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Soil organic carbon accumulation in dry tropical mountainous zone of Cameroon

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Abstract

The present work was designed to study a soil sequence in Mount Mandara, in order to identify the influence of altitude, soil characteristics, and land use on the accumulation of soil organic carbon (SOC). The study was conducted in four sites in the Far-North region of Cameroon, including Zamai, Kossohone, Sir, and Rhumsiki. Three pits were dug down to the weathering horizons in three positions (upslope, mid-slope, and footslope) along a representative toposequence in each site. Samples were taken from each pit at regular increment of 25 cm from the soil surface. The total SOC stock (T-SOCS) contents are 128.63 ± 5.25 Mg ha⁻¹ in Arenosols at Zamai (608 m a.s.l.), 158.248 ± 10.52 Mg ha⁻¹ in Leptosols at Kosohone (865 m a.s.l.), 158.99 ± 13.25 Mg ha⁻¹ in Luvisols at Sir (970 m a.s.l.), and 144.79 ± 24.23 Mg ha⁻¹ in Regosols at Rhumsiki (1050 m a.s.l.). The main secondary minerals are smectite, kaolinite, sepiolite, lepidocrocite, hematite, and calcite. Clay minerals and iron oxides are good receptacle for SOC and might constitute a major asset for the accumulation and the sequestration of SOC. Increase in elevation leads to decrease in the annual temperature which affect microbial activity, leading thus to a slow rate of soil organic matter (SOM) decomposition, which thereby affected SOCS. This is confirmed by the significant correlation between altitudinal gradient and T-SOCS (r=0.70), with altitude contributing to the accumulation of SOC for 49.68%. Texture also plays a central role in carbon sequestration in the studied area, confirmed by the significant and positive correlation between silt fraction and SOM. Under Regosols, there is a decrease in T-SOCS value as a result of a reduction of the quantity of organic matter returned to the soil and more rapid SOM decomposition due to ploughing. This research provides a preliminary assessment for SOC stock at Mount Mandara. It suggests that altitudinal gradient, land use, and soil characteristics should be included in SOCS models and estimations at local and regional scales.

Keywords Soil organic carbon · Altitudinal gradient · Land use · Mount Mandara · Cameroon

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Introduction

Pedogenesis involves all the processes relating to the formation and evolution of soils on the earth surface. It is the consequence of climate (temperature, precipitation), parent materials, biota, time, and topography (slope angle, slope shape, altitude) (Jenny 1941; Amundson 2014; Singer 2015). In general, there is a vertical organization of soil from the base of massifs to their summit. This variation along the altitudinal gradient has repercussions on soil characteristics and soil organic carbon (SOC) stocks (Fiener et al. 2015; Fissore et al. 2017; Tsozué et al. 2019). This is even more accentuated when the bedrock is made of volcanic rocks containing glasses which can easily be transformed into amorphous materials which further facilitate the retention of soil organic matter (SOM) (Tsozué et al. 2019). It is demonstrated that altitude alone accounted for 73% of the variation in SOC storage in tropical to temperate part of Indian Himalayan region (Tashi et al. 2016) and 75% in Mount Bambouto in the humid part of Cameroon (Tsozué et al. 2019).

Soil organic carbon stocks (SOCS) play an essential role in the climate regulation through atmospheric carbon dioxide (CO₂) storage (Olson et al. 2016; Greiner et al. 2017; Mayer et al. 2019; Tsozué et al. 2021). It is proved that vegetation fixes about 120 Pg of carbon (C) per year from the atmosphere through photosynthesis process and half of this C is returned to the atmosphere by plants (Bernoux and Chevalier 2013; Tsozué et al. 2021). Part of the atmospheric C drawn by plants is stored in biomass and soil which have the potential to constitute a C sink, in the form of SOM, slowing down the increase of CO₂ in the atmosphere (Mayer et al. 2019; Tsozué et al. 2019, 2021).

Soils are the largest continental reservoir of organic C (Smith 2012; Chenchounia and Neffar 2022). This reservoir contains nearly 800 GtC in the first 30 cm of soil and between 1500 and 2000 GtC over the first meter (Ciais et al. 2013; Chenchounia and Neffar 2022), representing about twice that of the atmospheric reservoir (805 GtC) and three times what is stored in vegetation (550 GtC) (Houghton 2007). Vegetation and earthly animals constitute a stock of 610 GtC (FAO 2015; Chenchounia and Neffar 2022). SOCS varies in space and time. It is in response of several biotic and abiotic factors among which the most important is the climate (Boudjabi and Chenchouni 2022; Chenchounia and Neffar 2022). Climate controls both net primary production and SOC dynamics, through the combined and synergetic effects of temperature and precipitation (Deb et al. 2015; Chenchounia and Neffar 2022). High precipitation increases soil moisture and soil microbiological community activity, which induces an increase in plant primary production and greater inputs of C in the soil (Razafimbelo et al. 2010; Chenchounia and Neffar 2022). Soil microbiological communities are key drivers of biogeochemical cycling and numerous important ecosystem processes such as decomposition and soil respiration (Xu et al. 2020), which have important effect on the sequestration of C in the soil. Temperature affects the sequestration time of C in the soil (Razafimbelo et al. 2010; Chenchounia and Neffar 2022). High temperatures stimulate soil biological activity and accelerate the SOM degradation, leading thus to a decrease in microbial biomass and dissolved SOC, which is also related to the low quantity of litter produced by plants (Boudjabi and Chenchouni 2022; Chenchounia and Neffar 2022). In fact, an increase of 10 °C in air temperature decreased the sequestration time of SOC by a factor of 2-3 (Raich and Schlesinger 1992). In addition to climate, other environmental and anthropogenic conditions which have an impact on the C cycle dynamics include soil mineralogy, soil texture and structure, chemical properties, landscape, landscape position, slope, latitude, anthropogenic factors, soil management, and natural disturbances such as wind, fire, drought, insects, and diseases (Zinn et al. 2007; Lal 2009a; Lal 2009b; Wang et al. 2010; Novara et al. 2011; Corral-Fernández et al. 2013; Cerdá et al. 2014; Fernandez-Romero et al. 2014; Lozano-García and Parras-Alcantara 2014a, b; Bruun et al. 2015; Tsozué et al. 2019, 2021).

In dry areas, SOC content is low, representing only less than 1% of the soil mass (Lehmann and Kleber 2015; Plaza et al. 2018; Tsozué et al. 2021). In the temperate zone on contrary, it reaches 4 to 5% in grassland soils or under forest (Lehmann and Kleber 2015; Plaza et al. 2018; Tsozué et al. 2021). In humid tropical mountainous zone of Cameroon, it was observed an increase in SOCS along with the altitudinal gradient from 158 Mg ha^{-1} at the base of the massif (1400 m a.s.l.) to 302 Mg ha⁻¹ in its summit (2740 m a.s.l.) (Tsozué et al. 2019). This summit was characterized by the presence of market gardening crops where a wide range of European vegetables are cultivated (cabbage, sweet pepper, potato, leek, garlic, onion, tomato, pepper, parsley, beets, carrots, celery, and peas) (Tsozué et al. 2011). A similar organization is observed in the Mount Mandara that located in the dry tropical zone of Cameroon, from its base at 400 m a.s.l. to its summit at 1400 m a.s.l. There is a gradual appearance of European vegetables, largely represented by potatoes from 800 m a.s.l. The presence of European vegetables is synonymous of land use change in the area which is modifying the SOCS that will positively or negatively impact local and global climate change.

In general, works devoted to SOC in dry tropical mountainous zone are very rare. In Central Africa, works related to this environment are completely absent. The present work proposes to study the influence of altitude, soil characteristics, and land use on SOCS at dry tropical mountainous zone, particularly in the Mount Mandara. Specifically, it will involve to (1) study the mineralogical, geochemical, and physicochemical characteristics of these soils; (2) study the variation of the SOCS along the altitudinal gradient; and (3) analyze the influence of soil characteristics, human activities, and altitudinal gradient on SOC sequestration. To achieve this, the Mayo-Tsanaga Division in general and more specifically the sites of Zamai, Kossohone, Sir, and Rhumsiki in the Far-North region of Cameroon were chosen. We hypothesized that altitude, soil characteristics, and land use are main factors controlling the accumulation of SOC in the Mount Mandara. The results are expected to improve our understanding on the distribution of SOCS in Mount Mandara and provide important additional information for the scientific evaluation of SOCS at local and regional scales.

Materials and methods

Description of the study area

The Far-North region of Cameroon stretches on about 325 km from the Sudanese countries in the south to the shores of Lake Chad in the north. It is located between $10^{\circ}00'$ and $13^{\circ}00'$ North and between $13^{\circ}30'$ and $15^{\circ}30'$ East. The study took place in the Mayo-Tsanaga Division (Fig. 1). The different study sites are Zamai, Kossohone, Sir, and Rhumsiki (Fig. 1). The climate is characterized by an alternation of two seasons, a dry season from October to May and a short rainy season from June to September. The average annual precipitation varies between 800 and 1000 mm and the mean annual air temperature range between 28 and 25.8 °C (Table 1). The relative humidity is generally low, reaching 10% in the dry season. The landscape is undulate and made of a succession of hills, with altitude ranging between 600 and 1400 m a.s.l. (Fig. 1). The different soil units are developed on alluvial materials (Zamaï), gneiss (Kossohone and Sir), and granit (Rhumsiki) (Ngounounou et al. 2000; Tamen et al. 2015; Gountié Dedzo et al. 2019). The different soil types present in the region are Vertisols, Arenosols, Leptosols, Regosols, Acrisols, Lixisols, Luvisols, Fluvisols, and Gleysols (Tsozué et al. 2017; Silatsa et al. 2021). Due to the relief and the climate, the hydrographic network has two flow regimes: a low water regime in the dry season and a flood regime in the rainy season. The main collectors are Mayo-Tsanaga and Mayo-Louti. The vegetation is strongly anthropized in the whole Mount Mandara (Table 1). The most represented species are Acacia hockii, Albizia-chevaleri, Balanites aegytiaca, Bauhinia rufescens, Combretuma culeatum, Dichrosta chyscinera, Ziziphus mauritiana, Piliostigma reticulatum, Strychnos spinosa, Ximenia americanass, Andropogon pinguipes, Bothriochloa bladhii, Chrysochloa hindsii, Eragrostis viscosa, Heteropogon melanocarpus, Leersia drepanothrix, Oryza glaberrima, Pogonarthria squarrosa, Schizachyrium nodulosum, and Sehima ischaemoides (Letouzey 1985; Van der Zon 1992). The main human activities are agriculture and livestock breeding. All the upper part of the massif above 1000 m a.s.l. is cultivated.

Field work and soil sampling

A detailed study of the topographic maps of the study area help to choose four study sites. The study was carried out in the dry season. Precipitation of the year reported in the study is the mean annual precipitation according to the historic. At each site, meticulous landscape analyses were conducted. The main parameters examined are the relief,



Fig. 1 Location of the study area

Table 1 Physical and mineralogical characteristics of the studied sites

Site	Altitude	de Climate		Bedrock	Mineralogical	Surface state	Soil classification	n ^b
	(m a.s.l. ^a)	Precipitation (mm)	Tem- perature (°C)		composition			
Rhumsiki	1050	1006.6	25.8	granite	Quartz, kaolinite, smec- tite, sépiolite, plagio- clase, alkaline feldspar, calcite, hematite	Vegetation highly anthro- pogenized; cultivation of maize and sorghum but recently colonized	Eutric Regosols	12
Sir	970	829.5	26.7	gneiss	Quartz, kaolinite, smectite, sépiolite, lépidocrocite, calcite, plagioclase, alkaline feldspar	Virgin forest reserve	Haplic Luvisols	15
Kosohon	865	802	27.5	gneiss	Quartz, kaolinite, smec- tite, sépiolite, hematite, plagioclase, alkaline feldspar	Virgin Forest reserve	Dystric Leptosols	13
Zamai	608	800	28.0	Sandy alluvial deposits	Quartz, kaolinite, smec- tite, sépiolite, calcite, hematite plagioclase, alkaline feldspar, amphibole	Virgin forest reserve	Dystric Arenosols	15

^aMeters above sea level

^bNumber of sample

bedrock, hydrography, and vegetation. Four transects were previously established in each site and one most representative toposequence was retained per site. Along each toposequence, three pits with a depth up to more than 150 cm were made, including one at the top of the interfluve, a second at mid-slope, and a third at the footslope. Twelve pits in total were dug down to the weathering horizons. Soils description was made on the walls of the profiles and the colors were determined in the field using the Munsell colors charts. Samples were thereafter taken from each pit at regular increment of 25 cm from top to bottom in each soil profile. Depth increments were used for a uniform comparison between studied soils. A total of 55 samples were collected in the polythene bags labeled with corresponding defined codes for laboratory analysis. For soil classification, the IUSS Working Group WRB (2015) was used.

Soil analysis and SOC estimation

The mineral phase analysis was carried out by X-ray diffraction (XRD) on disoriented powders and on oriented aggregates, and by Fourier transform infrared spectroscopy (FTIR) on total powder. Samples were scanned in the 20 range from 2° to 45° with step size of 0.02° and time per step 2 s, with CuK α 1 radiance, $\lambda = 1$, 5418 Å, at 40 kV, and 30 mA. Detailed identification was done through air-drying (24 h), glycolation (22 h), and heating (500 °C for 4 h) with

further tests on clay fraction (Moore Duane and Reynolds Robert 1989).

Diffuse reflectance infrared spectra were recorded between 4000 and 500 cm⁻¹ using a FTIR spectrometer (IFS 55). The spectrum resolution was 4 cm⁻¹ and spectra were obtained by accumulation of 200 scans (Nguyen et al. 1991).

Elemental composition analysis was performed by X-ray fluorescence spectroscopy (XRF) (Beckhoff et al. 2006).

For physicochemical analysis, different methods used in this study were already used in the same laboratory by Tsozué et al. (2019, 2021). Except the bulk density, all the physical and chemical analysis were carried out after sieving the soil through 2-mm meshes. For bulk density (BD), clods were first roughly sized in order to fit into the cell $(8 \times 6 \times 6 \text{ cm})$ of the clod box, and thereafter, it was measured by the paraffin-coated clods method (USDA 2004). The soil texture was determined by Robinson's pipette method using sodium hexametaphosphate $[Na_6(PO_3)_6]$ as dispersing agent (USDA 2004) and the textural class was determined using the USDA texture triangle. Soil pH was carried out with a pH meter equipped with a glass electrode in soil-water suspensions (1:2.5) (Guitián and Carballas 1976). Exchangeable cations were extracted by 1 N NH₄OAc at pH 7 and their concentrations were determined by atomic absorption spectrometry Perkin Elmer for Ca and Mg, and by flame emission spectrometry for K and Na. Cation exchange capacity (CEC) was also determined using the ammonium acetate method at pH 7, by a direct continuation using a 1 N potassium chloride (KCl) saturation solution. SOC was determined by dichromate oxidation using the Walkley–Black method (Walkley and Black 1934). SOM content was obtained using a conversion factor of 1.724 (Walkley and Black 1934). Total nitrogen (TN) was determined with the Kjeldahl method (Bremner 1996) and the C:N ratio calculated by dividing the SOC concentration by the N concentration.

SOCS, expressed in Mg ha⁻¹, was calculated according to IPCC (2003) as follows:

$$SOCS = \sum_{\text{horizon=1}}^{\text{horizon=n}} ([SOC] \times BD \times d \times (1 - \delta 2mm\%) \times 10)_{\text{horizon}}$$

where SOC is the organic C content (g kg⁻¹), d the thickness of the control section (m), δ_{2mm} is the ratio of gravel larger than 2 mm in size in the soil, and BD the soil bulk density (Mg m⁻³). Total SOCS (T-SOCS), refer to the entire soil profile thickness (Mg ha⁻¹), was calculated according to IPCC (2003) as follows:

$$T - SOCS = \sum_{Soilsection} SOCstock_{Soilsection}$$

Statistical analysis

Data were subjected to the Anderson-Darling normality test which evaluate the normal distribution of analyzed variables. Soil properties were statistically analyzed for each of the soil types, including the mean values, standard deviations (SD), and correlation for each soil property investigated in order to characterize their general trends. A Spearman rank correlation coefficient was carried out in order to assess the possible connection between the soil properties and SOC in each site. The statistical significance of the differences in each soil variable was identify using the Kruskal-Wallis statistical test. Significance was considered at p < 0.05. Linear regression analyses were carried out to evaluate the relationship between geochemical elements, altitude, and T-SOCS. For the regression analysis, only the mean values of each site were used. All analyses were performed using XSLSAT 2008.6.03 software for Excel.

Results

Physicochemical characteristics of soils

The texture of the soil was sandy loam to loam sandy. Sand is the most important particle size fraction. With the exception of Luvisols in which there is an increase in clay content from the surface $(7.23 \pm 1.15\%)$ to the middle part of the soil profile $(20.23 \pm 1.52\%)$, then a decrease to the

base of the soil profile $(8.23 \pm 1.52\%)$, all the particle size fractions show a zigzag evolution with depth. The highest contents of sand were observed in the Arenosols at Zamai $(80.33 \pm 12.01\%)$ while the lowest were observed in the Luvisols at Sir $(59.03 \pm 0.77\%)$ (Table 2). The silt and clay contents are low and vary generally within the same ranges, with 7.56 ± 4.72 to $22.36 \pm 5.22\%$ for silt and 8.10 ± 5.56 to $22.6 \pm 9.19\%$ for clay. Soil pH is weakly acidic in Regosols and Luvisols (6.10 ± 0.17 to 6.90 ± 0.26), but weakly acidic to neutral in Arenosols and Leptosols $(6.3 \pm 0.00 \text{ to})$ 7.55 ± 0.91) (Table 2). The BD varies between 0.75 ± 0.00 and 2.33 ± 0.22 g/cm³. It increases with depth in Regosols, decreases from the surface to the middle zone before increasing thereafter to the base of the soils in Luvisols, but presents a zigzag evolution in Leptosols and Arenosols. A weak and low significant positive correlation exist between bulk density and sand (r = 0.472, p < 0.05), and negative significant correlation with silt (r = -0.418, p < 0.05) and nitrogen (r = -0.482, p < 0.05). The gravel contents vary greatly at the scale of the massif. The lowest content is observed in Arenosols $(3.66 \pm 3.05\%)$ while the highest is perceived in Leptosols $(25.00 \pm 0.00\%)$. A weak positive and significant correlation exist between gravel and TN (r = -0.348, p < 0.05). Overall, with the exception of silt and pH which values show significant differences, no significant difference was noted between other physical parameters of the studied soil in the entire massif (Table 3). There is a significant and negative correlation between silt and sand (r = -0.674, p < 0.05), clay and sand (r = -0.811, p < 0.05), and between pH and clay (r = -0.306, p < 0.05). In addition, there is a weak significant and positive correlation between pH and sand (r=0.312, p<0.05) (Table 4).

SOM, TN, and C:N ratio

In each soil type, the SOM content is high in the first 25 cm, and there is no significant statistic difference between layers (H(2)=3.563; p=0.515) (Table 2). The highest content is observed in Leptosols $(6.09 \pm 1.96\%)$ while the lowest is observed in Arenosols $(2.33 \pm 0.81\%)$. These contents generally decrease with depth, but their evolution varies from one soil type to another. In the Regosols, Luvisols, and Arenosols, there is a decrease from the surface to the middle part of the soil profiles, then an increase down to the base of the soils (Table 2). No significant difference is noted in the variations of SOM content at the scale of the Mount Mandara (H(3)=7.815;p=0.632) (Table 3). There is a significant correlation between SOM and silt (r = -0.406, p < 0.05) (Table 4). The TN contents are very low in the studied soils. They vary between 0.12 ± 0.01 in Leptosols and $0.06 \pm 0.02\%$ in Luvisols (Table 2). There is a significant difference between the TN contents in Arenosols at Zamaï, Leptosols at Kossohone and Luvisols at Sir. However, no significant difference was noted between the TN contents

Site	Section	Detph (cm)	Sand %	Silt %	Clay %	0M %	N %	C/N	рН	$BD g cm^{-3}$	Gravel %
Rhumsiki (Regosols)	S1	0–25	73.33±5.07a	16.6±5.84a	10.06±0.76a	4.31±3.73a	0.08±0.1a	28.51±26.79a	6.16±0.30a	1.6±1.20a	21.66±7.63a
	S2	25-50	$71.66 \pm 2.62a$	9.93±3.44a	$18.4 \pm 3.96a$	$1.11 \pm 1.03a$	$0.02 \pm 0.0a$	$27.94 \pm 20.77a$	$6.20 \pm 0.52a$	$1.96 \pm 0.62a$	7.66±6.42a
	$\mathbf{S3}$	50-75	$71.33 \pm 6.67a$	$11.93 \pm 3.10a$	16.73±4.19a	$0.83 \pm 0.22a$	$0.01 \pm 0.0a$	$45.88\pm6.80a$	6.30 ± 0.30 a	$2.04\pm0.89a$	$5 \pm 2.00a$
	$\mathbf{S4}$	75-100	$78.33 \pm 2.28a$	$10.6 \pm 1.38a$	$11.06 \pm 2.02a$	1.19±0.55a	$0.01\pm0.0a$	68.64±38.79a	$6.53 \pm 0.23a$	2.33±0.22a	5±2.00a
Sir (Luvisols)	S1	0-25	$74.7 \pm 1.00a$	$18.06 \pm 0.57a$	$7.23 \pm 1.15a$	$3.83 \pm 0.70a$	$0.06 \pm 0.02a$	44.50±19.69a	$6.83 \pm 0.25a$	$1.36 \pm 0.22a$	18.33±5.77a
	S2	25-50	$63.03 \pm 5.77a$	$17.4 \pm 1.00a$	19.56±4.93a	$1.37 \pm 1.02a$	$0.03 \pm 0.02a$	$21.18 \pm 5.88a$	6.10±0.17a	$1.33 \pm 0.55a$	4.33±2.30a
	S3	50-75	$59.03 \pm 0.77a$	$20.73 \pm 1.15a$	$20.23 \pm 1.52a$	$1.03 \pm 0.90a$	$0.02 \pm 0.00a$	$27.77 \pm 28.11a$	$6.26 \pm 0.40a$	$1.10 \pm 0.05a$	9.66±13.31a
	$\mathbf{S4}$	75-100	$71.03 \pm 8.62a$	$18.4 \pm 8.88a$	$10.56 \pm 1.15a$	$1.66 \pm 0.22a$	$0.021 \pm 0.01a$	$56.42 \pm 34.85a$	$6.66 \pm 0.15a$	$1.31 \pm 0.30a$	$9.00 \pm 13.85a$
	S5	100 - 125	$78.36 \pm 2.30a$	$13.4 \pm 1.00a$	$8.23 \pm 1.52a$	$1.65 \pm 1.36a$	$0.01 \pm 0.01a$	55.55±30.76a	$6.90 \pm 0.26a$	$1.49\pm0.58a$	$9.00 \pm 13.85a$
Koshon (Leptosols)	S1	0–25	$69.2 \pm 10.22a$	$22.36 \pm 5.22a$	$8.43 \pm 5.13a$	6.09 ± 1.96a	$0.12 \pm 0.01a$	$30.36 \pm 12.82a$	$6.53 \pm 0.20a$	$0.88\pm0.25a$	9.00±5.29a
	S2	25-50	$60.33 \pm 20.00a$	$20.9 \pm 6.08a$	$18.76 \pm 14.84a$	1.31 ± 1.44a	$0.05 \pm 0.03a$	10.99±9.94a	$6.76\pm0.15a$	$1.01 \pm 0.19a$	$16.66 \pm 10.40a$
	S3	50-75	$68.13 \pm 15.49a$	$13.43\pm6.51a$	18.43 ± 13.01	$1.76 \pm 0.40a$	$0.06 \pm 0.02a$	17.27±10.30a	$6.76 \pm 0.05a$	$1.09\pm0.54a$	$25.00 \pm 5.00a$
	S4	75–100	$61.9 \pm 4.38a$	$15.5 \pm 4.80a$	$22.6\pm9.19a$	1.10±0.06a	$0.07 \pm 0.04a$	10.63±9.96a	$7.55 \pm 0.91a$	$0.99 \pm 0.06a$	$13.00 \pm 16.97a$
	S5	100-125	$69.00 \pm 0.00a$	$18.9 \pm 0.00a$	$12.1 \pm 0.00a$	$1.57 \pm 0.00a$	$0.08 \pm 0.00a$	$10.40 \pm 0.00a$	$6.3 \pm 0.00a$	$0.75\pm0.00a$	$25.00 \pm 0.00a$
	S6	125-150	$70.00 \pm 0.00a$	$18.9 \pm 0.00a$	$11.1 \pm 0.00a$	$1.23 \pm 0.00a$	$0.02 \pm 0.00a$	$29.13 \pm 0.00a$	$7.0 \pm 0.00a$	$0.96 \pm 0.00a$	$5.00 \pm 0.00a$
Zamai (Arenosols)	S1	0–25	$75.33 \pm 13.61a$	$14.9 \pm 9.84a$	$9.76 \pm 3.78a$	2.33±0.81a	$0.07 \pm 0.01 a$	19.18±5.66a	$6.93\pm0.56a$	$1.24 \pm 0.37a$	7.66±7.02a
	S2	25-50	$83.0 \pm 10.58a$	$8.90 \pm 5.19a$	$8.10 \pm 5.56a$	1.96±0.15a	$0.07 \pm 0.02a$	19.71±14.14a	$6.83\pm0.85a$	$1.26\pm0.44a$	$11.0 \pm 12.49a$
	S3	50-75	$80.33 \pm 12.01a$	7.56±4.72a	$12.1\pm7.550a$	$0.86 \pm 0.81a$	$0.03 \pm 0.02a$	$12.09 \pm 9.34a$	$7.1 \pm 1.25 a$	$1.27 \pm 0.08a$	11.0 ± 2.49
	$\mathbf{S4}$	75-100	$70.56 \pm 8.50a$	$11.3 \pm 5.15a$	$18.13 \pm 9.30a$	$1.62 \pm 0.24a$	$0.04 \pm 0.05a$	$27.71 \pm 18.99a$	$7.1 \pm 1.21a$	$1.12 \pm 0.37a$	$10.66\pm12.50a$
	S5	100-125	$69.56 \pm 20.15a$	$13.73 \pm 12.94a$	$16.7 \pm 9.18a$	$2.61 \pm 2.43a$	$0.03 \pm 0.00a$	32.00±23.19a	7.13±1.06a	$1.38 \pm 0.79a$	3.66±3.05a
<i>BD</i> bulk density, <i>OM</i> considering the same t	organic m topograph	natter, SD stanc ic position	dard deviation. Nu	umbers followed	by different lower	rcase letters wit	thin the same co	olumn have signifi	icant difference	ss (<i>p</i> < 0.05) at	different depths,

Table 2Physical and chemical properties of the soils evaluated (mean \pm SD) in Mount Mandara

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Table 3Mean soilcharacteristics (mean \pm SD)variation along the MountMandara

Variables	Sites								
	Zamai (Arenosols)	Kossohone (Leptosols)	Sir (Luvisols)	Rhumsiki (Regosols)					
Sand %	75.76±12.69a	65.83±11.93a	69.23 ± 8.47a	73.66±4.85a					
Silt %	11.28±7.42b	18.37±5.72b	17.6±6.18a	$12.26 \pm 4.22a$					
Clay %	12.96±7.42a	15.79±10.22a	13.16±6.18a	$14.06 \pm 4.55a$					
OC %	0.97 ± 0.62a	1.45±1.33a	1.10±0.74a	$1.08 \pm 1.29a$					
OM %	1.88 ± 1.29a	$2.50 \pm 2.29a$	1.91 ± 1.28a	$1.86 \pm 2.23a$					
N %	$0.05 \pm 0.02b$	0.07 ± 0.03 ab	$0.02 \pm 0.03a$	$0.03 \pm 0.03a$					
C/N	22.12±15.05a	$18.20 \pm 11.71a$	$41.09 \pm 26.50a$	$42.74 \pm 28.13a$					
рН	$7.02 \pm 0.85b$	$6.81 \pm 0.46b$	6.55 ± 0.39 ab	$6.3 \pm 0.34a$					
SCOS	$25.72 \pm 18.35a$	$31.53 \pm 26.64a$	$31.80 \pm 21.58a$	$36.20 \pm 24.70a$					

SOCS soil organic carbon stock, SD standard deviation. Numbers followed by different lowercase letters within the same column have significant differences (p < 0.05) at different depths, considering the same topographic position

Table 4 Spearman correlation matrix for relationships between selected soil parameters in the study area

Variables	Sand	Silt	Clay	pH	ОМ	N	C/N	BD	Gravel	SOCS
Sand	1									
Silt	-0.674	1								
Clay	-0.811	0.197	1							
pН	0.312	-0.163	-0.306	1						
OM	-0.133	0.406	-0.185	-0.059	1					
N	-0.090	0.280	-0.217	-0.003	0.525	1				
C/N	0.002	0.052	-0.055	-0.127	0.488	-0.377	1			
BD	0.472	-0.418	-0.245	-0.029	-0.180	-0.482	0.261	1		
Gravel	0.089	0.078	-0.186	0.186	0.163	0.348	-0.150	0.033	1	
SOCS	0.043	0.259	- 0.298	-0.177	0.859	0.332	0.591	0.227	0.149	1

SOCS soil organic carbon stock. Values in bold are different from 0 with a significance level alpha=0.05

at Sir and Rhumsiki sites (Table 3). There is a weak positive and significant correlation between TN and silt (r = -0.280, p < 0.05) and between TN and SOM (r = -0.525, p < 0.05). The C/N ratios are very high. They vary between 12.09 ± 9.34 and 32.00 ± 23.19 in Arenosols at 608 m a.s.l., 10.40 ± 0.00 and 30.36 ± 12.82 in Leptosols at 865 m a.s.l., 21.18 ± 5.88 and 56.42 ± 34.85 in Luvisols at 970 m a.s.l., and between 27.94 ± 20.77 and 68.64 ± 38.79 in Regosols at 1050 m a.s.l. (Table 2). No significant difference is noted between the values of this ratio either as a function of the depth in each soil type or at the scale of the massif (Table 3). There is a weak significant and positive correlation between the C/N ratio and SOM (r = -0.488, p < 0.05) and negative between the C/N ratio and TN (r = -0.377, p < 0.05) (Table 4).

Evolution of SOC and SOCS contents in the studied soils

SOC always presents the highest contents in the upper part of the soils, in the 0-25-cm interval. The highest content

is observed in Leptosols $(35.36 \pm 11.39 \text{ g kg}^{-1})$ while the lowest is noted in Arenosols $(13.54 \pm 4.69 \text{ g kg}^{-1})$. The evolution of the SOC content is modeled on that of the SOM (Table 2). The highest total SOC (T-SOC) value is obtained in Leptosols at 865 m a.s.l. $(75.89 \pm 3.74 \text{ g kg}^{-1})$, but the lowest is obtained in Regosols at 1050 m a.s.l. $(43.27 \pm 8.03 \text{ g kg}^{-1})$, which are entirely cultivated (Table 5). In general, the T-SOC content increases from Zamai to Kossohone, then decreases up to the summit of the massif. As for the SOCS, its evolution with depth is almost reflect that of SOC content, varying between 13.67 ± 5.65 and 38.40 ± 18.01 g kg⁻¹ in Arenosols at the base of the massif, 12.81 ± 0.0 and 69.64 ± 15.90 g kg⁻¹ in Leptosols at 865 m a.s.l., 15.13 ± 14.88 , and 60.46 ± 3.41 g kg⁻¹ in Luvisols at 970 m a.s.l., and between 23.27 ± 12.56 and 53.063 ± 31.14 g kg⁻¹ in Regosols at 1050 m a.s.l. (Table 5). Overall, there were no significant differences in SOC and SOCS contents both as function of depth and at the scale of the massif (Tables 3 and 5). There is a weak negative and significant correlation between SOCS and clay (r = -0.298,

Table 5Soil organic carbon content and soil organic carbon stock (mean \pm SD) in the study area

Site	Section	Depth cm	SOC g·kg ⁻¹	T-SOC $g\cdot kg^{-1}$	SOCS Mg·ha ⁻¹	T-SOCS Mg·ha ⁻¹
Rhumsiki (Regosols)	S1	0–25	$25.05 \pm 21.63a$	43.27 ± 8.03	$53.063 \pm 31.14a$	144.79 ± 24.23
	S2	25-50	6.43±5.98a		28.916±31.03a	
	S 3	50-75	$4.84 \pm 1.28a$		23.27 ± 12.56a	
	S4	75-100	$6.95 \pm 3.23a$		39.55 ± 22.19 a	
Sir (Luvisols)	S 1	0–25	$22.23 \pm 4.10a$	55.45 ± 4.9	$60.46 \pm 3.41a$	158.99 ± 13.25
	S2	25-50	7.96±5.93a		$20.33 \pm 11.21a$	
	S 3	50-75	$6.01 \pm 5.25a$		15.13 ± 14.88a	
	S4	75-100	9.63±1.28a		$28.13 \pm 5.21a$	
	S5	100-125	9.62±7.93a		34.94±31.54a	
Kossohone (Leptosols)	S 1	0–25	35.36±11.39a	75.89 ± 3.74	69.64±15.90a	158.248 ± 10.52
	S2	25-50	7.63 ± 8.39a		$20.68 \pm 23.85a$	
	S 3	50-75	$10.25 \pm 2.32a$		28.74±19.77a	
	S4	75-100	$6.4 \pm 0.35a$		13.51±3.61a	
	S5	100-125	9.11±0.0a		$12.81 \pm 0.0a$	
	S6	125-150	$7.14 \pm 0.0a$		$12.85 \pm 0.0a$	
Zamai (Arenosols)	S1	0–25	13.54±4.69a	48.90 ± 6.12	$38.40 \pm 18.01a$	128.63 ± 5.25
	S2	25-50	$11.40 \pm 6.72a$		$35.43 \pm 26.43a$	
	S 3	50-75	$5.0 \pm 5.53a$		$13.67 \pm 5.65a$	
	S4	75-100	$7.10 \pm 4.90a$		15.88±12.69a	
	S5	100–125	11.84±8.78a		$25.23 \pm 13.49a$	

SOC soil organic carbon, *T-SOC* total SOC, SOCS soil organic carbon stock, *T-SOCS* total SOCS, SD standard deviation. Numbers followed by different lowercase letters within the same column have significant differences (p < 0.05) at different depths, considering the same topographic position

p < 0.05) but a significant and weak positive correlation with TN (r = 0.332, p < 0.05) and the C/N ratio (r = -0.591, p < 0.05), and strong negative correlation with SOM (r = -0.859, p < 0.05) (Table 4). The T-SOCS contents is quite important. It is 128.63 ± 5.25 Mg ha⁻¹ in Arenosols at the base of the massif, 158.248 ± 10.52 Mg ha⁻¹ in Leptosols at Kossohone (865 m a.s.l.), 158.99 ± 13.25 Mg ha⁻¹ in Luvisols at Sir (970 m a.s.l.), and drops down to 144.79 ± 24.23 Mg ha⁻¹ in Regosols at Rhumsiki (1050 m a.s.l.) (Table 5), with a marginal significant difference between Zamai and Sir. There is a strong correlation between altitude and T-SOCS (r = 0.70) (Fig. 4).

Mineralogical and geochemical characteristics of soils

The mineralogical composition of the studied soils slightly varied along the Mount Mandara slope. The main secondary minerals are smectite, kaolinite, sepiolite, calcite, lepidocrocite, and hematite (Table 1). They are associated with quartz, plagioclase, and alkaline feldspars. Sepiolite is observed in the summit of the interfluve. Lepidocrocite on contrary is observed in the mid- and footslope. Except the Kossohone site, calcite is present in all the three other sites. Also, hematite was not detected in Sir. In the X-ray diffractions, quartz is identified by peaks at 4.25 Å and 3.34 Å, kaolinite by peaks at 7.14 Å and 3.58 Å, plagioclase at 3.18 Å, alkaline feldspar at 3.24 Å, calcite at 3.03 Å, hematite at 2.75 Å, and smectites at 9.81 Å and 4.48 Å. In the infrared spectra, smectites are detected by his peak at 1645.58 cm⁻¹, sepiolite by peak at 527.2 cm⁻¹, and quartz by peaks at 1000.34 cm⁻¹, 461.12 cm⁻¹, and 417.23 cm⁻¹. Kaolinite which the characteristic absorption band ranges between 3600 and 3700 cm⁻¹ was identified at 3695.34 and 907.81 cm⁻¹. Lepidocrocite was detected at 741.02 cm.⁻¹

The geochemistry of the various soil samples at the scale of the whole slope is dominated by silicon. Its content oscillates between 50.05 and 73.13% (Fig. 2). Apart from silicon, only aluminum and iron are significantly expressed and have contents range respectively between 13.56 and 23.60% Al₂O₃ and 1.88 and 8.73% Fe₂O₃. Potassium contents varied between 3.35 and 7.78% K₂O. The other alkalis and alkaline-earths contents are low, almost below 2%. The other major elements contents such as TiO₂, P₂O₅, and MnO are almost below 0.4% (Fig. 2). The Si/Al ratios varied between 2.00 and 4.6. This confirms the presence of phyllosilicates and quartz minerals already noted in the mineralogical analysis. The loss on ignition is between 2.33 and 13.71%, and might characterize the destruction of SOM in the studied sample, the dehydration, and decarbonation reactions.

Fig. 2 Box plots showing the summary of major elements characteristics in the studied soils along the Mount Mandara slope



Discussion

Soil properties in the Mount Mandara

The studied area is composed of Arenosols, Leptosols, Luvisols, and Regosols. These soils are characterized by the sandy loam and loam sandy texture, as consequence of the large predominance of the sandy fraction which contents range between $59.03 \pm 0.77\%$ in Luvisols at Sir and $80.33 \pm 12.01\%$ in Arenosols at Zamai. The predominance of sandy texture is related to the nature of the bedrocks, constituted of sandy alluvial materials, gneiss, and granite which the mineralogy is mainly dominated by quartz mineral. pH is weakly acidic in Regosols and Luvisols $(6.10\pm0.17 \text{ to } 6.90\pm0.26)$ in the upper part of the Mount from 970 to 1050 m a.s.l., but weakly acidic to neutral in Arenosols and Leptosols $(6.3 \pm 0.00 \text{ to } 7.55 \pm 0.91)$ towards the base of the Mount from 600 to 865 m a.s.l., in line with the significant difference noted between the two groups of soil. The low soil pH value observed at high altitude might be due to the strong presence and dissociation of carboxyl functional groups below pH 7 (Boguta and Sokołowska 2020), leading thus to the acidification of the soil solution, as consequence of progressive increase in SOC content. The BD varies between 0.75 ± 0.00 and 2.33 ± 0.22 g/ cm³. High values are observed in Regosols. In Arenosols, Leptosols, and Luvisols, all values are below 1.38 g/cm³. The low values obtained below 970 m a.s.l. are closed to those obtained by Chenchounia and Neffar (2022) in arid and semi-arid steppe rangelands of northeastern Algeria. According to Pieri (1989) and Li et al. (2006), any decrease in BD generally reflects a loosening effect of the plant against a compaction effect. High BD values at the summit of the massif might be due to their regosolic nature, in line with the significant correlation between the BD and the sand fraction. However, a low value was obtained in the topsoil $(1.6 \pm 1.20 \text{ g/cm}^3)$ compared to the subsoil $(2.33 \pm 0.22 \text{ g/} \text{cm}^3)$ in this Regosols. Generally, tillage which is the main management system in the upper part of the Mount Mandara is known to reduce BD in topsoils (Strudley et al. 2008; Palm et al. 2014).

N, SOC, and SOCS in the Mount Mandara

SOM is a key component of any terrestrial ecosystem, and any variation in its amount and composition has important effects on many of the processes that occur within the system (Batjes 2014). In different studied soil types, its content is high in the first 25 cm (2.33 ± 0.81) to $6.09 \pm 1.96\%$), in line with observations already made by Plaza et al. (2018). These contents generally decrease with depth, but their evolution varies from one soil type to another (Batjes 1996; Parras-Alcántara et al. 2015; da Silva et al. 2019). A significant correlation was noted between SOM and silt. This means that a part of SOM in the studied soil is linked to the silt fraction. This might be due to the existence of expansive 2:1 clay minerals in the silt fraction (Iturri and Buschiazzo 2014; Liu et al. 2020). It is well known that expansive 2:1 clay minerals (e.g., smectite) have larger specific surface area (Yukselen-Aksoy and Kaya 2010; Tang et al. 2015; Liu et al. 2020). The TN contents are very low in the studied soils. It varies between 0.06 ± 0.02 and $0.12 \pm 0.01\%$. Generally, the concentration of TN increases with SOM and SOC level, indicating that TN nutrition of plants greatly depends on the maintenance of SOM and SOC level (Sakin 2012; Tsozué et al. 2021). The relationship between TN and SOM is confirmed by the significant positive correlation between TN and SOM obtained from the Spearman correlation matrix between soil parameters in the studied area. Here, TN content is very low compared to that of SOM, leading thus to high C/N ratios which reach 68.64 ± 38.79 in Regosols in the upper part of the Mount Mandara, in line with the significant and positive correlation between the C/N ratio and SOM. The correlation between TN and SOC indicates that these two soil components act together along the Mount Mandara slope, contrary to observations made in the latitudinal variation of SOCS by Tsozué et al. (2021). According to Batjes and Dijkshoorn (1999), a C/N ratio above 12-14 is often considered indicative for a shortage of TN in the studied soil. The decrease in C/N ratio with depth in subsoil reflects a greater degree of breakdown and older age of the humus stored in the lower parts of the soil profile (Batjes 2014). The increase of C/N ratio thereafter until the base of each soil type might indicate the accumulation of this humus in this part of the soil profile. As for the SOCS, its evolution with is almost modeled on that of soil organic carbon content. This is confirmed by the strong negative correlation (r = -0.859, p < 0.05) between soil SOCS and soil SOM as already observed by many authors (Tsozué et al. 2019, 2021). There is a weak negative and significant correlation between SOCS and clay. This confirmed the importance of clay stabilization mechanisms on SOC in the studied soils (Parras-Alcántara et al. 2015). The T-SOCS contents is 128.63 ± 5.25 Mg ha⁻¹ in Arenosols, 158.248 ± 10.52 Mg ha⁻¹ in Leptosols, 158.99 ± 13.25 Mg ha⁻¹ in Luvisols, and drops down to 144.79 ± 24.23 Mg ha⁻¹ in Regosols. The T-SOCS values obtained here in Leptosols and Regosols are higher than those obtained by De Vos et al. (2015) and Parras-Alcántara et al. (2015) in Leptosols and Regosols $(53.8 \pm 18.3 \text{ to } 158.0 \pm 15.8 \text{ Mg ha}^{-1})$ in the Despeñaperros Natural Park under temperate semi-arid climate in southern Spain, in similar altitude gradient interval (607 to 1168 m a.s.l.). The high T-SOCS in the dry tropical mountainous zone where the decomposition rate is naturally low in the dry season and increases in the wet season might be due to the quality of the organic matter (OM), the soil microbial community, and particularly the soil mineralogy. It might also be due to more productive vegetation as consequence of the high precipitation in the study area (Tsozué et al. 2021). The T-SOCS obtained in Arenosols in the study area is also higher than that document by De Vos et al. (2015) in Europe (102 Mg ha⁻¹). As for Luvisols, the T-SOCS obtained in the studied area (158.99 \pm 13.25 Mg ha⁻¹) is similar to that obtained under semi-arid climate in Tunisia (109.2 Mg ha⁻¹), situated in north of Africa and in south of Mediterranean Sea, but in the 0–100-cm depth (Brahim et al. 2014). Those T-SOCS values are globally below those obtained in the humid tropical mountainous zone in soils developed on volcanic rocks where T-SOCS values reached 302 Mg ha⁻¹ (Tsozué et al. 2019).

Effect of soil mineralogy on SOCS

The studied soils contain several clay minerals, which include smectite, kaolinite, and sepiolite. Generally, fractions rich in kaolinite often showed less C contents and the smectite-rich fractions contain organic C within a wide (Manjaiah et al. 2010). The presence of kaolinite thus indicates the presence of low reactive clay-sized silicates which offer fewer binding site to SOC contrary to smectite which indicate the presence of high reactive clay-sized silicates (White et al. 2005; Manjaiah et al. 2010). However, both 2:1 and 1:1 clay layers contain pH-dependent charges corresponding to surface charge generated by protonation-deprotonation reactions of surface hydroxyl groups (Singh et al. 2017). Soils rich in 2:1 clay minerals, for example, smectite and sepiolite, have a greater ability to protect the C in soil than 1:1 clay minerals such as kaolinites (Six et al. 2002; Ibrahim et al. 2022), due to their higher specific surface area and cation exchange capacity (Hassink 1997; Wattel-Koekkoek et al. 2001). In addition to clay minerals, hematite and lepidocrocite are present in the studied soils. These iron oxides are reactive constituents (Singh et al. 2017). Many works show that they adsorbed more SOC in comparison to phyllosilicate clay minerals such as kaolinite and smectite (Kaiser and Guggenberger 2003; Tombácz et al. 2004; Singh et al. 2017). Clay minerals and iron oxides are good receptacle for SOC and might constitute a major asset for the accumulation and the preservation of SOC leading to thus an increase of SOCS along the Mount Mandara slope.

Effect of soil geochemistry on SOCS

Iron and aluminum are the main metal-(oxyhydr)oxides in the studied soils. Their contents range respectively between 13.56 and 23.60% Al_2O_3 and 1.88 and 8.73% Fe_2O_3 . Under oxidizing conditions, these oxides play an important role in soil sorption capacity, as they usually occur as coatings on clay particles (Rieuwerts et al. 1998). It is demonstrated that the formation of oxides and oxyhydroxides of Al and Fe might play a role in the occlusion of SOC at the level of clay microstructure (Mayer et al. 2004; Reichenbach et al. 2021). When these oxides are exposed to water, their surfaces become hydroxylated, and consequently, they have variable charge (Ngole-Jeme 2019). This might increase the soil sorption capacity and subsequently the global aptitude of soil to sorb SOC. These characteristics are thus favorable to the sequestration and stabilization of SOC resulting in an increase of SOCS in the studied area (Reichenbach et al. 2021). They could be anionic, neutral, or cationic depending on the degree of protonation they undergo, which in itself is determined by the prevailing pH conditions (Ngole-Jeme 2019). They have a net positive surface charge at low pH, a net negative surface charge at high pH, and are neutral at circumneutral pH conditions (Hyun et al. 2003). These conditions are met in the study area where the pH are generally weakly acidic to neutral. The contribution of iron and aluminum oxides and even the other major oxides to the accumulation and stabilization of SOCS in the studied soils is very low (<7%) (Fig. 3). Except SiO₂ and Fe₂O₃ whose effects on SOCS are respectively 5.92 and 6.49%, the effects of the other major oxides are less or equal to 3.60% (Fig. 3).





Effect of land use and environmental factors on SOCS

Altitudinal change is very common in mountainous environments and this change influence the microclimate, which generally affect the SOC content (Parras-Alcántara et al. 2015; Chang et al. 2016; Tsozué et al. 2019). In the studied area, T-SOCS increase with altitudinal gradient from 128.63 ± 5.25 Mg ha⁻¹ in Arenosols at the base of the massif to 158.99 ± 13.25 Mg ha⁻¹ in Luvisols at 970 m a.s.l. Increases in elevation lead to decreases in the annual temperature as observed in the studied area, which affect microbial activity. This leads to a slow rate of SOM decomposition, which thereby affected SOCS (Trumbore et al. 1996; Garten 2004; Garten and Hanson 2006). The decrease in temperature occurs simultaneously to an increase in precipitation in the studied area. These two climate components, temperature and precipitation, are important factors affecting C sequestration in soil because both play important roles in biomass production (Singh et al. 2017). However, less microbial activity occurs in dry soils as in the Mount Mandara, and greater decomposition of SOC is observed in wetter soils, provided there is good movement of air (Chang et al. 2016). Altitudinal gradient thus affects the climatic characteristics of the studied area which in turn affect the biomass production and consequently the T-SOCS. This is confirmed by the significant correlation between altitudinal gradient and T-SOCS (r=0.70), with altitude contributing to the accumulation of SOC for 49.68% (Fig. 4). Moreover, since the climatic characteristics of Zamai and Sir are not contrasting enough, it can be assumed that the climatic characteristics in response to changes in altitude are not the only principal factor of interest at these sites, the soil characteristics might be another main factor of greater importance affecting the C sequestration.



Fig. 4 Plots of the total soil organic carbon stock (T-SOCS) versus altitudinal gradient in Mount Mandara

Among soil characteristics which affect C sequestration, soil texture plays a central role in C sequestration (Singh et al. 2017). This is confirmed along the altitudinal gradient by the significant and positive correlation between silt fraction and SOM (r=0.406, p<0.05). It is well known that mineralization of C is highly expressed in coarse-textured soils compared to fine-textured soils (Hassink 1992; Singh et al. 2017). These results could be due to the greater ability of fine-textured soils to physically protect the SOC from microbial attack (Hassink 1995; Singh et al. 2017). It also plays an important role in C stabilization in the lower soil depths under all land uses (Fontaine et al. 2007; Albaladejo et al. 2013; Singh et al. 2017). In fact, SOC is not easily accessible to soil microbes in clay dominant soils compared with sandy soils. This is due to the chemical adsorption of C on clay minerals surfaces and also its physical occlusion within soil microaggregates (Sissoko and Kpomblekou-A 2010; Singh et al. 2017).

Under Regosols in Rhumsiki at 1050 m a.s.l., there is a decrease in T-SOCS value respectively reaching 144.79 ± 24.23 Mg ha⁻¹. All this upper part of the massif is intensely cultivated. Land use management is known to have a strong influence on SOC (Muñoz-Rojas et al. 2012; Schlüter et al. 2022). It can affect C mineralization through changes in thermal properties, water retention and consumption, and through biomass production (Schlüter et al. 2022). Conversion of natural vegetation to cropland as observed in Regosols in the upper part of the Mount Mandara is assumed to cause a decrease of SOM content (Degryze et al. 2004; Gerzabek et al. 2005). In addition, ploughing is the main management system in the upper part of the Mount Mandara. Ploughing was found to be the management that resulted in the lowest SOM levels worldwide due to the lack of vegetation and increased C emissions as a consequence of repeated ploughing (Lozano-Garcia and Parras-Alcantara 2014a; Carr et al. 2015; de Moraes et al. 2015; Parras-Alcantara et al. 2016; Tsozué et al. 2019). It might also be due to a decrease in the amount of OM returned to the soil and more rapid SOM decomposition, as a consequence of the intensification of ploughing (Dalal et al. 1991; Nweke and Nnabude 2014).

Conclusion

The present work studied the properties of a soil sequence from a dry tropical zone in Mount Mandara, Cameroon. The study area is composed of Arenosols, Leptosols, Luvisols, and Regosols. They are sandy loam to loam sandy. Soil pH is weakly acidic in Regosols and Luvisols, but weakly acidic to neutral in Arenosols and Leptosols. The C/N ratios are very high, ranging between 12.09 ± 9.34 and 68.64 ± 38.79 . The T-SOCS contents are quite substantial. It is 128.63 ± 5.25 Mg ha⁻¹ in Arenosols at the base of the massif, 158.248 ± 10.52 Mg ha⁻¹ in Leptosols at Kossohone $(865 \text{ m a.s.l.}), 158.99 \pm 13.25 \text{ Mg ha}^{-1}$ in Luvisols at Sir (970 m a.s.l.), and drops down to 144.79 ± 24.23 Mg ha⁻¹ in Regosols at Rhumsiki (1050 m a.s.l.). There is a strong correlation between altitude and T-SOCS (r=0.70). The main secondary minerals are clay minerals (smectite, kaolinite, sepiolite) and iron oxides (lepidocrocite, hematite). Clay minerals and iron oxides are good receptacle for SOC and might constitute a major asset for the accumulation and the preservation of SOC leading to thus an increase of SOCS along the Mount Mandara slope. Altitudinal gradient affects the climatic characteristics of the studied area which in turn affect the biomass production and consequently the T-SOCS. This is confirmed by the significant correlation between altitudinal gradient and T-SOCS (r=0.70), with altitude contributing to the accumulation of SOC for 49.68%. Among the soil factors, texture plays a central role in C sequestration, confirmed in the studied area by the significant and positive correlation between silt fraction and SOM. Under Regosols in Rhumsiki at 1050 m a.s.l., there is a decrease in T-SOCS value $(144.79 \pm 24.23 \text{ Mg ha}^{-1})$ as a consequence of a decrease in the amount of OM returned to the soil and more rapid SOM decomposition due to ploughing in the upper part of the Mount Mandara. Therefore, cultivation has a negative consequence on SOCS, and if this continues at the other sites, it will increase greenhouse gas emissions contributing to global climate change. This research provides a preliminary assessment for SOCS at Mount Mandara. It suggests that altitudinal gradient, land use, and soil characteristics should be included in SOCS models and estimations at local and regional scales.

Declarations

Competing interests The authors declare no competing interests.

References

- Albaladejo J, Ortiz R, Garcia-Franco N, Navarro AR, Almagro M, Pintado JG, Martínez-Mena M (2013) Land use and climate change impacts on soil organic carbon stocks in semi-arid Spain. J Soil Sedim 13:265–277. https://doi.org/10.1007/ s11368-012-0617-7
- Amundson R (2014) Soil formation. Treatise on Geochemistry, 2nd Edition, University of California, Berkeley, CA, USA, Elsevier, 1–26. https://doi.org/10.1016/B978-0-08-095975-7.0050-5
- Batjes NH (1996) Total carbon and nitrogen in the soils of the world. Eur J Soil Sci 47:151–163. https://doi.org/10.1111/j.1365-2389. 1996.tb01386.x
- Batjes NH (2014) Total carbon and nitrogen in the soils of the world. Eur J Soil Sci 65:4–21. https://doi.org/10.1111/ejss.12114_2
- Batjes NH, Dijkshoorn JA (1999) Carbon and nitrogen stocks in the soils of the Amazon Region. Geoderma 89:273–286. https://doi. org/10.1016/S0016-7061(98)00086-X

- Beckhoff B, Kanngießer B, Langhoff N, Wedell R, Wolff H (2006) Handbook of Practical X-Ray Fluorescence Analysis. Springer
- Bernoux M, Chevallier T (2013) Le carbone dans les sols des zones sèches. Des fonctions multiples indispensables, Les dossiers thématiques du CSFD, No. 10, CSFD/Agropolis International, Montpellier France
- Boguta P, Sokołowska Z (2020) Zinc binding to fulvic acids: assessing the impact of pH, metal concentrations and chemical properties of fulvic acids on the mechanism and stability of formed soluble complexes. Molecules 25:1297. https://doi.org/10.3390/molec ules25061297
- Boudjabi S, Chenchouni H (2022) Soil fertility indicators and soil stoichiometry in semi-arid steppe rangelands. Catena 210 (3), 105910. https://doi.org/10.1016/j.catena.2021.105910
- Brahima N, Ibrahima H, Hatira A (2014) Tunisian Soil Organic Carbon Stock - Spatial and Vertical Variation 24th DAAAM International Symposium on Intelligent Manufacturing and Automation. Procedia Engineering 69:1549–1555. https://doi. org/10.1016/j.proeng.2014.03.154
- Bremner JM (1996) Nitrogen total, in: Methods of soil analysis Part 3: Chemical Methods, edited by: Sparks, D. L., SSSA Book Series 5, Soil Science Society of America, Madison, Wisconsin, 1085– 1122. https://doi.org/10.2136/sssabookser5.3.c37
- Bruun TB, Elberling B, de Neergaard A, Magid J (2015) Organic carbon dynamics in different soil types after conversion of forest to agriculture. Land Degrad Dev 26(3):272–283. https://doi.org/10. 1002/ldr.2205
- Carr PM, Brevik EC, Horsley RD, Martin GB (2015) Long-term no-tillage sequesters soil organic carbon in cool semiarid regions. Soil Horizons 56(6). https://doi.org/10.2136/ sh15-07-0016.
- Cerdá A, Morera GA, García OF, Morug_an A, Pelayo GO, Pereira P, Novara A, Brevik EC (2014) The impact of abandonment of traditional flood irrigated citrus orchards on soil infiltration and organic matter. In: Arnaez J, Gonzalez-Samperiz P, Lasanta T, Valero-Garces, BL Geoecología, Cambio Ambiental Y Paisaje: Homenaje Al Profesor Jose María García Ruiz. Instituto Pirenaico de Ecología, Zaragoza, pp. 267–276
- Chang EH, Chen TH, Tian G, Chiu C (2016) The effect of altitudinal gradient on soil microbial community activity and structure in moso bamboo plantations. Appl Soil Ecol 98:213–220. https:// doi.org/10.1186/s40529-016-0123-0
- Chenchounia HB, Neffar S (2022) Soil organic carbon stock in arid and semi-arid steppe rangelands of North Africa. Catena 211:106004. https://doi.org/10.1016/j.catena.2021.106004
- Ciais P, Sabine C, Bala G, Bopp L, Brovkin V, Canadell J, Chhabra A, DeFries R, Galloway J, Heimann M (2013) Carbon and other biogeochemical cycles. In: Climate change 2013: The physical science basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change. Cambridge University Press, pp. 465–570
- Corral-Fernández R, Parras-Alcántara L, Lozano-García B (2013) Stratification ratio of soil organic C, N and C: N in Mediterranean evergreen oak woodland with conventional and organic tillage. Agric Ecosyst Environ 164:252–259. https://doi.org/10.1016/j. agce.2012.11.002
- da Silva LF, Fruett T, Zinn YL, Inda AV, do Nascimento PC, (2019) Genesis, morphology and mineralogy of Planosols developed from different parent materials in southern Brazil. Geoderma 341:46–58. https://doi.org/10.1016/j.geoderma.2018.12.010
- Dalal RC, Henderson PA, Glasby JM (1991) Organic matter and microbial biomass in a vertisols after 20 years of zero tillage. Soil Biol Biochem 23:435–441. https://doi.org/10.1016/0038-0717(91) 90006-6
- de Moraes SJC, Seguy L, Tivet F, Lal R, Bouzinac S, Borszowskei PR, Briedis C, dos Santos JB, da Cruz Hartman D, Bertoloni CG,

Rosa J, Friedrich T (2015) Carbon depletion by plowing and its restoration by no-till cropping systems in oxisols of subtropical and tropical agro-ecoregions in Brazil. Land Degrad. Dev. 26: 531–543. 10.002/ldr.2218

- De Vos B, Cools N, Ilvesniemi H, Vesterdal L, Vanguelova E, Carnicelli S (2015) Benchmark values for forest soil carbon stocks in Europe: results from a large scale forest soil survey. Geoderma 251–252:33–46. https://doi.org/10.1016/j.geoderma. 2015.03.008
- Deb S, Bhadoria PBS, Mandal B, Rakshit A, Singh HB (2015) Soil organic carbon: towards better soil health, productivity and climate change mitigation. Clim Change Environ Sustain 3(1):26–34. https://doi.org/10.5958/2320-642X.2015.00003.4
- Degryze S, Six J, Paustian K, Morris SJ, Paul EA, Merckx R (2004) Soil organic carbon pool changes following land-use conversions. Glob Chang Biol 10:1120–1132. https://doi.org/10.1111/j.1529-8817.2003.00786.x
- FAO (2015) Understanding mountain soils: à contribution from mountain areas to the International Year of Soils. Rome, 169p.
- FAO (2017) Soil organic carbon: the hidden potential. Food and Agriculture Organisation of the United Nations, Rome, 203p
- Fernandez-Romero ML, Lozano-García B, Parras-Alcántara L (2014) Topography and land use change effects on the soil organic carbon stock of forest soils in Mediterranean natural areas. Agric Ecosyst Environ 195:1–9. https://doi.org/10.1016/j.agee.2014. 05.015
- Fiener P, Dlugoβ V, Van Oost K (2015) Erosion-induced carbon redistribution, burial and mineralization – is the episodic nature of erosion process important? CATENA 133:282–292. https:// doi.org/10.1016/j.catena.2015.05.027
- Fissore C, Dalzell BJ, Berhe AA, Vogtle M, Evans M, Wu A (2017) Influence of topography on soil organic dynamic in Southern California. CATENA 149:140–149. https://doi.org/10.1016/j. catena.2016.09.016
- Fontaine S, Barot S, Barre P, Bdioui N, Mary B, Rumpel C (2007) Stability of organic carbon in deep soil layers controlled by fresh carbon supply. Nature 450:277–280. https://doi.org/10.1038/ nature06275
- Garten CT (2004) Potential net soil N mineralization and decomposition of glycine-13C in forest soils along an elevation gradient. Soil Biol Biochem 36:1491–1496. https://doi.org/10. 1016/j.soilbio.2004.04.019
- Garten CT, Hanson PJ (2006) Measured forest soil C stocks and estimated turnover times along an elevation gradient. Geoderma 136:342–352. https://doi.org/10.1016/j.geoderma.2006.03.049
- Gerzabek MH, Strebl F, Tulipan M, Schwarz S (2005) Quantification of organic carbon pools for Austria's agricultural soils using a soil information system. Can J Soil Sci 85:491–498. https://doi. org/10.4141/S04-083
- Gountié Dedzo M, Asaah ANE, Fozing EM, Tchamabé BC, Zangmo GT, Dagwai N, Seuwui DT, Kamgang P, Aka FT, Ohba T (2019) Petrology and geochemistry of lavas from Gawar, Minawao and Zamay volcanoes of the northern segment of the Cameroon volcanic line (Central Africa): Constraints on mantle source and geochemical evolution. J Afr Earth Sci 153:31–41. https://doi. org/10.1016/j.jafrearsci.2019.02.010
- Greiner L, Keller A, Grêt-Regamey A, Papritz A (2017) Soil function assessment: review of methods for quantifying the contributions of soils to ecosystem services. Land Use Pol 69:224–237. https:// doi.org/10.1016/j.landusepol.2017.06.025
- Grossman RB, Reinsch TG, (2002) Bulk density and linear extensibility. In: Dane, J.H., Topp, G.C. (Eds.), Methods of soil analysis, Part 4. Physical Methods. Soil Sci. Am. Book Series No. 5 ASA and SSSA, Madison, WI, pp. 201–228

- Guitian F, Carballas T (1976) Tecnicas de Analisis de suelos, 2, a ed. Pico Sacro, Santiago de Compostela
- Hassink J (1992) Effects of soil texture and structure on carbon and nitrogen mineralization in grassland soils. Biol Fertil Soils 14:126–134. https://doi.org/10.1007/BF00336262
- Hassink J (1995) Density fractions of soil macroorganic matter and microbial biomass as predictors of C and N mineralization. Soil Biol Biochem 27:1099–1108. https://doi.org/10.1016/0038-0717(95)00027-C
- Hassink J (1997) The capacity of soils to preserve organic C and N by their association with clay and silt particles. Plant Soil 191:77– 87. https://doi.org/10.1023/A:1004213929699
- Houghton RA (2007) Balancing the global carbon budget. Annu Rev Earth Planet Sci 35:313–347. https://doi.org/10.1146/annurev. earth.35.031306.140057
- Hyun S, Lee LS, Rao PSC (2003) Significance of anion exchange in pentachlorophenol sorption by variable-charge soils. J Environ Qual 32:966–976. https://doi.org/10.2134/jeq2003.9660
- Ibrahim MM, Guo L, Wu F, Liu D, Zhang H, Zou S, Xing S, Mao Y (2022) Field-applied biochar-based MgO and sepiolite composites possess CO₂ capture potential and alter organic C mineralization and C-cycling bacterial structure in fertilized soils. Sci Total Environ 813:152495. https://doi.org/10.1016/j. scitotenv.2021.152495
- IPCC (2003) In: Penman J, Gytarsky M, Hiraishi T, Krug T, Kruger D, Pipatti R, Buendia L, Miwa K, Ngara T, Tanabe K, Wagner F. (Eds.), Intergovernmental Panel on Climate Change: Good Practice Guidance for Land Use, Land Use Change and Forestry. IPCC/OECD/IEA/IGES, Hayama, Japan
- Iturri LA, Buschiazzo DE (2014) Cation exchange capacity and mineralogy of loess soils with different amounts of volcanic ashes. CATENA 121(5):81–87. https://doi.org/10.1016/j.catena. 2014.04.021
- IUSS Working Group WRB (2015) World Reference Base for Soil Resources 2014, update 2015. International soil classification system for naming soils and creating legends for soil maps. Word Soil Resources Reports No106, FAO, Rome.
- Jenny H (1941) Factors of Soil Formation. McGraw-Hill, New York, NY
- Kaiser K, Guggenberger G (2003) Mineral surfaces and soil organic matter. Eur J Soil Sci 54:219–236. https://doi.org/10.1046/j. 1365-2389.2003.00544.x
- Koen TB (2013) Variations in soil organic carbon for two soil types and six land uses in the Murray Catchment, New South Wales, Australia. Soil Res 51:631–644. https://doi.org/10.1071/SR123 53
- Lal R (2009a) Challenges and opportunities in soil organic matter research. Eur J Soil Sci 60:158–169. https://doi.org/10.1111/j. 1365-2389.2008.01114
- Lal R (2009b) Sequestering carbon in soils of arid ecosystems. Land Degrad Dev 20:441–454. https://doi.org/10.1002/ldr.934
- Lehmann J, Kleber M (2015) The contentious nature of soil organic matter. Nature 528:60–68. https://doi.org/10.1038/nature16069
- Letouzey R (1985) Carte phytogéographique du Cameroun au 1/500 000. Institut de Recherche agronomique. Yaoundé-Cameroun et Institut de la Carte Internationale de la Végétation Toulouse-France, Paris, p.2–9
- Li XR, Jia XH, Dong GR (2006) Influence of desertification on vegetation pattern variations in the cold semi-arid grasslands of Qinghai-Tibet Plateau North-west China. J Arid Environ 64:505– 522. https://doi.org/10.1016/j.jaridenv.2005.06.011
- Liu J, Wang Z, Hu F, Xu C, Ma R, Zhao S (2020) Soil organic matter and silt contents determine soil particle surface electrochemical properties across a long-term natural restoration grassland. Catena 190:104526. https://doi.org/10.1016/j.catena.2020. 104526

- Lozano-García B, Parras-Alcantara L (2014a) Variation in soil organic carbon and nitrogen stocks along a toposequence in a traditional Mediterranean olive grove. Land Degrad Dev 25:297–304. https://doi.org/10.1002/ldr.2284
- Lozano-García B, Parras-Alcantara L (2014b) Land use and management effects on carbon and nitrogen in Mediterranean cambisols. Agric Ecosyst Environ 179:208–214. https://doi.org/ 10.1016/j.agee.2013.07.009
- Manjaiah KM, Sarvendra Kumar MS, Sachdev PS, Datta SC (2010) Study of clay-organic. Curr Sci 98:915–918
- Mayer LM, Schick LL, Hardy KR (2004) Organic matter in small mesopores in sediments and soils. Geochim Cosmochim Acta 68:3863–3872. https://doi.org/10.1016/j.gca.03.019
- Mayer S, Kühnel A, Burmeister J, Kögel-Knabner I, Wiesmeier M (2019) Controlling factors of organic carbon stocks in agricultural top soils and subsoils of Bavaria. Soil Till Res 192:22–32. https://doi.org/10.1016/j.still.2019.04.021
- Moore Duane M, Reynolds Robert C (1989) X-ray diffraction and the identification and analysis of clay minerals
- Muñoz-Rojas M, Jordán A, Zavala LM, De la Rosa D, Abd-Elmabod SK, Anaya-Romero M (2012) Organic carbon stocks in Mediterranean soil types under different land uses (Southern Spain). Solid Earth 3:375–386. https://doi.org/10.5194/ se-3-375-2012
- Ngole-Jeme VM (2019) Fire-induced changes in soil and implications on soil sorption capacity and remediation methods. Appl Sci 9:3447. https://doi.org/10.3390/app9173447www.mdpi.com/ journal/applsci
- Ngounouno I, Durelle B, Demaiffe R (2000) Petrology of the biomodal Cenozoic volcanism of the Kapsiki plateau Northern most Cameroun, Central Africa. J Volcanol Geoth Res 12:21–44. https://doi.org/10.1016/S0377-0273(00)00180-3
- Nguyen TT, Janik LJ, Raupach M (1991) Diffuse Reflectance Infrared Fourier Transform (DRIFT) spectroscopy in soil studies. Aust J Soil Res 29:49–67. https://doi.org/10.1071/SR9910049
- Novara A, Gristina L, Bodì MB, Cerdá A (2011) The impact of fire on redistribution of soil organic matter on a Mediterranean hillslope under maquia vegetation type. Land Degrad Dev 2:530–536. https://doi.org/10.1002/ldr.1027
- Nweke IA, Nnabude PC (2014) Organic carbon, total nitrogen and available phosphorous concentration in aggregate fractions of four soils under two land use systems. Impact Int. J Res Appl Nat Soc Sci 2:273–288
- Olson KR, Al-Kaisi M, Lal R, Cihacek L (2016) Impact of soil erosion on soil organic carbon stocks. J Soil Water Conserv 71:61A-67A. https://doi.org/10.2489/jswc.71.3.61A
- Palm C, Blanco-Canqui H, De Clerck F, Gatere L, Grace P (2014) Conservation agriculture and ecosystem services: an overview. Agric Ecosyst Environ 187:87–105. https://doi.org/10.1016/j. agee.2013.10.010
- Parras-Alcantara L, Lozano-García B, Brevik EC, Cerdá (2015) Soil organic carbon stocks assessment in Mediterranean natural areas: a comparison of entire soil profiles and soil control sections. J. Environ. Manage. 155: 219–228. https://doi.org/10.1016/j.jenvm an.2015.03.039
- Pieri C (1989) Fertilité des terres de savanes. Bilan de trente ans de recherche et de d'eveloppement agricoles au sud du, Sahara. Ed. MINCOP et developpement CIRAD–IRAT, Paris
- Plaza C, Gabriel G, Ana MM, Claudio Z, Fernando TM (2018) Soil organic matter in dryland ecosystems. The Future of Soil Carbon. https://doi.org/10.1016/B978-0-12-811687-6
- Raich JW, Schlesinger WH (1992) The global carbon dioxide flux in soil respiration and its relationship to vegetation and climate. Tellus B 44(2):81–99. https://doi.org/10.1034/j.1600-0889.1992. t01-1-00001

- Razafimbelo T, Albrecht A, Feller C, Ravelojaona H, Moussa N, Razanamparany C, Rabeharisoa L (2010) Stockage de carbone dans les sols sous systèmes de culture en semis direct sous couvert végétal, SCV) dans différents contextes pédoclimatiques à Madagascar. Etude Et Gestion Des Sols 17:143–158
- Reichenbach M, Fiener P, Garland G, Griepentrog M, Six J, Doetterl S (2021) The role of geochemistry in organic carbon stabilization against microbial decomposition in tropical rainforest soils. SOIL 7:453–475. https://doi.org/10.5194/soil-7-453-2021
- Rieuwerts JS, Thornton I, Farago ME, Ashmore MR (1998) Factors influencing metal bioavailability in soils: preliminary investigations for the development of a critical loads approach for metals. Chem Speciat Bioavailab 10:61–75. https://doi.org/ 10.3184/095422998782775835
- Sakin E (2012) Relationships between of carbon, nitrogen stocks and texture of the Harran Plain soils in southeastern Turkey. Bulg J Agric Sci 18:626–634
- Schlüter S, Roussety T, Rohe L, Vusal G, Evgenia B, Thomas (2022) Land use impact on carbon mineralization in well aerated soils is mainly explained by variations of particulate organic matter rather than of soil structure. SOIL 8: 253–267. https://doi.org/ 10.5194/soil-8-253-2022
- Silatsa FBT, Nanda SS, Yemefack M, Ngomeni AF (2021). Cameroon: soil organic carbon sequestration potential national map. National Report. Version 1.0. https://doi.org/10.13140/RG.2.2. 18170.77762
- Singer MJ (2015) Basic principles of pedology. Reference module in earth systems and environmental sciences. University of California-Davis, Davis, CA, USA. https://doi.org/10.1016/ B978-0-12-409548-9.09290-3
- Singh M, Sarkar B, Sarkar S, Churchman J, Bolan NK, Mandal MM, Tapan JP, David JB (2017) Stabilization of soil organic carbon as influenced by clay mineralogy. Adv Agron. https://doi.org/10. 1016/bs.agron.2017.11.001
- Sissoko A, Kpomblekou-A K (2010) Carbon decomposition in broiler litter-amended soils. Soil Biol Biochem 42:543–550. https://doi. org/10.1016/j.soilbio.2009.10.007
- Six J, Conant RT, Paul EA, Paustian K (2002) Stabilization mechanisms of soil organic matter: implications for C-saturation of soils. Plant Soil 241:155–176. https://doi.org/10.1023/A: 1016125726789
- Smith P (2012) Soils and climate change. Curr Opin Environ Sustain 4:539–544. https://doi.org/10.1016/j.cosust.2012.06.005
- Strudley MW, Green TR, Ascough JC (2008) Tillage effects on soil hydraulic properties in space and time: state of the science. Soil Till Res 99:4–48. https://doi.org/10.1016/j.still.2008.01.007
- Tamen J, Nkoumbou C, Reusser E, Tchoua F (2015) Petrology and geochemistry of mantle xenoliths from the Kapsiki Plateau (Cameroon Volcanic Line): implications for lithospheric upwelling. J Afr Earth Sci 101:119–134. https://doi.org/10. 1016/j.jafrearsci.2014.09.008
- Tang Y, Li H, Liu XM, Zhu HL, Tian R (2015) Unraveling the size distributions of surface properties for purple soil and yellow soil. J Environ Sci 32:81–89. https://doi.org/10.1016/j.jes.2014. 12.011
- Tashi S, Singh B, Keitel C, Adams M (2016) Soil carbon and nitrogen stocks in forests along an altitudinal gradient in the eastern Himalayas and a meta-analysis of global data. Glob Chang Biol 22:2255–2268. https://doi.org/10.1111/gcb.13234
- Tombácz E, Libor Z, Illes E, Majzik A, Klumpp E (2004) The role of reactive surface sites and complexation by humic acids in the interaction of clay mineral and iron oxide particles. Org Geochem 35:257–267. https://doi.org/10.1016/j.orggeochem. 2003.11.002

- Trumbore SE, Chadwick OA, Amundson R (1996) Rapid exchange between soil carbon and atmospheric carbon dioxide driven by temperature change. Science 272:393–396. https://doi.org/10. 1126/science.272.5260.39
- Tsozué D, Bitom D, Nghonda JP (2011) Andosols of the eastern slope of the Bambouto Mountains (West Cameroon): general characteristics, impact human activities and environmental risks. Syllabus Review 2:57–68
- Tsozué D, Nghonda JP, Tematio P, Basga DS (2019) Changes in soil properties and soil organic carbon stocks along an elevation gradient at Mount Bambouto, Central Africa. CATENA 175:251–262. https://doi.org/10.1016/j.catena.2018.12.028
- Tsozué D, Noubissie MMN, Mamdem TLE, Basga DS, Bitom D (2021) Effects of environmental factors and soil properties on soil organic carbon stock in a natural dry tropical area of Cameroon. SOIL 7:677–691. https://doi.org/10.5194/soil-7-677
- Tsozué D, Nzeukou NA, Azinwi TP (2017) Genesis and classification of soils developed on gabbro in the high reliefs of Maroua region, North Cameroon. Eurasian J Soil Sci, 6(2): 168 – 177. https:// doi.org/10.18393/ejss.286631
- USDA (2004) Soil Survey Laboratory Methods Manual, Soil Survey Investigation Report No. 42, Version 4.0. USDA-NCRS, Lincoln, NE.
- Van der Zon APM (1992) Graminées du Cameroun. Phytogéographie et pâturages. Wageningen Agricultural University Papers I
- Wang Z, Chappellaz J, Park K, Mak JE (2010) Large variations in southern hemisphere biomass burning during the last 650 years. Science 330:1663–1666

- Walkley A, Black IA (1934) Determination of organic matter in soil. Soil Sci 37:549–556
- Wattel-Koekkoek EJW, van Genuchten PPL, Buurman P, van Lagen B (2001) Amount and composition of clayassociated soil organic matter in a range of kaolinitic and smectitic soils. Geoderma 99:27–49
- White AF, Schulz MS, Vivit DV, Blum AE, Stonestrom DA, Harden JW (2005) Chemical weathering rates of a soil chronosequence on granitic alluvium: III. Hydrochemical evolution and contemporary solute fluxes and rates. Geochim Cosmochim Ac 69:1975–1996. https://doi.org/10.1016/j.gca.2004.10.003
- Yukselen-Aksoy Y, Kaya A (2010) Method dependency of relationships between specific surface area and soil physicochemical properties. Appl Clay Sci 50(2):182–190. https://doi.org/10. 1016/j.clay.2010.07.020
- Zinn Y, Lal R, Bigham J, Resck D (2007) Edaphic controls on soil organic carbon retention in the Brazilian Cerrado: texture and mineralogy. Soil Sci Soc Am J 71:1204–1214

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