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European rabbit (*Oryctolagus cuniculus***) as seed disperser in arid ecotones of Argentina: non‑native herbivore facilitation of native and non‑native plants**

S. Yasmin Bobadilla1 [·](http://orcid.org/0000-0003-4133-2872) E. T. Olivares1 · F. M. Jaksic2 · R. A. Ojeda3 · M. F. Cuevas1

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Abstract

The European rabbit plays a major role in seed dispersal in its native environment. We evaluated its ecological interactions with plants through the dispersal of seeds by endozoochory in two invaded arid ecotones of Argentina. We found 855 whole seeds in 1283 fecal pellets, belonging to one non-native (*Sesuvium portulacastrum*) and fve native plants (*Arjona* sp., *Fabiana denudata*, *Frankenia juniperoides*, *Lycium chilensis*, *Poa* sp.). Our results indicate that the European rabbit is a legitimate disperser of *F. juniperoides*, *L. chilensis,* and *S. portulacastrum* by the consumption and dissemination of viable seeds. Contrastingly, the rabbit is an illegitimate disperser of *Arjona* sp., *F. denudata*, and *Poa* sp. Our study identifes new interactions between an invasive herbivore and sympatric plants in the arid ecosystems of Argentina. We put forth that the mutualistic interaction between the European rabbit and both native and non-native plants highlights the complexity of trophic networks in invaded environments.

Keywords Argentina · Endozoochory · Germination · Mutualism · Trophic interactions

Introduction

The presence of a new (exotic) species in an ecosystem can alter its ecological network directly, by establishing new interactions with the native species of the invaded community, or indirectly, by infuencing their abundance, distribution, or behavior (McConkey et al. [2012](#page-10-0); Blackburn et al. [2014](#page-8-0)). Although negative consequences of invasions are often emphasized, some studies show that mutualisms

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 \boxtimes S. Yasmin Bobadilla ybobadilla@mendoza-conicet.gob.ar

- ¹ Ecología de Mamíferos de Tierras Secas, Instituto Argentino de Investigaciones de Zonas Áridas (IADIZA), CCT-Mendoza CONICET, Avenida Ruiz Leal S/N - Parque General San Martín, M5500 Mendoza, Argentina
- Center of Applied Ecology and Sustainability (CAPES), Pontifcia Universidad Católica de Chile, 8331150 Santiago, CP, Chile
- ³ Laboratorio de Filogeografía, Taxonomía Integrativa Y Ecología (LFTIE), IADIZA, CCT-Mendoza CONICET, Avenida Ruiz Leal S/N - Parque General San Martín, M5500 Mendoza, Argentina

may occur between non-native and native species, or among non-native species themselves, thus facilitating the establishment of exotic species in invaded communities (Simberlof and Von Holle [1999;](#page-10-1) Traveset and Richardson [2006](#page-10-2)). Positive interactions frequently occur among non-native species, indicating that mutualisms may be important in varied invasion scenarios (Braga et al. [2018](#page-8-1)). Among the types of plant–animal mutualisms that facilitate invasions is seed dispersal by endozoochory (i.e., internal seed dispersal: Howe [1986](#page-9-0); Herrera and Pellmyr [2002](#page-9-1)). This type of mutualism shows a lack of specifcity in animal-dispersed plants and provides reciprocal benefts between them (Howe and Smallwood [1982](#page-9-2)). The consumption of seeds by herbivores can beneft plants because the dissemination of viable seeds carried in feces allows them to escape the often high mortality rates close to the parental plant (Janzen [1970](#page-9-3)), and also because it renders it possible to explore new vacant areas for recruitment, thus allowing the expansion of those plant species (Howe and Smallwood [1982](#page-9-2); Jordano et al. [2011\)](#page-9-4). Furthermore, under certain conditions, seed dispersal makes it possible to colonize novel sites suitable for germination and survival (Howe and Smallwood [1982\)](#page-9-2). On the other hand, animal dispersers obtain a nutritional reward from the fruits and seeds they consume and transport (Jordano [1987](#page-9-5)).

Another important factor to consider is the efectiveness of seed dispersal (Schupp et al. [2010\)](#page-10-3). That is, the contribution of each disperser to the reproduction of a given plant species (Reid [1989\)](#page-10-4), as well as to the contribution in dispersal that a plant population receives from its dispersers (Schupp et al. [2010](#page-10-3)). Quantitative components (e.g., number of seeds dispersed by a dispersal agent, number of visits a dispersal agent makes multiplied by the number of seeds dispersed per visit) and qualitative ones (i.e., probability that a dispersed seed will produce a new adult) are to be considered when evaluating the efectiveness of the dispersal process (Schupp et al. [2010\)](#page-10-3). Focusing on the qualitative component of the process, a given disperser may modify the viability and/or the germination of seeds when they are ingested and pass through its digestive tract (quality of treatment in the mouth and gut), as well as their survival depending on the quality of the sites where the seeds are deposited once laid in the feces (quality of deposition) (Schupp [1993](#page-10-5)). Now, considering the quality of the treatment, when an animal ingests seeds of a given species and disposes of them in viable conditions for their germination, it is considered a legitimate disperser of that species (Herrera [1995](#page-9-6)). On the contrary, when an animal delivers unviable seeds, it is considered an illegitimate seed disperser or a seed predator (Castro et al. [2008](#page-8-2)).

Non-native herbivorous mammals can act as important legitimate dispersers of both non-native (Holmgren [2002](#page-9-7); Dacar et al. [2019\)](#page-9-8) and native plant species (Campos et al. [2008](#page-8-3); Calviño-Cancela [2011](#page-8-4); Muñoz-Gallego et al. [2019\)](#page-10-6) and in many cases provide plants with the same services as native dispersers (Westcott and Fletcher [2011\)](#page-10-7). Alternatively, they can also act as illegitimate dispersers (Foster et al. [2014](#page-9-9); Dénes et al. [2018](#page-9-10)). In particular, lagomorphs (including hares and rabbits) are dispersers of viable grass and herb seeds accidentally consumed while feeding on foliage and fruits (Schupp et al. [1997](#page-10-8); Henríquez et al. [2014](#page-9-11)). Among the seed-dispersing lagomorphs is the European rabbit (*Oryctolagus cuniculus*), an herbivore introduced to several regions around the world and considered as one of the 100 most harmful invasive species due to its impact on biological diversity and human activities (Lowe et al. [2000](#page-9-12); Long [2003\)](#page-9-13). The European rabbit generates changes not only in plant community structure through herbivory but also in vertebrate community structure through competition or via its role as a keystone prey species (Courchamp et al. [2003;](#page-8-5) Davey et al. [2006;](#page-9-14) Barbar et al. [2018](#page-8-6); Correa-Cuadros et al. [2022;](#page-8-7) Gubelin et al. [2023](#page-9-15)). Particularly in their role as endozoochorous seed dispersers, rabbits facilitate the establishment of non-native plants in areas where they colonize (Fernández and Saiz [2007;](#page-9-16) Bobadilla et al. [2020](#page-8-8)). These positive interactions among exotic species may potentially cause an invasional meltdown (Simberloff and Von Holle [1999](#page-10-1)), leading to an accelerated increase in the number of invading species and their impact on native communities. On the other hand, rabbits yield positive ecosystem services as seed-dispersing agents for some native plants (Castro et al. [2008](#page-8-2); Jaksic and Castro [2021\)](#page-9-17).

The European rabbit was frst recorded in the arid ecosystems of Argentina in the 1970s (Jaksic et al. [2002;](#page-9-18) Bonino and Borrelli [2006;](#page-8-9) Bobadilla et al. [2021](#page-8-10)). The diverse ecological mosaic in this region constitutes an important scenario for the evolution of its biota and supports more species and endemic genera than other macrohabitats or biomes (Ojeda and Tabeni [2009\)](#page-10-9). These arid ecosystems are undergoing rapid habitat conversion due to human activities (agriculture, grazing, and lumber and frewood extraction), desertifcation, salinization, and changes caused by shifting climate and expansion of invasive non-native species (Ojeda and Mares [1982\)](#page-10-10), including the European rabbit (Bonino and Amaya [1985](#page-8-11); Cuevas et al. [2011;](#page-8-12) Bobadilla et al. [2021](#page-8-10)). In the arid ecosystems of Argentina, the colonization and dispersal of European rabbit populations are known to be associated with riverbanks, streams, or moist areas that provide them with feeding and sheltering sites (Bonino and Soriguer [2009;](#page-8-13) Cuevas et al. [2011](#page-8-12); Bobadilla et al. [2022\)](#page-8-14). Nonetheless, there are no specifc studies about the rabbit–plant interactions established in these recently invaded ecosystems.

Characterizing the attributes of the European rabbit's role as a seed dispersal agent should enable understanding not only its efects on the plant communities but also if such interactions facilitate its colonization and expansion in the arid ecosystems of Argentina. We hypothesize that the European rabbit is a legitimate disperser of native and non-native plant species by consuming and disseminating whole seeds, so it could be expected that their feces will yield viable seeds of both types of plants. Therefore, we aim to evaluate the European rabbit's ecological interactions with sympatric plants in its expansion range in the arid landscapes of Argentina. Specifcally, (a) we identify, quantify, and characterize (fruit type and seed size) the seeds of native and non-native species found in European rabbit feces at two sites with diferent establishment times of rabbit populations; (b) we evaluate and compare the germination pattern (percentage and rate of germination and viability) of whole seeds found in feces with those collected directly from parental plants.

Materials and methods

Study area

We selected 2 sampling sites with diferent times of settlement by European rabbit populations within the expansion range in the arid ecosystems of Argentina: (1) Bardas Blancas village (35° 52′ S—69° 48′ W, 1420 to 2800 m elevation), the core area of the rabbit's expansion range (residence time over 45 y); and (2) Laguna de Llancanelo Provincial Reserve (35° 45´ S—69° 08´ W, 1270 to 1500 m elevation), the easternmost edge of the rabbit's expansion range (residence time under 15 y), hereafter called Llancanelo Reserve (Fig. [1](#page-2-0)). These study sites lie in the arid ecotones of the Andean–Patagonia and the Monte–Patagonia ecoregions, respectively (hereafter simply called arid ecotones) (Méndez [2005](#page-10-11), [2014](#page-10-12)). Bardas Blancas site is a dry steppe with an annual mean temperature of 11.9 °C, absolute maximum in December of 36.5 °C, and absolute minimum of−14.6 °C in August (De Fina et al. [1964](#page-9-19)), and with a mean annual precipitation no greater than 300 mm (Norte [2000\)](#page-10-13). The vegetation is a mosaic composed of grasslands dominated by *Sporobolus*, *Pappostipa,* and *Poa*; of shrublands, *Chuquiraga*, *Neosparton*, *Larrea*, and *Prosopis*; and of wetlands, called "vegas", dominated by *Cortaderia*, *Juncus, Baccharis*, *Tessaria,* and *Glycyrrhiza* (Méndez [2014](#page-10-12)). Llancanelo Reserve is a Ramsar site encompassing approximately 90,000 ha with public and private land and includes one of the largest endorheic lagoons in the region, with permanent rivers and streams as well as temporary water inflows caused by fluctuations in nearby water bodies that food the areas closest to the lagoon (Cabrera [1994](#page-8-15); Palma Leotta et al. [2019](#page-10-14)). The Llancanelo Reserve constitutes the main natural wetland in the Monte–Patagonia arid ecotone of Argentina (Blendinger and Alvarez [2002\)](#page-8-16). Here, the climate is arid, with a mean temperature of 19.5 \degree C in warmer months (December to February) and of 3 °C in cooler ones (June to August), with a mean annual rainfall ranging from 215 to 240 mm. The vegetation of the reserve contains 76% of the endemisms of the Mendoza province, and is characterized by fve main environment types: (i) shrublands, dominated by *Prosopis*, *Bougainvillea*, and *Chuquiraga*; (ii) pichanal, dominated by *Baccharis spartioides*; (iii) sand dunes, characterized by *Sporobolus*, *Suaeda*, and *Atriplex*; (iv) wetlands, dominated by *Distichlis*, *Frankenia*, and *Cortaderia*; and (v) tamarindal, dominated by the invasive species *Tamarix* spp. (Méndez [2005\)](#page-10-11).

Feces collection

Sampling was conducted during the wet season of 2018 (February–March) at the two study sites. We used a stratifed random sampling model for 115 fxed strip transects of 1000 m² (5 m \times 200 m) laid out across the study areas on the basis of the habitat types recognized in each site (Bobadilla et al. [2022](#page-8-14)). We covered a total area of approximately 20 $km²$ in Bardas Blancas (45 transects) and 66 km² in Llancanelo Reserve (70 transects). We collected fresh rabbit feces along those transects and checked them once during the wet season. The latrines were avoided because rabbits urinate in them, and this might afect seed germination (Dellafore et al. [2009](#page-9-20)). Rabbit fecal pellets are easy to identify

Fig. 1 Map of Argentina (left) and zoom in of the study area showing the 2 sampling sites with diferent settlement times by European rabbits within their expansion range (depicted by the contour area delineated in black lining) in arid ecotones of Argentina (based on Bonino and Amaya ([1985](#page-8-11)), Bonino and Soriguer ([2009](#page-8-13)), and Bobadilla et al. ([2022](#page-8-14)))

in the feld by their distinctive size, color, shape, and rugosity (Salgado [2016\)](#page-10-15). All feces collected in one transect were pooled into 1 sample (considered independent of the other transects) to be analyzed in the laboratory. In the middle of each study transect, we established one vegetation transect 50 m long, where we measured the specifc composition of plants (Passera et al. 1986). We also collected mature fruits and seeds from plants present in each transect during same season and phenological stage as those obtained from feces. These were used to build a reference collection and for control treatment on germination assays.

Seeds extraction

All feces in each sample were placed in test tubes with an aqueous solution for 30 min to facilitate pellet disintegration. Once the material had softened, the pellets were macerated and disintegrated without exerting pressure. Each sample was sieved to retain fragments on meshes of 1000 μ m and then of 200 µm, washing the material with water several times. Each Petri dish was carefully observed under a binocular magnifying glass $(10\times)$ to separate all seeds. Three categories of seeds were distinguished: (a) whole not damaged after gut passage, (b) partially digested (whole but deemed damaged or unviable), and (c) fragments. In this latter case, a relationship was previously established for seed quantifcation. For this, the number of fragments that constitute one whole seed was determined depending on the species and the mean size of its seeds (Fernández and Saiz [2007\)](#page-9-16). All seeds were recorded, quantifed, and characterized. They were identifed according to the reference collection-built ad hoc during the sampling period and to the reference collection of seeds at the Ruiz Leal Herbarium (IADIZA-CONICET, Argentina). Seeds were identifed to species level when possible. To characterize the seeds of the species found in the feld and identifed in the feces, a bibliographical search was carried out to build a table containing their ecological and morphological characteristics. All whole seeds were kept in a refrigerator at 8 °C until used for the germination tests.

Germination assay

To evaluate the germination capacity of all whole seeds found in rabbit feces, we performed a laboratory assay in standard conditions common to all species, following the recommendations of the International Seed Testing Association (ISTA [2005](#page-9-21)) manual. Previous to the assay, we determined if any of the whole seeds of the species extracted from feces required pre-germination treatments to break dormancy, following the recommendations of the ISTA for each species. The seeds were sowed in Petri dishes (maximum of 40 per dish) with humid flter paper for a maximum of 30 days. We also sowed seeds belonging to the same species, but obtained directly from the mature fruits within the study sites, as a control treatment. Seeds from fruits were extracted under a binocular magnifying glass with the help of tweezers and needles, and they were dried with absorbent paper. During the assay, seeds were kept in a germination room (20–30 °C, 12–12 h light–darkness, watered daily) and observed every day to check for root emergence, indicating seed germination.

Viability assay

To evaluate the viability of the seeds that did not germinate, we used a standard bioindicator (2, 3, 5 triphenyltetrazolium chloride (TTC)) that detects seed viability by staining the embryo tissue pink/red (Moore [1985](#page-10-16)). For it, all seeds were cut open leaving the embryo exposed. Only the seeds with a healthy embryo were incubated in the TTC solution, at room temperature and under darkness for 24 h. We determined seed viability by observing seed coloration under microscope amplifcation, in comparison to positive control seeds collected from the study sites. We determined the number of unproductive versus viable seeds among those non-germinated, by the TTC assay.

Data analyses

We calculated the number and proportion of feces containing whole, digested, and/or fragmented seeds. Both for seeds in feces and for the control, we calculated the germination percentage (%*G*) with the equation: $%G = (n)$ germinated seeds/total seeds sown) \times 100. We analyzed diferences in the proportion of germinated seeds obtained from feces and the control treatment using a chi-square test of homogeneity (Zar [2010\)](#page-10-17) for each species by site. The cumulative percent germination was calculated and was drawn in order to demonstrate diferences in germination patterns among treatments. The evaluation of seed germination was based on two additional parameters: germination start (GS) and germination rate (GR). The germination start was defned as the time interval (days) between sowing and the emergence of 1/6 of the fnal germination percentage, and was calculated with the equation: $GS = 1/6 \times \%G$ (Izhaki and Ne'eman [1997](#page-9-22)). We calculated the germination rate with the equation: $GR = (5/6 \times \%G)$ / (*T*2−*T*1), where *T*1 is the interval (in days) between sowing and emergence of 1/6 x %*G* of the seedlings, and *T*2 is the interval (in days) between sowing and emergence of 5/6 x %*G* of the seedlings (Izhaki and Ne'eman [1997](#page-9-22)). The parameter GR was calculated in order to describe the rapidity in the % germination per day for each species by site and treatments (Izhaki and Ne'eman [1997](#page-9-22)).

Results

Feces collection

We collected a total of 1283 pellets. Of these, 71% were from the wetland areas of both study sites. The number of feces per site was 397 pellets at Bardas Blancas (number of transects=6) and 886 at Llancanelo Reserve (number of $transects=8$).

Seeds extraction

We found a total of 855 whole seeds of 6 identifed plant species, and 46 whole seeds of one unidentified plant (Table [1](#page-4-0)). The number of whole seeds per pellet was higher in Bardas Blancas than in Llancanelo Reserve (Table [1](#page-4-0)). Most of the whole seeds found in feces were of the shrub *Frankenia juniperoides* (88% in Bardas Blancas and 57% in Llancanelo Reserve). The highest number of fully or partially digested seeds was also observed for the same *F. juniperoides* (98% and 68%, respectively). The highest number of seed fragments identifed was from the shrub *Lycium chilensis* (55%) and they were detected only in Llancanelo Reserve (Table [1\)](#page-4-0). Of the six identifed species, only one was non-native: *Sesuvium portulacastrum*. Two species were identifed at the genus level, with one morphotype for each genus (*Arjona* sp. and *Poa* sp.). Three species had incomplete values because data were not available. The whole seeds found in feces had a maximum size of ca. 1.6×2.0 mm and corresponded mostly to *L. chilensis* (Table [1](#page-4-0) and Fig. [2](#page-5-0)).

Germination assay

Three of the seven plant species with whole seeds obtained from feces germinated after 30 days: *F. juniperoides* (at both sites), *L. chilensis*, and *S. portulacastrum* (only at the Llancanelo Reserve) (Fig. [2\)](#page-5-0). Of the seed species obtained from the fruits (control), only *F. juniperoides* germinated both in Bardas Blancas as in Llancanelo Reserve. Fourteen percent of the 901 whole seeds from feces germinated after 30 days and 0.9% died due to fungal attack. Forty one percent of the total control seeds germinated after 30 days and 1.8% were attacked by fungi. The germination percentage was signifcantly higher in the control treatment than in the digested seeds for *F. juniperoides* both in Llancanelo Reserve (51% and 31%, respectively; χ 2=26.86, *p* < 0.0001, df = 1) and in Bardas Blancas (58% and 7%, respectively; χ 2=177.10, $p < 0.0001$, df = 1) (Fig. [3](#page-5-1)). Three percent of *L. chilensis* seeds extracted from feces showed radicle protrusion during the assay, while none of the seeds from the fruits germinated $(\chi^2 = 4.07, p < 0.05, df = 1)$ (Fig. [3\)](#page-5-1). The seeds of *S*. *portulacastrum* obtained from fruits did not germinate after 30 days, while those obtained from feces did (5%; χ 2=4.1, $p < 0.05$, df = 1) (Fig. [3\)](#page-5-1). The *F. juniperoides* control seeds germinated frst and had the highest germination rate than those seeds from feces at both Bardas Blancas and Llancanelo Reserve (Table [2](#page-5-2) and Fig. [4](#page-5-3)).

throughout the wet season of 2018 at two study sites in arid ecotones of Argentina								
Plant species	Fruit type	Seed size $(mm)^*$	Bardas Blancas village			Llancanelo Reserve		
					Whole seeds Digested seeds Damaged seeds Whole seeds Digested seeds Damaged seeds			
Arjona sp.	Achene (dry)					$\overline{2}$	26	29
Fabiana denudata	Capsule (dry)	$1.3 - 1.6 \times 0.5 - 3$ 0.6		Ω				
Frankenia juniperoides	Capsule (dry)	$0.5 - 0.8 \times 1.0 - 293$ 1.6		241	35	326	526	10
Monnina dictyocarpa	Samara (dry)	$\overline{}$	0	$\mathbf{0}$				
Lycium chilensis		Berry (fleshy) $1.3 - 2.0 \times 1.0 -$ 1.5				120	198	91
Poa sp.	Caryopsis (dry) -		35	6	22			
Sesuvium portulacastrum	Capsule (dry) $1.2-1.5$			$\mathbf{0}$	9	75	19	16
Undetermined 1	\sim		$\mathbf{0}$	$\mathbf{0}$	Ω	46	Ω	11
Undetermined 2 -			0	$\mathbf{0}$	96	Ω	$\mathbf{0}$	8
Total seeds			332	247	164	569	769	165
Seeds/pellet			0.84	0.62	0.41	0.64	0.87	0.19
SD			0.98	0.43	0.55	0.63	1.05	0.18

Table 1 Number, fruit type, and average size (length×width) of seeds of diferent plant species found in European rabbit pellets collected throughout the wet season of 2018 at two study sites in arid ecotones of Argentina

 $*$ Based on Bernardello [\(1983](#page-8-17)), Lonard and Judd [\(1997](#page-9-23)), Lüdtke et al. ([2009\)](#page-9-24), and Abalos [\(2016](#page-8-18))

Fig. 2 Germinated seeds extracted from European rabbit feces of: **a** non-native *Sesuvium portulacastrum*, **b** native *Frankenia juniperoides*, and **c** native *Lycium chilensis*

Fig. 3 Germination percentage of *F. juniperoides* $(n_{\text{treatment}} = 326,$ $n_{\text{control}} = 326$, *L. chilensis* ($n_{\text{treatment}} = 120$, $n_{\text{control}} = 120$), and *S. portulacastrum* ($n_{\text{treatment}} = 75$, $n_{\text{control}} = 75$) in Llancanelo Reserve; and of *F. juniperoides* in Bardas Blancas $(n_{\text{treatment}}=293, n_{\text{control}}=293)$. $n_{\text{treatment}}$: whole seeds found in European rabbit feces and exposed to the assay; $n_{control}$: whole seeds extracted from the plants and exposed to the assay. **p*<0.05; ****p*<0.0001

Viability assay

Of the 721 seeds from the feces that did not germinate during the assay, 12% were unproductive seeds (Table [3](#page-6-0)). All non-germinated seeds from the feces that presented a healthy embryo were deemed non-viable by the TTC assay (Table [3](#page-6-0)). Of the 468 non-germinated control seeds, 19% had an empty seed or a poorly developed embryo. Eleven percent

Table 2 Average germination start (GS, time interval in days between sowing and emergence of $\frac{1}{6} \times P$), germination rate (GR, % germination per day) of *Frankenia juniperoides* seeds at two study sites

Fig. 4 Cumulative germination percentage over time of native *F. juniperoides* seeds obtained from European rabbit feces (segmented lines) and the control (continuous lines) for Bardas Blancas (grey) and Llancanelo Reserve (black)

of the control seeds corresponding to *F. juniperoides* and *L. chilensis* were positive by the TTC assay (Table [3](#page-6-0)). All seeds belonging to *A. patagonica*, *F. denudata*, and *Poa* sp. were judged non-viable (Table [3](#page-6-0)).

Discussion

Our results show that the European rabbit is a legitimate disperser of at least three plant species in its expansion range in the arid ecosystems of Argentina. In these study areas, the non-native plant species represent less than 10% of the total species (54 in Bardas Blancas and 52 in Llancanelo Reserve). According to our results, the rabbit disperses 2 to 4% of the native plant species and 2% of the non-native plant species recorded in the feld. In its native range, the rabbit disperses 8% of the plant species in a coastal dune (Dellafore et al. [2009](#page-9-20)) and 50% of the plants in a Mediterranean dehesa (Malo and Suárez [1995](#page-9-25)). These diferences in the percent of dispersed plants by the same vector in diferent **Table 3** Number of unproductive seeds (empty seed or poorly developed embryo) found among all nongerminated seeds by species, and viable seeds over the total of seeds with healthy embryos exposed to the TTC test in each of the two study sites in arid ecotones of Argentina

environments may be related to both the number of species likely to be dispersed (survival capacity of seeds when passing through the rabbit's digestive tract) and their abundance in the originating community (supply of fruits and/or seeds) (Traveset [1998](#page-10-18)). For invasion ranges, the eco-evolutionary experience of both the introduced species and the recipient community (Saul et al. [2013\)](#page-10-19) and the possibility of establishment of novel interactions (ecological ftting, Janzen [1985](#page-9-26)), could also be factors that explain these diferences. In this sense, it is expected that the incidence and extent of the novel interactions increase with residence time. The analysis of these factors is beyond the scope of the present work and reinforces the need for further studies considering the abundance of rabbits and of plant resources, through diferent phenological stages in the arid landscapes of Argentina.

The load of whole seeds per rabbit pellet recorded in Argentina is below the values reported in sand dunes in Spain, 0.94–0.97 (Calviño-Cancela [2002;](#page-8-19) Larrinaga [2010](#page-9-27)), but higher than the lowest values reported in an acid grassland in England, 0.1–0.5 seeds/pellet (Pakeman et al. [1999](#page-10-20)). Regarding the relationship between numbers of plant species dispersed versus those eaten, rabbits dispersed from 6 to 17% of the species recorded in their diet at the same two study sites in Argentina (Bobadilla et al. [2022\)](#page-8-14). In dune environments as well as in woods, scrub, and pastures, some of the species dispersed by rabbits also have been cited as part of their diet (Dellafore et al. [2007](#page-9-28), [2009](#page-9-20)). Indeed, few studies have determined the number of species consumed by rabbits versus the number of species dispersed by endozoochory at any given site (Godó et al. [2022\)](#page-9-29).

It is generally assumed that seeds are accidentally ingested by herbivores when consuming vegetative parts of plants (the "foliage is fruit" hypothesis of Janzen [1984\)](#page-9-30). This could be the case of *L. chilensis,* because its foliage represents an important food item in the rabbit's diet, but not for *F. juniperoides* or *S. portulacastrum*, whose seeds but not their vegetative parts are present in the diet (Bobadilla et al. [2022](#page-8-14)). Similarly, Pakeman et al. ([1999\)](#page-10-20) observed in England that rabbits play a more frugivorous role than an indirect removal of seeds by herbivory, actually selecting the fruits of some of the species that they disperse. Such evidence is consistent with our results and with other studies showing that the European rabbit actively selects small seeds (Larrinaga [2010](#page-9-27)) or mainly consumes small fruits (Guerrero-Campos et al. [2023](#page-9-31)) both in native and non-native areas.

The proportion of fragmented seeds in rabbit feces ranged 6–22% between the two study sites. In addition, we found>40% digested whole seeds. Thus, our results highlight the complex relationship between endozoochory and predation (involving mastication and digestion) (Malo and Yanes [1999\)](#page-9-32). According to Dellafore et al. ([2007\)](#page-9-28), in a sand dune in Spain the plants dispersed by endozoochory were also preyed on by rabbits. Indeed, the passing of seeds through an herbivore's digestive tract can increase their mortality if seed coats are weak, and increase their germination by scarifcation during digestion (Janzen et al. [1985](#page-9-26)). Our germination tests allowed to determine that the rabbit's gut passage of seeds favored the germination of *L. chilensis* and *S. portulacastrum*, while for *F. juniperoides* the seeds collected directly from the plant germinated earlier and in greater quantity. Therefore, as observed by Dellafore et al. ([2007,](#page-9-28) [2009\)](#page-9-20) in Spain, we found that for some plant species the passage of seeds through the digestive tract of rabbits is only a dispersion means, while for others it also favors subsequent germination, which indicates that rabbits act as legitimate dispersers. Contrariwise, in Argentina, rabbits act as illegitimate dispersers for those plant species whose whole but non-viable seeds were found in the feces (*Arjona* sp., *F. denudata*, *Poa* sp.).

Viability tests indicate that all seeds ingested by rabbits that did not germinate were not viable. Nevertheless, viable seeds from the controls were found for *F. juniperoides* and *L. chilensis*. Our results indicate that the control seeds of *L. chilensis*, were viable but did not germinate during the time that the test lasted, while those in the feces did, showing that their passage through the rabbit's gut accelerated their germination rate. It is important to point out that during water deficit periods, seeds are naturally bound to germinate quickly to ensure species survival in arid ecosystems (Contreras Quiroz et al. [2015\)](#page-8-20). Therefore, the improved germination time of *L. chilensis* seeds dispersed by rabbits could be relevant in such environments.

Herbivory efects involve digestion, intestinal retention period, seed size, and seed coat hardness (Campos et al. [2008](#page-8-3)). Considering that rabbits carry out caecotrophy (ingestion of their own soft feces), seeds may pass more than once through their digestive tract, afecting their germination pattern according to the number of intestinal passes (Castro et al. [2008](#page-8-2); Mancilla-Leyton et al. [2013\)](#page-9-33). Previous studies showed signifcant increases in the germination percentage of seeds that passed twice through rabbit gut (Castro et al. [2008](#page-8-2); Mancilla-Leytón et al. [2013](#page-9-33)). Also, negative associations have been observed between seed size and the percentage of whole seeds recovered in rabbit feces: seeds<4 mm were more likely to pass whole through their digestive tract (Malo et al. [2000;](#page-9-34) Pakeman et al. [2002\)](#page-10-21). Consistent with these reports, in our study, whole seeds in rabbit feces did not exceed 2×1.5 mm. This suggests a maximum seed size for their intact passage through the rabbit gut and reinforces the need for further studies that assess morphometric variables of whole and damaged seeds in rabbit feces.

Of the three species of seeds dispersed by rabbits in our study, two were native (*L. chilensis* and *F. juniperoides)* and one was non-native (*S. portulacastrum)*. Our results show a signifcant increase in the germination percentage of seeds of *L. chilensis*. This native shrub is a good forage species, especially important in arid and semi-arid ecosystems where most plants have low nutritional value due to their high fber content (Noy-Meir [1973](#page-10-22); Cuevas et al. [2013](#page-8-21)). Indeed, *L. chilensis* is the ffth most frequent food item in the rabbit diet during the wet season at Llancanelo Reserve (Bobadilla et al. [2022\)](#page-8-14), and its feshy and juicy fruits constitute a nutritious, accessible, and abundant food during summer (Giorgetti et al. [2000](#page-9-35)). Therefore, the rabbit–*L. chilensis* interaction may depict a scenario of positive ecological efects between a non-native disperser and a native plant in this study site. Regarding the other native plant, *F. juniperoides*, our results show a signifcant reduction in the percentage and rate of seed germination after the passage through the rabbit's gut. This is thus, a legitimate but inefficacious or inefficient dispersal service. However, the rabbit–*F juniperoides* interaction provides another dispersal mechanism for this native plant that could provide an advantage if its seeds are deposited in suitable sites, far from the parental plant (Schupp et al. [2010](#page-10-3)).

For the non-native *S. portulacastrum*, our results show a signifcant increase in the germination percentage of its seeds. This is a perennial herb with a wide geographic distribution on all continents and is considered non-native in our study sites (Méndez [2005;](#page-10-11) Jocou and Gandullo [2020](#page-9-36)). It presents typical anatomical features of wetland plants, combined with xeromorphic characters (Apóstolo [2005\)](#page-8-22) that could act as adaptive strategies for acclimatizing to these wetlands in arid environments. Its success as a colonizing species in other regions has been associated to its ability to be propagated by vegetative fragments, while presenting a continuous fowering and fruiting cycle (Lonard and Judd [1997](#page-9-23)). This reproductive strategy enables fruits and seeds to be available throughout the year, which is especially important in arid landscapes with periods where food is less abundant for herbivores (dry season). There are no records of endozoochorial seed dispersers for this species (Lonard and Judd [1997;](#page-9-23) Bohley et al. [2017](#page-8-23)), so the European rabbit would be its frst described animal vector. Therefore, the incipient new interaction of rabbit–*S. portulacastrum* could have a reciprocal effect on one another, potentially aiding their mutual invasive process (and subsequent impacts) in the wetlands of Bardas Blancas and Llancanelo Reserve. These habitat types are especially important in arid biomes (Blendiger and Alvarez [2002](#page-8-16)). In particular, the wetland in Llancanelo Reserve has a high level of endemism, made up of diferent plant communities distributed in two vegetation units: hygrophilic (humid) and halophilic (saline) (Roig et al. [2000](#page-10-23)).

Our fndings add to other studies documenting the European rabbit as a relevant disperser of several plant species both native and non-native in indigenous and invaded ranges in diferent ecosystems (Malo et al. [2000](#page-9-34); Castro et al. [2008](#page-8-2); Dellafore et al. [2009](#page-9-20); Salas-Pascual et al. [2009](#page-10-24); Bobadilla et al. [2020](#page-8-8)). From a conservation viewpoint, these interactions may be deemed negative if they facilitate the invasion of non-native plant species, or positive if the dispersal of native species occurs (McConkey et al. [2012\)](#page-10-0). Nevertheless, it is known that exotic lagomorphs may be less efficient in seed dispersal than are native frugivores which may be specialized dispersers of certain plants, considerably changing seed removal rate, seed availability, as well as seed and seedling survival (Godó et al. [2022](#page-9-29)). In our study, rabbits legitimately dispersed two native and one non-native plant species by consuming and disseminating whole seeds, supporting our hypothesis. On the order hand, due to Llancanelo Reserve being at the edge of the expansion range of the European rabbit in arid ecotones of Argentina, the ecological interaction of rabbits with diferent plant species takes special importance for two reasons. First, because of the seed dispersal service that rabbits perform for recruiting native plants into the wetlands inside the protected area. Second is for their spreading of non-native plants, aiding their invasive

process. Our study highlights the complex ecological role played by the European rabbit in simultaneously facilitating recruitment of native plants but also that of non-native species, thus conducive to invasional meltdown (Simberloff and Von Holle [1999](#page-10-1)).

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Author contribution All authors contributed to the study conception and design. Material preparation, data collection, and analysis were performed by S.Y. Bobadilla, E. T. Olivares, F. M. Jaksic, R. A. Ojeda, and M. F. Cuevas. The frst draft of the manuscript was written by S.Y. Bobadilla and all authors commented on previous versions of the manuscript. All authors read and approved the fnal manuscript.

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Data availability The datasets generated and/or analyzed during the current study are available from the corresponding author on reasonable request.

Declarations

Competing interests The authors declare no competing interests.

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