15 g of PAni/MCM-41 during 90 min of equilibrium (pH 2). As Fig. 9 presents, removal reduced by increasing the dye dosage and enhanced by increasing the temperature to 323 K; this finding might be attributed to the increased activity on the surface [5, 6]. Therefore, AB62 adsorption on nanocomposite adsorbents is both a spontaneous and an endothermic reaction.

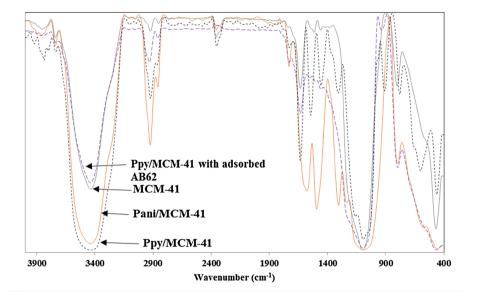
Contact time

Dye adsorption by PPy/MCM-41 and PAni/MCM-41 was evaluated at 20 °C; the findings are presented in Fig. 10. At the beginning of the adsorption process, rapid dye removal





Fig. 6 FTIR spectra of MCM-41, PAni/MCM-41, and PPy/MCM-41 before and after AB62 adsorption



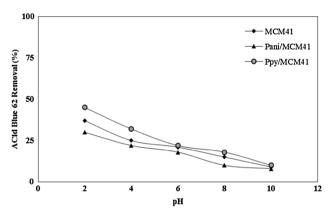


Fig. 7 Effect of pH on the percentage of dye removal by MCM-41, PAni/MCM-41, and PPy/MCM-41 (adsorbent dosage, 0.02 g, 100 mL of dye solutions, initial dye concentration, 40 mg $\rm L^{-1}$, contact time, 90 min, and T 20 °C)

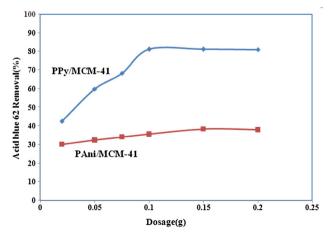


Fig. 8 Effect of adsorbent dosage on the percentage of dye removal (initial dye concentration 40 ppm; pH 2; contact time 90 min; T 20 °C)

occurred, given more access to the active sites. Afterward, adsorption gradually slowed down until equilibrium at 90 min. In the equilibrium state, the adsorption sites of the composites were filled [33]. Since PPy/MCM-41 nanocomposite showed higher removal efficiency than PAni/MCM-41 in all the experiments, 0.15 g of PPy/MCM-41 was considered as the optimal dosage and the subsequent experiments were carried out with 0.15 g of PPy/MCM-41.

Adsorption thermodynamics

We used thermodynamic analysis to understand the characteristics and mechanisms of adsorption. ΔG° , ΔH° , and ΔS° values were measured using the following formulae:

$$k_{\rm c} = C_{\rm Ae}/C_{\rm e},\tag{3}$$

$$\Delta G^{0} = -RT \ln kc, \tag{4}$$

$$\Delta G^{\rm o} = \Delta H^{\rm o} - T \Delta S^{\rm o}, \tag{5}$$

$$\log k_{\rm c} = \Delta S^{\rm o}/2.303R - \Delta H^{\rm o}/2.303RT,\tag{6}$$

where ΔH (kJ mol⁻¹) is measured with respect to the slope of $\log k_{\rm c}$ vs. 1/T plot and ΔS (J mol⁻¹ K⁻¹) is measured relative to the plot's intercept. Also, $\Delta G^{\rm o}$ was measured using Eq. (6). The results of thermodynamic analysis are shown in Table 2.

According to Table 2, increasing the temperature could enhance adsorption. Therefore, dye adsorption on PPy/MCM-41 was endothermic, as confirmed by the positive value of $\Delta H^{\rm o}$. On the other hand, the heat of physical adsorption was less than 21 kJ mol⁻¹. Based on the $\Delta H^{\rm o}$ values, AB62 adsorption on PPy/MCM-41 was in the form of physical adsorption [5, 34, 35]. The negative $\Delta G^{\rm o}$ values confirmed the feasibility of dye adsorption. Also, positive



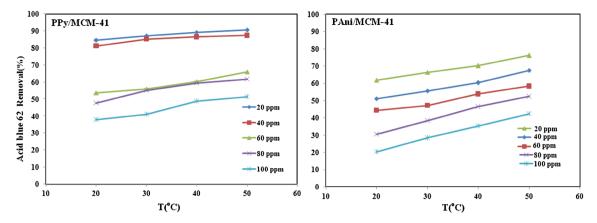


Fig. 9 Effect of temperature on dye removal (initial dye concentration, 40 up to 100 ppm; nanocomposite dosage; 0.1 g of PPy/MCM-41 and 0.15 g of PPy/MCM-41; pH of 2; and contact time 90 min)

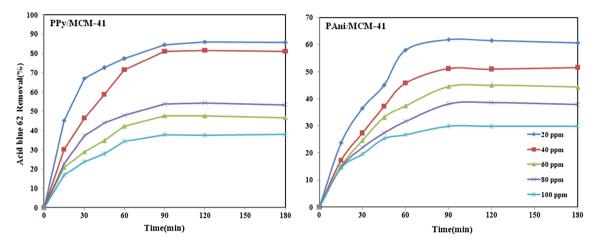


Fig. 10 Effect of contact time on the percentage of dye removal (initial dye concentration, 40 up to 100 ppm; 0.15 g of PAni/MCM-41 and 0.1 g of PPy/MCM-41 in 100 mL of dye solutions; pH 2; and T 20 °C)

Table 2 Thermodynamic parameters for the adsorption of Acid Blue 62 onto the PPy/MCM-41

Initial dye concentration (mg L ⁻¹)	$\Delta H^{\rm o} ({\rm kJ~mol}^{-1})$	$\Delta S^{o} (J \text{ mol}^{-1} K)$	20 °C	$\Delta G^{\rm o}$ (kJ mol ⁻¹) 30 °C	40 °C	50 °C
20	14.80	64.75	-4.14	-4.82	-5.51	-6.07
40	13.44	54.01	-3.55	-4.41	-4.83	-5.19
60	12.12	46.77	-0.36	-0.60	-1.08	-1.78
80	11.88	40.16	-0.22	-0.52	-1.00	-1.27
100	10.12	38.21	-0.20	-0.41	-0.88	-0.91

 ΔS^{o} values indicated the reversible adsorption of dye on PPy/MCM-41 [5, 11].

Adsorption isotherm equations

Adsorption isotherm equations describe the adsorbent-adsorbate interactions. Isotherm models are applied for designing the adsorption process. To evaluate dye adsorption isotherms, equilibrium data were prepared with 0.1 g of PPy/MCM-14 nanocomposite at 50 °C (pH 2). In this

study, five important isotherms were applied to investigate the dye adsorption.

Langmuir isotherm equation

This equation, which is based on monolayer adsorption, was applied to determine the adsorption efficiency. Adsorption was observed on the same active sites on the surface, without any interactions among the adsorbate molecules [11].





The nonlinear Langmuir isotherm equation is as follows:

$$q_{\rm e} = q_{\rm m} k_{\rm L} C_{\rm e} / (1 + k_{\rm L} C_{\rm e}),$$
 (7)

where $q_{\rm e}$ denotes the adsorption efficiency in equilibrium (mg g⁻¹), $C_{\rm e}$ the equilibrium dosage (mg L⁻¹), $q_{\rm m}$ the highest adsorption potential (mg g⁻¹), and $k_{\rm L}$ the Langmuir constant. The binding affinity of dye molecules is indicated by high $k_{\rm L}$ values.

Freundlich isotherm equation

This equation is derived from empirical information and is generally applied for the expression of heterogeneous surfaces [36]:

$$q_{\rm e} = K_{\rm f} C_{\rm o}^{1/n},\tag{8}$$

where K_f represents the adsorption capacity (L g⁻¹), n (a constant) the intensity of adsorption, and 1/n the heterogeneity of the composite surface; n values between 2 and 10 exhibit favorable adsorption.

Redlich-Peterson model

In this three-parameter equation, Langmuir and Freundlich isotherms are combined [37]:

$$q_{\rm e} = k_{\rm R} C_{\rm e} / \left(1 + \alpha_{\rm R} C_{\rm e}^{\beta} \right), \tag{9}$$

where $k_{\rm R}$ (L g⁻¹) and $\alpha_{\rm R}$ (L mg⁻¹) denote the Redlich–Peterson constants and β represents the equation exponent. The Freundlich isotherm is the supreme isotherm when β is

near 0, while the Langmuir model is predominant when β is close to the unit.

Dubinin-Radushkevich equation

It is written as [38]:

$$q_{\rm e} = q_{\rm m}e^{-\beta\varepsilon 2},\tag{10}$$

where $q_{\rm m}$ denotes the monolayer efficiency (mg g⁻¹), β the adsorption energy constant, and ε the Polanyi adsorption potential:

$$\varepsilon = RT \ln(1 + 1/C_c). \tag{11}$$

In this equation, R represents the gas constant (8.314 J^{-1} mol K^{-1}), T the absolute temperature, β the mean free energy, and ε the dye molecules moved to the solid surface. The relation of ε with β is described as follows:

$$\varepsilon = 1/[(2\beta)0.5]. \tag{12}$$

Temkin isotherm

For investigating multilayer adsorption, the Temkin isotherm was applied [39, 40]:

$$q_{\rm e} = B \ln(AC_{\rm e}). \tag{13}$$

Parameters of the isotherms were measured, based on the linear regression of nonlinear forms in the models, as shown in Table 3. Comparison was performed between the

Table 3 The values of parameters for isotherm models

Isotherm	Nonlinear equation	Linear form	Parameter	R^2
Langmuir 1	$q_{\rm e} = q_{\rm m} K_{\rm L} C_{\rm e} / (1 + K_{\rm L} C_{\rm e})$	$C_{\rm e}/q_{\rm e} = 1/k_{\rm l}q_{\rm m} + (1/q_{\rm m}) C_{\rm e}$	$q_{\rm m} = 55.55 \text{ mg L}^{-1}$	0.988
		$C_{ m e}/q_{ m e}$ vs. $C_{ m e}$	$K_{\rm l} = 0.227 \; {\rm L \; mg^{-1}}$	
Langmuir 2		$1/q_{\rm e}=1/k_{\rm l}q_{\rm m}C_{\rm e}+1/q_{\rm m}$	$q_{\rm m} = 55.55 \text{ mg L}^{-1}$	0.972
		$1/q_{\rm e} \ {\rm vs.}1/C_{\rm e}$	$K_1 = 0.818 \text{ L mg}^{-1}$	
Langmuir 3		$q_{\rm e} = q_{\rm m} - q_{\rm e} / k_{\rm L} C_{\rm e}$	$q_{\rm m} = 52.47 \text{ mg L}^{-1}$	0.878
		$q_{\rm e}$ vs. $q_{\rm e}/C_{\rm e}$	$K_1 = 0.878 \text{ L mg}^{-1}$	
Langmuir 4		$q_{\rm e}/C_{\rm e} = q_{\rm m}/k_{\rm l} - q_{\rm e}/k_{\rm l}$	$q_{\rm m} = 54.47 \text{ mg L}^{-1}$	0.878
		$q_{\rm e}/C_{\rm e}$ vs. $q_{\rm e}$	$K_1 = 3.73 \text{ L mg}^{-1}$	
Freundlich	$q_{ m e}=K_{ m f}C_{ m e}^{1/n}$	$\ln q_{\rm e} = \ln K_{\rm f} + 1/n \ln C_{\rm e}$	n = 3.41	0.887
	1	Ln $q_{\rm e}$ vs. ln $C_{\rm e}$	$K_{\rm f} = 17.38 \; ({\rm mg \; g^{-1}}) \; ({\rm L \; mg^{-1}})^{1/n}$	
Redlich-Peterson	$q_{\rm e} = k_{\rm R} C_{\rm e} / \left(1 + a_{\rm R} C_{\rm e}^{\beta}\right)$	$\operatorname{Ln}(K_{\mathrm{R}}*(C_{\mathrm{e}}/q_{\mathrm{e}})-1) \operatorname{vs.}\beta \ln C_{\mathrm{e}} + \ln \alpha_{\mathrm{R}}$	$K_{\rm R} = 17.85 \; {\rm L} \; {\rm g}^{-1}$	0.911
	,		$\alpha_{\rm R} = 0.38 \; (L \; mg^{-1})^{\beta}$	
			$\beta = 0.35$	
Temkin	$q_{\rm e} = B \ln(AC_{\rm e})$	$q_{\rm e}$ vs. ln $C_{\rm e}$	$B = 19.63 \text{ J mole}^{-1}$	0.901
			$A = 11.78 \text{ L g}^{-1}$	
Dubinin-Radushkevich	$q_{ m e}=q_{ m m}e^{-etaarepsilon 2}$	Ln $q_{\rm e}$ vs. $^2_{\scriptscriptstyle \mathcal{E}}$	$q_{\rm m} = 46.3 \ {\rm mg \ g^{-1}}$	0.881
			$\beta = 0.13$	
			$E = 1/\sqrt{2\beta} = 1.9 \text{ kJ mole}^{-1}$	



Initial DY86	Pseudo-first order			Pseudo-second-order R^2				Intraparticle diffusion		
Concentration (mg L^{-1})	$k_{\rm ad} $ (1 min ⁻¹)	$q_{ m e,cal}$. (mg g ⁻¹)	R^2	$k_{\rm h}$ (g mg ⁻¹ min)	$q_{ m e,cal}$. (mg g ⁻¹)	$q_{\rm e,ex}$. (mg g ⁻¹)	R^2	$\frac{K_{\rm p}}{(\text{mg g}^{-1} \text{min}^{0.5})}$	$C \text{ (mg g}^{-1}\text{)}$	R^2
20	0.036	22.20	0.97	0.0028	20.40	16.91	0.99	1.33	5.00	0.90
40	0.036	38.70	0.96	0.0003	52.63	32.45	0.99	3.24	1.87	0.98
60	0.032	45.38	0.99	0.0007	43.47	32.66	0.99	3.73	2.12	0.94
80	0.034	56.68	0.94	0.0004	55.55	38.17	0.98	3.93	2.80	0.98
100	0.034	54.24	0.93	0.0005	52.63	40.14	0.98	4.14	2.95	0.97

Table 4 Pseudo-first order, pseudo-second-order, and intraparticle diffusion model parameters for Acid Blue 62 adsorption on PPy/MCM-41 adsorbent

experimental information and model findings with respect to R^2 values:

Langmuir(type 1) > Redlich-Peterson >

Temkin > Dubinin-Radushkevich > Freundlich.

Based on the obtained results, Langmuir equation showed the greatest correlation coefficient indicative of monolayer adsorption. Regarding the Freundlich isotherm, n was measured to be 3.41, indicating suitable adsorption. The Redlich–Peterson isotherm correlation coefficient (0.911) was below the value obtained by the Langmuir model; as a result, the Langmuir model was introduced as the supreme isotherm.

To continue the experiments, the Dubinin–Radushke-vich model was used. The correlation coefficient of this model (0.881) was lower than that of Langmuir. We also applied Temkin isotherm to evaluate multilayer adsorption. The correlation coefficient was measured to be 0.901, which is below that reported in the Langmuir model. Therefore, it can be concluded that monolayer adsorption happened, and the Langmuir model is the predominant one.

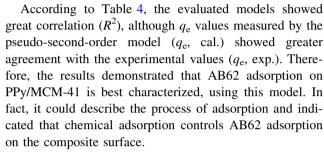
Adsorption kinetics

Using different models, the mechanism of adsorption was evaluated. The pseudo-first and pseudo-second-order linear models are commonly used to fit the kinetic information via Eqs. (14) and (15), respectively:

$$\ln(q_e - q) = \ln(q_e) - k_1 t, \tag{14}$$

$$t/q = 1/k_{\rm h}q_{\rm e}^2 + t/q_{\rm e},\tag{15}$$

where q_t denotes the adsorbate amount in each adsorbent unit at time t (mg g⁻¹), k_1 the pseudo-first-order constant, and k_h the second-order constant. The rate constants, as well as the adsorption efficiencies in the pseudo-first-order model, were measured at various doses with respect to the $\ln(q_e - q_t)$ vs. t plot. In addition, in the other model, the rate constants and adsorption capacities were measured at various doses, based on the t/q_t versus t plot. The findings are demonstrated in Table 4.



The adsorption rate can be managed using intraparticle diffusion, along with other kinetic mechanisms including boundary layer or film diffusion effects. Weber and Morris first presented the model of intraparticle diffusion [41]:

$$q_{\rm t} = k_{\rm p} t^{0.5} + C. (16)$$

In this equation, C denotes the intercept, k represents the intraparticle rate constant, and k_p is calculated with respect to the slope of q_t (mg g⁻¹) versus $t^{0.5}$ plot. The k_p values are presented in Table 4. Based on the findings, given the great driving force, the increase in diffusion rate was associated with an increase in the primary dye dosage.

Intraparticle diffusion characterizes the rate-control stage if the linear regression of q_t versus $t^{0.5}$ plot crosses

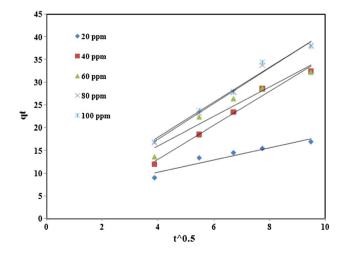


Fig. 11 Intraparticle diffusion model for the adsorption of Acid Blue 62 onto PPy/MCM-41

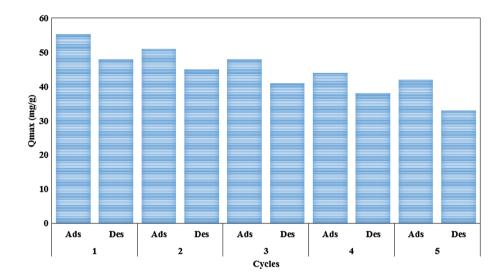




Table 5 A comparison between the results of the present work and some reported results in literature

Adsorbent	Dye	$q_{ m e}$	Refs.
Activated carbon	Acid Red 97	52.08	[43]
Activated red mud	Acid Blue 113	83.33	[44]
Spinel ferrite nanoparticles	Acid Red 88	111.1	[45]
copper oxide nanoparticle loaded on activated carbon	Acid Blue 129	65.36	[46]
Ppy/MCM-41	Acid Blue 62	55.55	This wo

Fig. 12 Adsorption and desorption behaviors of AB62 dye on PPy/MCM-41



the origin. Nevertheless, it is obvious from Fig. 11 that the linear graphs did not cross the origin. As a result, other mechanisms could have affected the adsorption rate (it is not only dependent on intraparticle diffusion) [11, 42]. Table 5 presents the comparison of AB62 adsorption on PPy/MCM-41 with other adsorbents used in previous studies [43–46]. As can be seen, the presented adsorbent generally has a good adsorption capacity for AB62. The prepared adsorbent showed a higher adsorption capacity than other composites.

Adsorption mechanism

With respect to AB62 adsorption on different adsorbents (i.e., MCM-41/PPy and MCM-41/PAni), various mechanisms can be convoluted including ionic bonding between cationic functional groups of the adsorbents and anionic functional group(s) in dissolved dye molecules. During adsorption, different phenomena may occur. First, in the aqueous medium, AB62 is dissolved and sulfonate groups of dye transform into anionic dye ions (17). In the second step, the adsorbent amino groups (PPy and PAni) are protonated in the acidic medium (18). According to Eq. (19), adsorption continues, given the electrostatic bond between the counter ions:

$$Dye-SO_3Na \rightarrow DyeSO_3^- + Na^+, \tag{17}$$

$$(PAni \,\&\, PPy) - NH + H^+ \rightarrow (PAni \,\&\, PPy) - NH_2^+, \eqno(18)$$

$$(PAni \& PPy) - NH_2^+ + Dye SO_3^-$$

$$\rightarrow (PAni \& PPy) - NH_2 + -3OSDye.$$
(19)

Desorption studies

By performing desorption assessments, we can analyze the adsorption mechanism and the possibility of adsorbent reuse. The interaction of the adsorbent surface with dye molecules (either a strong or a weak bond) determines the reversibility of adsorption. Alkaline solutions (pH 10) were applied to desorb AB62 dye from PPy/MCM-41. The results of desorption study for five cycles are presented in Fig. 12. As can be seen, PPy/MCM-41 showed desorption ability.

Based on our findings, weak forces mainly accounted for the bonding between the adsorbent surface and AB62 molecules and could lead to high desorption capacity and produce adsorbents; the activation energy and thermodynamic parameters also support this finding. The adsorbent showed acceptable adsorption efficiency for AB62 and might be an acceptable adsorbent for treatment purposes.



Conclusion

PPy/MCM-41 and PAni/MCM-41 nanocomposites were successfully synthesized through chemical polymerization and were described using different techniques. The adsorption of AB62 on both synthesized nanocomposites was evaluated in aqueous solutions. Batch experiments demonstrated that the dye removal efficiency of PPy/ MCM-41 nanocomposite was more favorable than PAni/ MCM-41. The smaller size of PPy molecules in comparison with PAni makes it more capable of removing dyes. Moreover, the significance of pH, adsorbent dosage, duration of contact, and temperature in dye removal was studied. The removal capacity showed adverse dependence on pH, which enhanced by increasing the duration of contact, adsorbent dosage, and temperature. The highest removal efficiency was 90.57% at a pH of 2 and contact time of 90 min, using 20 mg L⁻¹ of AB62 and 0.1 g of PPv/MCM-41 nanocomposite. Thermodynamic studies showed that adsorption is both an endothermic and a spontaneous reaction. Moreover, the findings reported in equilibrium were compatible with the Langmuir isotherm, adsorption capacity on PPy/MCM-41 55.55 mg g⁻¹ at a primary dye dosage of 20 mg L⁻¹. Based on the intraparticle diffusion model, adsorption is influenced by a boundary layer, as well as intraparticle diffusion. Also, the pseudo-second-order model showed the greatest agreement with the experimental findings.

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