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Use of water treatment residuals for colour removal from real textile dye wastewater

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Abstract

The use of readily available water treatment residuals (WTR) as a low-cost material for removal of colour from real textile wastewater was investigated. WTR was used in three forms, namely in raw wet form as a coagulant, in the dried form as an adsorbent and as a fltration media in column tests. The results showed a maximum colour removal of 55 and 36% by coagulation and adsorption, respectively, and the corresponding COD removals were 35 and 37%. Coagulation and batch sorption tests showed the efect of initial pH on the colour removal, and maximum colour removal was obtained at an initial pH of 3.0. Long-duration continuous-fow column test using WTR as a fltration/sorption media showed that a maximum colour removal of 60% can be achieved. In column studies, complete exhaustion of the media occurred at 180 and 120 bed volumes, respectively, for initial pHs of 3.0 and 6.2. The study thus shows the potential of WTR for primary treatment of real textile dye wastewater.

Keywords Textile dye wastewater · Water treatment residuals · Coagulation · Adsorption · Fixed-bed column test

Introduction

With the increasing demand for textile products, management of huge quantity of textile and dyeing wastewater generated from textile industries has become a major concern as it contains a variety of pollutants and toxic substances such as dyes (Kadam et al. [2015\)](#page-7-0). Nowadays, the synthetic origin of dyes to produce strong fastness has produced several detrimental effects on the environment and human health. Firstly, very low concentration of dyes in water makes water highly coloured and unpleasant (Paz et al. [2017](#page-7-1); Sultan [2017\)](#page-7-2), and the discharge of coloured wastewater and their metabolites in aquatic ecosystems reduces sunlight penetration to cause inhibitory efects on photosynthesis (Hassine et al. [2016](#page-6-0)). Several treatment processes used for colour removal are based on physiochemical and biological processes. Compounds produced when textile wastewater undergoes anaerobic degradation for colour removal are reported to be toxic, carcinogenic and mutagenic (Paz et al. [2017](#page-7-1); Ponraj et al. [2017\)](#page-7-3). Thus, the usual approaches used

 \boxtimes M. Mansoor Ahammed mansoorahammed@gmail.com for colour removal are coagulation, sedimentation, adsorption, photocatalysis, degradation and ion-foatation (Ponraj [2017](#page-7-3); Mashkoor et al. [2018;](#page-7-4) Inamuddin [2019\)](#page-7-5).

To achieve the sustainable development goal, reuse and recycle of waste materials are essential. In view of the same, several research efforts have been made particularly to recover and reuse waterworks sludge in benefcial ways. Waterworks sludge is a by-product generated as a result of coagulation and focculation process due to the addition of coagulant (aluminium or ferrous salts) in water treatment plants (Jiao et al. [2016](#page-7-6)) and is generally known as water treatment residuals (WTR). In India, alum is commonly used as a coagulant. As alum is added to water, it precipitates as aluminium hydroxide as amorphous and semi-crystalline form together with which suspended impurities are removed in settling unit (Benjamin and Lawler [2013](#page-6-1); Keeley et al. [2014](#page-7-7)). However, due to lack of sustainable approaches towards utilisation of WTR in benefcial and economical way WTR are still disposed of in water bodies in developing countries, while another option used for disposal is landfll (Wang et al. [2016\)](#page-7-8). As million of tons of WTR are generated annually around the world, scarcity of landfll sites and stringent environmental legislation makes disposal of WTR more problematic and costlier. Thus, reuse of WTR for potential applications has received much attention in recent years.

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Water treatment residuals can be used by frst recovering the coagulant. Direct reuse with or without alteration is also possible. Recovery of coagulant from WTR involves process such as acidifcation, basifcation, ion exchange and membrane separation (Nair and Ahammed [2014b;](#page-7-9) Keeley et al. [2012](#page-7-10)) which are efective if they are economical. The WTR management approaches include reuse in construction materials (Babatunde and Zhao [2007;](#page-6-2) Verma et al. [2012](#page-7-11); Ahmad et al. [2016](#page-6-3)), wastewater treatment (Wang et al. [2016](#page-7-8); Nair and Ahammed [2015](#page-7-12); Maghaddam et al. [2010](#page-7-13)) and for soil improvement (Soleimanifar et al. [2016](#page-7-14)). WTR have been used in two diferent forms in wastewater treatment. It can be used in its raw wet form as a coagulant or as an adsorbent in its dried form. The efective reuse of WTR as a coagulant has been reported for colour removal from acid dye (Maghaddam et al. [2010](#page-7-13)) and disperse dye (Gadekar and Ahammed [2016](#page-6-4)) and for posttreatment of up-fow anaerobic sludge blanket reactor (UASBR)-treated wastewater (Nair and Ahammed [2014a\)](#page-7-15). The WTR have been widely used as an adsorbent for removal of anionic/cationic compounds such as phosphorus (Wang et al. [2016;](#page-7-8) Soleimanifar et al. [2016](#page-7-14); Quiñones et al. [2016\)](#page-7-16) and heavy metals (Jiao et al. [2016;](#page-7-6) Soleimanifar et al. [2016;](#page-7-14) Quiñones et al. [2016](#page-7-16); Ociński et al. [2016](#page-7-17); Zhou and Hyenes [2011\)](#page-7-18). WTR have been used for colour removal by Gadekar and Ahammed ([2016](#page-6-4)), Moghaddam et al. ([2011](#page-7-19)) and Moghaddam et al. [\(2010](#page-7-13)) in its wet form. No studies have been reported on the use of WTR as a sorbent for colour removal. Also, no studies have been reported on their use for treatment of real textile dye wastewater. Further, studies comparing the application of WTR in diferent forms such as dry and wet form and also as a fltration material for colour removal from real textile wastewater are also lacking. This work aimed to use water treatment residuals in its available form without modifcation for treatment of coloured textile wastewater.

In the present study, water treatment residuals were characterised. The raw wet form of WTR as a coagulant and dry form as an adsorbent were used for colour removal from a real textile dye wastewater. The efects of two parameters, namely WTR dose and initial pH, were studied. Further, continuous-fow fxed-bed column studies were also carried out.

Materials and methods

Textile dye wastewater

Textile dye wastewater was collected from a common efuent treatment plant run by Palsana Enviro Care, Surat, Gujarat, India. This plant treats wastewater collected from diferent textile dye industries. These industries mostly use disperse dyes in their processes. The collected wastewater

was stored at 4 °C and was used as required. Colour was measured at a wavelength corresponding to the maximum absorbance of 350 nm for the wastewater used. Table [1](#page-1-0) shows the characteristics of the wastewater used in the study.

Water treatment residuals

Water treatment residuals were collected from the coagulation/focculation unit of a water treatment plant in Bhandup, Mumbai, India, where polyaluminium chloride (PACl) is used as a coagulant. A portion of the collected sludge was stored at room temperature (27–30 °C) in its original form, and this was used in tests of WTR as a coagulant. Another portion of the WTR was sun-dried and ground, and the fraction passing through 216-µm sieve was used as a sorbent and in the column tests without further modifcation.

Characterisation of water treatment residuals

The elemental composition of WTR was analysed using inductively coupled plasma atomic emission spectroscopy (ICP-AES) (ARCOS Spectro, Germany) after acid digestion as per USEPA method 3050B (USEPA [1996](#page-7-20)). Scanning electron microscope (SEM) with energy-dispersive spectroscopy was used to analyse the constituents and nature of particles present in the WTR.

The point zero charge (pHzpc) of WTR was determined adopting the following procedure of Mehdi et al. [\(2014](#page-7-21)): 0.01 M·NaCl solution with diferent pHs in the range of 2–10 was prepared by adding HCl/NaOH. Thereafter, 0.20 g of the WTR was added to 50 mL of 0.01 M NaCl solution of diferent pHs in reagent bottles. The samples were centrifuged after allowing them to react for 48 h at room temperature, and the pH of each solution was measured.

Analysis

The colour removal was calculated using the following equation (Gadekar and Ahammed [2016](#page-6-4)).

*Based on analysis of four samples

Column

\n
$$
\text{Column}(\%) = \frac{C_o - C_e}{C_o} \times 100\tag{1}
$$

where C_o and C_e are colour concentration measured as absorbance at maximum absorbing wavelength of raw and treated wastewater, respectively.

The chemical oxygen demand of samples was analysed using the closed refux volumetric analysis, while pH and electrical conductivity (EC) were determined by potentiometric method. Biochemical oxygen demand was analysed by titrimetric method and total solids by gravimetric method. All of methods performed were in accordance with the Standard Methods ([1998\)](#page-7-22).

Coagulation/focculation tests

To simulate coagulation/focculation process at laboratory scale, a six-beaker jar-test apparatus (DBK Instruments, Mumbai) was used. Raw wet-form WTR as a coagulant was added to 250 mL of wastewater. The WTR dosages were calculated based on dry weight. The standard procedure consisted of 2-min fash mixing at 150 rpm, 30-min slow mixing at 25 rpm and 20-min settling. Efects of two parameters, namely efect of WTR dose and initial pH, were studied. Samples of desired initial pHs were obtained by addition of 1 N H_2SO_4 or NaOH. The wastewater was tested with an initial pH range of 3.0–7.0 and WTR dose of 2 to 12 g/L. The supernatant sample was centrifuged at 5000 rpm for 3 min and was analysed for colour removal.

Batch sorption tests

Batch sorption studies were carried out using different doses of WTR with 50 mL of wastewater. Dye solution with WTR was mixed on the orbital shaker at 150 rpm for 60 min. Based on the equilibrium studies an equilibrium time of 60 min was found enough to reach the equilibrium. The wastewater was tested in initial pH range of 3.0–7.0 and WTR dose of 10–50 g/L. Treated samples were centrifuged and analysed for colour removal.

Fixed‑bed continuous‑fow column study

Continuous-fow column experiments were performed in a transparent cylindrical acrylic column (2.5 cm internal diameter and 30 cm height) with a medium bed height of 15 cm. Glass wool was used at inlet and outlet to avoid the loss of the media. The column was operated in upward mode by pumping the wastewater solution through the column at a fow rate of 50 mL/h using a peristaltic pump. Samples were collected at diferent time intervals and were analysed for colour removal.

Results and discussion

Characteristics of water treatment residuals

Chemical composition of WTR is shown in Table [2](#page-2-0). Composition of WTR depends on the source water quality and the coagulant used (Nair and Ahammed [2014a\)](#page-7-15). Aluminium and iron concentrations were found to be 64.60 and 93.01 mg/g, respectively, while cations like calcium, magnesium and potassium were also detected in ample amounts. The literature reports typical aluminium content of 27.09–171.77 mg/g dry WTR (Jiao [2016;](#page-7-6) Dassanayake et al. [2015](#page-6-5); Agyin-Birikorang and O'Connor [2009,](#page-6-6) Gadekar and Ahammed [2019](#page-6-7)). Most of the studies reported are based on alum-based WTR, but the present study was carried out employing PACl-based WTR. Scanning electron micrographs (SEM) (Fig. [1](#page-3-0) a, b) show the amorphous nature of WTR, obtained as a result of aggregation, rearrangement and hydrolyses to amorphous hydroxide precipitate (Shi et al. [2007](#page-7-23)).

The point zero charge (pHzpc) of WTR was found to be 6.9. The surface of WTR becomes positively charged at pH < p Hzpc favouring the adsorption of anionic pollutants, whereas cationic pollutant sorption would be favourable at pH>pHzpc (Peláez-Cid et al. [2016\)](#page-7-24). Point zero charge for WTR reported in the literature varies between 5.5 and 6.8 (Zhou and Haynes [2011](#page-7-18); Putra and Tanaka [2011\)](#page-7-25). Moreover, morphology of sorbent is important in the sorbent process (Ociński et al. [2016\)](#page-7-17).

Coagulation study

Efect of pH

In coagulation/focculation process initial solution pH is a key parameter (Merzouk et al. [2011;](#page-7-26) Verma et al. [2012\)](#page-7-11), and maximum pollutant removal will be obtained at optimal pH with optimal coagulant dosage. Figure [2](#page-3-1) presents the effect of initial solution pH on colour removal efficiency at different WTR dosages. It can be noticed that maximum colour removal occurred at an initial pH value of 3.0 for all the

Table 2 Characteristics of water treatment residuals used

Parameter	Value	Unit
pH	6.4	
Al	64.60	(mg/g of dry sludge)
Fe	93.01	(mg/g of dry sludge)
Ca	27.27	(mg/g of dry sludge)
Mg	17.34	(mg/g of dry sludge)
K	20.06	(mg/g of dry sludge)

Fig. 1 SEM–EDS of water treatment residuals

Fig. 2 Effect of initial pH on colour removal at different WTR dosages **Fig. 3** Efect of WTR dose on fnal pH at diferent initial pH values

three WTR dosages. Colour removal was signifcantly low at all other pHs and was nearly constant for a particular WTR dosage. Final solution pH is a major infuencing parameter in coagulation and focculation phase (Gadekar and Ahammed [2016](#page-6-4)). Figure [3](#page-3-2) presents the efect WTR dose on fnal pH at diferent initial pH values. In general, the fnal pH increased with increase in WTR dosage. However, this increase was more pronounced at lower initial pH values. The aluminium hydroxide in WTR balances the anionic charge of dye and thus increases the wastewater pH. Maximum colour removal recorded at an initial pH value of 3.0 and WTR dose of 10 g/L which had a fnal pH around 5.30. This implies that maximum colour removal can be obtained using WTR at system pH around 5.0–5.5. The results presented show similar trends as reported by Gadekar and Ahammed [\(2016\)](#page-6-4), and Moghaddam et al. [\(2011](#page-7-19)) for disperse dye removal and acid dye removal, respectively. Maximum colour removal was obtained at an acidic pH value 3.0 (Gadekar and Ahammed

[2016](#page-6-4)) and at 3.42 (Moghaddam et al. [2011\)](#page-7-19). This behaviour may be explained based on the efect of fresh coagulant on pollutant removal. Shi et al. ([2007\)](#page-7-23) investigated the colour removal performance with diferent aluminium species and found that coagulation efficiency significantly increased with decrease in pH and complete dye removal was obtained at sufficiently lower pH of 5.6. The formation of $Al(OH)_{3}$ responsible for pH change in coagulation process, which precipitates when the fnal pH is around 4.0 for alum (Merzouk et al. [2011](#page-7-26)). Thus, WTR as coagulant can be efective when addition of WTR can increase the initial pH of solution to around 5.0.

Efect of WTR dose

Figure [4](#page-4-0) shows the efect of WTR dose on colour and COD removals. Maximum colour removal obtained was 56% at an initial WTR dose of 10 g/L. Maximum removal reported for

Fig. 4 Efect of dose of WTR on colour removal during coagulation **Fig. 5** Efect of WTR dose on colour and COD removal for wastewa-

removal of colour from synthetic disperse dye solution using WTR was 88% for an initial dye concentration of 25 mg/L, which was reduced to 76% when dye concentration was increased to 75 mg/L which was obtained at a WTR dose of 3 g/L (Gadekar and Ahammed [2016](#page-6-4)). With fresh coagulant (alum) a colour removal of 78.9% was observed for synthetic disperse dye solution at an alum dose of 200 mg/L (El-Gohary and Tawfk [2009\)](#page-6-8), while Merzouk et al. ([2011\)](#page-7-26) reported a colour removal of 94% with an alum dose of 40 mg/L. It is seen that (Fig. [4\)](#page-4-0) colour removal increased with WTR dose and similar trends have been reported with WTR (Gadekar and Ahammed [2016](#page-6-4)), alum (Merzouk et al. 2011 ; El-Gohari and Tawfik 2009), FeCl₃ and alum combined with polysaccharides (Huang et al. [2015\)](#page-6-9) for disperse dye wastewater. Higher colour removal was observed at a WTR dose of 10 g/L, and further increase in WTR results reduction in the colour removal. This behaviour may be due to the fact that addition of excessive WTR might reduce the zeta potential to zero and thus produce adverse efect on the colour removal (Merzouk et al. [2011\)](#page-7-26). Work carried out by Moghaddam et al. ([2010](#page-7-13), [2011](#page-7-19)) using ferrous-based and aluminium-based WTR on acid dye removal showed similar trends of increase in colour removal with increase in WTR dose. In the present study, COD removal increased with increase in WTR from 18 to 30% as WTR dose increased from 2 to 10 g/L and further increase in the WTR dose reduced the COD removal.

Batch sorption study

Efect of WTR dose

The effects of the sorbent dosage were investigated, and the data are shown in Fig. [5.](#page-4-1) The colour removal increased with increase in adsorbent dosage from 10 to 30 g/L, and no further increase in removal was observed beyond 30 g/L. The

ter at initial pH 3.0 in sorption study

percentage removal at the adsorbent dosage of 30 g/L was 36%, and further addition of WTR did not have a signifcant effect on removal of colour. Maximum colour removal obtained was 52.6% for disperse navy blue 3G dye at WTR dose of 30 g/L (Gadekar and Ahammed [2019](#page-6-7)). Similar trends were observed for COD removal also. Maximum COD removal obtained was 37%.

Efect of pH

Figure [6](#page-5-0)a presents the effect of initial pH on adsorption of colour onto WTR. It was observed that the colour removal was decreased with increase in initial pH which implies that the dye sorption was largely controlled by initial solution pH. This could be due to dissolution of aluminium at low pHs (Eq. [2](#page-5-1), [3](#page-5-2)) (Keeley et al. [2014;](#page-7-7) Nair and Ahammed 2016b), which later precipitates in hydroxide form to produce a flocculation effect and subsequent colour removal. However, at lower pHs of 3.0 and 4.0 signifcant colour removal was observed. This implies that aluminium leaching is not the only parameter which affects the colour removal process. Another reason behind this could be explained by the charge-neutralising capacity of the WTR. The point zero charge of WTR was 6.9, and decreasing pH below pH_{zpc} indicates surface of the material would attain a net cationic charge and thus adsorption of anionic dye is favoured. In charge neutralisation, the increase in pH leads to change of surface charge from positive to negative (Yang et al. [2006\)](#page-7-27) and thus reduction in dye removal. Figure [6b](#page-5-0) shows the COD removal of wastewater at diferent initial pHs. A drastic reduction in COD removal was observed as pH increased from 3.0 to 4.0. It was observed that COD removal was high due to the surface charge neutralisation efect and thus at lower pH maximum colour removal was obtained. However, as pH was increased to 4.0 the charge neutralisation was not

Fig. 7 Breakthrough curve for colour removal at **a** initial pH 6.2, **b** initial pH 3.0

signifcant to adsorb dyes and organic matter in wastewater, which resulted in reduction in COD removal.

$$
Al^{3+} + H_2O \rightleftharpoons Al(OH)^{2+} + H^+ \tag{2}
$$

$$
Al^{3+} + 2H_2O \rightleftharpoons Al(OH)_2^{2+} + 2H^+ \tag{3}
$$

Fixed bed continuous flow column study

Results of the column tests are presented in the form of breakthrough curve for unchanged initial pH of 6.2 (Fig. [7](#page-5-3)a) and modifed initial pH 3.0 (Fig. [7](#page-5-3)b). The breakthrough curves show that a good colour removal of 60–70% was obtained in the beginning of the column operation for both the initial pH values. A comparison of the column performance at two initial pHs indicates that at the lower initial pH of 3.0, the column could be operated for longer duration. In this, the colour removal remained above 60% for about 80 bed volumes, while in the other column, it was only about 50 bed volumes. Further, complete exhaustion of the column occurred at about 180 and 120 bed volumes, respectively, for initial pH of 3.0 and 6.2. This diference was not

Fig. 8 Spectral variation of wastewater at diferent bed volume treated

surprising as the colour removal of wastewater using WTR is majorly controlled by the pH of solution as discussed in earlier on effect of pH. Figure [8](#page-5-4) shows the spectral variation of influent and effluent colour at different time intervals. It is observed that peak obtained for colour removal was at 350 nm wavelength. Moreover, there was no signifcant peak change observed in the effluent. Thus, colour removal using sorption was signifcant and constant over complete visual spectrum. Further, it can be concluded from spectral variation that WTR can be used for colour removal as no colour was imparted by WTR.

Comparison of the diferent options

In this study, the use of WTR in their diferent forms was compared for the colour removal from a real textile dye wastewater. The results show that WTR has the potential to be used for colour removal. A comparison of colour removal by diferent processes shows that while coagulation/focculation process gave a maximum removal of 57%, sorption process showed a maximum colour removal of 36%. Even in column operation, which generally shows a greater removal could not produce complete removal of colour, and gave colour removal in the range of 60–70%. All the processes were controlled by the initial solution pH which showed that WTR was more efective for colour removal in acidic solutions. This indicates that WTR can be utilised as a primary treatment for colour removal from textile wastewater. Though the capacity of the WTR would be less compared to commonly available sorbents, its availability in large quantities at free of cost can be used advantageously. In this case regeneration or recycling of WTR used is not recommended.

One of the major concerns with the use of WTR for industrial application is the transportation cost of the material. If WTR has to be used in wet form for coagulation, due to high moisture content in WTR, transportation cost from production site will be high. Otherwise WTR produced has to be pretreated to reduce moisture content with sludge thickeners. Another issue with the use of WTR as a coagulant is that it would result in sludge generation and subsequent load on the drying beds. On the other hand, applicability of WTR in dry form for sorption process would be cheaper due to low transport cost of WTR.

Conclusions

Water treatment residuals (WTR), a by-product of water treatment plants, which are available in huge quantities were used in diferent forms. The use of WTR as a coagulant gave a maximum colour and COD removal of 55 and 35%, respectively, at a WTR dose of 10 g/L and an initial pH of 3.0. In batch sorption tests, maximum colour removal obtained was 36% at a WTR dose of 30 g/L. Initial pH had a signifcant efect on the colour removal. Continuous-fow column studies showed the potential of WTR for treatment of real textile dye wastewater. Up to 180 and 120 bed volume could be treated before exhaustion of the media at initial pH

values of 3.0 and 6.2, respectively, and the maximum colour removal obtained was 65%. Overall, the study showed the potential of WTR for treating real textile dye wastewater.

Compliance with ethical standards

Conflict of interest The authors declare that they have no confict of interest.

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