ORIGINAL ARTICLE



Spatial and inter-annual variability of proto- and metazooplankton during summer around the Kneiss Islands (Tunisia, Central Mediterranean Sea)

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Abstract

We studied the distribution of proto- and metazooplankton coupled with environmental factors in the coast area around Kneiss Islands (Central Mediterranean Sea). Zooplanktonic communities were sampled during summer 2009 and summer 2010 at three stations. Our results showed difference in suspended matter concentrations between summer 2009 (92.88 \pm 7.15 mg L^{-1}) and summer 2010 (47.37 ± 23.12 mg L^{-1}). Large variations in the N/P ratio were recorded (6.94–36.76) due to the direct influence of the variability in concentration of both the dissolved inorganic nitrogen and dissolved inorganic phosphate components of the ratio. Ciliates abundance peaked in summer 2009 and was 3 times more abundant than summer 2010. Ciliates community composition was dominated by loricate ciliates (75% of total ciliates) in summer 2009 and naked ciliates (56% of total ciliates) in summer 2010. Copepods were the most abundant metazooplankton present during the entire study period, comprising 30–96% of the total metazooplankton community. Small planktonic copepods reached important abundance, particularly oithonids, were found to largely dominate copepods community in both summer 2009 (Oithona nana, 45% of total copepods) and summer 2010 (Oithona similis, 22% of total copepods). The small planktonic species Paracalanus parvus (54% of total copepods) was abundant during summer 2010. The results also indicate that (1) ciliates abundance was very low, showing a possible predation by copepods and also by heterotrophic dinoflagellates, (2) copepods capable to complete a top-down control on phytoplankton and ciliates, with preference to ciliates more than diatoms of similar size and shape and (3) the resistance of loricate ciliates compared to naked ciliates may be explained by their capacity to escape grazing due to the existence of a protective lorica.

 $\textbf{Keywords} \;\; Kneiss \; Islands \cdot Nutrients \cdot Ciliates \cdot Zooplankton$

Introduction

The coast of Sfax is a part of the eastern Mediterranean sea and is located in south of Tunisia (Rekik et al. 2012). It is bounded by the Kneiss Islands in the northeast, which is composed of 4 little islands (El Bessila, El Hajar, El Laboua and El Gharbia) with very low landforms of calcareous sand stone. El Bessila is occupied by sebkha, chotts and, especially, maritime marshes (Gueddari and Oueslati 2002). Everywhere the soft material of the island's shores is the

victim of erosion. These islands are uninhabited (Mosbahi et al. 2015). The islands encourage the presence of many halophilous species, covering over 80% of the soil surface. These vast stretches are a favourite site for aquatic avifauna, basically migratory and wintering birds, which occupy all the niches in the various available water levels. Because it lies in pre-Saharan Tunisia, this site is a major stopover for migratory birds. There are a great many birds: 70% of Tunisia's birds winter in the Kneiss Islands, and there may be more than 100,000 of them (Hamza et al. 2015). Over 75% of the marine plant populations are marine phanerogams, including several thousand hectares of Cymodocea lawns (Mosbahi et al. 2016a). The islands are colonized by the seagrass Zostera noltei, protected species listed in the "IUCN Red List" of threatened species in Mediterranean Sea, as characterizing a diversified habitat requiring monitoring and protection (Mosbahi et al. 2016b). The originality of the



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site's appearance is a result of the existence of a very big sandy-silty estran, crossed by channels. The Kneiss Islands were declared as a "Nature Reserve" in 1993. These islands have been a SPAMI "Specially Protected Area of Mediterranean Importance" since 2001 because it is having one of the Mediterranean's biggest estran ecosystems with exceptional avifauna (Mosbahi et al. 2016b). The archipelago was stated as an "Important Bird Area" (IBA) in 2003 and a "RAMSAR Site" in 2007 (Mosbahi et al. 2016b). Yet, the Kneiss Islands environment is suffering from considerable pollution through hypertrophication, basically due to the consequences of the phosphate processing industry and the dumping of hydrocarbons (Rekik et al. 2012). The presence of hundreds of boatless fishermen in the islands, especially el Bessila, seriously harms the conservation of the environment, since it greatly disturbs the nesting birds or even prevents them making any attempt at nesting (M'Rabet 1995). Marine turtles present in the area are caught sporadically. For the entire marine coastal area, the scraping of the seabed is very likely to lessen the area's biodiversity (Galil 2000).

Mosbahi et al. (2015, 2016a, b) presented a research that investigated the spatial and seasonal structure of the intertidal macrozoobenthic communities of the area around the Kneiss Islands. However, data concerning the spatial and inter-annual distribution of ciliates and zooplankton assemblages coupled to environmental factors in the Kneiss Islands coast were scarce. Our objective was (1) to investigate the dynamic and diversity of the planktonic ciliates and metazooplankton through summer 2009 and summer 2010; (2) to identify the role of the main abiotic factors that determine the protozooplankton and metazooplankton communities structure and functioning; and (3) to study the spatial

distribution of ciliates and metazooplankton during summer in relation to its potential prey such as phytoplankton. To the best of our knowledge, there have been no previous comprehensive field studies that include all of the same parameters found in our study.

Materials and methods

Study site and sampling procedures

Nutrients, phytoplankton, ciliate and metazooplankton samples were collected in July 2009 and July 2010. Water samples were collected on 3 stations around the Kneiss Islands coast (Fig. 1). Seawater samples for physico-chemical analyses, phytoplankton and ciliate examination were collected from the surface water with a Van Dorn-type closing bottle at each station. Metazooplankton was collected using a cylindro-conical net (30 cm aperture, 100 cm high, and 100 µm mesh size). Nutriment samples (120 mL) were stored immediately in the dark at 20 °C. Phytoplankton samples (1 L) were preserved with Lugol iodine solution (4%) for enumeration (Bourrelly 1985). Metazooplankton samples were preserved in 2% buffered formaldehyde solution and were stained with rose Bengal to facilitate dissection. Plankton samples were kept at low temperature (4 °C) in the dark until analysis.

Physico-chemical variables

Physical parameters (temperature, salinity, and pH) were measured immediately after sampling using a

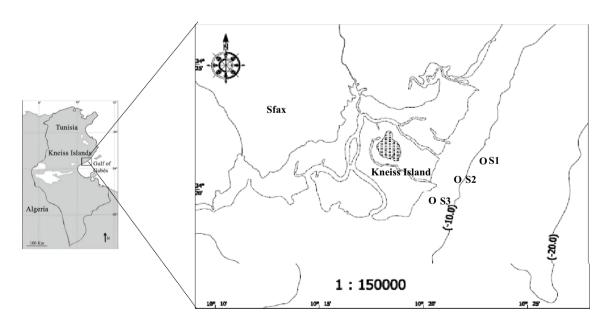


Fig. 1 Location of sampling stations on the Kneiss Islands



multi-parameter kit (Multi 340 i/SET). Suspended matter concentrations were measured using the dry weight of the residue after filtration of 0.5 L of seawater onto Whatman GF/C membrane filters. Chemical parameters (nitrite, nitrate, ammonium, orthophosphate, silicate, total nitrogen and total phosphate) were analysed with a Bran and Luebbe type 3 autoanalyzer.

Plankton enumeration

Phytoplankton and ciliates samples (50 mL) were analysed under an inverted microscope after 24–48 h settling using the Utermöhl method (Utermöhl 1958). Zooplankton enumeration was performed under a vertically mounted deepfocus dissecting microscope (Olympus TL 2). Plankton species identification was made according to various keys (Rekik et al. 2012, 2013a, b, 2015a, b, c).

Statistical analysis

The data recorded in this study were submitted to a normalized principal component analysis (PCA) using XLstat (Dolédec and Chessel 1989). Simple $\log (x+1)$ transformation was applied to data in order to correctly stabilize variance (Frontier 1973). Means and standard deviations (SD) were reported when appropriate. The potential relationships between variables were tested with Pearson's correlation coefficient. One-way ANOVA employing XLstat software followed by a post hoc evaluation using Tukey's analysis was used to recognize major variations among summer 2009 and summer 2010.

Results

Environmental parameters

The physical characteristics of the study are summarized in Table 1. The average of the water temperature ranged between 31 and 38 °C. The lowest temperature was detected at station 3 in summer 2009 and in stations 1 and 3 in summer 2010. The highest temperature was recorded at station 1 during summer 2009. The mean temperature values, which varied between 35.33 ± 3.79 °C and 31.33 ± 0.58 , showed similar variations in both summer 2009 and summer 2010 during the survey period. The salinity ranged from 37.5 to 39 psu recorded in stations 2 and 1 during summer 2010 and summer 2009, respectively (Table 1). The pH was relatively stable and distributed homogeneously throughout the monitoring stations varying from 8 to 8.4 observed in all sampling area (Table 1). The mean pH values were usually alkaline, suggesting a pronounced photosynthetic activity. The mean suspended matter concentrations showed an important difference between summer 2009 (92.88 \pm 7.15 mg L⁻¹) and summer 2010 (47.37 \pm 23.12 mg L⁻¹). The highest value was recorded in station 2 from summer 2009 and the lowest in summer 2010 (28.10 mg L⁻¹ in summer 2010) (Table 1).

The chemical variables analysed during this study are summarized in Table 1. Nitrate and nitrite concentrations did not show marked changes between summer 2009 and summer 2010. On the contrary, during summer 2010, ammonium concentration showed a significant increase in stations 1 (8.04 µM) (Table 1). Overall, the concentration of total nitrogen was not changed from summer 2009 to summer 2010 of the coastal area. Summer 2009 total nitrogen concentration ranged from 12.35 (station 2) to 15.20 µM (station 3). In summer 2010, total nitrogen concentration was normally, $\sim 14 \, \mu M$ throughout the study period (Table 1). The mean value of orthophosphate concentration was 0.29 µM, showed similar variations in both summer 2009 and 2010 (Table 1). When considering total phosphate concentrations, values were 8 times that of orthophosphate concentration in the summer 2009 and up to 21 times in the summer 2010 (Table 1). The N/P: DIN (DIN = $NO_2^- + NO_3^- + NH_4^+$) to DIP (DIP = PO_4^{3-}) ratio ranged from 6.94 in summer 2009 (station 3) to 36.76 in summer 2010 (station 1). However, the DIN/DIP ratio was generally higher than the Redfield ratio (16), with the exception of station 3 (6.94 in summer 2009 and 15.60 in summer 2010) (Table 1). These averages were more important than the Redfield ratio (16), suggesting a potential P limitation. The largest variation of silicate concentration $(4.95 \pm 2.48 - 12.05 \pm 1.24 \mu\text{M})$ was observed during summer 2009 and summer 2010, respectively (Table 1).

Ciliates, community diversity and dynamics

Ciliates abundance ranged from 2.00×10^2 to 15.00×10^2 cells L^{-1} (3.00 ± 1.00 × 10² cells L^{-1} in summer 2010, $9.33 \pm 4.93 \times 10^2$ cells L⁻¹ in summer 2009). Ciliates proliferated exclusively at station 3; the highest ciliates abundance $(15.00 \times 10^2 \text{ cells L}^{-1})$ was recorded during summer 2009 (Fig. 2). In terms of inter-annual abundance distribution, the majority of ciliates were found in summer 2009. Loricate ciliates summer 2009 abundance varied from 4 to 11×10^2 cells L⁻¹ recorded at stations 1 and 3, respectively. In summer 2010, the variation of loricate ciliates abundance followed a similar trend to that of summer 2009, with an average value of $1.33 \pm 1.53 \times 10^2$ cells L⁻¹ less important than summer 2009, $7.00 \pm 3.61 \times 10^2$ cells L⁻¹. Loricate ciliates abundance peaked at the same time, in summer 2009 $(11.00 \times 10^2 \text{ cells L}^{-1})$ and summer 2010 $(3.00 \times 10^2 \text{ cells})$ L^{-1}), at station 3 (Fig. 2). Loricate ciliates made the most abundant group during summer 2009 (75% of total ciliates), but, in summer 2010, the contribution of loricate ciliates is only 44% of total ciliates. Naked ciliates were observed at all stations, whereas they were poorly represented in terms of



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Table 1 Minimum, maximum and mean ± SD of physical-chemical and biological variables on the Kneiss Islands and results of ANOVA analysis to identify the significant differences between summer 2009 and summer 2010

Variables	Summer 2009			Summer 2010			Summer 2009/2010	
	Minimum	Maximum	Mean ± SD	Minimum	Maximum	Mean ± SD	F values	P
Physical variables								
T (°C)	31.00	38.00	35.33 ± 3.79	31.00	32.00	31.33 ± 0.58	3.27	0.14
S (p.s.u.)	38.50	39.00	38.73 ± 0.25	37.50	38.00	37.77 ± 0.25	22.13	$9.28 \times 10^{-3} **$
pH	8.15	8.40	8.28 ± 0.13	8.00	8.10	8.03 ± 0.06	9.59	0.04 *
Suspended matter (mg l ⁻¹)	84.66	97.66	92.88 ± 7.15	28.00	73.00	47.37 ± 23.12	10.61	0.03 *
Chemical variables								
NO_3^- (μM)	1.38	2.53	2.12 ± 0.64	2.42	2.77	2.65 ± 0.20	1.81	0.25
$NO_2^-(\mu M)$	0.22	1.03	0.59 ± 0.41	0.18	0.63	0.35 ± 0.24	0.74	0.44
$NH_4^+(\mu M)$	0.78	1.60	1.10 ± 0.44	1.82	8.04	4.02 ± 3.49	2.07	0.22
Total-N (µM)	12.35	15.20	13.54 ± 1.48	11.11	17.19	13.78 ± 3.11	0.01	0.91
$PO_4^{3-}(\mu M)$	0.16	0.48	0.29 ± 0.17	0.25	0.33	0.29 ± 0.04	4×10^{-3}	0.95
Total-P (µM)	0.98	3.38	2.31 ± 1.22	4.42	7.54	6.31 ± 1.66	11.32	0.03 *
N/P ratio	6.94	21.63	16.55 ± 8.33	15.60	36.76	23.95 ± 11.26	0.84	0.41
$Si(OH)_4 (\mu M)$	2.46	7.41	4.95 ± 2.48	11.11	13.45	12.05 ± 1.23	19.76	0.01 *
Biological variables								
Phytoplankton ($\times 10^2$ cells l ⁻¹)	11.00	33.00	20.00 ± 11.53	30.00	128.00	63.00 ± 56.29	1.68	0.26
Total ciliates ($\times 10^2$ cells l^{-1})	6.00	15.00	9.33 ± 4.93	2.00	4.00	3.00 ± 1.00	4.75	0.09
Loricate ciliates ($\times 10^2$ cells l^{-1})	4.00	11.00	7.00 ± 3.61	0.00	3.00	1.33 ± 1.53	6.82	0.07
Naked ciliates ($\times 10^2$ cells l ⁻¹)	1.00	4.00	2.33 ± 1.53	1.00	2.00	1.67 ± 0.58	0.50	0.52
Total metazooplankton ($\times 10^2$ ind m ⁻³)	21.50	123.35	66.75 ± 51.87	53.44	83.47	67.64 ± 15.08	8.20×10^{-4}	0.98
Copepods ($\times 10^2$ ind m ⁻³)	20.71	37.47	29.72 ± 8.45	43.28	71.93	56.89 ± 14.38	7.96	0.50 *
Copepods nauplii ($\times 10^2$ ind m ⁻³)	0.00	2.01	1.29 ± 1.12	5.46	8.74	7.40 ± 1.72	26.52	$6 \times 10^{-3} **$
Cyclopoids ($\times 10^2$ ind m ⁻³)	9.88	17.69	12.75 ± 4.30	6.83	23.52	17.12 ± 9.00	0.58	0.49
Calanoids ($\times 10^2$ ind m ⁻³)	2.30	18.47	10.02 ± 8.11	20.40	36.97	28.52 ± 8.29	7.63	0.05 *
Harpacticoids ($\times 10^2$ ind m ⁻³)	2.11	8.36	4.77 ± 3.23	1.83	10.30	4.72 ± 4.84	$2.21 \!\times\! 10^{-4}$	0.98
Poecilostomatoids ($\times 10^2$ ind m ⁻³)	0.00	1.89	0.88 ± 0.95	0.68	1.06	0.90 ± 0.20	1.27×10^{-3}	0.97
Other metazooplankton ($\times 10^2$ ind m ⁻³)	0.79	85.88	37.03 ± 43.92	10.16	11.54	10.75 ± 0.71	1.07	0.36

F values were determined by a one-way ANOVA test, P values for differences among seasons within each variable

abundance. In terms of inter-annual distribution, the majority of other ciliates were found in summer 2010 (56% of total ciliates abundance). Naked ciliates abundance varied from 10^2 to 4.00×10^2 cells L^{-1} ; the highest naked ciliates abundance was observed in station 3 during summer 2009. The inter-annual mean of naked ciliates abundance was $2.33 \pm 1.53 \times 10^2$ and $1.67 \pm 0.58 \times 10^2$ cells L^{-1} in summer 2009 and summer 2010, respectively (Fig. 2).

The ciliates community consisted of 15 taxa (12 species in summer 2009 and 5 species in summer 2010) belonging to 10 genera. Figure 3 reports the abundances of all the observed taxa during summer 2009 and summer 2010. Clear differences in ciliates diversity and composition were found from year to year and between stations. Naked ciliates were the most diverse ciliates group with 8 species (5 and 4 taxa

summer 2009 and summer 2010). The genus *Tintinnopsis* was dominant among ciliates (4 taxa), followed by *Euplotes* and *Strombidium* and (8 taxa) (Fig. 3). *Tintinnopsis beroidea* was the most important species and peaked in station 3 at both sampling periods (5.00×10² cells L⁻¹, summer 2009 and 3.00×10² cells L⁻¹, summer 2010). Among the species collected, *Favella ehrenbergi* and *Tintinnopsis* sp. were found with important abundance. There were 10 species specific of summer 2009 (*Favella ehrenbergi*, *Helicostomella subulata*, *Tintinnopsis aperta*, *Tintinnopsis radix*, *Tintinnopsis* sp., *Undella* sp., *Colpoda* sp., *Euplote* sp., *Leegardiella sol* and *Lohmanniella oviformis*) and 3 species for summer 2010 (*Aspidisca lynceus*, *Strombidium capitatum* and *Strombidium conicoides*). Some ciliates taxa were omnipresent at all study, and among these ciliates, *Euplote charon* (Fig. 3).



^{*}P<0.05; **P<0.01

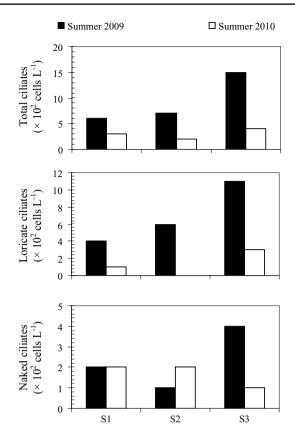
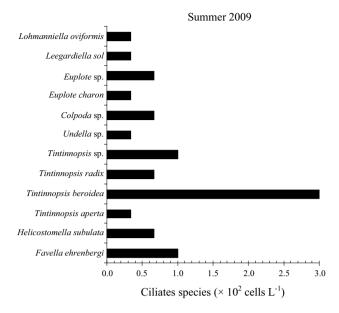


Fig. 2 Spatial and inter-annual variation of total ciliates and different ciliates groups abundance at sampled stations

Zooplankton, community diversity and dynamics

The total metazooplankton abundance varied from 21.50×10^2 (station 3, Summer 2009) to 123.35×10^2 ind m⁻³ (station 2, summer 2009), the highest mean abundance being recorded in summer 2010 $(67.64 \pm 15.08 \times 10^2)$ ind m⁻³) (Fig. 4). Metazooplankton assemblages along the coast of the Kneiss Islands were dominated by copepods accounting for 45 and 84% of the total metazooplankton abundance during summer 2009 and 2010, respectively. The summer 2009 distribution of copepods $(29.72 \pm 8.45 \times 10^2)$ ind m⁻³) illustrated a high cyclopoids density (17.69×10^2) ind m⁻³) at station 1 and was associated with Oithonidae aggregations. Abundances of copepods, copepods nauplii, cyclopoids and calanoids peaked at the same time, in summer 2010 at station 1. High abundances of calanoids were recorded during summer 2010. Calanoids clearly dominated the copepods during summer 2010, accounting for 50% of total copepods abundances. Low numbers of copepods nauplii were observed, ranging between 0 and 8.74×10^2 ind m⁻³ (1.29 ± 1.12 × 10² ind m⁻³, summer 2009 and $7.40 \pm 1.72 \times 10^2$ ind m⁻³, summer 2010). The interannual distribution of Harpacticoids and Poecilostomatoids was slightly greater in summer 2009. Harpacticoids



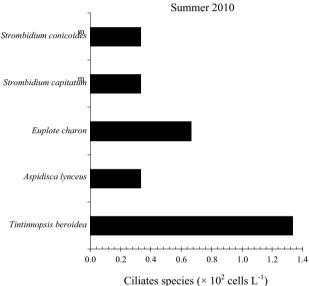


Fig. 3 Inter-annual variation of ciliates species abundance at sampled stations

varied from 1.83 to 10.30×10^2 ind m⁻³ $(4.77 \pm 3.23 \times 10^2)$ ind m⁻³ in summer 2009 and $4.72 \pm 4.84 \times 10^2$ ind m⁻³ in summer 2010) and Poecilostomatoids from 0 to 1.89×10^2 ind m⁻³ $(0.88 \pm 0.95 \times 10^2)$ ind m⁻³ in summer 2009 and $0.90 \pm 0.20 \times 10^2$ ind m⁻³ in summer 2010) (Fig. 4), following the copepods community throughout the study period, reaching only low abundances. Other metazooplankton clearly dominated the metazooplankton community during summer 2009, accounting for 55% of total metazooplankton abundances. Other metazooplankton abundance varied from 0.79×10^2 (station 3) to 85.88×10^2 ind m⁻³ (station 2), the highest mean abundance being recorded in summer 2009 $(37.03 \pm 43.92 \times 10^2)$ ind m⁻³ (Fig. 4).



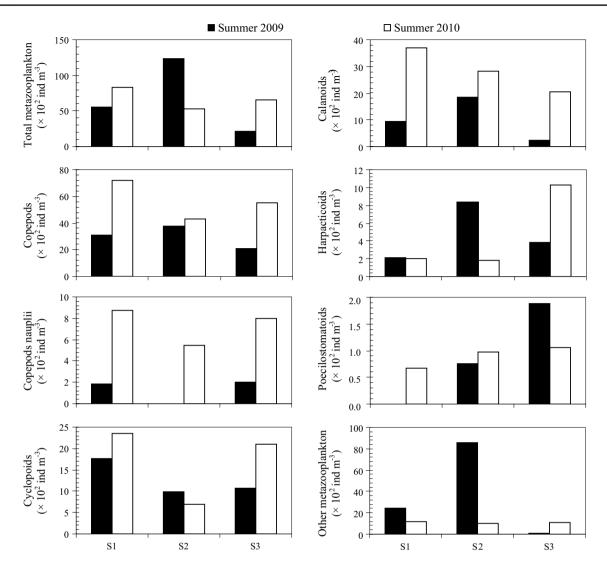


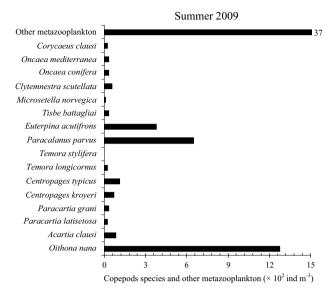
Fig. 4 Spatial and inter-annual variation of total metazooplankton and different metazooplankton groups abundance at sampled stations

A summary of the copepods species and other metazooplankton groups observed throughout the study period is given in Fig. 5. Metazooplankton composition was significantly different among the 2 years and stations. A total of 14 copepods families were found at all stations, with Oithonidae dominating the copepods community, among which Oithona nana and Oithona similis were the most abundant species, representing 45 and 22% of the total copepods presence in summer 2009 and summer 2010, respectively. Copepods richness was more pronounced in summer 2009 (16 species) than in summer 2010 (13 species) (Fig. 5). Copepods consisted particularly of Euterpina acutifrons (station 2 in summer 2009 and station 3 in summer 2010), Centropages typicus (station 1, summer 2009) and *Paracalanus parvus* (station 1, summer 2010) (Fig. 5).

Statistical analysis

Principal component analysis (PCA) allowed discrimination of four groups around the components of the F1 and F2 axes (Fig. 6), explaining 100% of the variance. The F1 axis, explaining 65.10% of the variability of abiotic and biotic parameters, positively selected group G1 composed of NO₃⁻, NH₄⁺, N/P ratio, total phytoplankton, total zooplankton, copepods, other zooplankton, calanoids and harpacticoids. The F2 axis, representing 34.90% of the variability, positively selected group G2 comprising physical factors such as temperature and salinity correlated with cyclopoids abundance. G3 comprised T-N, copepods nauplii and other ciliates. The group G4 was formed by pH, suspended matter, NO₂⁻, PO₄³⁻, T-P, Si(OH)₄, total





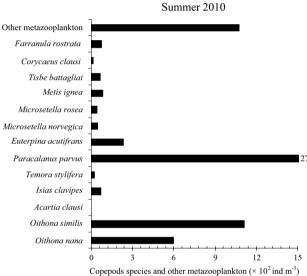


Fig. 5 Inter-annual variation of copepods species and other metazooplankton abundance at sampled stations

ciliates, tintinnids and poecilostomatoids. This combination was selected in summer 2009 (Fig. 6).

During summer 2010, the PCA distinguished between four groups surrounding the F1 and F2 component axes thus explaining 100% of the variance. The axes selected a group G1 comprising the biological parameters (total phytoplankton, total zooplankton, copepods, other zooplankton, copepods nauplii, cyclopoids and calanoids) and several physico-chemical variables (pH, NH₄⁺, T-P and N/P ratio). F1 component axis, which extracted 61.05% of the variability, selected positively the group G2, with total ciliates, tintinnids and harpacticoids correlating to abiotic parameters (salinity, NO₃⁻, PO₄³⁻ and Si (OH)₄). The gradient along the F1 axis is mainly due to suspended matter and poecilostomatoids (selected negatively in G3)

and to temperature, NO₂⁻, T-N and other ciliates (selected negatively in G4) (Fig. 6).

Discussion

The present study is the first examining the spatial distribution of ciliates and zooplankton communities in the shallow coastal waters around Kneiss Islands coupling to nutrients and phytoplankton abundance during summer 2009 and summer 2010. Our results demonstrated distinct inter-annual contrasts. Ciliates and zooplankton assemblages in coastal ecosystems are controlled by a combination of environmental and biological factors.

During this study, the high values of temperature and salinity were typical of arid to semiarid zones (Elloumi et al. 2015). Suspended matter concentrations recorded in summer $2009 (92.88 \pm 7.15 \text{ mg L}^{-1})$ were higher than those founded in summer 2010 (47.37 \pm 23.12 mg L⁻¹). The striking difference of the suspended matter concentrations levels can logically be attributed to the shallowness of the sampled stations and the spatial distribution of polluting elements and that may result in resuspension of tiny particles with, for instance, tidal action (Ben Salem et al. 2015). Large variations in the N/P ratio were recorded (6.94–36.76) due to the direct influence of the variability in concentration of both the dissolved inorganic nitrogen and dissolved inorganic phosphate components of the ratio. The average value of the N/P ratio was generally higher than the Redfield (16) ratio in both summer 2009 and summer 2010. However, large variations in the N/P ratio were observed due to the direct influence of the variability in concentration of both the N and P components of the ratio. These results may be explained by phytoplankton's rapid consumption (Rekik et al. 2016), and the importance of nitrogen availability may be caused by atmospheric deposition (Liu et al. 2015).

In this study, there was an important difference in ciliates abundance between years; on the other hand, the difference influences the ciliates composition and distribution. For example, in terms of annual scale, ciliates abundance peaked in summer 2009 and was 3 times more abundant than summer 2010. With regard to the structural composition, the ciliates community at Kneiss Islands coast was dominated by loricate ciliates (75% of total ciliates) in summer 2009 and naked ciliates (56% of total ciliates) in summer 2010. At the coastal waters around Kneiss Islands, loricate ciliates demonstrated a clear inter-annual pattern: high values in summer 2009 and low values in summer 2010 with an obvious peak in station 3 throughout the study period. Loricate ciliates counted for a large percentage (0-73%) of the total ciliates. Their density seemed to be associated to their morphology that they can escape grazing due to the existence of a protective lorica (Abboud-Abi Saab 2008), their tolerance



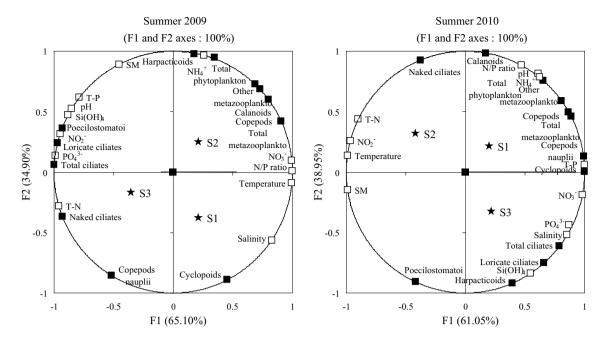


Fig. 6 Principal component analysis (Axis I and II) of biological parameter abundance and selected environmental variables at sampled stations in the Kneiss Islands

of marine habitats stress and high flexibility to eutrophic conditions. With regard to the structural composition, loricate ciliates population at the coast of Kneiss Islands was dominated by the genus Tintinnopsis. This genus has usually been reported to be the dominant planktonic ciliates in both temperate and tropical coastal systems (Dolan 2006). The most abundant species was Tintinnopsis beroidea; this species is relatively common in coastal ecosystems and in particular in the north coast (Rekik et al. 2015b) and south coast (Rekik et al. 2015c) of Sfax. The dominance of the agglutinated species T. beroidea appears to be related to the availability of particles to construct the lorica (Rakshit et al. 2015). This result suggests that this species performed better than other taxa probably owing to more flexible adaptive strategies, to particular environmental factors (temperature, salinity) and the shallowness of the water sampling (Kchaou et al. 2009).

Small planktonic copepods reached important abundance throughout the study period. The advantage of the small taxa bearing egg sacs is to keep eggs against predators until they hatch and make nauplii (Ben Ltaief et al. 2015). Small copepods, particularly oithonids, were found to largely dominate copepods community in both summer 2009 (*Oithona nana*, 45% of total copepods) and summer 2010 (*Oithona similis*, 22% of total copepods). This appears to be a common feature in the north coast (Rekik et al. 2012) and south coast (Ben Salem al. 2015) of Sfax, in the Gulf of Gabes (Ben Ltaief et al. 2015) and in the north of Tunisia (Daly Yahia et al. 2004). Other studies reported a very important contribution

of small copepods over the Balearic Sea (Fernandez de Puelles et al. 2003), the Adriatic Sea (Vidjak et al. 2007) and the Aegean Sea (Zervoudaki et al. 2007). The dominance of small species, particularly the genus Oithona, in salty and warm coastal waters has also been reported in the bay of Tunis (Daly Yahia et al. 2004) and in the Lagoon of Tunis (Annabi-Trabelsi et al. 2006) (north of Tunisia). The dominance of *Oithona nana* in various marine ecosystems was inferred owing to its adaptive strategies (Riccardi and Mariotto 2000) allowing it to tolerate environmental perturbations (Gallienne and Robins 2001). Oithona similis has been recognized as the most ubiquitous copepod species in the world ocean (Gallienne and Robins 2001). This small species revealed clear eurythermal and euryhaline distribution (Bel Hassen et al. 2008). The Kneiss Islands coast was also distinguished by the prevalence of the small planktonic species Paracalanus parvus $(27.52 \times 10^2 \text{ ind m}^{-3})$ in summer 2010. This species was known by omnivorous diet (Turner 1991).

During this study, the ciliates abundance was very low, showing a possible predation by copepods and also by heterotrophic dinoflagellates (Jeong et al. 2010). Copepods represent an important metazooplankton group capable to complete a top-down control on phytoplankton (r=0.91, n=24, p=0.05) and ciliates (r=0.68, n=24, p=0.05) communities (Zervoudaki et al. 2007). Similar results reported the strong top-down control of copepods in regulating phytoplankton and ciliates populations in the eastern Mediterranean Sea (Siokou-Frangou et al. 2002). Moreover, the resistance of



loricate ciliates compared to naked ciliates may be explained by their capacity to escape grazing due to the existence of a protective lorica (Abboud-Abi Saab 2008). *Oithona nana* is a species known to be omnivorous (Lampitt and Gamble 1982), opportunistic (Lam-Hoai and Rougier 2001) and capable of feeding on small prey. *Oithona similis* is omnivorous feeding on heterotrophic dinoflagellates, ciliates and naupliar stages of copepods (Turner 2004). Indeed, this copepod has preference to ciliates more than diatoms of similar size and shape (Castellani et al. 2008).

Conclusion

The proto- and metazooplankton abundance showed clear variations along the coastal stations during summer cruises conducted in years 2009 and 2010. This study indicates that ciliates were largely dominated by loricate ciliates in summer 2009 and naked ciliates in summer 2010, while copepods were the most numerous metazooplankton, including Oithona nana and Oithona similis which was spread abundantly along the coast during summer 2009 and summer 2010, respectively. The trophic interplay between phytoplankton, ciliates and copepods suggests that copepods were capable to complete a top-down control on phytoplankton and ciliates communities. On the other hand, trophic interplay between phytoplankton, ciliates and copepods suggests that abiotic parameters and nutrients were also involved in the environmental forcing of the summer proto- and metazooplankton dynamics around Kneiss Islands.

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Compliance with ethical standards

Conflict of interest We declare that we have no conflict of interest.

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