Promising cellulolytic fungi isolates for rice straw degradation

Diana Catalina Pedraza-Zapata¹⁷, Andrea Melissa Sánchez-Garibello^{1,2†}, Balkys Quevedo-Hidalgo³, Nubia Moreno-Sarmiento¹, **,** and Ivonne Gutiérrez-Rojas^{2*}

1 Grupo de Bioprocesos y Bioprospección, Instituto de Biotecnología, Universidad Nacional de Colombia, Sede Bogotá, 111321, Bogotá, D.C., Colombia

2 Laboratorio de Biotecnología Aplicada, Grupo de Biotecnología Ambiental e Industrial (GBAI), Departamento de Microbiología, Facultad de Ciencias, Pontificia Universidad Javeriana, 110231, Bogotá, D.C., Colombia

3 Grupo de Investigación en Agricultura Biológica. Facultad de Ciencias, Pontificia Universidad Javeriana, 110231, Bogotá, D.C., Colombia

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The objective of this study was to evaluate the potential of eight fungal isolates obtained from soils in rice crops for straw degradation *in situ***. From the initial eight isolates,** *Pleurotus ostreatus* **T1.1 and** *Penicillium* **sp. HC1 were selected for further characterization based on qualitative cellulolytic enzyme production and capacity to use rice straw as a sole carbon source. Subsequently, cellulolytic, xylanolytic, and lignolytic (***Pleurotus ostreatus***) activity on carboxymethyl cellulose, oat xylan, and rice straw with different nitrogen sources was evaluated. From the results obtained it was concluded both isolates are capable to produce enzymes necessary for rice straw degradation. However, their production is dependent upon carbon and nitrogen source. Last, it was established that** *Pleurotus ostreatus* **T1.1 and** *Penicillium* **sp. HC1 capability to colonize and mineralize rice straw, in mono-and co-culture, without affecting nitrogen soil content.**

*Keywords***:** rice straw, biodegradation, *Pleurotus ostreatus*, *Penicillium* sp., ligninolytic enzymes, cellulolytic enzymes

Introduction

The increase in world population and its effect on food demand, in addition to reducing environmental impact derived from food production, has raised a challenge for the agricultural industry in order to improve resource utilization and diminish food industry impact on non-renewable resources (Godfray *et al.*, 2010). Within this context, crop residues play an important role in agricultural sustainability. Thus, it is

necessary not to consider them as waste, but to acknowledge them as a renewable resource (Misselbrook *et al.*, 2012).

 Rice straw is one of the most produced agricultural residues worldwide. For each kg of cropped grain 1 to 1.5 kg rice straw is produced. Therefore, a worldwide estimate of byproduct is about 650–975 million tons/year (Binod *et al.*, 2010). Furthermore, rice straw is an important source of soil nutrients. Approximately 40% of the nitrogen (N), 30 to 35% of the phosphorus (P), 80 to 85% of the potassium (K), and 40 to 50% of the sulphur (S) taken up by the rice plant remains in the vegetative parts after the rice grain is harvested (Dobermann and Fairhurst, 2002). None the less, given its chemical composition, cellulose (32–47%), hemicellulose (19–27%), and lignin (5–24%), its degradation in soil is very slow. Therefore, farmers prefer to remove it from the field as a faster alternative, taking away nutrients (Dobermann and Fairhurst, 2002), incurring in expenses and demanding space, time, and technology (Matsumura *et al.*, 2005). *In situ* burning is another option increasing air pollution and consequently affecting public health (Binod *et al*., 2010).

 To avoid burning and field removal, harvest residues can be incorporated into field soil. It has been reported this practice improves the soil's physico-chemical conditions and organic matter content, in the long term (Pathak *et al.*, 2006; Das *et al.*, 2008; Ortiz Escobar and Hue, 2008; Tejada *et al.*, 2008). However, it has also been described crop productivity can significantly decrease immediately after straw incorporation to soil, as a consequence of nitrogen immobilization by competing microbial population during the degradation process (Shindo and Nishio, 2005). In the same manner, CH4 production can be tripled after a field is flooded in comparison with emissions generated during burning, when residue decomposition is performed in an anaerobic atmosphere (Kakua *et al.*, 2000; Glissmann *et al.*, 2001; He *et al.*, 2009).

 Direct straw incorporation disadvantages can be lowered incorporating it partially or completely degraded. Bio-augmentation with lignocellulolytic fungi is an alternative. It has been demonstrated microorganism use and/or their enzymes increase recalcitrant compound degradation kinetics (lignin, cellulose, and hemicellulose) in these plant materials (Taniguchi *et al.*, 2005; Eun *et al.*, 2006; Harada *et al.*, 2008; Dinis *et al.*, 2009; Kausar *et al.*, 2010; Singh *et al.*, 2011). Additionally, it has been found fungi growth improves soil's physical structure (Yao *et al.*, 2011).

 In Colombia, the National Rice Federation (Federación Nacional de Arroceros-FEDEARROZ) has evaluated the use of the fungus *Trichoderma viride*, a soil borne pathogens control agent, with excellent results as a straw decomposer; although this organism was not isolated or produced for this purpose. Performing crop residue handling and nutrient

[†] These authors contributed equally to this work.

^{*}For correspondence. E-mail: ivonne.gutierrez@javeriana.edu.co; Tel.: +5 7-1-3208320 (Ext 4073); Fax: +57-1-3208320 (Ext 4022)

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recycling using *T. viride*, these investigators achieved to increase rice yield between 1.0 to 1.4 t/ha, demonstrating this alternative is ecological and economically sustainable (Garcés and Ospina, 2009; Castilla, 2012). In the present study we evaluated the lignocellulolytic activity of eight fungi, previously isolated from rice crops, on defined cellulosic substrates and rice straw; thus establishing their potential for *in situ* straw degradation. Based on *Trichoderma viride* successful results, we hypothesized better results could be attained if microorganisms for this specific activity were selected.

Materials and Methods

Microorganisms

Evaluated microorganisms in this work were previously isolated from soils where rice-harvesting took place, and preselected for their cellulolytic and/or lignolytic activity (Gutiérrez-Rojas *et al.*, 2012). They had been previously identified as *Penicillium* sp. (codes HC1, HC7.4, HC7.5, and HC7.5'), *Trichoderma* sp. (HC8), *Paecilomyces* sp. (S7.1), *Streptomyces* sp. (S7.1), and *Pleurotus ostreatus* (T1.1) and are conserved on agar plates at 4°C at the Fermentation Laboratory, Biotechnology Institute, Universidad Nacional de Colombia. Isolates reactivation was performed in potato dextrose agar (PDA) for 8 days incubation period at 25°C.

Microorganism selection

To select microorganisms for subsequent phases of the study, cellulolytic enzyme production was qualitatively evaluated, as well as the capacity to utilize rice straw as the only carbon source. For qualitative evaluation Congo red hydrolysis halos in carboxymethyl cellulose (CMC) were employed (Kasana*, et al.*, 2008). As negative and positive controls *Escherichia coli* ATCC 25922 and *Trichoderma reesei* ATCC 26921 were used, respectively. All assays were performed in triplicates. Microbial respiration activity on rice straw was measured using as a carbon and nitrogen source 8 g of cut straw (2–3 cm long) adding 5 ml salts solution (in g/L: 0.3 KH_2PO_4 , 0.83 MgSO_4 . 7H₂O, 0.3 CaCl₂, 0.005 FeSO₄·7H₂O, 1.56 mg MnSO₄, 1.4 mg ZnSO4, 0.2 ml Tween 80, pH 6.0), in tightly sealed glass flasks. A container with 30 ml of 0.2 N NaOH was placed in the center of the flask to capture $CO₂$ emanating from the respiration process. Flasks were inoculated with 10 ml conidial suspension $(10^6 \text{ conidia/ml})$, for T1.1 isolate suspension was adjusted to 10^6 UFC/ml. An abiotic control was left without inoculation. Flasks were incubated for 60 days at 25°C. All assays were performed in triplicates. Every two days the container with NaOH was removed to evaluate $CO₂$ production by titration, and replaced with a new sterile flask containing NaOH. NaOH excess was titrated with a 0.2 N HCl standard solution until the final reaction point was established with phenolphthalein as a pH indicator (Critter *et al.*, 2004).

Cellulolytic and xylanolytic activity evaluation in selected microorganisms

Evaluation was performed in 250 ml Erlenmeyer with 120 ml Mandel's mineral salts solution (Mandels and Weber, 1969) adding the carbon source to be evaluated at 1% (w/v): carboxymethyl cellulose (CMC), straw or xylan, and 0.5% (w/v) nitrogen source: yeast extract or $(NH₄)₂SO₄$ (Table 1). pH was adjusted to 6.0 and 12 ml inoculum was used at 10° conidia/ml for cellulolytic isolates and 10⁶ UFC/ml for lignolytic isolate. Erlenmeyer flasks were incubated at 28°C with 100 rpm shaking for 10 days. After the fourth day of incubation, samples were collected every two days to determine enzyme activity with protocol described in enzyme determination. All assays were performed in triplicates.

Rice straw degradation capacity evaluation in soil microcosm

These assays were carried-out in 6 L plastic boxes with lids, in which 600 g of sterile soil was placed with 50 g rice straw cut into 3 to 6 cm long pieces. Treatments were performed in triplicate in the following manner: 1) Lignolytic microorganism, 2) Cellulolytic microorganism, 3) Lignolytic-cellulolytic co-culture, and 4) Abiotic control. All treatments were performed in triplicates. Inoculum for all treatments was 62 ml suspension at a concentration of 10^6 conidia/ml for cellulolytic isolates and 10^6 UFC/ml for lignolytic isolate. The experiment was maintained for 50 days under greenhouse conditions with CO₂ follow-up and enzyme activity evaluation. Scanning electron microscopy (SEM) was carriedout on days 0, 30, and 50. Last, a soil sample was taken at the beginning and at the end of the experiment to analyze nitrogen and organic carbon content by Kjeldahl (Bremner, 1960) and Walkley and Black (1934) methodologies, respectively.

 For enzymatic activity 2 g of rice straw was sampled every eight days to which 20 ml citrate buffer pH 5.0 was added (Brijwani *et al.*, 2010), and agitated at 150 rpm for 6 h at 25°C. To obtain an extract, the sample was centrifuged twice at $2,500 \times g$ for 20 min. With the resulting extract, endoglucanase, cellobiohydrolase, β-glucosidase, xylanase, laccase, and manganese peroxidase activity were assayed as described in a latter section. To evaluate $CO₂$ production a polyvinyl chloride tube was placed as a container of a glass vial with 20 ml 1 N NaOH to trap $CO₂$, following the methodology previously described. 1 N HCl was used for titration, measurements were carried-out every three day during the 50 days of the assay.

Enzyme activity determination

All enzyme activities were spectrophotometrically determined (Evolution 60 UV-VIS Thermo Scientific Co.). Endo-1,4 β-glucanase activity was assayed from 500 μl sample mixed

with 500 μl 2% (w/v) low viscosity CMC in 50 mM citrate buffer (pH 5.0), at 40°C for 60 min (Ghose, 1987). The reaction was stopped by 5 min incubation on ice, followed by 5 min centrifugation at 2,500 × *g* (Qinnghe *et al.*, 2004). Quantification reducing sugars equivalent to glucose was determined by 3,5 dinitrosalicylic acid (DNS) (Miller, 1959). A unit of endoglucanase was defined as the quantity of enzyme required to release 1 μmol reducing sugars per min under assay conditions.

 Cellobiohydrolase (exo-1,4-β-glucanase) was evaluated employing ρ-nitrophenyl-β-D-cellobioside as substrate (PNPC, Sigma-Aldrich). The reaction contained 320 μl PNPC in 50 mM sodium acetate buffer (pH 5.0) and 100 μl sample. This mix was incubated at 40°C for 60 min. The reaction was stopped adding 250 μl 0.5 M sodium carbonate. Absorbance was read at 400 nm (Valásková and Baldrian, 2006). Enzyme activity was calculated from ρ-nitrophenol standard curve. One unit cellobiohydrolase activity was defined as the quantity of enzyme required to release 1 μmol ρ-nitrophenol per min.

 1-4-β-Glucosidase enzyme activity was quantified using as substrate ρ-nitrophenyl-β-D-glucoside (PNPG, Sigma-Aldrich), following the same method as for cellobiohydrolase.

 Endoxylanase activity was determined mixing 500 μl sample with 500 μl 2% (w/v) beech tree xylan in 50 mM citrate buffer (pH 5.0). The reaction was incubated at 40°C for 60 min (Driss *et al.*, 2011). A unit of xylanase was defined as the quantity of enzyme releasing 1 μmol reducing sugars equivalent to xylose per min.

 Laccase activity was measured monitoring the change of absorbance at 436 nm (ξ_{436} = 29,300 L/mol/cm) caused by ABTS oxidation [2,2' azino-*bis*-(3 ethylbenzthiazoline sulphonic) acid] in 60 mM sodium acetate buffer (pH 4.5). The assay was carried out by mixing 100 μl buffer, 100 μl ABTS, and 800 μl sample at 20°C for 3 min. One unit laccase activity was defined as the quantity of enzyme oxidiz-

Fig. 1. Zone of clearance diameter in CMC agar for eight fungal isolates evaluated. Negative control: *Escherichia coli* ATCC 25922. Positive control: commercial enzyme *Trichoderma reesei* ATCC 26921. Bars with different letters indicate groups that are significantly different in *post hoc* testing ($\alpha = 0.05$).

ing 1 μmol ABTS per min (Tinoco *et al.*, 2001).

 Manganese peroxidase activity was determined by oxidation of 2,6-dimethoxyphenol (2,6-DMP) (Sigma-Aldrich) measured at 468 nm (ξ_{468} = 49,600 M/cm). Reaction mixture contained 450 μl culture supernatant, 500 μl 10 mM 2,6-DMP in 100 mM sodium acetate buffer (pH 5.0), 50 μl 0.4 mM MnSO₄, and 30 µl 22 mM H_2O_2 . The mixture was incubated for 3 min (Santoyo *et al.*, 2008). Given that both manganese peroxidase and laccase can oxidize 2,6-DMP, laccase activity was subtracted using data from control experiments performed in the absence of H_2O_2 and manganese (Quevedo-Hidalgo *et al.*, 2012). One unit of manganese peroxidase activity was defined as 1 μmol of 2,6-DMP oxidized per min.

Statistical analysis

Statistical analysis was performed using one way analysis of variance (ANOVA), with Tukey *post-hoc* test (α = 0.05), using IBM SPSS-Statistics (version 19, IBM Corp.).

Results and Discussion

Microorganism selection

Qualitative evaluation results in CMC agar evidenced the highest responses were obtained from isolates S7.1 (13.93 ±

Fig. 2. CO₂ production results using rice straw as carbon and nitrogen **source in the laboratory for eight fungal isolates evaluated.** Production kinetics (A) and rate of production (B). Bars with different letters indicate groups that are significantly different in *post hoc* testing ($\alpha = 0.05$).

2.64 mm halo) identified as *Streptomyces* sp. and isolate HC1 (9.97 ± 1.11 mm halo) identified as *Penicillium* sp. (Fig. 1). Isolate HC8, *Trichoderma* sp., presented the lesser response $(1.37 \pm 0.55$ mm halo). The rest of the evaluated isolates corresponding to *Penicillium* sp. and *Paecilomyces* sp. had intermediate zone of clearance formation. In contrast to CMC agar results, $CO₂$ production results using rice straw as carbon and nitrogen source demonstrated isolate T1.1 (*P. ostreatus*) had the highest $CO₂$ production, 233.64 \pm 2.03 mg $CO₂/g$ rice straw on day 60, with a production rate of 4.15 \pm 0.05 mg CO₂/g rice straw/day, followed by isolate HC1 (*Penicillium* sp.), 140.33 ± 2.60 mg CO₂/g rice straw on day 60 and 2.35 \pm 0.05 mg CO₂/g rice straw/day and isolate S7.1 (*Streptomyces sp.*), 117.69 ± 2.19 mg CO₂/g rice straw on day 60 and 1.92 \pm 0.09 mg CO₂/g rice straw/day (Fig. 2A and B).

 Differences in results obtained could be due to substrate nature. Enzyme production in the cellulolytic complex is highly regulated. In addition, this mechanism depends on microorganism and substrate type (Gutiérrez-Rojas *et al.*, 2015). It was expected a microorganism such as *P. ostreatus*, iso-

late T1.1, in addition to cellulases, would produce xylanases, laccase, and manganese peroxidase (Membrillo *et al.*, 2008), thus, becoming more efficient in lignocellulosic substrate degradation, measured as $CO₂$ production on rice straw. Even though, it did not present the highest CMC agar activity. On the other hand, *Streptomyces* sp. S7.1 demonstrated an opposite behavior, i.e. better CMC agar activity than $CO₂$ production on rice straw. MacKenzie and collaborators (1987) evaluated cellulases and xylanases production in *Streptomyces flavogriseus* and *Streptomyces olivochromogenes*, culturing on different substrates. They found for both microorganisms growth on cellulose primarily induce cellulase activity, but failed to produce appreciable levels of xylanases. Whereas, growth on xylan and wheat bran induced high levels of endoxylanase; suggesting enzyme expression mechanisms of regulation are independent between the two enzymes, but substrate dependent (Mackenzie *et al.*, 1987).

 Regarding *Penicillium* sp. HC1 isolate, consistent results were obtained for CMC agar activity as well as $CO₂$ production on rice straw. It has been reported a great number of species in the *Penicillium* genus are capable of producing a

Means with different letters in each enzyme activity (for each fungi) were significantly different in *post hoc* testing (α = 0.05).

multienzymatic extracellular complex efficiently degrading cellulose and hemicellulose in plant biomass. For some cases this complex has resulted more efficient than *Trichoderma reesei* mutants traditionally used in industry. The reasons for such efficiency are high β-glucosidase activity, high specific cellobiohydrolase activity, low sensitivity to product inhibition of the later one, and low bond between these enzymes to biomass lignin component (Gusakov and Sinitsyn, 2012). However, our results suggest these characteristics are distinct of the species, since not all *Penicillium* sp. isolates evaluated demonstrated the same efficiency on CMC agar activity as well as $CO₂$ production on rice straw.

 Given the results obtained on both substrates did not show an association, i.e., isolates with the best CMC agar activity did not correspond to the best $CO₂$ production on rice straw, it was decided to prioritize activity on rice straw, thus the selected microorganisms for the following stages of this study were isolates T1.1 (*Pleurotus ostreatus*) and HC1 (*Penicillium* sp.).

Cellulolytic and xylanolytic activity evaluation on selected isolates

Selected microorganisms cellulolytic activity was evaluated on CMC and for xylanolytic activity on oat-xylan or rice straw. Additionally, since enzyme expression can also be regulated by nitrogen source (Qinnghe *et al.,* 2004), carbon sources were combined with nitrogen sources: none, yeast extract, ammonium sulphate, and a mixture of both. Results of this evaluation are described in Table 2.

 The highest values for *Penicillium* sp. HC1 β-glucosidase $(174.7 \pm 3.9 \text{ U/L})$ were achieved when a mixture of organic and inorganic nitrogen source was used. In contrast, the highest values for cellobiohydrolase (364.0 \pm 2.9 U/L) and endoglucanase (136.9 \pm 7.25 U/L) were observed when ammonium sulphate was the nitrogen source. Xylanases were not affected by the source of nitrogen and the highest value was 644.6 ± 2.3 U/L without any nitrogen source added. Even though, during the last years an increase in reports describing cellulases and xylanases produced form different *Penicillium* genus has been observed, result comparison is difficult due to distinct conditions for each culture. Substrate, culture condition, and enzyme activity assay conditions, all represent determinant factors. As a case in point, β-glucosidase values have been reported from 200 U/L, when using a mix of lactose (0.25%) and cellulose (0.75%) as a substrate, up to 500 U/L, when only lactose at 1% is used for *P. echinelatum* 9A02S1 (Sehnem *et al.*, 2006). Krogh and collaborators (2004) evaluated β-glucosidase activity and endoglucanase using 2% Solka-Floc cellulose with reported values of 970 U/L (β-glucosidase) and 12,000 U/L (endoglucanase) for *P. verruculosum* IBT 18366, 1,700 U/L (β-glucosidase) and 900 U/L (endoglucanase) for *P. minioluteum* IBT 21486, 1,090 U/L (β-glucosidase) and 980 U/L (endoglucanase) for *P. brasilianum* IBT 20888, and 2,450 U/L (β-glucosidase) and 600 U/L (endoglucanase) for *P. pinophilum* IBT 10872. These same authors evaluated xylanase production in 12 *Penicillium* isolates, using as a substrate oat xylan finding activities ranging from 1,500 to 105,000 U/L, with *P. persicinum* IBT 13226 presenting the highest activity (105,000 U/L) and *P. funiculosum* IBT 5816 the second highest ac-

tivity (42,000 U/L) (Krogh *et al.*, 2004), all of which were higher compared with the ones reported in this work.

 Even though, the values are not comparable for the aforementioned reasons, their ratio β-glucosidase:Endoglucanase is comparable. For the results reported by Krogh *et al*. (2004) they were as follows: 0.080 for *P. verruculosum* IBT 18366, 1.89 for *P. minioluteum* IBT 21486, 1.11 for *P. brasilianum* IBT 20888, and 4.08 for *P. pinophilum* IBT 10872. For our case 1.28, however we did not find the same behavior for the β-glucosidase:Cellobiohydrolase ratio, our was 0.48. For most cases the ratio was greater than one, characteristic of the *Penicillium* genus, representing an advantage over other cellulolytic microorganisms such as *Trichoderma reesei*. *T. reesei* is identified by a low β-glucosidase secretion, resulting in a β-glucosidase:Endoglucanase or β-glucosidase:Cellobiohydrolase ratio lower than one (Gusakov and Sinitsyn, 2012). A high level of β-glucosidase is necessary to promote a rapid and complete conversion of cellobiose to glucose. Cellubiose is an intermediate metabolite of cellulose hydrolysis and is an endoglucanase inhibitor, and to a greater extent of cellobiohydrolases. Thus, it represents a bottleneck in the cellulose enzymatic hydrolysis (Gruno *et al.*, 2004).

 In regards to *P. ostreatus* the highest values for β-glucosidase (158.2 ± 22.33 U/L), cellobiohydrolase (68.01 ± 4.18 U/L), and endoglucanase (94.7 \pm 21.23 U/L) can be obtained using yeast extract as the nitrogen source. A similar behavior was observed for xylanases, a maximum activity value of 874.51 ± 2.7 U/L was obtained when yeast extract was the sole nitrogen source. These results agree with those reported by Qinnghe *et al*. (2004), who evaluated *P. ostreatus* SYJ042 xylanase production in submerged fermentation using different nitrogen sources including ammonium nitrate (NH_4NO_3) , ammonium sulfate ((NH₄)₂SO₄), phosphatic diamine ((NH₄)₂) HPO4), yeast extract, tryptone, peptone, and meat extract. They found an activity range between 1,900 to 1,980 U/L with inorganic sources. In addition, activities increased when organic sources were employed (6,530 to 8,430 U/L), with peptone yielding the highest activity (Qinnghe *et al*., 2004). On the other hand, it has been reported nitrogen source does not affect *P. dryinus* IBB 903 endoglucanase or xylanases production when tangerine peelings were used as a carbon source. Furthermore, the maximal endoglucanase activity was revealed after 10 days growth in medium containing ammonium sulphate (53,700 U/L), while maximal xylanase activity (95,200 U/L) was recorded after 8 days of submerged fermentation (Elisashvili *et al.*, 2006). Many previous studies have proved that both the nature and concentration of nitrogen sources are powerful factors regulating ligninolytic enzyme production by wood-rotting basidiomycetes (Songulashvili *et al.*, 2007). However, nitrogen source influence is related to wood-rotting basidiomycetes and/or the source of carbon and/or the condition of the culture. Therefore, it is not possible to generalize regarding this aspect.

 Results obtained in this evaluation allow to conclude both native isolates evaluated *Penicillium* sp. HC1 and *Pleurotus ostreatus* T1.1 are capable of producing enzymes necessary for rice straw degradation, although the evaluated values are subject to optimization if either one would be used for enzyme production.

Fig. 3. Rice straw residue SEM at day 30 of inoculation. *Penicillium* sp. HC1 (A), *Pleurotus ostreatus* T1.1 (B), and co-culture (C).

Evaluation of microorganism capability to degrade rice straw

Scanning electron micrographs during residue degradation process illustrate for monoculture or co-culture microorganism growth and colonization on rice straw residue (Fig. 3). As it was expected, treatment exclusively inoculated with *Penicillium* sp. HC1 displayed scarce mycelium growth and abundant conidia production (Fig. 3A). In contrast, mycelium growth was observed for *Pleurotus ostreatus* T1.1 inoculated treatment (Fig. 3B). Last, co-culture treatment exhibited copious mycelium and limited conidia (Fig. 3C).

 Moreover, CO2 measurements evidenced *Penicillium* sp. HC1 treatment had the lowest $CO₂$ production (30.34 \pm 0.94 mg $CO₂/g$ straw for day 47 during evaluation) compared with *P. ostreatus* T1.1 with a CO_2 production of 68.76 \pm 3.73 mg $CO₂/g$ rice straw (Fig. 4). Likewise, rate of $CO₂$ production for *Penicillium* sp. HC1 was the lowest $(0.72 \pm 0.09 \text{ mg } CO_2/g$ rice straw/day) compared with *P. ostreatus* T1.1 (1.47 ± 0.16 mg $CO₂/g$ straw/day), behavior comparable to evaluations performed in the laboratory (Fig. 2A and B). Contrary to what was expected, microorganism co-culture did not produce an increase in $CO₂$ production. This remained similar to the *P. ostreatus* T1.1 treatment $(65.81 \pm 3.37 \text{ mg } CO₂/g)$ straw; rate of production 1.32 ± 0.08 mg CO₂/g straw/day), indicating microorganism co-culture did not increase straw degradation. In regards to enzyme activity, no differences were observed among the three treatments for endoglucanase, cellobiohydrolase, and xylanase maximum values. Although, β-glucosidase activity was greater for *Penicillium* sp. HC1 treatment, no differences were evidenced between *P. ostreatus* T1.1 and microorganism co-culture treatment (Table 3). With respect to lignolytic activity, *Penicillium* sp. HC1 was unable to produce detectable levels of laccase and manganese peroxidase (Table 3). These results agree with reports in the literature, since the *Penicillium* genus is reported as cellulolytic and xylanolytic, and not capable of generating the afore-

Fig. 4. Greenhouse CO₂ production (A), production kinetics and (B) rate **of production.** Bars with different letters indicate groups that are significantly different in *post hoc* testing ($\alpha = 0.05$).

mentioned enzymes (Chávez *et al.*, 2006).

 Compatible lignocellulolytic fungal consortium (fungal compatible co-culture) is frequently associated with increased substrate degradation in comparison with activity obtained by each fungus (He *et al*., 2009). Dwivedi *et al*. (2011) evaluated xylanase and laccase production by a mutant *Penicillium oxalicum* SAUE-3.510 strain in mono-culture or in co-culture with *Pleurotus ostreatus* MTCC 1804, on different lignocellulosic substrates. Their findings evidenced xylanase and laccase activities were higher in co-culture than in axenic cultures (Dwivedi*, et al.*, 2011). Furthermore, Monahan *et al*. (2014) compared fungal monoculture efficiency with co-cultures in simultaneous delignification and saccharification of kitchen waste and *Eichhornia crassipes* leaves. These authors described *Trichoderma* spp. co-culture with other cel-

Means with different letters were significantly different in *post hoc* testing (α = 0.05).

Fig. 5. Percent of soil organic carbon (A) and total nitrogen (B), after rice straw treatment (50 days). Bars with different letters indicate groups that are significantly different in *post hoc* testing ($\alpha = 0.05$).

lulolytic fungi improved lignocellulose degrading enzyme activity compared with *Trichoderma* monoculture. Nevertheless, co-culture with other fungi (*Aspergillus, Pycnopurus*, and undidentified strain F113) did not result in significant degradation activity compared to corresponding monocultures (Mohanan *et al.*, 2014). Hence, it can be concluded coculture is not always implied in increased degradation activity. In fact, improved degradation is dependent on the organisms involved and used conditions during the degradation process.

 With respect to the results obtained in the present work, these demonstrate behavior of two fungi in co-culture is practically identical to *P. ostreatus* T1.1 monoculture, indicating that *P. ostreatus* T1.1 inhibited *Penicillium* sp. HC1 growth as observed in our SEM results (Fig. 3), even though it had been previously verified these were not antagonistic microorganisms (data not shown). This could be due to *P. ostreatus* T1.1 growth rate being greater than *Penicillium* sp. HC1, generating a negative effect, since both fungi were inoculated at the same time. Future studies should evaluate inoculation conditions such as time and concentrations to grant *Penicillium* sp. HC1 an advantage over *P. ostreatus* T1.1.

 Despite the system is still susceptible for optimization, high carbon mineralization rates (measured as $CO₂$ production, Fig. 4) were observed when comparing to reports in the literature. Devevre and Horwath (2000) evaluated the effects of flooding and temperature on rice straw degradation. Under their best conditions (non-flooded, 25°C) they obtained 2.5 mg $CO₂/g$ dry soil, approximately, after 160 days of observation (Devevre and Horwath, 2000), results far less lower

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when comparing with the data of the present study. This indicates lignocellulolytic microorganisms efficiently contribute to an increase in the rate of rice straw degradation.

 Last, we determined total organic carbon and nitrogen as an indicator of nutrient return to the soil proceeding from rice straw degradation (Fig. 5). In a crop this return of nutrients results in decrease of fertilizers used by the farmers (Garcés and Ospina, 2009). For all samples analyzed no differences in organic carbon content were observed, most likely due to short evaluation time. Nonetheless, differences were observed regarding total nitrogen, since all treatments presented higher values compared with control (Fig. 4B). One of the inconveniences of rice straw incorporation into the soil is nitrogen availability for the plant, due to nitrogen immobilization in microbial biomass as a result of a high C/N ratio in this residue (Shindo and Nishio, 2005). Pathak *et al.* (2006) demonstrated rice straw could be successfully managed *in situ*. Microorganism inoculation avoids problems of net *N*-immobilization and yield reduction of the subsequent crop due to rice straw amendment (Pathak *et al.,* 2006). In this work we found an increase in nitrogen content, even though we cannot distinguish organic (non-available) or inorganic (available), it is an interesting finding that must be proven in future studies in the field.

 In conclusion, *Pleurotus ostreatus* T1.1 and *Penicillium* sp. HC1 isolates selected out of eight initially chosen demonstrated their potential to be utilized for *in situ* rice straw degradation, and posterior incorporation into the soil. Both were capable of colonizing the residue and produce high rates of mineralization without diminishing the soil nitrogen content. However, future work should optimize co-culture condition if both microorganisms are to be used at the same time.

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References

- **Binod, P., Sindhu, R., Singhania, R., Vikram, S., Devi, L., Nagalakshmi, S., Kurien, N., Sukumaran, R., and Pandey, A.** 2010. Bioethanol production from rice straw: An overview. *Bioresour. Technol.* **101**, 4767–4774.
- **Bremner, J.** 1960. Determination of nitrogen in soil by the Kjeldahl method. *J. Agric. Sci*. **55**, 11–33.
- **Brijwani, K., Oberoi, H., and Vadlani, P.** 2010. Production of a cellulolytic enzyme system in mixed-culture solid-state fermentation of soybean hulls supplemented with wheat bran. *Process Biochem.* **45**, 120–128.
- **Castilla, L.** 2012. Manejo productivo de los residuos de la cosecha de arroz. *Revista Arroz - FEDEARROZ.* **60**, 10–18.
- **Critter, S., Freitas, S., and Airoldi, C.** 2004. Comparison of microbial activity in some Brazilian soils by microcalorimetric and respirometric methods. *Thermochim. Acta.* **410**, 35–46.
- **Chávez, R., Bull, P., and Eyzaguirre, J.** 2006. The xylanolytic enzyme system from the genus *Penicillium*. *J. Biotechnol.* **123**, 413–433.
- **Das, A., Patel, D., Munda, G., Hazarika, U., and Bordoloi, J.** 2008.

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Nutrient recycling potential in rice–vegetable cropping sequences under *in situ* residue management at mid-altitude subtropical Meghalaya. *Nutr. Cycl. Agroecosys.* **82**, 251–258.

- **Devevre, O. and Horwath, W.** 2000. Decomposition of rice straw and microbial carbon use efficiency under different soil temperatures and moistures. *Soil Biol. Biochem.* **32**, 1773–1785.
- **Dinis, M., Bezerra, R., Nunes, F., Dias, A., Guedes, C., Ferreira, L., Cone, J., Marques, G., Barros, A., and Rodrigues, M.** 2009. Modification of wheat straw lignin by solid state fermentation with white-rot fungi. *Bioresour. Technol.* **100**, 4829–4835.
- **Dobermann, A. and Fairhurst, T.H.** 2002. Rice straw management. *Better Crops International* **16**, 7–11.
- **Driss, D., Bhiri, F., Elleuch, L., Bouly, N., Stals, I., Miled, N., Blibech, M., Ghorbel, R., and Chaabouni, S.E.** 2011. Purification and properties of an extracellular acidophilic endo-1,4-β-xylanase, naturally deleted in the "thumb", from *Penicillium occitanis* pol6. *Process Biochem.* **46**, 1299–1306.
- **Dwivedi, P., Vivekanand, V., Pareek, N., Sharma, A., and Singh, R.** 2011. Co-cultivation of mutant *Penicillium oxalicum* saue-3.510 and *Pleurotus ostreatus* for simultaneous biosynthesis of xylanase and laccase under solid-state fermentation. *New Biotechnol.* **28**, 616–626.
- **Elisashvili, V., Penninckx, M., Kachlishvili, E., Asatiani, M., and Kvesitadze, G.** 2006. Use of *Pleurotus dryinus* for lignocellulolytic enzymes production in submerged fermentation of mandarin peels and tree leaves. *Enzyme Microb. Technol.* **38**, 998– 1004.
- **Eun, J., Beauchemin, K., Hong, S., and Bauer, M.** 2006. Exogenous enzymes added to untreated or ammoniated rice straw: Effects on in vitro fermentation characteristics and degradability. *Anim. Feed Sci. Tech.* **131**, 87–102.
- **Garcés, G. and Ospina, J.** 2009. Estrategias para el aprovechamiento de los residuos de cosecha del arroz. *Boletín informativo de la Federación Nacional de Arroceros - Colombia.* **217**, 4–5.
- **Ghose, T.K.** 1987. Measurement of cellulase activities. *Pure Appl. Chem.* **59**, 257–268.
- **Glissmann, K., Weber, S., and Conrad, R.** 2001. Localization of processes involved in methanogenic degradation of rice straw in anoxic paddy soil. *Environ. Microbiol.* **3**, 502–511.
- **Godfray, H., Beddington, J., Crute, I., Haddad, L., Lawrence, D., Muir, J., Pretty, J., Robinson, S., Thomas, S., and Toulmin, C.** 2010. Food security: The challenge of feeding 9 billion people. *Science* **327**, 812–818.
- **Gruno, M., Valjamae, P., Pettersson, G., and Johansson, G.** 2004. Inhibition of the *Trichoderma reesei* cellulases by cellobiose is strongly dependent on the nature of the substrate. *Biotechnol. Bioeng.* **86**, 503–511.
- **Gusakov, A. and Sinitsyn, A.** 2012. Cellulases from *Penicillium* species for producing fuels from biomass. *Biofuels* **3**, 463–477.
- **Gutiérrez-Rojas, I., Matiz-Villamil, A., Aguirre-Morales, M., Reyes-Pineda, E., Lemos-Gordo, S.N., Méndez-Pedraza, J.M., Núñez-Arbeláez, Á.J., Parra-Fajardo, L.N., Alfonso-Piragua, A., Avendaño-Herrera, D.,** *et al.* 2012. Estimación de poblaciones de microorganismos ligninolíticos y celulolíticos y actividad de β-glucosidasa en agrosistemas de arroz, p. 89–109. *In* Uribe-Vélez, D., and Melgarejo, L.M. (eds.), Ecología de microorganismos rizosféricos asociados a cultivos de arroz de Tolima y Meta, Universidad Nacional de Colombia, Bogotá D.C.
- **Gutiérrez-Rojas, I., Moreno-Sarmiento, N., and Montoya, D.** 2015. Mecanismos y regulación de la hidrólisis enzimática de celulosa en hongos filamentosos: Casos clásicos y nuevos modelos. *Rev. Iberoam. Micol.* **32**, 1–12.
- **Harada, K., Tanaka, K., Fukuda, Y., Hashimoto, W., and Murata, K.** 2008. *Paenibacillus* sp. strain HC1 xylanases responsible for degradation of rice bran hemicellulose. *Microbiol. Res.* **163**, 293–298.
- **He, J., Liu, X., Zheng, Y., Shen, J., and Zhang, L.** 2009. Effects of soil water content and rice straw incorporation in the fallow season

on CH4 emissions during fallow and the following rice-cropping seasons. *Biol. Fertil. Soils* **335**, 283–291.

- **Kakua, N., Uekia, A., Fujiib, H., and Ueki, K.** 2000. Methanogenic activities on rice roots and plant residue and their contributions to methanogenesis in wetland rice field soil. *Soil Biol. Biochem.* **32**, 2001–2010.
- **Kasana, R., Salwan, R., Dhar, H., Dutt, S., and Gulati, A.** 2008. A rapid and easy method for the detection of microbial cellulases on agar plates using gram's iodine. *Curr. Microbiol.* **57**, 503– 507.
- **Kausar, H., Sariah, M., Saud, H., Alam, M., and Ismail, M.** 2010. Development of compatible lignocellulolytic fungal consortium for rapid composting of rice straw. *Int. Biodeter. Biodegr.* **64**, 594–600.
- **Krogh, K., Mørkeberg, A., Jørgensen, H., Frisvad, J., and Olsson, L.** 2004. Screening genus *Penicillium* for producers of cellulolytic and xylanolytic enzymes. *Appl. Biochem. Biotechnol.* **113-116**, 389–401.
- **Mackenzie, C., Bilous, D., Schneider, H., and Johnson, K.** 1987. Induction of cellulolytic and xylanolytic enzyme systems in *Streptomyces* spp. *Appl. Environ. Microbiol.* **53**, 2835–2839.
- **Mandels, M. and Weber, J.** 1969. The production of cellulases. *Adv. Chem.* **95**, 391–414.
- **Matsumura, Y., Minowa, T., and Yamamoto, H.** 2005. Amount, availability, and potential use of rice straw (agricultural residue) biomass as an energy resource in Japan. *Biomass Bioenergy* **29**, 347– 354.
- **Membrillo, I., Sanchez, C., Meneses, M., Favela, E., and Loera, O.** 2008. Effect of substrate particle size and additional nitrogen source on production of lignocellulolytic enzymes by *Pleurotus ostreatus* strains. *Bioresour. Technol.* **99**, 7842–7847.
- **Miller, G.** 1959. Use of dinitrosalicilic acid reagent for determination of reducing sugar. *Anal. Chem.* **31**, 426–428.
- **Misselbrook, T., Menzi, H., and Cordovil, C.** 2012. Preface recycling of organic residues to agriculture: Agronomic and environmental impacts. *Agric. Ecosyst. Environ.* **160**, 1–2.
- **Mohanan, K., Ratnayake, R., Mathaniga, K., Abayasekara, C., and Gnanavelrajah, N.** 2014. Effect of co-culturing of cellulolytic fungal isolates for degradation of lignocellulosic material. *J. Yeast Fungal Res.* **5**, 31–38.
- **Ortiz Escobar, M. and Hue, N.** 2008. Temporal changes of selected chemical properties in three manure–amended soils of Hawaii. *Bioresour. Technol.* **99**, 8649–8654.
- **Pathak, H., Singh, R., Bhatia, A., and Jain, N.** 2006. Recycling of rice straw to improve wheat yield and soil fertility and reduce atmospheric pollution. *Paddy Water Environ.* **4**, 111–117.
- **Qinnghe, C., Xiaoyu, Y., Tiangui, N., Chenga, J., and Qiugang, M.** 2004. The screening of culture condition and properties of xylanase by white-rot fungus *Pleurotus ostreatus*. *Process Biochem.* **39**, 1561–1566.
- **Quevedo-Hidalgo, B., Narvaéz-Rincón, P.C., Pedroza-Rodríguez, A.M., and Velásquez-Lozano, M.E.** 2012. Degradation of chrysanthemum (*Dendranthema grandiflora*) wastes by *Pleurotus ostreatus* for the production of reducing sugars. *Biotechnol. Bioprocess Eng.* **17**, 1103–1112.
- **Santoyo, F., Gonzalez, A., Terron, M., Ramirez, L., and Pisabarro, A.** 2008. Quantitative linkage mapping of lignin-degrading enzymatic activities in *Pleurotus ostreatus*. *Enzyme Microb. Technol.* **43**, 137–143.
- **Sehnem, N., de Bittencourt, L., Camassola, M., and Dillon, A.** 2006. Cellulase production by *Penicillium echinulatum* on lactose. *Appl. Microbiol. Biotechnol.* **72**, 163–167.
- **Shindo, H. and Nishio, T.** 2005. Immobilization and remineralization of N following addition of wheat straw into soil: Determination of gross n transformation rates by $\mathrm{^{15}N}$ -ammonium isotope dilution technique. *Soil Biol. Biochem.* **37**, 425–432.
- **Singh, D., Zeng, J., Laskar, D., Deobald, L., Hiscox, W., and Chen, S.**

2011. Investigation of wheat straw biodegradation by *Phanerochaete chrysosporium*. *Biomass Bioenerg.* **35**, 1030–1040.

- **Songulashvili, G., Elisashvili, V., Wasser, S., Nevo, E., and Hadar, Y.** 2007. Basidiomycetes laccase and manganese peroxidase activity in submerged fermentation of food industry wastes. *Enzyme Microb. Technol.* **41**, 57–61.
- **Taniguchi, M., Suzuki, H., Watanabe, D., Sakai, K., Hoshino, K., and Tanaka, T.** 2005. Evaluation of pretreatment with *Pleurotus ostreatus* for enzymatic hydrolysis of rice straw. *J. Biosci. Bioeng.* **100**, 637–643.
- **Tejada, M., Gonzalez, J., García-Martínez, A., and Parrado, J.** 2008. Application of a green manure and green manure composted with beet vinasse on soil restoration: Effects on soil properties. *Bioresour. Technol.* **99**, 4949–4957.
- **Tinoco, R., Pickard, M., and Vazquez-Duhalt, R.** 2001. Kinetic differences of purified laccases from six *Pleurotus ostreatus* strains. *Lett. Appl. Microbiol.* **32**, 331–335.
- **Valásková, V. and Baldrian, P.** 2006. Estimation of bound and free fractions of lignocellulose-degrading enzymes of wood-rotting fungi *Pleurotus ostreatus*, *Trametes versicolor* and *Piptoporus betulinus*. *Res. Microbiol.* **157**, 119–124.
- **Walkley, A. and Black, I.A.** 1934. An examination of the Degtjareff method for determining soil organic matter and a proposed modification of the chromic acid titration method. *Soil Sci*. **37**, 29–38.
- **Yao, S., Zhang, B., and Hu, F.** 2011. Soil biophysical controls over rice straw decomposition and sequestration in soil: The effects of drying intensity and frequency of drying and wetting cycles. *Soil Biol. Biochem.* **43**, 590–599.