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Hydrobiogeochemical evolution along the regional groundwater fow systems in volcanic aquifers in Kumamoto, Japan

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Abstract

Kumamoto is known as the largest groundwater city in Japan. Geochemical modeling (saturation indices and mineral stability diagram) was applied in this area for better understanding a hydrogeochemical evolution in volcanic aquifers in regional scale with additional constraint from stable isotopic dataset. Geochemical evolutionary model was interpreted along the water flow dynamics. In total 136 water samples were collected from wells, springs, and rivers in and around the major groundwater fow lines for geochemical analyses. Our results indicated that plagioclase is the major weathering reactant in aquifers with secondary important weathering minerals of pyroxenes. These reactions facilitate current hydrochemical signatures and produce secondary minerals of kaolinite or halloysite and smectite in later stage. Observed hydrogeochemical processes can be distinguished into two distinct criteria along fow regime. The frst processes are material loads from the surface (mixing of contaminants and river waters) and initial stage of silicate weathering, resulting in the formation of Ca- $HCO₃$, Ca-NO₃–HCO₃, and Ca–Mg–Na-HCO₃ type waters. These processes prevail in aquifers at the recharge to lateral flow zones with relatively shorter residence time of $< ca$. 40 years. The second processes are the precipitation of clay minerals, i.e., smectite, and cation exchange reaction of Ca^{2+} and Mg^{2+} to Na⁺ in downslope aquifers (Na-HCO₃ type) with relatively longer groundwater residence time of > 55 years. Microbiological reduction reactions dominate over these areas and salinization occurs at the coast that changes aquifer waters to be Na-Cl type. Proposed models and approach shown in this study may be useful and applicable in interpreting systems in other volcanic aquifer systems at similar climate conditions and for sustainable water resources management.

Keywords Groundwater · Volcanic aquifer · Saturation indices · Silicate weathering · Cation exchange · Flow dynamics

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Introduction

Kumamoto is the largest groundwater user area in Japan (Fig. [1\)](#page-1-0). About one million people in and around the area completely depend on groundwater for their all purposes (e.g., Taniguchi et al. [2019\)](#page-18-0). The area comprises Kumamoto City and its surrounding districts and towns, extended to the western fank of Aso caldera mountains in central Kyushu (Fig. [1\)](#page-1-0), which is one of the largest volcanic calderas in Japan. The abundant groundwater resources in this area are a blessing of combination of both high local precipitations and highly permeable surface geological structures that comprise of Quaternary pyroclastic fow deposits of the Aso volcanic eruptions. Since the people in Kumamoto relies on groundwater for drinking purpose, its quality is one of major concerns to the local citizens and policy makers in a management point of view. However, the groundwater quality is becoming a serious issue in this area due to some

Fig. 1 Location of the study area showing major geographic features, groundwater region and sampling locations along major groundwater fow lines A–A′ and B–B′, which hydrogeologic sections are shown in Fig. [3](#page-3-0). Division of fow zones are shown along each arrow. Solid contour lines represent groundwater potential for high water season

during September and October (m, above sea level, after Kagabu et al. [2017](#page-17-13)). Green and yellow squares represent sampling location of spring waters from Aso caldera watershed and Kumamoto area, respectively. Yellow and white stars indicate sampling location of river waters from Shira and Midori Rivers, respectively

geogenic (arsenic and fuoride) (Hossain et al. [2016a,](#page-17-0) [b\)](#page-17-1) and anthropogenic (nitrate and volatile organic compounds) contaminations (Kumamoto Prefecture and Kumamto City [2005;](#page-17-2) Kumamoto City Waterworks and Sewerage Bureau [2008](#page-17-3); Tomiie et al. [2009;](#page-18-1) Hosono et al. [2013,](#page-17-4) [2014;](#page-17-5) Zeng et al. [2016](#page-18-2); Okumura et al. [2018](#page-18-3)).

In general, chemical evolution of groundwater and thus its quality is greatly controlled by the geochemical weathering processes (Hem [1989;](#page-17-6) Nordstrom et al. [1989](#page-18-4); Drever [1997](#page-17-7)). Improved knowledge on these processes could help to understand a whole hydrochemical system of an area and for efective management and proper utilization of the water resources (Fisher and Mullican [1997;](#page-17-8) Cloutier et al. [2006](#page-16-0)). However, such attempt has not yet been made in the Kumamoto area although flow dynamics of groundwaters and their contamination processes are well studied. Therefore, the primary objective of this study is to depict these principle geochemical processes that control the chemical evolution of groundwaters along their flow systems. After a pioneering work of Garrels and Mackenzie ([1967\)](#page-17-9), a good number of researchers explained geochemical evolution of aquifer waters using major ion chemistry and mineral weathering reactions with mass balance approach in diferent hydrogeological systems (White et al. [1980](#page-18-5), [1998](#page-18-6); Wood and Low [1986](#page-18-7); Thomas et al. [1996](#page-18-8) and references therein). However, studies for carbonate-free, silicate dominated volcanic rock aquifers are still in limited numbers (e.g., Join et al. [1997;](#page-17-10) Nesbitt and Wilson [1992](#page-18-9); Koh et al. [2016\)](#page-17-11). Moreover, in most of the previous studies detailed fow dynamics were unclear in regional scale (Kenoyer and Bowser [1992](#page-17-12)). Due to the ambiguity in groundwater fow delineation, it was somehow difficult to observe geochemical evolutional characteristics precisely.

Kumamoto basin is volcanic terrains composed of pyroclastic fow deposits with coastal vicinity. The recharge and discharge areas of groundwaters with their flow paths are well defned based on hydrogeological assessments (e.g., Taniguchi et al. [2003;](#page-18-10) Ono et al. [2013;](#page-18-11) Kagabu et al. [2017](#page-17-13); Okumura et al. [2018\)](#page-18-3) and their residence times are well characterized using 85 Kr groundwater age tracer (Kagabu et al. [2017](#page-17-13)). In addition, the efects of material loads from recharge areas and anion reduction processes due to microbiological succession on groundwater chemistry are well documented along fow paths using multiple isotope ratios $(\delta^{13}C, \delta^{15}N$ and $\delta^{34}S)$ (Hosono et al. [2013,](#page-17-4) [2014\)](#page-17-5), trace elements (Hossain et al. [2016a,](#page-17-0) [b\)](#page-17-1), and PCR analysis (Zeng

et al. [2016](#page-18-2)), which information allow us to depict a combined hydro-bio-geo-chemical evolution of groundwaters from recharge areas to the coast in regional scale. This particular integrated study is associated with systematic use of the major ion chemistry, stable isotope ratios, and geochemical modeling with available mineralogical data to construct a model showing hydro-bio-geochemical evolution in volcanic aquifer systems in a context of regional groundwater fow systems with their known residence time.

Study area

Topography, climate, geology and mineralogy

Kumamoto area is located in the central part of Kyushu Island in southern Japan with the longitude between 130°30′ and 131°05′ E and the latitude between 32°40′ and 33°00′ N (Fig. [1\)](#page-1-0). The area is consisting of the western mountain feet of the Aso caldera mountains, surrounding plateaus and hilly mountains (expressed as 'highlands'), and Kumamoto plain. The study area, termed as 'Kumamoto groundwater region' (Hosono et al. 2013) occupies 945 km² in an area, lying between the Shira River watershed to the north and Midori River watershed to the south, the mountain foot of western caldera somma to the east and Kinpo Mountain (665 m) and Ariake Sea to the west (Fig. [1\)](#page-1-0). The topography generally inclines from north-east to the south-west. Gradients are drastically changed near the Ezu Lake (Fig. [1](#page-1-0)), which

is spring fed lake as described later. Three major highlands were formed during the Aso volcanic activity with altitude around 100–200 m. Among them Kikuchi and Takayubaru highlands are situated northeastern and eastern side of the study area, respectively, whereas Ueki highland is in the north. Kumamoto area has humid, subtropical climate. Average annual temperature and precipitation of Kumamoto are 16.9 °C and 1986 mm during 1981–2010, respectively [\(https](https://www.jma.go.jp/jma/menu/report.html) [://www.jma.go.jp/jma/menu/report.html\)](https://www.jma.go.jp/jma/menu/report.html). Around 40% of total precipitation occurs during summer rainy season in June–July.

Land-use in the study area is shown in Hosono et al. [\(2013](#page-17-4)). Briefy, the western mountain feet of the Aso caldera mountains are covered by forest and grassland. Vegetable and livestock farming are the main land-use in the Ueki, Kukuchi, and Takayubaru highlands. On the contrary, paddy is widely distributed at lowland areas along rivers and plains.

In the Kumamoto groundwater region, the Aso volcanic rocks of Quaternary age and Holocene alluvium deposits overlie hydrogeological basement of the Paleozoic to Mesozoic metamorphic rocks and Tertiary to Quaternary Pre-Aso volcanic rocks (Figs. [2](#page-2-0), [3](#page-3-0)). These metamorphic rocks are exposed only in the southeastern and northwest-western parts of the Kumamoto area (Fig. [2](#page-2-0)). The Pre-Aso volcanic rocks, which are composed of intermediate to acidic lavas and tuff breccia, present around the Mt. Kinpo, Mt. Tatsuta and Aso caldera somma (Watanabe [1978](#page-18-12)). The Aso volcanic rocks that form the main body of the Quaternary deposits in Kumamoto widely present between the Kinpo Mountain

Fig. 2 Surface geology map of the study area after Kumamoto Prefecture Geological Map Compilation Committee Report ([2008\)](#page-17-14). Dotted line in the fgure shows elevation line of ca. 5 m (above sea level). Ariake clay sediments are expected to be distribute at elevation lower than ca. 5 m (see text for details)

Fig. 3 Simplifed hydrogeological cross sections along A–A′ and B-B' flow lines shown in Fig. [1.](#page-1-0) Arrows in the aquifers represent the groundwater fow direction and the size of arrows corresponds to flow velocity calculated by groundwater fow simulation (after Hosono et al. [2019](#page-17-21))

and western fank of Aso caldera mountains (Fig. [2\)](#page-2-0). Aso volcanic rocks are classifed into 4 units by 4 gigantic eruption cycles: Aso-1 (eruption volume $>$ 30 km³, eruptive age 370–270 ka), Aso-2 ($>$ 25 km³, 170 ka), Aso-3 ($>$ 40 km^3 , 100 ka) and Aso-4 (> 80 km³, 80 ka) (Aramaki [1984\)](#page-16-1) (Fig. [3\)](#page-3-0). There were many minor eruptions occurring in between these large eruption cycles (Ono et al. [1977;](#page-18-13) Matsumoto et al. [1991](#page-17-15)). Detailed geology and eruption history in and around caldera are summarized, e.g., in Ono et al. ([1977](#page-18-13)), Kamata ([1997\)](#page-17-16), Watanabe ([1978\)](#page-18-12), and Ono and Watanabe ([1985](#page-18-14)). The Aso volcanic products are generally permeable due to its porous structures. On the other hand, impermeable clay layers of lacustrine sediments are widely distributed in between Aso-3 and Aso-4 units (Kumamoto Prefecture and Kumamoto City [1995\)](#page-17-17) (Fig. [3\)](#page-3-0). In addition, a thick clay layer of marine origin of the last regression time, Ariake clay layer is deposited over the Kumamoto plain area, overlying the Aso-4 pyroclastic fow deposits at elevation generally lower than 5 m in the west of Ezu Lake in

the coastal side (Figs. [2,](#page-2-0) [3\)](#page-3-0). Beside this, fuvial alluvium sediments are distributed along the rivers at elevation higher than 5 m.

Mineralogy of the Kumamoto area has been studied by various researchers (Watanabe [1978](#page-18-12), [1979;](#page-18-15) Ono and Watanabe [1985;](#page-18-14) Hunter [1998](#page-17-18); Kaneko et al. [2007;](#page-17-19) Miyoshi et al. [2009](#page-18-16)). In general, plagioclase, clinopyroxene, orthopyroxene, and titanomagnetite are the major phenocrysts in Aso volcanic rocks (Hunter [1998\)](#page-17-18). Dominant secondary clay minerals are kaolinite, illite and smectite with some other alteration products. Detailed information for mineral abundances, compositions and major rock type for each unit of the Aso volcanic rocks are provided in Table S1. In general, the Aso volcanic rocks mainly consist of pyroclastic fall and flow deposits. Aso-1 to -3 consist of a range of welded to non-welded basaltic to deictic pyroclastic fow materials including lava flows, while Aso-4 consist of airfall, ignimbritic deposits, non-welded to welded dacitic ash and pumice, grading up to basaltic scoria (Lipman [1967](#page-17-20);

Watanabe [1979;](#page-18-15) Hunter [1998](#page-17-18)) and contain large amounts of lithic fragments (Ono et al. [1977](#page-18-13)). Petrographic analyses of Pre-Aso volcanic rocks have found that most common rock type is pyroxene andesite, tracked by subordinate hornblende andesite, biotite rhyolite, and basaltic to dacitic pyroclastic materials (Hunter [1998;](#page-17-18) Watanabe [1978,](#page-18-12) [1979\)](#page-18-15).

Hydrogeology and groundwater fows

According to stratigraphy and subsurface geology of Kumamoto area, groundwater mainly occurs in the aquifers that consist of Aso pyroclastic fow and alluvial sedimentary deposits (Fig. [3](#page-3-0)). Two main aquifers are delineated in this area based on the lithologic properties and hydrodynamic conditions: frst (shallow) and second (deep) aquifers from top to the bottom (Kumamoto prefecture and Kumamoto City [1995\)](#page-17-17). In general, depth of the first aquifer is ca. \sim 90 m from the surface whereas second aquifer is situated at depth of ca. 20–250 m (Kumamoto Prefecture and Kumamoto City [1995](#page-17-17)). The frst aquifer is mainly composed of Aso-4 pyroclastic fow and alluvium sedimentary deposits (Figs. [1](#page-1-0), [2,](#page-2-0) [3](#page-3-0)). On the other hand, Aso-1 to -3 pyroclastic fow deposits and volcanic fow lavas consist of the second aquifer. An impermeable clay layer called 'Futa' and 'Hanabusa' layers of lacustrine sediments is distributed between the frst and second aquifers (Fig. [3](#page-3-0)). However, this layer is discontinuous especially under the mid-stream area of the Shira River and partially absent around Ezu Lake area (Figs. [1,](#page-1-0) [2,](#page-2-0) [3\)](#page-3-0) (Aramaki [2002](#page-16-2)). Hence, frst aquifer is unconfned aquifer in nature whereas second aquifer can be considered as semi confned to confned aquifers.

With hydrogeological point of view, the groundwater in Kumamoto is thought to be largely recharged through infltration of precipitation at highlands surrounding the plain areas due to the presence of highly permeable pyroclastic flow of Aso-4 (Kumamoto Prefecture and Kumamoto City [1995](#page-17-17); Taniguchi et al. [2003](#page-18-10); Okumura et al. [2018](#page-18-3)). It is also confrmed the water can be recharged by surface waters through both natural percolation and artifcial infltration by rice paddy ponding around mid-stream area of the Shira River (Figs. [1,](#page-1-0) [3\)](#page-3-0). It is suggested in this area that surface water can directly be transported into the second aquifer as there is no aquiclude present and soil infltration capacity is very high \sim 500 mm/day (Tanaka et al. [2010\)](#page-18-17). Thus the river loses waters particularly at this area. Presence of the Togawa lava, highly porous andesitic lava of Aso-2, makes the confned aquifer more conductive (Fig. [3\)](#page-3-0). At the end of this lava distribution, spring waters emerged at Ezu Lake with the discharge rate of $1.5-1.8 \times 10^8$ m³/year (Ono et al. [2013\)](#page-18-11). Waters from both aquifers are discharging in this area (Ono et al. [2011](#page-18-18)). Other than Ezu Lake, spring waters from mountain aquifers discharge through more than 30 major spring sites at mountain feet (Fig. [1](#page-1-0)) with the discharge

rate ranging from 1.0×10^3 to 1.6×10^7 m³/year. Estimated hydraulic conductivities are 1.0×10^{-5} to 1.0×10^{-2} m/s and 1.0×10^{-5} to 2.5×10^{-1} m/s for the first and second aquifers, respectively (Mori et al. [2016](#page-18-19); Hosono et al. [2019\)](#page-17-21).

There are two major geographically distinct groundwater flow systems prevailing in this area (Hosono et al. [2013,](#page-17-4) [2014](#page-17-5) and see groundwater fow simulation results shown in Fig. S1). First one is A–A' flow that is the most significant flow system from the perspective of groundwater volume as well as city water supply (Figs. [1](#page-1-0), [3\)](#page-3-0). This flow line can be divided into three diferent zones: recharge and mixing zones $(A-1)$, lateral flow to discharge zone $(A-2)$ and stagnant zone (A-3) (Hosono et al. [2013](#page-17-4), [2014\)](#page-17-5). The water is recharged by precipitation mainly through Kikuchi and Takayubaru highlands (Taniguchi et al. [2003;](#page-18-10) Kagabu et al. [2017](#page-17-13); Okumura et al. [2018](#page-18-3)) and also through surface water to the second aquifer in the mid-stream area of the Shira River as mentioned previously. A groundwater residence time was determined by applying ⁸⁵Kr groundwater age tracer based on piston fow model and exponential piston fow model for specifc samples (Kagabu et al. [2017](#page-17-13)). According to this report, its residence time in A-1 zone was identifed as younger than 20 years. The flow path A-2 is the most major flow line over the studied region. Recharged water from A-1 zone is laterally fowing through highly porous Togawa lava formation towards the discharge zone (Ono et al. [2013\)](#page-18-11) with mean flow velocity of ~40 m/day (Kumamoto Prefecture and Kumamoto City [1995\)](#page-17-17). The 85 Kr age of this groundwater at the end of A-2 is estimated as ca. 40 years (Kagabu et al. [2017](#page-17-13)). Remaining fow towards coastal plain presents in A-3 zone. This groundwater is almost stagnant with its age determined as>55 years (Kagabu et al. [2017](#page-17-13)).

The B–B' flow line is geographically separated from the major flow line A–A' and can be divided into three zones (Figs. [1](#page-1-0), [3\)](#page-3-0). The size of groundwater systems in this fow line is volumetrically minor compared to that in A–A′ (Fig. S1) but it's also important for water supply in the vicinity areas. The waters are recharged by precipitation through Ueki highland in the recharge zone (B-1). Waters in this zone generally show younger groundwater age (>20 years; Kagabu et al. [2017](#page-17-13)) and this feature is similar to that found in A-1 zone. Then waters pass through basement uplifting area towards the entrance of Kumamoto plain, which area we defned as B-2 zone. The groundwater residence time in this zone show relatively old $(> 55$ years), due to attenuation in groundwater fow velocity (Kagabu et al. [2017](#page-17-13)). B-3 is south-westward flow to the coast. Most of groundwaters in A–A′ line emerge at Ezu Lake; those fows of B–B′ line extend towards coast (Fig. S1). In both A-3 and B-3, i.e., the stagnant zones, groundwater residence time is estimated $as > 55$ years (Kagabu et al. [2017](#page-17-13)).

As mentioned before the groundwaters can be mixed with surface water in mid-stream of the Shira River where downward fuxes predominate. In addition, mixing of waters from other sources may be occurred in midsection of the major groundwater fow lines including contribution of mountain aquifer waters and water percolations from the frst aquifer to deeper second aquifer at where aquicludes between these two aquifers are not distributed. However, in the lateral fow and stagnant zones active groundwater interactions are not expected across two diferent aquifers. Rather, groundwaters in these areas flow along major flow lines without signifcant additional water mixing in general that are controlled by regional geographical and geological conditions. Groundwater geochemical evolutions are assessed qualitatively along these well known flow structures.

Methods

Sampling

Groundwater samples were collected from 95 wells along the two major flow lines A–A' and B–B', both for the first (well depths 4–80 m) and second aquifers (well depths 44–211 m) (Fig. [1\)](#page-1-0). 58 samples were collected from the second aquifer while 37 samples were collected from the frst aquifer. From the flow line point of view, 64 wells are distributed along the A–A′ fow line and 31 wells are distributed along the B-B' flow line. Most of the wells are observation wells governed and managed by Ministry of Land, Infrastructure and Transport of Japan, Kumamoto Prefectural Government and Kumamoto City. Because of city water supply, most of the observation wells are located along A–A′ fow line and B–B' flow line has limited number of wells. Besides these observation wells, 35 samples were collected from private wells to complement geographical sampling distributions which are used for domestic purposes and owned by local peoples. Most observation wells have single screen at levels 5–10 m higher than the well depth with its length of ca. 5–10 m, but some wells have 2 or 3 screens in the diferent depths. Private wells for the second aquifer follow the same well structures but those for the frst aquifer are dug wells without screens.

Samples from private wells were collected during July 2011 while all other samples were collected during October–November 2011. Beside these, a total of 30 spring and 11 river samples were also collected (Fig. [1](#page-1-0)) at around the study area during the same periods. 8 springs are located in the plateaus of Kikuchi and Takayubaru highlands, while 6 springs are situated around Kinpo Mountain. Other 7 spring samples were collected within the Aso caldera watershed. 7 river samples were collected from Shira River from upstream to the downstream and 4 samples were collected from Midori River. Groundwater samples were collected using either existing electric pump or portable pump (MP1, Grundfos, Denmark). During pumping groundwater was continuously monitored with feld meter (HORIBA D-54, Japan) for temperature, electrical conductivity (EC), pH, dissolved oxygen (DO), and oxidation reduction potential (ORP). Sample was collected according to standard protocol (Wood [1981\)](#page-18-20), when all the physical parameters were almost at constant values generally in 30–60 min after fushing out of stagnant water in the boreholes.

Laboratory analysis

All samples were fltered through 0.2 μm cellulose acetate flters and collected in a pre-rinsed high-density polyethylene bottles without any headspace. Samples for cations and trace metals were collected in a 100 ml bottle and immediately acidified with 6 N ultrapure $HNO₃$ to produce $pH < 2$ for stabilization of trace metals. Samples for alkalinity, anion and dissolved silica $(SiO₂)$ measurement were collected into a separate 100 ml bottles without preservation. Samples for stable isotopic analysis (δ^{18} O and δ^2 H) were collected in a 50 ml bottle. The samples were stored in a cooler box at a temperature of $<$ 4 $^{\circ}$ C and transferred to the laboratory and stored in a refrigerator $(<$ 4 °C) before analysis $(usually < 1 week).$

All inorganic and stable isotope ratios analyses were carried out at the Hydrology Laboratory of Kumamoto University, Japan following standard analytical procedures (APHA-AWWA-WEF [2012](#page-16-3)). Alkalinity was measured within the same day of sampling by volumetric titration using 0.05 N H_2SO_4 and bromocresol green methyl red indicator. The redox potential values in this study have not been corrected to the standard hydrogen electrode, hence, can be used as relative values. Major cations $(Na^+, K^+, Ca^{2+}$ and $Mg^{2+})$ and anions (CI^- , $SO_4^2^-$ and NO_3^-) were measured by ion chromatography (Compact IC 761, Metrohm, Switzerland) within few days of sampling. The analytical precision of major ion analysis is around 3% . Dissolved $SiO₂$ was measured by molybdenum ammonium colorimetric method at a wavelength of 410 nm with a microplate spectrophotometer (Multiskan™ GO, Thermo Fisher Scientifc). The accuracy of the chemical analysis was validated by calculating ion balance errors when ion concentration expressed in meq/L (Appelo and Postma 2005) and it was better than $\pm 5\%$ for all samples except 7 samples in which it was around 6–7%. Trace metals were analyzed by ICP-MS (NexION 300, Perkin-Elmer Co., Ltd, USA).

The stable isotope ratios of water molecule $(\delta^{18}O$ and δD) were measured by an isotope ratio mass spectrometer (Delta V Advantage, Thermo Fisher Scientifc, USA). Samples were prepared using H_2O -CO₂ equilibrium method for δ^{18} O and H₂O-H₂ equilibrium followed by chromium reduction method for δD. Isotopic ratios are expressed as relative diferences (δ values) from Vienna standard mean ocean water (V-SMOW) in parts per mill (‰) (Craig [1961\)](#page-17-22) which are represented by δ (% e) = ($R_{\text{sample}}/R_{\text{V-SMOW}} - 1$) × 1000 and R is the ratio of ²H/¹H or ¹⁸O/¹⁶O in either sample or standard. Based on replicate measurements of standards and samples, the analytical precisions for $\delta^{18}O$ and δD were better than ± 0.05 and $\pm 0.5\%$, respectively. All analytical results are provided in Tables S2 and S3. Note that $\delta^{18}O$ and δD data except for river water samples from Midori River were presented in the recent article (Hosono et al. [2020\)](#page-17-23) that discussed coseismic hydrological change mechanisms after the 2016 Kumamoto earthquake, which research aim was diferent from our current study.

To identify type of fuid-mineral reactions, the geochemical computer code PHREEQCI (version 2, Charlton and Parkhurst [2002\)](#page-16-5), a complete Windows-based graphical user interface to the geochemical computer program PHREEQC (version 2, Parkhurst and Appelo [1999\)](#page-18-21) operating with accompanying WATEQ4F database, was used to calculate mineral saturation indices and aqueous speciation.

Results and discussion

Origin of groundwater

Figure [4](#page-6-0) shows the variation of δ^{18} O and δ D compositions of diferent water samples collected from the study area. Such diagram can provide valuable information on the origins of groundwater (e.g., Gat and Gonfantini [1981\)](#page-17-24). The most groundwaters show compositions close to the summer local meteoric water line (LMWL) indicating their meteoric origin and implying that summer rainfall is the dominant source of the groundwaters. Kudo et al. ([2013](#page-17-25)) studied in a forested catchment in western slope of the Aso caldera mountains and inferred that major recharge occurs during the summer rainfall, corresponding to our results. In detail, regional aquifer waters fall on compositional areas within certain ranges (Fig. [4\)](#page-6-0) refecting mixing of waters with different origins.

On Fig. [4,](#page-6-0) pore water samples collected from unsaturated soils in recharge areas are plotted on the compositional area to the right of local summer meteoric line, which isotopic shift can be explained by evaporation at soil surface before infltration (Okumura et al. [2018\)](#page-18-3). In contrast, spring waters collected outside from Aso caldera show compositions that are on the left of this line. Furthermore, all river waters and spring waters within Aso caldera watershed show most depleted δ^{18} O and δ D compositions reflecting their recharge sources with relatively higher elevation. Groundwater samples generally show intermediate compositions among these three end-components, soil pore waters, surrounding mountain aquifer waters, and the Shira River waters (Fig. [4\)](#page-6-0). It

Fig. 4 Bivariate plot of δ D and δ ¹⁸O for river, spring and aquifer waters collected from the study area. Local meteoric lines and compositions of soil pore waters are after Okumura et al. [\(2018](#page-18-3)). The global meteoric water line is after Gat and Gonfantini ([1981\)](#page-17-24). Average compositions of 100 core soil waters (10 m deep, 10 cm interval) are shown for four soil cores (Okumura et al. [2018](#page-18-3)), which sampling locations are shown in Fig. [1](#page-1-0)

is thus reasonable to think that aquifer waters are generally originated from summer rainfall and through multiple recharge pathways, soil water infltrations, mixing with mountain aquifer waters, and surface water percolations (Hosono et al. [2020\)](#page-17-23).

At a scale of flow line, a major focus of this study, there is no signifcant diference in terms of range of isotopic compositions between A–A′ and B–B′ lines (Tables S2 and S3): average values of $\delta^{18}O$ and δD for A–A' and B–B' flows are −7.1 and −46.3‰ and −7.0 and −46.9‰, respectively. This similarity could be attributed to common origins for both regional flow systems. However, $\delta^{18}O$ and δD values of the water samples change along the fow lines that was typically found for A–A′ fow line at mid-stream area of the Shira River in A-1 zone and some part of deep aquifer in A-2 zone and sporadically for groundwaters in B–B' line. This is due to a mixing process of waters from diferent origins in the midsection of entire fow lines as discussed in later section.

Hydrochemical characteristics

The principal characteristics of geochemical compositions for river, spring and aquifer waters, in terms of both concentrations and stable isotopes, are listed in Tables S2 and S3, where groundwater results are arranged in order of distance

from recharge through discharge to stagnant zones (Fig. [1](#page-1-0)). Groundwater temperatures ranged from 18.0 to 25.0 °C in A–A' flow and 17.4 to 25.3 \degree C in B–B' flow lines. On the other hand, spring water temperatures are lower than those of groundwaters with a range of 13.5–21.0 °C. An average temperature of river waters was 22.4 °C. Piper diagrams (Piper [1944\)](#page-18-22) show that groundwaters from the study area have a wide range of chemical compositions (Fig. S2). All waters from the study area are broadly classifed into four dominant and two mixed water types (Table S2) including $Ca-HCO₃$, $Ca-SO₄$, Na-HCO₃, Na-Cl, Ca–Mg–Na-HCO₃ (mixed cation) and Ca–(Na)-NO₃–HCO₃ (mixed anion) types. In general, groundwater chemistry of the frst and second aquifers is similar each other (Fig. S2 and Tables S4 and S5); however, it difers along the fow paths.

Spatial distribution of hydrochemistry types in the study area is shown in Fig. [5.](#page-7-0) In the A–A′ fow line groundwaters in the A-1 zone are mainly of $Ca-HCO₃$ type. In general, this type of waters is considered as juvenile waters in groundwater evolution model which are at their early stage of geochemical evolution (e.g., Plummer et al. [1990](#page-18-23); Edmunds and Smedley [2000](#page-17-26)). Groundwaters with rapid circulation process and short residence time, which have not experienced a prolonged water–rock interaction, may also have similar characteristics (e.g., McKenzie et al. [2001\)](#page-17-27). However, groundwaters in active flow line A-2 are mainly of Ca–Mg–Na-HCO₃ type, and in the discharge area $(A-3)$, at the end of $A-A'$ flow line, they became $Na-HCO₃$ type.

Similarly, groundwaters in the B–B' flow line begin as Ca-HCO₃ type in recharge zone $(B-1)$ followed by Ca–Na- $HCO₃$ type, then flows as Na-HCO₃ type in B-2 and B-3 zones (Fig. [5](#page-7-0) and Table S2). However, the last three samples in B–B′ fow line are Na-Cl type (Table S2; K16, K35 and K15) which are situated in a reclaimed land at the end of the flows near the coast (Figs. $1, 2$ $1, 2$) with TDS values ranging 500 to 2,230 mg/L. It has been revealed from nitrate tracers that efect of contaminations through surface soil water infltrations predominates in recharge areas (Fig. [5](#page-7-0), shown as "nitrate dominant" type water) and is limited in lateral flow and discharge zones for both fow lines (Hosono et al. [2013](#page-17-4)). Thus, observed hydrochemical evolutions from $Ca-HCO₃$ type in recharge areas to other chemistry types in downstream areas are results of several possible processes including silicate mineral weathering, cation exchange, anaerobic degradation of organic matters, deep fuid contributions, and mixing of saline waters.

Water discharging from springs as well as groundwaters in and around recharge zones (A-1 and B-1) are characterized by low electrical conductivities (\sim 180 μ S/cm) and a predominating Ca- (Mg) -(Na)-HCO₃ water type (Tables S2) and S3). However, some waters are enriched in $NO₃⁻$, due to contaminations from farms, crop lands and orchards in the Ueki, Kikuchi and Takayubaru highlands (Fig. [5\)](#page-7-0). On the contrary, spring waters in Aso Mountain area are dominated with $Ca-SO₄$ –(HCO₃) type. High sulfate in spring and groundwaters in the Aso area are mainly of hydrothermal

Fig. 5 Map showing spatial distribution of groundwater types expressed by stif diagrams

origin (Hosono et al. [2018\)](#page-17-28). Since the Aso springs are the main source of Shira River head water, sulfate is one of the dominant anions in this river apart from HCO_3^- (Ca-SO₄ type) though the Midori River is $Ca-HCO₃$ type water.

Changes in each measurement parameter along the fow lines

The trend of chemical constituents as well as isotopes in the Kumamoto groundwaters follows a progressive sequence of hydrochemical evolution. Here, changes of each measurement parameter along the fow lines A–A′ and B–B′ are described and possible causes to explain these changes are documented based on the results from this study and taking previous reports on anaerobic microbiological reactions deduced by multiple isotope ratios ($\delta^{13}C$, $\delta^{15}N$ and $\delta^{34}S$) (Hosono et al. [2013,](#page-17-4) [2014\)](#page-17-5), trace metals (Hossain et al. [2016a,](#page-17-0) [b\)](#page-17-1), and PCR analysis (Zeng et al. [2016](#page-18-2)) for anions.

The depletion in both $\delta^{18}O$ and δD of groundwater at last half part of A-1 zone (Fig. [6](#page-9-0)) implied that waters were largely recharged from isotopically depleted Shira River while rapid increase in both compositions in the entrance of A-2 indicate predominant contribution of mountain aquifer waters (Fig. [4](#page-6-0)). Then slight decrease in both isotope ratios afterwards implies increasing contribution of the mixing of Shira River water along the flow. The effect of river water infltration on hydrochemistry was most efectively traced in SO_4^2 ⁻ (Fig. [6\)](#page-9-0): it increased in the end of A-1 zone due to strong mixing and after the sudden depression due to the mixing with diluted mountain waters then increased again along the flow in A-2 zone with additional mixing with high SO_4^2 ⁻ river waters, corresponding well with isotopic results. Similar depleted isotopic and high SO_4^2 ⁻ anomalies are detected locally at the end of B-2 zone along B-B' flow near the Shira River, suggesting the occurrence of the same river water infltration in this limited location. Another isotopic anomaly, an enriched $\delta^{18}O$ compared to relatively depleted δD (Fig. [6](#page-9-0)), confrmed in A-2 zone and some spots on B–B′ flow line, possibly corresponds to soil porewaters (Fig. [4](#page-6-0)). This may be attributed to surface soil water percolations at local scale via preferential pathways (Kawagoshi et al. [2019](#page-17-29)); however, simultaneous anomalies were not detected in other geochemical parameters.

A very homogeneous $\delta^{34}S_{SO4}$ feature (8–10‰) was reported through A–A' flow line (Hosono et al. [2014](#page-17-5)) except mixing zone and in the end of stagnant zone where anaerobic sulfate reduction occurs with some fractionated $\delta^{34}S_{SO4}$ of up to 16‰. The source of SO_4^2 ⁻ in the recharge waters is a mixture of surface water (Shira River water), ecosystemrecycled sulfur near surface and of anthropogenic (chemical fertilizers) origins (Hosono et al. [2014](#page-17-5)) and we think an another SO_4^2 ⁻ source input is not prominent through the pathways. Cl− followed almost steady pattern (Fig. [6](#page-9-0)).

Similarly, $\mathrm{HCO_3}^-$ showed constant pattern through the flow (ranging ca. 70–100 mg/L) except one sample from the A-3 zone with its value $>$ 200 mg/L (Fig. [6](#page-9-0)). In the same A–A' flow line, homogeneous $\delta^{13}C_{\text{DIC}}$ compositions (–21 to −17‰) were confrmed for majority waters except at the end of stagnant zone where methanogenesis reaction occurs with increased HCO_3^- and fractionated $\delta^{13}C_{\text{DIC}}$ of up to -14% (Hosono et al. [2014](#page-17-5)). Observed $\delta^{13}C_{\text{DIC}}$ signature for majority waters indicates their overall origin that is ecosystem-recycled $CO₂$ with relatively young fresh mineral materials (e.g., Schulte et al. [2011\)](#page-18-24), but contribution from deep seated and volcanically originated $CO₂$ (e.g., Yamada et al. [2011](#page-18-25); Koh et al. [2017\)](#page-17-30) is negligible. Nitrate concentrations (Fig. 6 ; \sim 50 mg/L, originated from chemical fertilizers applied in the recharge zone; Hosono et al. [2013](#page-17-4)) are the highest in the A-1 recharge zone but decreased gradually toward the end of A–A' flow. Significant $\delta^{15}N_{NQ3}$ isotopic fractionation was not found through the main flow path except in the end of stagnant zone in A-3 where denitrification occurs with slightly fractionated $\delta^{15}N_{NQ3}$ of up to 10‰ (Hosono et al. [2013](#page-17-4)). $NO₃⁻$ decline along A-A' main flow path seems to be caused by dilution with both mountain aquifer and Shira River waters. Anion patterns observed in this study can reasonably be explained with a help of previously proposed information on their origins and processes.

In B–B′ flow line, nitrate concentrations sharply decreased along the fow (Fig. [6\)](#page-9-0). It has been confrmed in the same flow line that high $NO₃⁻$ (~60 mg/L) in the recharge zone is attributed to nitrogen loads from farm lands and it attenuated by denitrifcation reaction progressed along the flow path (Hosono et al. [2013;](#page-17-4) Zeng et al. [2016](#page-18-2)). The same authors confirmed that decreased SO_4^2 ⁻ through B-1 to B-2 zone (Fig. [6\)](#page-9-0) is associated with anaerobic sulfate reduction, in contrast, HCO_3^- increased along the flow B-B' line (Fig. [6\)](#page-9-0), which can be interpreted as it was caused by microbiological respiration through denitrifcation, sulfate reduction and methanogenesis. More signifcant anaerobic microbiological activities observed along B-B' flow line may be facilitated by relatively higher dissolved organic carbon content in the aquifer (Hosono et al. [2015](#page-17-31)) and/ or slower fow rate than A–A′ groundwater fow (Hosono et al. [2014\)](#page-17-5). Cl− concentrations are generally constant in these zones. However, both Cl^- and $SO_4^2^-$ concentrations increased in stagnant zone (B-3) resulting from mixing of pore waters of marine origin hosted in the Ariake clay sediments (Hosono et al. [2014](#page-17-5); Hossain et al. [2016a,](#page-17-0) [b\)](#page-17-1). Exceptionally, Cl− increased sharply (both up to 70 mg/L) in the end of B-1 zone (Fig. [6\)](#page-9-0). The highest groundwater temperature anomaly (\sim 23 °C) was depicted only in the end of B-1 zone over the study area (AIST [2014](#page-16-6)) and many hot spring facilities are developed in this spot. We thus assume that localized deep fuid seeps in this location, which possibility is supported by simultaneous increase in Na⁺ (Fig. [6\)](#page-9-0).

Fig. 6 Changes in measurement parameters along groundwater flow A–A' and B–B'. *X* axis represents the relative distance of sampling location from principal recharge area (Table S2). $Na_{corr}⁺$ means

sodium corrected with input of marine salts or atmospheric spray and is calculated by the formula Na $-$ (Cl⁻ \times 10,600/19,000)

Fig. 6 (continued)

For cations, significant concentration changes are observed partly associated with the degree of water mineralization along the both flow lines (Fig. 6). For instance, Na⁺ shows a gradual increasing tendency along the B–B' flow with the median values of 19, 23 and 121 mg/L in B-1, B-2 and B-3 zones, respectively. $Na⁺$ concentrations are relatively constant in A-1 and A-2 zones while little increased at the end of A-3 zone for A-A' line. Ca^{2+} shows steady state concentrations in initial two zones of both A–A′ and B–B' flow lines and decreasing trends in the final part of the both flow lines except last three samples in B–B' flow. Mg^{2+} generally follows similar trend with Ca^{2+} except in A-3 zone where it shows reverse trend. Roughly but gradual decreasing in $SiO₂$ along both flow lines (Fig. [6\)](#page-9-0) is probably due to precipitation of silica phases along the fow lines. At the end, K^+ showed increasing concentrations in the recharge areas (Fig. [6](#page-9-0)), possibly due to the impact of both mineral weathering and dissolution of agrochemicals applied.

Overall, it can be summarized that groundwater characteristics along A–A' flow line are mainly of aerobic environment and mixed cation $(Ca^{2+}, Mg^{2+}$ and Na⁺) dominated with no drastic variations along the flow except in A-1 zone where groundwater chemistry is heterogeneous due to mixing of different source waters. In contrast, B-B' flow can be characterized evolving as Na-dominated, Ca-poor, high pH and high alkalinity groundwater along the fow line with strong reducing activity. In contrast to the variations in the groundwater chemistry along flow directions, no obvious gap was observed between frst and second aquifers.

Mineralization processes

Alumino-silicate minerals are the most reported rock-forming minerals in the study fled (Table S1). On Fig. S3, most data tend to plot within the domain of global average of silicate weathering, suggesting that incongruent dissolution or leaching of the siliceous volcanic rocks is the major controlling process responsible for the formation of groundwater chemistry. Additional analysis using mass-balance calculations were provided in supplementary material (see 'massbalance calculations'). The extent of weathering and leaching depends on the chemical and mineralogical compositions of the host rocks. Following Herczeg ([2001](#page-17-32)) and Appelo and Postma ([2005](#page-16-4)), incongruent dissolution reactions of silicate weathering in the study area can be hypothesized as follows that are applicable for entire aquifers:

plagioclase weathering and leaching tends to be occurred in order of $Ca^{2+} > Na^{+} > Mg^{2+}$ (Mukherjee et al. [2009\)](#page-18-26) as the reaction proceeds. Therefore, $Ca-HCO₃$ and $Ca Mg-Na-HCO₃$ would be the major groundwater type in such area. According to Garrels and Mackenzie ([1967](#page-17-9)), hydrochemical evolutions in igneous rocks are dominated by hydrolysis of plagioclase by $CO₂$ and kaolinite is often an early weathering product which was accounted for ca. 80% of the total weathering capacities in Sierra Nevada spring systems. In addition, a presence of any volcanic glasses enhances weathering rate of the rocks in ca. 10 times faster than the crystalline basaltic rocks (e.g., Gislason and Eugster 1987) and Na⁺ releases preferentially over Ca^{2+} . In the study area, plagioclase phenocryst has the highest modal abundance, around 3–43% (Table S1), followed by pyroxenes $(1-25\%)$, and crystalline glass is the most common groundmass (see ["Topography, climate,](#page-2-1) [geology and mineralogy](#page-2-1)" and Table S1).

Enrichment of $Na⁺$ is one of the principal geochemical features of the Kumamoto groundwater. These excess Na⁺ are shown in Na_{corr}^+/Ca^{2+} and Na_{corr}^+/Mg^{2+} changes along flow paths (Fig. 6), which could be explained in two different ways, either weathering of silicate minerals or cation exchange reaction (Maybeck [1987;](#page-18-27) Plummer et al. [1990](#page-18-23)). It was proposed that the molar ratio of dissolved Na^+/Ca^{2+} in groundwater falls approximately equal to 1 in the common plagioclase phenocryst dominated volcanic regions (Garrels and Mackenzie [1967](#page-17-9)). Figure [6](#page-9-0) shows all the groundwaters from A–A' flow line have the molar ratio equal or less than 1 except few samples from A-3 zone with the molar ratio of >1 , which results entirely consistent with congruent dissolution of plagioclase. In the same zones (A-1 and

$$
Albite: 2NaAlSi3O8 + 2CO2 + 11H2O \rightarrow Al2Si2O5(OH)4 + 2Na+ + 2HCO3- + 4H4SiO4
$$
\n(1)

$$
Another: CaAl_2Si_2O_8 + 2CO_2 + 3H_2O \rightarrow Al_2Si_2O_5(OH)_4 + 2Ca^{2+} + 2HCO_3^-
$$
 (2)

Clinopyroxene :
$$
(\text{CaMg}_{0.7}\text{Al}_{0.6}\text{Si}_{1.7})\text{O}_6 + 3.4\text{CO}_2 + 4\text{H}_2\text{O} \rightarrow 0.3\text{Al}_2\text{Si}_2\text{O}_5(\text{OH})_4 + \text{Ca}^{2+} + 0.7\text{Mg}^{2+} + 3.4\text{HCO}_3^- + 1.1\text{H}_4\text{SiO}_4
$$
 (3)

Orthopyroxene :
$$
2CaMgFeAl_2Si_3O_{12} + 8CO_2 + 0.5O_2 + 15H_2O \rightarrow 2Al_2Si_2O_5(OH)_4
$$

+ $2Fe(OH)_3 + 2Ca^{2+} + 2Mg^{2+} + 8HCO_3^- + 2H_4SiO_4$ (4)

Weathering rate of plagioclase and mafic minerals is signifcantly faster than that of quartz and K-feldspar (White et al. [1998\)](#page-18-6). Hence, if plagioclase has a larger modal abundance than mafic minerals, groundwater chemical compositions can largely be explained by

A-2), groundwater chemistry evolved from $Ca-HCO₃$ type to Ca–Mg–Na-HCO₃ type along the flow path (Fig. [5\)](#page-7-0) that should be caused by the weathering of minerals containing Mg, i.e., pyroxenes, the second important host minerals over the aquifer systems, in addition to plagioclase weathering.

In the B–B' flow line, however, the molar ratio $\text{Na}^+_{\text{corr}}$ Ca^{2+} for the samples from B-1 and B-2 zones (with $pH < 8$) was approximately or slightly over than $1.$ Na⁺ is accumulating to the groundwater and Ca^{2+} is being depleted with increasing pH nearly or above 8 (Fig. [6\)](#page-9-0). Observing variations in $\text{Na}_{\text{corr}}^+/\text{Ca}^{2+}$ ratio along the flow lines B-B' and end of A–A′, we can postulate for another hydrochemical evolution process taking place especially in the stagnant zone. These phenomena can be described in two diferent ways, addition of Na+ originated other than plagioclase weathering or Ca^{2+} removal via cation exchange. Dissolution of volcanic glasses that are not described but maybe present in the study field may cause observed Na⁺ elevations. However, this phenomenon should occur in earlier stage of weathering sequence than plagioclase weathering. Thus former hypothesis is unlikely to explain observed hydrochemical signatures, although we agree the occurrence of volcanic glass dissociation itself but it might have been completed during time elapsing after the last volcanic eruptions.

Figure [7](#page-12-0) expresses the influence of cation exchange process. In this fgure, concentrations of divalent cations $(Ca^{2+} + Mg^{2+})$ that may have been involved in exchange reactions are corrected by equivalent concentrations of associated anions $(HCO_3^- + SO_4^{2-})$. Again, the amount of Na⁺ is corrected by Cl[−] to account for Na⁺ that is solely from the aquifer geology. If cation exchange is a signifcant controlling process, then the relation between these two sets of parameters should be linear with a slope of~−1 (Fisher and Mullican [1997](#page-17-8)). Figure [7](#page-12-0) shows that samples from A-3 zone and B-B' flow line (except two high salinity samples) having a trend line $(r^2 = 0.97)$ with a slope of −0.98, which is close to -1, indicating that cation exchange has an important control on those groundwaters. We further confrmed that reverse ion exchange processes would not

Fig. 7 Relation between $(Ca^{2+} + Mg^{2+}) - (HCO_3^- + SO_4^{2-})$ against $(Na⁺ – CI⁻)$ to indicate the cation exchange reactions for groundwater samples collected from A-3 flow zone of the A-A' line and B-1 to B3 zones of the B–B′ line

expect to be occurred based on $(Ca^{2+} + Mg^{2+})$ – $(Na^{+} + K^{+})$ vs HCO_3^- – $(SO_4^2^- + Cl^-)$ classification diagram (Fig. S6). To support cation exchange mechanism, there would need to exist sodium-charged cation exchange media within the rock unit to facilitate the process. It is already described that Kumamoto plain area is covered by marine and flood sediments which partly contains clays with saline pore waters with $Na⁺$. Moreover, in B-2 zone aquifers are covered by wet-land composed of clay materials. Both areas are characterized by relatively long (>55 years) groundwater residence time. It is plausible therefore that Ca^{2+} and Mg^{2+} in the water exchanged with previously absorbed $Na⁺$ on the surface of the clay minerals (Domenico and Schwartz [1990\)](#page-17-34) with relatively later stage of geochemical evolutions. Several researchers studied similar characteristics in groundwater systems and reported calcite and plagioclase dissolutions followed by cation exchange (Edmunds [1981](#page-17-35); Appelo and Postma [2005](#page-16-4); Mukherjee et al. [2009](#page-18-26)).

Thermodynamic interpretation of water–rock interactions

Saturation indices (SI) can be used to investigate thermodynamic control on groundwater chemistry and to represent the possible reacting mineral phases. It is helpful to evaluate the states of mineral equilibria in the groundwater system and to identify the reacting mineral phases in absence of accurate mineralogical data of the aquifer matrix. The SI of a particular mineral is defned as:

$$
SI = \log(IAP/K_T)
$$
 (5)

where, IAP is the ion activity product of the mineral–water reaction and K_T is the equilibrium solubility constant at a given temperature for particular samples. SI value of a specifc mineral nearly zero indicates that the groundwater system is in equilibrium with this specifc mineral. Whereas, $SI > 0$ or $SI < 0$ specifies groundwater is oversaturated or undersaturated that means minerals favorably precipitate or dissolve, respectively. The average state of saturation of the groundwater in diferent zones along the fow lines with respect to selective minerals is shown in Table [1](#page-13-0). Hydrochemical data except Al^{3+} shown in Table S2 were used to calculate the SI value. We do not have complete set of Al^{3+} data for all samples from present study and thus used average Al data as per fow lines from diferent set of samples as input data (A–A′ frst aquifer: 147 µg/L, A–A′ second aquifer: 79 µg/L, B–B′ frst aquifer: 51 µg/L, B–B′ second aquifer $47 \mu g/L$). Al³⁺ has a very low solubility in groundwater and tends to be conserved in solid phase (Freeze and Cherry [1979](#page-17-36)). There is a possibility of the presence of colloidal particles even though the samples were fltered through 0.2 μm membrane filter. Furthermore, in higher concentrations, $Al³⁺$

Table 1 Average saturation indices (SI) for diferent zones along the fow lines with respect to selected mineral phases

Mineral phases	Stoichiometry	First aquifer						Second aquifer					
Flow zone		$A-1$	$A-2$	$A-3$	$B-1$	$B-2$	$B-3$	$A-1$	$A-2$	$A-3$	$B-1$	$B-2$	$B-3$
Albite	NaAlSi ₃ O ₈	-3.0	-1.8	-1.0		$-1.8 -2.3$	0.2	-2.1	-1.9	-1.6	-1.4	-2.0	0.0
Anorthite	$CaAl2Si2O8$	-5.2	-4.3	-2.9		-4.3 -4.6 -1.8		-4.6	-4.4	-3.3	-4.2		$-4.5 -2.1$
Forsterite	Mg_2SiO_4	-12.9	-10.5		-10.6 -11.3 -9.8 -7.5			-11.2	-11.0	-10.3		-10.4 -10.4 -7.6	
Diopside	CaMgSi ₂ O ₆	-7.7	-5.0	-5.3		-5.7 -4.4 -2.4		-5.7	-5.5	-5.1	-4.7		$-5.0 -2.3$
Hornblende	$Ca2Mg5Si8O22(OH)2$	-17.5	-7.6	-8.3	$-10.5 - 5.9$			$1.6 - 10.1$	-9.4	-8.2	-6.9	-7.6	1.8
Phlogopite	$KMg_3AlSi_3O_{10}(OH)$	-9.9	-9.6	-8.0				$-8.6 -8.4 -3.0 -12.8$	-8.8	-8.1	-10.2	-8.0	-2.6
Chalcedony	SiO ₂	0.0	0.3	0.3	0.3	0.2	0.2	0.3	0.3	0.1	0.4	0.2	0.2
Cristobalite	SiO ₂	0.1	0.4	0.4	0.3	$0.2\,$	0.3	0.3	0.4	0.2	0.5	0.3	0.2
Ouartz	SiO ₂	0.5	0.8	0.8	0.7	0.6	0.7	0.7	0.8	0.6	0.9	0.7	0.6
Amorp. Silica	SiO _{2(a)}	-0.8	-0.5	-0.5		$-0.6 - 0.7$	-0.7	-0.6	-0.5	-0.7	-0.4	-0.7	-0.7
Calcite	CaCO ₃	-2.1	-1.1	-1.1		$-1.1 -0.7$	-0.1	-1.2	-1.2	-1.0	-1.1		$-0.8 - 0.1$
Dolomite	CaMg(CO ₃) ₂	-4.3	-2.3	-2.1		$-2.5 -1.5$	0.0	-2.6	-2.5	-2.0	-2.4	-1.7	-0.2
Gypsum	CaSO ₄ :2H ₂ O	-3.0	-2.7	-2.7		-3.0 -3.1 -3.1		-2.8	-2.7	-2.8	-3.1		$-2.4 - 3.2$
Halite	NaCl	-8.8	-8.6	-8.3		$-8.1 - 8.6 - 6.9$		-8.5	-8.6	-8.4	-8.0	-8.5	-6.7
Gibbsite	$Al(OH)_{3}$	0.9	0.5	1.3	0.6	0.3	1.3	0.5	0.6	1.2	0.4	0.5	1.0
Kaolinite	$Al_2Si_2O_5(OH)_4$	3.6	3.4	5.0	3.5	2.6	4.6	3.3	3.6	4.3	3.3	3.1	4.1
Ca-Smectite	$Ca_{0.165}Al_{2.33}Si_{3.67}O_{10}(OH)_{2}$	1.9	2.3	4.1	2.3	1.2	3.7	2.1	2.4	3.1	2.3	1.8	3.2
Illite	$K_{0.6}Mg_{0.25}Al_{2.3}Si_{3.5}O_{10}(OH)_{2}$	0.7	1.3	3.2	1.3	0.5	3.6	1.1	1.4	2.3	1.5	1.0	3.1
Halloysite	$Al_2Si_2O_5(OH)_4-H_2O$	-1.6	-1.7	-0.1		-1.6 -2.5 -0.5		-1.8	-1.6	-0.8	-1.8		$-2.0 -1.0$

is sensitive to complexation with other dissolved species especially with organic compounds. A number of samples from Kumamoto plain area contains relatively high Al^{3+} , indicating inorganic conservation of aluminum (Cleaves et al. [1970\)](#page-16-7). Therefore, a part of dissolved aluminum in solution may be organically bounded and should not be included in calculations of saturation index. However, calculated SI values were not signifcantly afected with a possible uncertainty of these Al^{3+} concentrations due to their trace concentrations relative to other major ions.

It is clear from Fig. [8](#page-14-0) and Table [1](#page-13-0) that groundwaters in the Kumamoto basin are undersaturated with respect to reported alumino-silicate minerals. The calculated SI values for albite and anorthite, diopside, and phlogopite (a micaceous mineral chemically similar to biotite) remain strongly negative along the fow lines showing that the groundwater is undersaturated with respect to these minerals (Table [1](#page-13-0); Fig. [8](#page-14-0)). It is notable that albite and anorthite tend to reach saturation along the fow paths at the end of A-3 and in B-3 zones, suggesting that the mineral saturation and subsequent precipitation of secondary minerals, i.e., smectite as explained later, could occur. Like most natural waters (e.g., Hem [1989](#page-17-6)), groundwaters are oversaturated with quartz $(0.05 < SI > 1.05)$ and undersaturated with amorphous silica $(-1.26 < SI > -0.27)$. In addition, they are in equilibrium to little oversaturated condition with respect to cristobalite and chalcedony phases. Thus silica in amorphous form is likely a major contributor in the groundwater systems. All groundwater samples are undersaturated with respect to gypsum $(-4.6 > SI < -1.98)$ and halite $(-9.18 > SI < -4.76)$ (Table [1\)](#page-13-0). The soluble sulfate and halite minerals are not present in the host rocks. Presence of high concentrations of sulfate in A-1 zone water and Na-Cl type groundwater in B-3 zone are originated by a mixing of Shira River water and high saline pore waters in marine clay sediments, respectively. Groundwater remains close to or somewhat above saturation with respect to gibbsite, whereas, it becomes signifcantly oversaturated with respect to secondary minerals of weathering products including kaolinite, smectite and illite (Fig. [8](#page-14-0); Table [1](#page-13-0)).

Figure [9](#page-15-0) shows mineral stability diagrams for the $Na₂O-Al₂O₃$ –SiO₂–H₂O and CaO–Al₂O₃–SiO₂–H₂O systems. These fgures were generated by assuming end-member compositions using the equilibrium relationship of Tardy ([1971](#page-18-28)) for standard temperature and pressure (25 °C and 1 atm). The groundwater samples from A–A' flow were plotted mainly on the kaolinite stability feld while some samples fall on boundary fled between kaolinite and smectite. On the contrary, most of the samples from B–B' flow line was placed in the stability feld of smectite and few samples followed boundary area. Groundwaters are oversaturated for both minerals for both fow lines and thus equilibrium with kaolinite or smectite is an important process that controls groundwater chemistry. However, some mineralogical study revealed that halloysite is the major clay mineral formed in resulting from plagioclase weathering in some outcrops of

Fig. 8 Variations in saturation indices for selected rock-forming minerals and clay minerals along the A–A′ and B–B′ fow lines. Thermodynamic values were calculated by using PHREEQCI with wateq4f.dat data base

the study field (Aomine and Wada 1962). It is difficult however to assess this issue thermodynamically because of the lack of stability diagram involving halloysite and the result showing groundwaters are generally undersaturated with respect to this type of clay minerals (Table [1\)](#page-13-0). Thus here we only state halloysite presented presumably as a major weathering product under humid and subtropical climate like in Kumamoto.

It has been reported that availability of excess silica from relatively high rate of chemical weathering of basaltic rocks prevailed in humid tropical climate favors kaolinite stability while sample from granitic area with arid climate tends to be plot in smectite feld which reaction occurs in later stage of weathering process with longer water–rock contact time (Appelo and Postma [2005](#page-16-4); Stallard and Edmond [1987](#page-18-29)). It is thus reasonable to think that more stability of A–A' flow samples in the kaolinite feld indicates that groundwater is well drained, whereas, samples plotting in smectite feld from B–B' flow suggest restriction in groundwater flow. This is in good agreement with groundwater fow regime in this area where waters in A-1 and A2 zones have shorter residence time (<40 years) with active hydrological cycle than B–B' flow line with longer residence time of > 55 years (Kagabu et al. [2017\)](#page-17-13). The diagrams (Fig. [9\)](#page-15-0) depicted greater maturity of waters in B–B′ than A–A′ lines in hydrochemical evolution. This assumption is supported by simultaneous saturation in plagioclase (Fig. 8) and decline in SiO₂ concentrations of matured waters (Fig. 6), which accompany

Fig. 9 Stability diagram of **a** $Na_2O-Al_2O_3-SiO_2-H_2O$ and **b** CaO– $Al_2O_3-SiO_2-H_2O$ systems showing minerals expected to be in thermodynamic equilibrium with groundwater samples. Dashed lines represent saturation line of quartz and amorphous silica, respectively. Thermodynamic data are taken from Tardy [\(1971](#page-18-28)). Legend: A–A′ frst aquifer (open triangle), A–A′ second aquifer (solid triangle), B–B′ frst aquifer (open circle), and B–B′ second aquifer (solid circle)

consumption of silica in forming smectite to be precipitated. These results correspond well with recent report in other volcanic areas (e.g., Ide et al. [2018](#page-17-37)).

Summary and implications

Geochemical weathering processes were studied to describe comprehensive hydrobiogeochemical evolutionary model for volcanic aquifer systems in Kumamoto area along two welldefned groundwater fow lines A–A′ and B–B′ from land to ocean (Fig. [10\)](#page-16-9). Groundwaters in this area are recharged largely by summer precipitation through soil infltration at plateaus and percolation of waters from surrounding mountains and locally through river waters. Impact of agrochemical applications predominates in recharge areas for both flow paths, most clearly shown by elevated NO_3^- concentrations (~60 mg/L). However, major water type (Ca-HCO₃ type) in these areas is controlled by mineral decomposition mainly of plagioclase to form secondary minerals, i.e., kaolinite (however, it is highly probable as it presents as halloysite), through atmospheric $CO₂$ mediated volcanic rock weathering.

In the A–A' flow path, recharged waters flow laterally with enhanced weathering mainly of plagioclase and secondary of pyroxenes but without signifcant anaerobic microbiological reactions, resulting $Ca-Mg-Na-HCO₃$ type waters that are discharged in the entrance of plain as springs in Ezu Lake. Relatively active groundwater fow with short residence time $(40 years) facilitates this type of water$ evolutions. However, some stagnant waters remain in the plain, which are defined as $Na-HCO₃$ type with depleted $NO₃⁻$, due to combined processes including ion exchange and anaerobic denitrifcation. At this stage of geochemical evolution, plagioclase decomposition is not important but smectite precipitation predominates in association with $SiO₂$ saturation. This type of waters is formed in reducing environments under inactive fow regime with long water–rock/ sediments contact time, at least, over 55 years.

Groundwaters in B–B′ line evolve their geochemical signatures to Na-HCO₃ type with depleted SiO_2 , NO₃⁻ and SO_4^2 [–] but elevated HCO_3^- in the middle of lateral flow zone due to progressive ion exchange reaction, smectite precipitation, and anaerobic microbial succession from denitrifcation until methanogenesis under extensive anoxic aquifer setting. Attenuation of the active groundwater flow (groundwater) residence time>55 years) in thinning aquifer systems at basement uplifting area enhanced water–rock-microbiological interaction in steady state. Then waters are transported towards the coast maintaining similar hydrobiogeochemical signatures in upgradient, however, with signifcant contribution of pore waters in marine clay sediments with high salinity. Mixing of these high-salinity and biochemically evolved pore waters of marine origin further alters original water characters to intensify ion exchange reaction that is triggered by clays with large amount of $Na⁺$ attachment. Finally, water changes to Na-Cl type due to strongest interaction with sea water components in the coast.

We understand that additional assessments with combined thermodynamics and kinetic control and inverse geochemical modeling may provide further insights into proposed scenario; these tests favorably require more data (e.g., concentrations of gas phases in water such as N_2 , NH₃ and H₂S and organic compounds like dissolved organic carbon) and additional analyses (e.g., defnition of confned fow systems without any water mixing along

Fig. 10 Conceptual cartoon showing hydrobiogeochemical evolution of groundwater in Kumamoto region along the two major fow lines. Microbiological reaction models are after Hosono et al. [\(2013](#page-17-4), [2014\)](#page-17-5) and Hossain et al. ([2016a,](#page-17-0) [b](#page-17-1))

the flow line, identification of initial and final groundwater chemistry) and thus could not be included in this study. Nevertheless, this study quantitatively illustrates how groundwater chemistry could evolve along its flow path in volcanic aquifers in regional scale at given time scales in consequence of multiple processes, i.e., diferent source inputs, water–rock interaction, contaminations and microbial activities. This study can serve as an important basis for understanding groundwater chemical evolutions in active groundwater fow systems, such as volcanic and sedimentary fan aquifers, and for management and protection of groundwater resources.

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