#### **ORIGINAL PAPER**



# Surface-functionalized spongy zinc ferrite as a robust visible-light driven nanocatalyst for wastewater remediation: characterization, kinetic, and mechanistic insight

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#### Abstract

Three different types of zinc ferrite nanoparticles ( $ZnFe_2O_4$  NPs) were successfully synthesized through the co-precipitation route. The synthesized ZnFe<sub>2</sub>O<sub>4</sub> NPs were structurally characterized by Fourier transform infrared (FT-IR), Scanning electron microscopy (SEM), X-ray diffraction (XRD), and Brunauer Emmet Teller spectroscopic analyses. FT-IR analysis confirmed that the (3-APTES) is well grafted on the ZnFe<sub>2</sub>O<sub>4</sub> NPs surface. XRD analysis shows the crystalline structure of calcined and modified  $ZnFe_2O_4$  NPs, while the amorphous nature of bare  $ZnFe_2O_4$  NPs was observed. SEM micrographs displayed a flat surface morphology with almost monodispersed grain size and irregular shape structure for all  $ZnFe_2O_4$  NPs. The grain size of the synthesized  $ZnFe_2O_4$  NPs was found to be 106.5, 23.3, and 13.9 nm for bare, calcined, and modified  $ZnFe_2O_4$ NPs, respectively. Optical bandgap energy (LUMO-HOMO gap) was experimentally calculated in acidic and basic mediums (2.47 and 2.70 eV) using a taulc plot. The photocatalytic performance of the synthesized ZnFe<sub>2</sub>O<sub>4</sub> NPs was investigated under visible irradiation in an aqueous medium against bromophenol blue dye. The influence of various parameters like irradiation time, photocatalyst dose, dye dose, pH effect, and H<sub>2</sub>O<sub>2</sub> on the degradation efficiency was investigated. Results showed that all the three types of  $ZnFe_2O_4$  NPs exhibited excellent photocatalytic properties, where the highest degradation (95.4%) was displayed by modified ZNFe<sub>2</sub>O<sub>4</sub> NPs, followed by calcined (90.2%) and bare ZnFe<sub>2</sub>O<sub>4</sub> 84.6\%). The synthesized  $ZnFe_2O_4$  NPs were reused and recycled, retaining the excellent photocatalytic property for four consecutive cycles. Kinetic models demonstrate that the degradation rate was pseudo 1st order of kinetics. The turnover number reached up to  $21.3 \times 10^6$ ,  $22.5 \times 10^{6}$ , and  $23.8 \times 10^{6}$  for bare, annealed, and functionalized ZNFe<sub>2</sub>O<sub>4</sub> nanoparticles, respectively. In short, the developed spongy zinc ferrite nanoparticles indicate a high potential of photo-driven catalytic remediation of tested bromophenol blue drained into the water systems.

 $\label{eq:constraint} \begin{array}{l} \mbox{Keywords} \ \mbox{Spongy zinc ferrite} \cdot \mbox{Surface morphology} \cdot \mbox{Grafting} \cdot \mbox{Optical bandgap} \cdot \mbox{Nanospheres} \cdot \mbox{Photocatalysis} \cdot \mbox{Remediation} \end{array}$ 

# Introduction

Dyes are among the most prominent organic pollutants polluting the aqueous environment (Fuku et al. March 2017; Rauf and Ashraf 2009). Various industries discharge large amounts of dyes into the water worldwide. Removing these dyes from water and wastewater is an environmental priority because mostly, these dyes are resistant to bacteriological

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decomposition (Nezamzadeh-Ejhieh and Banan 2012). Water pollution is a serious problem as it affects human health. More than 4-million people are annually dying from infections caused by microorganisms, and most of them are caused by water pollution (Nezamzadeh-Ejhieh and Moazzeni 2013; Buthiyappan et al. 2016). So, eliminating biological and chemical pollutants from contaminated water is a challenging task. To remove dyes from textile wastewater, it is essential to search for an effective wastewater treatment process (Magdalane et al. 2017). Various chemical and physical methods, such as adsorption, separation of pollutants, electrocoagulation, chemical precipitation, etc., (Bouanimba et al. 2011; Anchieta et al. 2014) are currently used for wastewater treatment. However, the fundamental limitations



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of these methods are that they are not effective because only contaminants relocate from one phase to another; therefore, different, and new kinds of contamination are faced, and additional treatment is deemed necessary (Rishikeshi et al. 2013). Advanced oxidation process (AOPs) (Buthiyappan et al. 2016), is an alternative conventional method, which works based on the generation of large reactive groups like OH, that non-selectively and quickly oxidizes a broad range of organic contaminants (Khan et al. 2015; Nezamzadeh-Ejhieh and Banan 2011).

The photocatalytic application of  $ZnFe_2O_4$  NPs is explored under visible-light irradiation against the decolorization of bromophenol blue. Due to its polar nature, it is a highly water-soluble dye and can be originated in the effluents of textiles, pharmaceuticals, and chemical industries (Dlamini et al. 2011; Haider et al. 2011). To find a close association among shape and various properties, ZnFe<sub>2</sub>O<sub>4</sub> NPs with diverse morphology, such as multi-porous bricks, nanorods, nanoplates, nanocubes, nanoparticles, etc., have been synthesized by multiple researchers (Srinivasan et al. 2016). Cao et al. (2009) prepared  $ZnFe_2O_4$  nanoparticles for the degradation of phenol by Microwave-hydrothermal ionic liquid (MHIL) method that exhibit up to 73% photocatalytic activity within 30 min. Li et al. (2010) developed ZnFe<sub>2</sub>O<sub>4</sub> nanospheres via the solvothermal route. Under Xelamp, the photocatalytic removal efficiency of Rhodamine B (RhB) was 84% after 4 h of irradiation. Xie, Taiping, et al. (Xie et al. 2013) synthesized ZnFe<sub>2</sub>O<sub>4</sub>/SrFe<sub>12</sub>O<sub>19</sub> through a one-step chemical co-precipitation route and studied the photocatalytic activity of composite against the removal of methylene blue under UV-irradiation. The removal rate of composites was greater than 70% after being reused for 4-times. Li, Xinyong, et al. (He et al. 2017) prepared ZnFe<sub>2</sub>O<sub>4</sub> nanospheres via solvothermal route, and under xenon-lamp irradiation, the photocatalytic degradation efficiency of RhB reached 85% after 3 h.

Neshin et al. (2021) reported Ni-doped BiFeO<sub>3</sub> NPs for degradation of BPB in aqueous media. Their photocatalyst efficiency was recorded as 85.76% (Mohsen et al. 2021). Similarly, Humayun Khan et al. (2016) reported Chitosan conjugated magnetic (Ni-CCMN) NPs for the degradation of BPB. They reported that the photocatalyst could degrade BPB up to 83% (Khan et al. 2016). Ameen et al. (2014) synthesized CeO<sub>2</sub>/TiO<sub>2</sub> nanocomposite as a visible-light photocatalyst for the degradation of BPB. They observed that the prepared photocatalyst can degrade BPB up to 72%(Ameen et al. 2014). Shah et al. (2019) reported TiO<sub>2</sub>/GNP nanoplatelets for the degradation of BPB. The observed efficiency was 86% (Shah et al. 2019). Emmanuel et al. (2021) reported Clinoptilolite/Activated Carbon for the degradation of BPB. The observed efficiency was 80.0% for the degradation/ removal of BPB (Emmanuel and Sudesh 2021). In comparison, we synthesized spongy ZnFe<sub>2</sub>O<sub>4</sub> NPs by



co-precipitation method, which is simple and cost-effective. We recorded that the spongy  $ZnFe_2O_4$  photocatalyst has a degradation efficiency of greater than 95%.

Different groups synthesized and investigated zinc ferrite ZnFe<sub>2</sub>O<sub>4</sub> NPs as photocatalysts for the degradation of different dyes. But more salts (organic) contents are used in reactions to having microstructure morphology at expensive costs. Secondly, lots of waste materials are also produced as a byproduct (Fan et al. 2009). Therefore, environmentfriendly, easy, and fast methods like sol-gel, hydrothermal, ball mill, sonication, co-precipitation, etc., are mostly preferred for the synthesis of ZnFe<sub>2</sub>O<sub>4</sub> NPs nanostructures. Herein, we prefer the method of co-precipitation for the synthesis of spongy zinc ferrite ZnFe<sub>2</sub>O<sub>4</sub> nanocatalyst/photocatalyst. Spongy ZnFe<sub>2</sub>O<sub>4</sub> photocatalyst is gaining more attention due to their excellent applications as gas sensor, mechanical properties, electric nature, and most importantly, their Photocatalysis character. Spongy ZnFe<sub>2</sub>O<sub>4</sub> photocatalyst shows good optical absorption due to its narrow bandgap of about 1.9 eV (Cao et al. 2009). Due to their magnetic properties, these spongy photocatalysts can be reused again and again as effective photocatalysts to eliminate industrial organic pollutants (Fu and Wang 2011).

The spongy nanostructures were passed through calcination at high temperatures, i.e., 1000 °C, and hydroxylation treated with 1 M HCl. It has been noticed that the surface properties of the photocatalyst become more efficient after treating with 3-(triethoxysilyl) propylamine. Different characterization confirmed the enhanced properties and morphology of the ZnFe2O4 nanostructures zinc ferrite ZnFe2O4 nanostructures. In this research study, the degradation of Bromophenol blue (BPB) dye was investigated using these synthesized spongy zinc ferrite photocatalyst. Bromophenol Blue (BPB) dye is used to examine the efficiency and photo-degradation of the synthesized photocatalyst. Marvelously, BPB shows 95.4% degradation at pH-12 in 2 h. Diverse parameters have been investigated; kinetics models were applied, demonstrating that the degradation process was pseudo 1<sup>st</sup> order kinetics. The catalytic performance concerning turnover frequency (TOF) or turnover number (TON are estimated to measure the activities of different catalysts in numerous disciplines of heterogeneous, homogeneous, and nanocatalysis.

# Materials and methods

## Reagents

Zinc sulfate ( $\geq$ 99%), iron sulfate ( $\geq$ 97%), triethylene glycol ( $\geq$ 99%), sodium hydroxide ( $\geq$ 95%), hydrochloric acid ( $\geq$ 97%), (3-APTES) ( $\geq$ 99%), ascorbic acid ( $\geq$ 99.9%), benzoquinone ( $\geq$ 98%), 1,4-diazabicyclo[2.2.2]octane (DABCO) ( $\geq$  98%), and bromophenol blue dye were purchased from Sigma Aldrich and used as received without further purification.

# Preparation of ZnFe<sub>2</sub>O<sub>4</sub> NPs

Zinc ferrite NPs were successfully prepared through the coprecipitation route by the reaction of iron sulfate (FeSO<sub>4</sub>) (1.822 g, 12 mM) and zinc sulfate  $(\text{ZnSO}_4)$  (0.538 g, 3 mM)in 100 ml of deionized water. Then 50 mM TEG solution was mixed for 2 h with vigorous stirring under ambient conditions. 10 M solution of NaOH was added dropwise into mixed salts mixture to maintain the pH up to 8.0. On this pH, a reddish-brown colored precipitate was achieved, centrifuged, and rinsed several times with distilled water followed by drying for 24 h at 80 °C and mechanically ground into powder. The bare ZnFe<sub>2</sub>O<sub>4</sub> NP was thermally heated at 1000 °C for 6 h in the furnace. It was noted that the triethylene glycol stabilized ZnFe<sub>2</sub>O<sub>4</sub> NPs at the interface due to the OH group of glycols, which may oxidize and make a complex with d-block elements, which may oxidize making a complex with the transition metal (Rishikeshi et al. 2013; Ruttink et al. 2012). The formation of  $[ZnFe_3O(OAc)_2(H_2O)_3)]$  complex following the formation and reduction of ZnFe<sub>2</sub>O<sub>4</sub> NPs on heating is also explained (Skrabalak et al. 2008).

#### Surface modification of ZnFe<sub>2</sub>O<sub>4</sub> NPs

The thermally treated calcined  $ZnFe_2O_4$  NPs were surface modified in two steps. In the initial step, the calcined  $ZnFe_2O_4$  NPs were immersed in one molar hydrochloric acid solution under vigorous stirring for 30 min, centrifuged, and washed several times with deionized water. Secondly, 1 g hydroxylated  $ZnFe_2O_4$  NPs were treated with 0.1 M solution of 3-(APTES) under vigorous stirring at 60 °C for 2 h, filter off, and drying at 80 °C 24 h and obtained the modified  $ZnFe_2O_4$  NPs.

#### Photocatalytic degradation of bromophenol blue

The optimum amount (i.e., 30 mg) of photocatalysts was dispersed in 50 mL of dye solution (20 ppm) in a flask in each reaction. The mixture was stirred in the dark for 30 min to establish adsorption/desorption equilibrium. A 300 W Xenon lamp equipped with a 420 nm UV-cutoff filter was used as the light source. The lamphouse was well-ventilated so that room temperature was maintained. After every certain time, 5 ml of the reaction solution was centrifuged, and the concentration of the dye was measured using Agilent Cary 60 UV–vis spectrophotometer. The pH values of the solutions were 6.1 and did not change during the degradation reactions.

#### Influence of irradiation time

The effect of irradiation contact time on the photocatalytic degradation of BPB was investigated. A known concentration solution of BPB was illuminated with a 300 W Xenon lamp equipped with a 420 nm UV-cutoff filter for various irradiation times ranging from 30 to 180 min. The illuminated solution was centrifuged at 4000 rpm for 20 min, and percent degradation was examined through Agilent Cary 60 UV–vis spectrophotometer.

#### Influence of photocatalyst

The influence of catalyst dose was measured against BPB dye using the various amount of catalyst doses ranging from 10, 15, 20, 25, 30, and 35 mg, respectively. For this purpose, the BPB dye solution was stirred in the dark for 30 min to establish an adsorption/desorption equilibrium and then illuminated with a 300 W Xenon lamp equipped with a 420 nm UV-cutoff filter for the optimized duration (i.e., 120 min). The illuminated solution was centrifuged at 4000 rpm for 20 min, and the percent degradation was examined through Agilent Cary 60 UV–vis spectrophotometer.

#### Influence of dye

The influence of BPB concentration on the photocatalytic decolorization of BPB was measured using different dye concentrations ranging from 2, 10, 15, 20, 25, 30, 35, and 40 ppm. Initially, 100 ppm stock solution of BPB was prepared. The  $\lambda$ max for BPB measured was at 590 nm. For this purpose, the BPB dye solution was treated in the dark for 30 min and then illuminated with a 300 W Xenon lamp equipped with a 420 nm UV-cutoff filter for the optimized duration. The illuminated solution was centrifuged at 4000 rpm for 20 min, and the percent degradation was examined through Agilent Cary 60 UV–vis spectrophotometer.

#### Influence of pH

To check the influence of pH on the photocatalytic decolorization of BPB dye. For this purpose, prepared BPB solutions with various pH (i.e., 1–14), the BPB dye solutions with various pH were treated for 30 min in the dark and then illuminated with 300 W Xenon lamp equipped with a 420 nm UV-cutoff filter for the optimized duration (with optimized photocatalyst and dye dose). The illuminated mixture was centrifuged at 4000 rpm for 20 min, and the degradation was examined through Agilent Cary 60 UV–vis spectrophotometer.



Influence of H<sub>2</sub>O<sub>2</sub>

The influence of solvent on the photocatalytic degradation of BPB was measured using various amount of  $H_2O_2$  (i.e., 1–6 ml). For this purpose, the BPB dye solutions with  $H_2O_2$ were treated for minutes in the dark and then illuminated with visible light under optimum conditions. The illuminated mixture was centrifuged at 4000 rpm for 20 min, and degradation was examined through Agilent Cary 60 UV–vis spectrophotometer.

#### Turnover frequency (TOF) and turnover number (TON)

In catalysis, it's normal to estimate the catalytic performance concerning turnover frequency (TOF) or turnover number (TON). The TOF and TON are estimated to measure the activities of different catalysts in numerous disciplines of heterogeneous, homogeneous, and nanocatalysis. The coupling product yields 85.4 percent, 90.2 percent, and 95.4 percent for the bare, annealed, and functionalized ZnFe<sub>2</sub>O<sub>4</sub> nanoparticles, respectively, with turnover frequencies (TOF) of 1186.11, 1252.77, and 1325. The turnover numbers for bare, annealed, and functionalized ZnFe<sub>2</sub>O<sub>4</sub> nanoparticles reached  $21.3 \times 10^6$ ,  $22.5 \times 10^6$ , and  $23.8 \times 10^6$ , respectively.

# **Results and discussion**

#### Fourier transform infrared (FT-IR) analysis

FT-IR analysis of the prepared  $ZnFe_2O_4$  NPs is shown in Fig. 1. In FT-IR spectra of bare  $ZnFe_2O_4$  NPs, three small characteristic peaks occur at 3175, 3015, and 1638 cm<sup>-1</sup>, attributed to the hydroxyl bond (OH) of triethylene glycol (TEG) (Akpan and Hameed 2009). The bond at 1505 cm<sup>-1</sup> corresponds to the stretching vibration of C–O of TEG,



Fig. 1 FTIR analysis of functionalized, annealed, and bare ZnFe<sub>2</sub>O<sub>4</sub>

which confirmed the existence of TEG group. These bonds further explain the contract of oxygen of C-O with the iron or zinc on the  $ZnFe_2O_4$  surface (Ali et al. 2020a), while the stretching vibration band of Zn-O and Fe-O is observed at 605 and 685 cm<sup>-1</sup>, respectively (Hankare et al. 2011). The disappearance of bands at 3175 and 3015 cm<sup>-1</sup> in FT-IR spectra of calcined ZnFe<sub>2</sub>O<sub>4</sub> NPs is a clear indication that there is no more hydroxyl group, which shows that there are no more water molecules (Ali et al. 2020a). A small band at 1412 cm<sup>-1</sup> corresponds to the bending vibration of the  $-COO^{-}$  group (Ali et al. 2020b), and the band at 861 cm<sup>-1</sup> corresponds to the C-H bending vibration (Yang et al. 2014). The bands at 644 and 527  $\text{cm}^{-1}$  are attributed to the bending vibration of ZnO (Sakthivel et al. 2003), while the band 446 cm<sup>-1</sup> corresponds to the Fe–O bond (Tourinho et al. 2007). The FT-IR spectra of modified ZnFe<sub>2</sub>O<sub>4</sub> NPs are entirely different from the bare and calcined ZnFe<sub>2</sub>O<sub>4</sub> NPs. The vibration peak of O-H of water molecule was observed at 3342 (Ali et al. 2020c), and 3281  $\text{cm}^{-1}$ , and the bands at 2937 and 2848 cm<sup>-1</sup> correspond to the vibration of CH<sub>2</sub>CH<sub>2</sub>CH<sub>2</sub>NH<sub>2</sub> (Ali et al. 2020c), which confirmed that the 3-(APTES) is well grafted on the ZnFe<sub>2</sub>O<sub>4</sub> NPs surface (Ali et al. 2020c). The binding and stretching vibration of N-H and C-H of -CO<sub>2</sub><sup>-</sup> group shows the bands at 1582 and 1462 cm<sup>-1</sup>, respectively (Shams-Ghahfarokhi and Nezamzadeh-Ejhieh 2015). The stretching vibration peaks of N–O of NO<sub>3</sub><sup>-</sup> group occurred at 987 and 909 cm<sup>-1</sup> (Ali et al. 2020d). The binding vibration bands C-H occurred at 748  $\text{cm}^{-1}$  (Bouanimba et al. 2011), while the bands at 679 and 532 cm<sup>-1</sup> correspond to the vibration of Zn–O (Posa et al. 2016) & Fe-O (Li et al. 2011), respectively.

#### X-ray diffraction (XRD) analysis

The structural morphology of the synthesized spongy  $ZnFe_2O_4$  NPs was investigated through the XRD pattern shown in Fig. 2. The peaks of XRD have been examined and inspected by using the software X'Pert High score (JCPDs No: 73-1963). The XRD diffraction pattern illustrated that all the recorded peaks are reasonably matched with the cubic structure of Zinc Ferrite nanoparticles (Ullah et al. 2018). By using the Scherer formula, average crystallite size has been calculated (Ali et al. 2020c).

$$S = \frac{K\lambda}{B\cos\theta} \tag{1}$$

Here, *S* is the grain size, *K* is constant, having the value of 0.94,  $\lambda$  is wavelength having the value of 1.5406 Å, *B* is the FWHM, where  $\theta$  is the diffraction peak position. Bare ZnFe<sub>2</sub>O<sub>4</sub> NPs do not show any intense diffraction peak at (2 $\theta$ ) as illustrated in Fig. 2, demonstrating the amorphous character. On the other hand, annealed and functionalized



Fig.2 XRD graphs of functionalized, annealed, and bare  $\rm ZnFe_2O_4$  nanoparticles

NPs show intense diffraction peaks at 25.4, 37.8, 48.2, 53.9, and 54.6 2theta degree, corresponding to the reflective plane (220), (311), (400), (440), and (121), respectively, show the crystallinity in modified NPs (Ali et al. 2020a). The calculated grain size for bare  $ZnFe_2O_4$ , annealed  $ZnFe_2O_4$ , and modified  $ZnFe_2O_4$  NPs are 106.5, 23.3, and 13.9 nm, respectively (Sehar et al. 2019). The XRD diffraction pattern for  $ZnFe_2O_4$  NPs provides various information about peak position, d-spacing values, maximum height values,

and full width and half maximum values. The average grain size for all three types of NPs was calculated by using the Scherrer calculator.

#### Scanning electron microscopy (SEM) analysis

SEM spectroscopy was used to get the definitive evidence of the shape and structural morphology of the prepared zinc ferrite ZnFe<sub>2</sub>O<sub>4</sub> NPs as illustrated in Fig. 3a–1. SEM micrograph of ZnFe<sub>2</sub>O<sub>4</sub> NPs (bare, calcined, and functionalized), respectively, exhibit the formation of distinct morphology of the prepared nanoparticles. The SEM micrographs of bare or blank ZnFe<sub>2</sub>O<sub>4</sub> NPs display the homogeneous dispersion of grain with an aggregate of big spheres of about 1 µm as shown in Fig. 3a-d. SEM results showed a spherical shape in a cluster, and these SEM micrographs of bare ZnFe2O4 NPs showed irregular shape and size. The surface morphology of calcined  $ZnFe_2O_4$  NPs appears more homogeneous, and the average surface roughness has been estimated to be slighter lesser than ZnFe<sub>2</sub>O<sub>4</sub> NPs as shown in Fig. 3e-h. The high thermal treatment of the bare ZnFe<sub>2</sub>O<sub>4</sub> NPs leads to the removal of water molecules, and the organic species, such as TEG and its oxidizing product, is the reduction of bare ZnFe<sub>2</sub>O<sub>4</sub> NPs (Jadhav et al. 2011). The high thermal treatment promotes the nucleation and growth of ZnFe<sub>2</sub>O<sub>4</sub> NPs, leading to the integration of smaller particles to produce a



Fig. 3 Scanning electron micrographs (SEM) of a-d Bare ZnFe<sub>2</sub>O<sub>4</sub>, e-g annealed ZnFe<sub>2</sub>O<sub>4</sub>, and i-l functionalized ZnFe<sub>2</sub>O<sub>4</sub>



greater aggregate of structure and agglomerations due to annealing occurred more in calcined NPs of ZnFe<sub>2</sub>O<sub>4</sub>, where the particles in the range of 0.7 µm and big spongy materials were formed (Dehghani et al. 2017). The surface modification procedure contains hydroxylation of calcined-ZnFe<sub>2</sub>O<sub>4</sub> NPs, which leads to the departure of roughly bound crystallite in calcined bunches of NPs and hydroxylation of the surface of the NPs. The ZnFe<sub>2</sub>O<sub>4</sub> NPs having OH sorts were treated with 3-(APTES). The SEM micrographs of modified ZnFe<sub>2</sub>O<sub>4</sub> NPs, as illustrated in Fig. 3i–l shows that the higher aggregation of nanospheres vanishes, the grain's overall dispersion is improved after modification, and spherical particles are more closely packed and interacted in modified NPs of  $ZnFe_2O_4$ . A reduction in the size of  $Fe_2O_4$  NPs after the oleic acid modification has been previously reported that having similar reasons (Dhiman et al. 2016).

#### Thermogravimetric analysis

Thermogravimetric analysis (TGA) of blank, annealed and functionalized  $ZnFe_2O_4$  NPs shown in Fig. 4. Thermo graph of blank and functionalized  $ZnFe_2O_4$  NPs shows two steps thermal degradation pattern, first weight loss (WL1)



Fig. 4 TGA analysis of the synthesized bare, annealed, and functionalized  ${\rm ZnFe_2O_4}\,{\rm NPs}$ 

Fig. 5 Taulc plots for calculation of bandgap energy of  $ZnFe_2O_4$  NPs



# The optical property of ZnFe<sub>2</sub>O<sub>4</sub> NPs

Diffuse reflectance spectroscopy was used to study the ultraviolet–visible absorption of  $ZnFe_2O_4$  NPs Fig. 5. BaSO<sub>4</sub> was used as a blank, and absorbance was expressed by treating the reflectance data with the Kubelka–Munk function. According to the UV absorption edges of 506, 530, and 545 nm for blank, annealed, and functionalized  $ZnFe_2O_4$ NPs, corresponding to the bandgap of 2.37, 2.28, and 2.22 eV for blank, annealed, and functionalized  $ZnFe_2O_4$ NPs, respectively.

# Photocatalytic activities ZnFe<sub>2</sub>O<sub>4</sub> NPs

## Effect of illumination time

The amount of  $ZnFe_2O_4$  NPs as 10 mg was fixed to investigate the relationship between illumination time and photocatalytic degradation efficiency of BPB, and the results are illustrated in Fig. 6. The results show that with increased illumination time, the degradation efficiency increase. In the heterogeneous catalyst, the OH ions are extremely accepted as primary oxidants (Jayaraman et al. 2012). The OH<sup>•</sup> radical have strong oxidizing power, which is enough to oxidize the adsorbed dye molecules on the surface of  $ZnFe_2O_4$  NPs into the water,  $CO_2$ , and other minerals acid (Khan et al. 2016; Jayaraman et al. 2012). Only a small increment of photocatalytic degradation efficiency was observed when the contact time was longer than 120 min. The probable reason is when enhancing the illumination time, a large





**Fig.6** Effect of irradiation time on the degradation of BPB using bare, annealed, and functionalized  $ZnFe_2O_4$  nanoparticles

number of little organic molecules are produced by photogeneration, and small organic molecules are degraded by the surface of  $ZnFe_2O_4$ , resulting in a decrease in the formation of hydroxyl radicals that attack the dye molecules; therefore, the dose does not lead to greater photo-degradation efficiency of BPB with an increase in irradiation time (Jia et al. 2011).

#### Influence of photocatalyst dose

The influence of catalyst dose on the photocatalytic decolorization of BPB was examined by using different amounts of photocatalyst ranging from 5 to 35 mg at the concentration of a fixed dye under UV–visible light. The catalyst degradation efficiency load on BPB is shown in Fig. 7, which shows that the efficiency is greatly increased by enhancing the concentration of catalyst from 10 to 35 mg. After this, the decolorization rate remains almost constant, and the highest degradation has occurred at 30 mg of ZnFe<sub>2</sub>O<sub>4</sub> NPs; similar results were achieved by other workers (Ali et al. 2020c). The number of the active site on the surface

of  $ZnFe_2O_4$  NPs increase with the increase in the amount of catalyst by increasing the number of  $O_2^{-\bullet}$  and  $OH^{\bullet}$  radicals (Hankare et al. 2011). At the same period with a large dose of photocatalyst, the turbidity of the mixture increase, and the photo-activated volume of suspension decrease because of a reduction in the penetration of VU-radiation (Yang et al. 2014). Thus, it can be determined that a high catalytic dose may not be valuable both because of accumulation is well as reduced irradiation field due to the radiation scattering (Jia et al. 2011).

#### Effect of dye concentration

The influence of dye concentration is one of the vital parameters, which affect the photocatalytic rate of BPB dye. On the removal of BPB dye, the influence of dye concentration showed that the highest decolorization occurs at 5 ppm of the dye solution, and then the degradation rate is reduced with increased dye dose, as shown in Fig. 8. The photocatalytic decolorization rate of BPB dye is because increased dye dose causes greater interaction of radiation with dye molecules, and radiation is unable to contact with catalyst surface; as a result, the rate of dye degradation decrease (Xie and Li 2006). High dye molecules produce many inorganic ions, such as nitrate & sulfate ions in a mixture that competes with dye molecules (Ali et al. 2020a). According to Bear-Lambert law (Ali et al. 2020c; Tashtoush 2021), when the amount of solution increase, the path length of incoming photons to solution molecules reduce, causing fewer numbers of photons to approach the photocatalyst surface and cause smaller degradation of dye molecules (Ali et al. 2020c).

#### Influence of solution pH

The influence of pH plays a key role in controlling the photocatalytic decolorization of dye on the photocatalyst surface. The pH values of the BPB solution were controlled by using



Fig. 7 Effect of catalytic dose on the degradation of BPB



Fig. 8 Effect of dye dose on the degradation of BPB

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Fig. 9 Effect of solution pH on the degradation of BPB

NaOH and HCl, and the achieved results are illustrated in Fig. 9. It was recorded that the photocatalyst showed the maximum photocatalytic activity at pH 12–14. In this, the Cl<sup>-</sup> anion can react with hydroxyl radical leading to inorganic smaller reactivity than OH<sup>•</sup>, so they do not participate in the degradation of BPB molecules. There is also a drastic competition among the dye molecules and anions concerning <sup>•</sup>OH radicals. A higher concentration of OH ions may lead to the deactivation of radicals (Ali et al. 2020c). Initially, the water and  $H_2O_2^{\bullet}$  were formed due to the reaction of hydroxyl-hydroxyl radicals. The reactivity of these radicals with dye molecules is very rare as compared to OH radicals. Secondly, due to the availability of a high amount of OH radicals; the radical–radical reaction occurs at a higher pH value (Ali et al. 2020a).

#### Influence of H<sub>2</sub>O<sub>2</sub>

The influence of  $H_2O_2$  on the degradation of BPB is summarized in Fig. 10. It exhibits that the photocatalytic degradation of BPB on  $ZnFe_2O_4$  NPs increases with increasing  $H_2O_2$ concentration. With a rise in the  $H_2O_2$  concentration, the degradation rate increases due to the presence of a greater number of hydroxyl radicals, which increase the degradation process by attacking the aromatic rings (Akpan and Hameed 2009). Due to the formation of  $HO_2^{\bullet}$ , the degradation rate is decreased when the hydroxyl radicals are maximum, it reacts with  $H_2O_2$  then  $HO_2^{\bullet}$  is formed.  $HO_2^{\bullet}$  is less effective a hydroxyl and when  $HO_2^{\bullet}$  is increasing, it has negligible contribution in dye destruction, which might result in lower degradation (Wu et al. 2016).

#### Mechanism of photocatalysis and recycling

To understand the degradation mechanism, the functions of superoxide  $(O_2^{\bullet-})$ , singlet oxygen  $({}^1O_2)$ , hole  $(h^+)$ , and hydroxyl radical ( ${}^{\bullet}OH$ ) in the degradation of BPB were investigated. Notably, benzoquinone is an  $O_2^{\bullet-}$  quencher,





Fig. 10 Effect of H<sub>2</sub>O<sub>2</sub> on the degradation of BPB

DABCO is a <sup>1</sup>O<sub>2</sub> quencher, ascorbic acid and isopropanol are the quenchers of both hole and <sup>•</sup>OH. Here, the influence of these quenchers on the photo-degradation of BPB was investigated. As shown in Fig. 11, all benzoquinone, ascorbic acid, and isopropanol caused the degradation rate of BPB to decrease drastically, while DABCO did not affect the rate. Hence, all  $O_2^{\bullet-}$  and hole (h<sup>+</sup>) and hydroxyl radical (\*OH) were involved in the photo-induced degradation of BPB, and meanwhile  $O_2^{\bullet-}$  should not be involved in the degradation reaction. By analysis of the degradation rates, it could be found that the degradation rates with benzoquinone, ascorbic acid, and isopropanol, respectively, were ca. 17%, 34%, and 45% of that without additive. Therefore, the degradation of the dyes under visible light irradiation should follow the sensitization mechanism, and  $O_2^{\bullet-}$ , hole (h<sup>+</sup>) and hydroxyl radical ( $^{\bullet}OH$ ) played a significant role in the degradation. The recycling experiments showed that ZnFe2O4 NPs still held excellent activity after five successive cycles. The removal efficiency decreased slightly to ca. 83.4% in the fifth run Fig. 12. The slight decrease of the activity may presumably be attributed to the loss of the catalyst.

#### Kinetic study of photocatalytic degradation of BPB

Kinetics studies are very important to know the mechanism of dye molecules. The reaction kinetics of photocatalytic degradation of BPB using bare, calcined, and modified  $ZnFe_2O_4$  NPs are shown in Fig. 13. The reaction kinetics were determined from various irradiation times. Then plot the 1 + log of optical density against radiation time, giving a straight line with R<sup>2</sup> value 0.931, 0.960, and 0.981 (bare, annealed, and modified)  $ZnFe_2O_4$  NPs, respectively, that implies the degradation reaction pseudo-first-order of kinetics, and the constant rate value is 2.303 × slope (Valizadeh et al. 2016).

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Fig. 12 The recycling experiment of  $ZnFe_2O_4$  NPs in the degradation of BPB dye



Fig. 13 Reaction kinetics of the degradation of BPB

The photocatalytic activity of all three types of photocatalysts can also be calculated using relationship (Ali et al. 2020c).

Fig. 14 Decolorization rate of ZnFe<sub>2</sub>O<sub>4</sub> photocatalyst

ZnFe2O4

Catalytic efficiency = 
$$k/C_{cat}$$
 (2)

ZnFe2O4

Catalyst type

ZnFe2O4

where k is rate constant, and  $C_{cat}$  is the amount of catalyst. The modified ZnFe<sub>2</sub>O<sub>4</sub> NPs have a smaller grain size and higher catalytic activity due to many active sites. The degradation rates of ZnFe<sub>2</sub>O<sub>4</sub> NPs (bare, calcined, and modified) were determined by using rate constants that were achieved from the slope of kinetics plots. The degradation rate of all three types of ZnFe<sub>2</sub>O<sub>4</sub> NPs is illustrated in Fig. 14. The photocatalytic degradation of BPB dye in the presence of modified ZnFe<sub>2</sub>O<sub>4</sub> was higher than bare and calcined ZnFe<sub>2</sub>O<sub>4</sub> NPs. The greater performance of modified ZnFe<sub>2</sub>O<sub>4</sub> NPs is due to the existence of a greater number of active sites.

# Conclusion

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Zinc ferrite (ZnFe<sub>2</sub>O<sub>4</sub>) NPs were successfully prepared via the co-precipitation route, and their surface was successfully modified with 0.1 M of 3-(APTES) solution. FT-IR analysis confirmed that the 3-APTES is well grafted on the surface



of ZnFe<sub>2</sub>O<sub>4</sub> NPs. The XRD analysis shows the crystalline structure in calcined and modified ZnFe<sub>2</sub>O<sub>4</sub> NPs, while the amorphous nature of bare ZnFe<sub>2</sub>O<sub>4</sub> NPs was observed. The SEM micrographs displayed a flat surface morphology with almost monodispersed grain size and irregular shape structure for all  $ZnFe_2O_4$  NPs. The grain size of the synthesized  $ZnFe_2O_4$  NPs was found to be 106.5, 23.3, and 13.9 nm for bare, calcined, and modified ZnFe<sub>2</sub>O<sub>4</sub> NPs, respectively. Optical band gap energy was calculated in both acidic (2.47 eV) and basic medium (2.7 eV) using a taulc plot. The photocatalytic performance of ZnFe<sub>2</sub>O<sub>4</sub> NPs exhibits that the modified ZnFe<sub>2</sub>O<sub>4</sub> degrade 95.4% of BPB dye at pH 12 in 120 min. The rate of BPB dye degradation on the surface of ZnFe<sub>2</sub>O<sub>4</sub> NPs followed pseudo 1st order of kinetics model  $(k = 6.7 \times 10^{-3}/\text{sec})$ . The photocatalytic degradation efficiency were 0.0011 min<sup>-1</sup> mg<sup>-1</sup> (bare), 0.0012 min<sup>-1</sup> mg<sup>-1</sup> (calcined) and 0.0015 min<sup>-1</sup> mg<sup>-1</sup> (modified) ZnFe<sub>2</sub>O<sub>4</sub> NPs, respectively. The degradation rate for the removal of BPB dye in the presence of modified ZnFe<sub>2</sub>O<sub>4</sub> catalyst was found to be greater than bare and calcined ZnFe<sub>2</sub>O<sub>4</sub> photocatalyst.

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Authors contribution FA: Conceptualization, Data analysis and curation, Project administration, Supervision, Validation, Writing—original draft, review & editing. LF, AS, NT, FR: Investigation, Methodology, Data analysis and curation. UA, MSA: Validation, Writing—review & editing. AR: Data analysis, Writing—review & editing. NA: Formal analysis, Writing—review & editing. NA, MB: Data analysis, Formal analysis, Writing—review & editing.

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# Declarations

**Conflict of interest** The authors declare that they have no conflict of interest.

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