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Response surface methodology‑based modeling and optimization of chromium removal using spiral‑wound reverse‑osmosis membrane setup

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Abstract

Contamination of groundwater sources due to the presence of chromium ions, a heavy metal, and the detrimental impact of it on the health of consumers have become a serious global concern. The present study was designed to understand the efects of three process variables, namely feed chromium concentration, feed pH and pressure, on chromium removal and permeate fux, by employing response surface methodology and central composite design techniques. Spiral-wound reverse-osmosis membrane was used for the removal of chromium from the groundwater samples in this study. Membrane performance, in terms of high permeate quality and fux, was evaluated by optimization of various operating conditions. The obtained experimental results were then matched with the predicted values obtained from the models. The models indicated that the highest chromium removal of 98.38% and permeate flux of 48.73 L/m^2 h could be achieved at the optimum feed pH of 3.00, pressure of 4 kg/cm² and feed chromium concentration of 0.431 ppm. Experimental validation confirmed that the model had higher predictive capabilities. The changes in pressure and the chromium concentration signifcantly afected the chromium removal and permeate fux, the former being positively correlated and the latter being negative to both the responses. Therefore, small-scale RO membranes can be used to treat aqueous solution with chromium contamination.

Keywords Central composite design · Chromium removal · Groundwater contamination · Reverse osmosis · Response surface methodology

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Introduction

Groundwater is acknowledged as essential source of drinking water, and one-third of the human population is reported to be dependent on it. Heavy metals present in groundwater are not biodegradable, hence, their treatment through the conventional means of treatment or natural processes is very unlikely. Other than that, heavy metal ions are toxic as well as carcinogenic (Kim et al. [2015;](#page-10-0) Saleh et al. [2019;](#page-10-1) Koedrith et al. [2013](#page-10-2)). Chromium (Cr), a commonly observed heavy metal contaminant, is found in surface water and groundwater due to numerous industrial activities, such as, wood preservation, textiles dyeing, leather tanning, electroplating (Kazakis et al. [2017](#page-10-3); Mishra and Bharagava [2016](#page-10-4); Saha et al. [2011](#page-10-5); Sharma et al. [2008](#page-10-6)). Naturally, Cr ions concentrations in groundwater are observed to be lower than 2 µg/L but, sometimes, concentrations as high as 120 µg/L have also been reported (World Health [1996;](#page-10-7) Kazakis et al. [2017](#page-10-3)). Contamination of groundwater with Cr is a prevailing prob-lem in countries like Brazil (Hirata and Suhogusoff [2019](#page-10-8)),

China (Guo et al. [2020\)](#page-10-9), Greece (Kazakis et al. [2017](#page-10-3)), India (Poonia et al. [2021\)](#page-10-10), Italy, Nigeria (Adeyemi and Ojekunle [2021](#page-9-0)), Pakistan (Raza et al. [2017](#page-10-11)), South Africa (Edokpayi et al. [2018\)](#page-9-1), and USA (Coyte et al. [2020](#page-9-2)), which are well acknowledged and documented in literatures.

Cr commonly exists in two states, based on valency in water; trivalent Cr (Cr^{3+}) and hexavalent Cr (Cr^{6+}) (World Health, 2020). Cr^{3+} , nontoxic and less absorbed, is an essential trace element from nutrition point of view, and its deficiency leads to glucose intolerance, and other metabolic disorders (Zheng et al. 2015). On the other hand, Cr^{6+} is considered harmful, even in small ingestion dose. The exposure routes to Cr^{6+} are ingestion through eating, drinking, skin contact, and inhalation (Tziritis et al. [2012\)](#page-11-1). The permissible limit for Cr^{6+} in drinking water is 0.05 mg/l (Standards [2012\)](#page-10-12). To conclude in broad spectrum, ingestion of Cr^{6+} causes liver and kidney damage (De Rossi et al. [2018;](#page-9-4) Marikkani et al. [2019](#page-10-13)), which may sometimes be fatal. So, it is required to devise an appropriate method/ technology for the removal of Cr from water.

Many conventional treatment technologies, like ion exchange (Wang et al. [2014](#page-11-2); Wang et al. [2020](#page-11-3)), electrolysis (Tao et al. [2014](#page-10-14); Wang et al. [2020](#page-11-3)), solvent extraction (Ying et al. [2020\)](#page-11-4), membrane separation (Muthumareeswaran et al. [2017a;](#page-10-15) Mnif et al. [2017;](#page-10-16) Taha et al. [2021](#page-10-17)) and biosorption process (Bashir et al. [2019](#page-9-5); Yunus et al. [2020;](#page-11-5) Yadav et al. [2021\)](#page-11-6), are used for the removal of heavy metal ions from the water. The Cr after treatment is generally managed by precipitating as hydrous oxide with caustic/lime soda or by incinerating it to form oxides to be disposed to landflls (Wilbur et al. [2012](#page-11-7)). All forms of inorganic ions as well as complexes of Cr are detached from water by using membrane in reverse-osmosis (RO) process. RO process involves the use of hydraulic pressure as driving force for the separation process (Rastogi [2018](#page-10-18)). Hydraulic pressure is applied in the feed side which drives the separation of ions in the feed solution and treated water (permeate) is passed through the pores in the membrane. The water through the membranes is calculated as water fux. The mechanism for Cr treatment employed in RO process is explained in Fig. [1](#page-1-0). Application of membranes is an efective method for the removal of the prevalent Cr^{6+} and Cr^{3+} ions from the water (Muthumareeswaran et al. [2017b;](#page-10-19) Hosseini et al. [2017](#page-10-20)). However, it has some limitations such as high initial investment, operational costs and membrane fouling (Jiang et al. [2017;](#page-10-21) Garg and Joshi [2015](#page-9-6)).

Response surface methodology (RSM) is an approach which uses mathematical and statistical techniques for development of empirical model (Montgomery [2006;](#page-10-22) Jiang et al. [2013\)](#page-10-23). RSM uses central composite design (CCD) of experiments to study the infuence of independent (input) variables and optimize dependent (output) responses (Montgomery [2006](#page-10-22); Schio et al. [2020\)](#page-10-24). RSM focuses on modeling

Fig. 1 Schematic fgure explaining the working mechanism of RO process

experimental responses with regards to input factors (Box and Draper [1987\)](#page-9-7).

Previously, RSM has been employed for the optimization of the treatment of heavy metals from water using various membrane based separation process (Rad et al. [2009;](#page-10-25) Al-Alawy and Salih [2016\)](#page-9-8). It was used to model the transmembrane pressure, rejection, and recovery for ultra-fltration, nanofltration and RO membranes. However, limited progress has been reported using small-scale RO membrane by employing RSM, which can save initial capital costs to treat groundwater with Cr contamination. Therefore, this study was conducted to model the performance and optimization of small-scale RO membrane for the operating parameters like Cr concentration, feed pH and pressure for maximizing Cr removal and permeate fux. Further, we have carried out kinetics study to determine the change in Cr concentration over time to validate the optimum results from the models. The experiments were carried out in Amity Institute of Environmental Sciences, Amity University-Noida, India.

Materials and methods

RO membrane setup and operating procedure

A laboratory-scale RO membrane system was used to model the removal of Cr from aqueous solution. Figure [2](#page-2-0) shows the schematic diagram of the laboratory-scale spiral-wound RO setup used in this study. It comprised of feed tank, pressure pump, membrane cell, and a RO membrane in spiralwound confguration. Commercially available domestic-RO

Fig. 2 Schematic diagram of small-scale spiral-wound reverse-osmosis (RO) membrane setup; 1—feed tank containing Cr aqueous solution, 2—Pressure pump, 3—Feed fow pressure gauge, 4—feed fow

meter, 5—RO spiral-wound membrane cell, 6—Permeate fow meter, 7—Brine fow pressure gauge, 8—Pressure control valve, 9—Computer for data logging

membrane (Make: CSM, Model: RE1812-80) was used in this study.

In this study, synthetically prepared Cr water (a known concentration of Cr) was applied as feed water to check the membrane performance. The fow rate at the feed side was kept uniform throughout the experiments. Pressure pump was used to provide the necessary operating pressure to the system which was regulated with the help of a pressure control valve. Pressure gauges fow meters were provided to monitor the pressure and flow rate of the system, respectively. Flow meters were connected to micro-processors for logging the data, to be used for the calculation of the permeate fux. After the experiments were conducted, initially, the membranes were back-fushed with distilled water followed by forward fushing with 1% sodium metabisulfte solution until the RO module was completely replaced with it.

Groundwater samples were collected from three locations $(L_1, L_2$ and L_3) in Delhi-NCR, India and were analyzed for its physical and chemical parameters. The details are shown in Table [1](#page-3-0) and have been compared with the acceptable and permissible limits from BIS 10500:2012 (Standards [2012](#page-10-12)). Though the Cr concentration from these locations was within the acceptable limit, higher Cr concentrations have been reported earlier (Shekhar and Sarkar [2013](#page-10-26)).

The feed and permeate stream concentrations of Cr were measured with the help of Spectrophotometer (Make: *Labman Scientifc Instruments Pvt. Ltd., India*), following standard methods (Baird and Bridgewater [2017\)](#page-9-9). Spectrophotometric analysis was conducted at the wavelength of 540 nm, using a light path passing through a 1 cm cuvette. For the analysis of Cr concentration, a calibration curve was prepared and plotted with observed absorbance values against known Cr concentration. The Cr ions concentrations of the water samples were then calculated using the linear equation obtained from the calibration curve. Cr removal (%) was estimated using Eq. (1) (1) (1) :

$$
\text{Cr removal}(\%) = \frac{(C_0 - C_t)}{C_0} \times 100 \tag{1}
$$

where C_0 is the initial Cr concentration in feed solution (mg/L) and C_t is the final Cr concentration in the permeate (mg/L) .

Permeate flux in L/m^2 h (LMH) was calculated as per Eq. ([2\)](#page-2-2) shown below:

Permeate Flux =
$$
\frac{q_p}{A_m}
$$
 (2)

where q_p is permeate flow rate (L/hr) and A_m is area of membrane $(m²)$.

Feed solution pH was maintained using 1 M NaOH and 1 M HCl solutions. It was measured with the help of calibrated pH meter (Make: *Hach Inc., USA*, Model: *HQ40d*).

CCD of experiments

A full-factorial CCD was used for developing mathematical equations, in terms of percentage removal of Cr, and permeate fux, providing quantifable assessment of membrane system to treat Cr contaminated water samples. Membrane performance, in terms of high permeate quality, and fow rate, was evaluated by optimizing various

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operating variables such as feed pressure, feed pH and Cr concentration in feed solution.

To study the efects of three process variables simultaneously, CCD of RSM was selected. Both linear and quadratic models were used to determine the Cr removal and permeate fux. The general form of second-order model can be expressed according to Eq. (3) (3) :

$$
y = \beta_0 + \sum_{i=1}^n \beta_i x_i + \sum_{i=1}^n \beta_{ii} x_i^2 + \sum_{i < j} \sum_{j < j} \beta_{ij} x_i x_j + \varepsilon \tag{3}
$$

where y is the response, x_i is coded levels of operating variables; β_o , β_i , β_{ii} and β_{ij} are the regression coefficients for constant term, linear, quadratic and interaction parameters, respectively; *n* is number of operating variables and ε is experimental error (K et al. [2021](#page-10-27); Aghilesh et al. [2021](#page-9-10)). To establish its statistical implication, analysis of variance (ANOVA) for the model was also carried out. Design-Expert

(*Stat-Ease Inc., USA*) software was used for design of experiments and for statistical and graphical data analysis.

In the present study, after a careful analysis of literature, concentration of Cr (Hafez and El-Mariharawy [2004\)](#page-10-28), feed pH and feed pressure (Gasemloo et al. [2019](#page-9-11); Hafez and El-Mariharawy [2004\)](#page-10-28) were chosen as input variables for the optimization of RO system. A full factorial design of the factors with various ranges were considered as shown in Table [2](#page-4-0). A total of 19 experimental sets were designed with the use of software according to Eq. [\(4](#page-3-2)). The experimental sets comprise of eight cubical or factorial points, six axial points, and fve repetition of center points.

$$
N = 2n + 2n + nc
$$
 (4)

where *N* is the required total number of experiments, *n* is the number of operating variables and n_c is replicate number at the center points (Sen et al. [2018;](#page-10-29) Garg and Joshi [2017](#page-9-12)).

Results obtained were subjected to ANOVA, later being utilized for the graphical analysis of the data. This gives the interaction between the responses and the variables.

and $+1$ range from Table [2](#page-4-0). Further, the optimum pressure and Cr concentration obtained from RSM optimization were used to validate the results.

Kinetics study

Kinetics study was carried out to investigate the change in feed and permeate stream over time. For this study, the experiments were carried out by recirculating the brine flow to the feed stream and collecting the permeate flow separately. The impacts of pressure and Cr concentration were studied for both the feed and permeate stream. A total of 5 L of feed Cr solution was prepared, and the experiments were carried out until the feed stream was fltered toward the permeate stream. The Cr concentrations in feed and permeate streams were continuously monitored. The pressure and Cr concentration were selected from − 1, 0

Results and discussion

Table [3](#page-4-1) shows the CCD matrix of experiments with its responses for Cr removal and permeate fux. The operating variables in Table [3](#page-4-1) are shown in terms of codded levels according to Table [2.](#page-4-0) All the experiments were carried out in accordance with the CCD of experiments and observed responses; Cr removal $(\%)$ and permeate flux (LMH) were calculated using Eqs. (1) (1) (1) and (2) (2) , respectively. Also, predicted responses obtained from the models are also shown in Table [3.](#page-4-1)

Table 3 CCD matrix with predicted and observed responses for Cr removal and permeate fux

> where *A* is Cr concentration in feed solution (ppm), *B* is pH of feed solution and *C* is feed pressure (kg/ cm^2)

Table 4 ANOVA results for Cr removal model and permeate fux mode**l**

Source	Sum of squares df Mean square F value				<i>p</i> value
Cr removal model					
Linear	1.33	1	1.33	8.69	0.022
Interaction	1.19	1	1.19	7.77	0.027
Square	1.05	1	1.05	6.89	0.034
Residual	1.07	7	0.15		
Permeate flux model					
Linear	2061.95	1	2061.95	1643.46	< 0.0001
Interaction	0.03	1	0.03	0.02	0.8857
Square	18.06	1	18.06	14.40	0.0043
Residual	11.29	9	1.25		

where d*f* is degrees of freedom

Response surface modeling for Cr removal

Suitability of the model was evaluated with the use of ANOVA (Gasemloo et al. [2019\)](#page-9-11). The experimental responses were used to analyze the data. Signifcance of the data and coefficient of determination (R^2) were estimated using ANOVA. Also, statistical indices like correlation coefficient (*R*), mean squared error (MSE), root-mean-squared error (RMSE) and mean average error (MAE) were calculated to determine the precision of the model. Table [4](#page-5-0) shows the ANOVA table for Cr removal and permeate fux.

(5) Chromium Removal =97.68 − 0.41*A* + 0.12*B* + 1.00*C* + 0.01*AC* + 0.04*BC* − 0.10*A*² − 0.06*B*² − 0.21*C*² − 0.55*A*²*C* + 0.29*AC*² − 0.36*BC*²

where *A*, *B* and *C* are coded factors, Cr concentration, feed pH and feed pressure. *AB*, *BC* & *AC* are the interactive terms. A^2C , AC^2 and BC^2 are the quadratic terms. From Table [4](#page-5-0), *F* value of 7.680 specifes that the model terms are signifcant and there is only 0.620% chance for the model to have larger *F* value which occurs due to noise. The *p* value of the linear, interaction and square terms were < 0.050 , which indicates the model to be signifcant, while *p* value of lack of fit was > 0.050 , which indicates that it is insignificant (Singh et al. [2019a\)](#page-10-30). This shows that the model can be used to predict the Cr removal.

Also, from Table [5](#page-5-1) R^2 and R^2 _{adjusted} values were calculated as 0.9238 and 0.8031, respectively. Higher values confrm the model agrees with the experimental results. MSE, RMSE and MAE was found to be 0.0560, 0.2367 and 0.2255, respectively. Lower values of MSE, RMSE and MAE can be attributed to better accuracy and precision of the model to predict the Cr removal.

Graphical analysis of the data was represented in terms of 3D response surface plots. These plots graphically describe

the interactive efects of independent variables (Gasemloo et al. [2019](#page-9-11); Khayet et al. [2011](#page-10-31); Kassem et al. [2017\)](#page-10-32) on the Cr removal (%). Figure [3d](#page-6-0) shows the predicted and actual plots for Cr removal. This indicates the accuracy and reliability of the model.

The response plot in Fig. [3](#page-6-0)b indicates that the Cr removal is afected very slightly with the change of feed pH. At higher feed pressure, the increase in pH has a slight decrease in Cr removal (%) (Thilagavathy and Santhi [2013](#page-11-8); Gasemloo et al. [2019](#page-9-11); Khalifa et al. [2019](#page-10-33)). However, the change of pH has lower change in the Cr removal (%) from water. However, Cr removal (%) increases with the increase in pressure (Fig. [3a](#page-6-0), c). This is because at lower pressure, the difusion of Cr ions is lower (Wei et al. [2013;](#page-11-9) Basaran et al. [2016](#page-9-13)). Therefore, as the pressure increases, we can observe a higher Cr removal (%). This indicates higher removal of Cr at higher pressures. But, as the Cr concentration in the feed increases, the Cr removal (%) slightly decreases for all pressure ranges. This happens because the osmotic pressure is a function of the concentration of salts or organics present in the feed solution. Figure [3](#page-6-0)d indicates that the predictions did hold true as actual values are either seen to fall on the same line or closely scattered around.

Response surface modeling for permeate fux

Tables [4](#page-5-0) and [5](#page-5-1) show the ANOVA results along with statistical indices like R , R^2 , MSE, RMSE and MAE with regard to Cr removal and permeate fux models. The analysis shows that the model terms are signifcant.

Permeate Flux =
$$
36.62 - 0.16A - 0.05B + 11.35C
$$

+ $0.01AB - 0.06AC - 0.05BC$ (6)
+ $0.05A^2 - 0.06B^2 - 0.87C^2$

where *A*, *B* and *C* are coded factors, Cr concentration, feed pH and feed pressure. *AB*, *BC* and *AC* are the interaction terms. From Table [4,](#page-5-0) *F* value of linear terms of the model shows that model terms are significant and there is $<0.01\%$ chance for the model to have larger *F* value which occurs due to noise (Singh et al. [2019b\)](#page-10-34). The *p* values of linear and square terms were < 0.050 , which indicates the model to be

Fig. 3 Response surface 3D plots for (**a**) interactive efect of feed Cr concentration and feed pressure, (**b**) feed pH and feed Cr concentration, (**c**) Contour plot for feed pressure and feed Cr concentration, (**d**) predicted vs. actual plots for Cr removal (%)

significant and p -vale of lack of fit was > 0.050 , which is insignifcant. Permeate fux model can the fux generated for the Cr removal.

Also, from Table [5](#page-5-1) R^2 and R^2 _{adjusted} values were calculated as 0.9946 and 0.9892, respectively and its higher values confrm the model with experimental results. Further, MSE, MSE and MAE was calculated as 0.5943, 0.7709 and 0.6158 0.6158 0.6158 , respectively as shown in Table 5 , which determines the precision of the model for predicting the permeate fux. Figure [4d](#page-7-0) shows the predicted and actual plots for Cr removal. This indicates the accuracy and reliability of the model. This plot shows an acceptable agreement between observed data and those predicted from the model.

The response plot (Fig. [4](#page-7-0)a) indicates that a decrease or increase in the pH of the feed solution do not affect the permeate fux. Whereas on increasing the pressure, the permeate fux increases gradually. The response plot (Fig. [4b](#page-7-0), c) shows that the efect of Cr concentration in the feed solution has a very slight effect on the flux. The increase in the concentration slightly decreases the permeate fux of the system at all values of pressure. This happens because higher Cr concentration in the feed results in lower membrane fux at a constant pressure. From Fig. [4](#page-7-0)d, it can be concluded that the actual and predicted values were in good agreement.

Multiparameter optimization of Cr removal

The main aim of this study was to use experimental design for optimization of Cr removal from aqueous solution. The optimization of was carried out using Design Expert software by setting the criteria of maximizing Cr removal $(\%)$ and permeate fux (LMH) while targeted values of three independent variables were selected within the range for the experiment.

The optimum conditions for maximum Cr removal and permeate fux with the use of RO membrane in this study are shown in Table [6](#page-7-1). These running conditions will provide with the best permeate quality in terms of Cr removal $(\%)$ and the permeate fux. To validate the results, additional experiments were carried out with the optimal parameters and the results are presented in Table [6](#page-7-1). The diference in the predicted and actual results for Cr removal and water fux was 0.2% and 2.92%, respectively. Also, Cr concentration was found to be within the permissible limit in drinking water i.e., 0.05 mg/l (Standards [2012](#page-10-12)). The fux will also be

Fig. 4 Response surface 3D plots for (**a**) interactive efect of feed pH and feed pressure, (**b**) feed pressure and feed Cr concentration, (**c**) contour plot for feed pH and feed Cr concentration, (**d**) predicted vs. actual plots for permeate fux (LMH)

maximized under these run conditions hence maximizing the output of the RO system. This will help us obtain the maximum amount of permeate in the given time thereby boosting the RO efficiency and reducing the input investment costs.

Kinetics study

The change in Cr concentrations in both feed and permeate streams at 2 and 4 kg/cm² of applied pressure is shown in Fig. [5](#page-8-0). It can be observed that in feed stream, Cr concentration increases at a faster rate at 4 kg/cm^2 pressure, compared to that of 2 kg/cm² pressure (Fig. [5a](#page-8-0)). From Fig. [5b](#page-8-0), it is shown that at 4 kg/cm^2 pressure Cr concentration was found below acceptable limit of 0.05 ppm (Standards [2012](#page-10-12)) than

the 2 kg/cm^2 applied pressure, this may be because the rejection of ions increases with increase in pressure (Garg and Joshi [2017](#page-9-12); Srivastava et al. [2021](#page-10-35)). Further, at the optimum pressure of 4 kg/cm², permeate stream was found to be well below the acceptable limit of Cr (Fig. [5b](#page-8-0)), thus validating the results obtained from Table [6](#page-7-1).

Also, experiments were carried out at various feed Cr Concentrations while keeping the other operating factors according to Table [6](#page-7-1). Figure [6](#page-8-1) shows the change of feed and permeate Cr concentrations at 0.3, 0.5, 0.7 and 0.431 ppm. The trend of Cr concentration in feed and permeate stream was found to be identical. The feed Cr concentration was highest at initial Cr concentration of 0.7 ppm (Fig. [6](#page-8-1)a). While in permeate stream, the Cr concentration was below the acceptable limit

Fig. 5 Change in Cr concentration in (**a**) feed and (**b**) permeate streams (at feed Cr concentration=0.431 ppm and pH=3)

Fig. 6 Change in Cr concentration in (a) feed and (b) permeate (at feed pressure = 4 kg/cm^2 and $pH = 3$)

at optimum feed Cr concentration of 0.431 ppm (Fig. [6](#page-8-1)b), thereby, justifying the optimization results (Table [6\)](#page-7-1) obtained by the RSM model.

Also, Table [7](#page-9-14) shows the comparison of Cr removal by various membrane types with the present study. This comparison data reveals that among various membrane types, the smallscale RO membranes used in the current study can facilitate highest removal of Cr from waters.

Conclusion

In this study, CCD of RSM was chosen to model the efects of three process variables: Cr concentration, feed pH and pressure, for two responses: Cr removal (%) and permeate fux. The optimal conditions for operating variables were at pH at 3, pressure of 4 kg/cm^2 and Cr concentration

Table 7 Comparison of Cr removal by various methods

of 0.431 ppm, resulting into a higher Cr removal rate of 98.38% and the higher permeate fux of 48.73 LMH. The results were compared with the predictions of the models and only had less than 3% diference. The predictive capability of the regression models was found to be higher and was statistically with R^2 values 0.9238 and 0.9946 for Cr removal and permeate fux models, respectively. From the interactive plots, it was observed that the increase in the operating pressure to 4 kg/cm^2 , increased Cr removal and permeate fux. Similarly, increase in the concentration of the feed was observed the decrease the Cr removal and a slight decrease in permeate fux. Based upon the observed results and the stability of the developed models, it could be concluded that the RO membranes can facilitate higher Cr removal with a high permeate fux, resulting into higher volume of water reclamation within the same input investment cost. Cr in groundwater can be remediated by reduction of Cr^{6+} to Cr^{3+} and precipitated as hydroxides. Natural attenuation can also be carried out by the similar principle if natural reductants are present in the groundwater. However, some future scope of actions are worth consideration for promoting its application, which are (a) inclusion of other operating parameters like temperature, recovery and energy consumption of membrane process for the performance optimization of RO system, (b) extension of similar approach for the treatment of aqueous solution contaminants including various other heavy metals like arsenic, cadmium, lead, and (c) similar optimization studies on various types of membranes separation techniques like ultrafltration, nanofltration to do a comparative study.

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Declarations

Conflict of interest The authors declare that they have no known competing fnancial interests or personal relationships that could have appeared to infuence the work reported in this paper.

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