# **Microfluidics-generated graphene oxide microspheres and their application to removal of perfluorooctane sulfonate from polluted water**

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### **KEYWORDS**

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# **ABSTRACT**

Monodisperse graphene oxide (GO) microspheres were synthesized via microfluidics technology as a novel adsorbent for rapid (in 2 min) and high efficiency (98%) removal of perfluorooctane sulfonate (PFOS) from water. This novel material is a potential solution for treatment of bioaccumulative organic polluted water. To achieve improved performance,  $Mg^{2+}$  was introduced into GO, and the metal composite exhibited significantly improved PFOS removal efficiency owing to bridging and interaction between Mg<sup>2+</sup> and the PFOS molecules, which was supported by density functional theory and X-ray photoelectron spectroscopy (XPS). This facile strategy may be extended to the synthesis of other spheres with unique structural features for application in water treatment.

# **1 Introduction**

Perfluorooctane sulfonate (PFOS), having a fully fluorinated 8-carbon alkane sulfonate, is a new, persistent, and bioaccumulative organic pollutant that exists in water and human bodies worldwide and is becoming a severe global environmental problem [1–8]. The release and control of PFOS is an emerging challenge in the environmental field [9–18]. Therefore,

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it is essential to develop an effective method for removal of PFOS from surface water to ensure the safety of water resources for human beings. To this end, adsorption technology is a proven and effective method for removal of PFOS from water using carbonaceous materials [19–24]. Among the carbonaceous adsorbent candidates, graphene oxide (GO), as a newly emerging member of the graphite family with a unique structure, exhibits great promise for potential application in the water treatment field [25–32]. However, to date, there is no reported investigation on the utilization of GO as an adsorbent to remove PFOS from water. It is a timely matter to find an effective adsorption medium to tackle this critical water pollution problem. Porous microspheres are one of the top choices for this purpose due to their high specific surface area, delivering ability, and surface permeability, and consequently higher adsorption rate.

Previously, microfluidic methods have been developed as a novel approach for the controllable synthesis of monodisperse particles. The microfluidic method also has many advantages such as stable yield, controllable structures, and uniform particle size [33–42]. Because of the ability to control the structures of the final emulsion, the microfluidic method has emerged as a promising and versatile technique to generate monodisperse emulsion droplets such as microgels, polymeric particles, compounds, and composite particles [43–49].

Herein, we describe a straightforward microfluidic approach for generating monodisperse hydrogel microspheres with GO microparticles as an effective adsorbent. The GO concentration in the resultant microspheres can be finely tuned and the microspheres have a uniform GO distribution. Magnesium ions are commonly present and detected in water, but there is little information available regarding exploitation of magnesium for PFOS removal by adsorption. Based on the potential and charge neutralization capacity of magnesium ions, these ions may strengthen the cross-linkage between the PFOS contaminants and affect the separation performance. Herein, we investigate the efficiency of these microspheres for PFOS removal from surface water samples with and without  $Mg^{2+}$  ions. The improved PFOS removal efficiency observed in the presence of  $Mg^{2+}$  is interpreted by simulating the geometry and bonding of Mg2+ and PFOS molecules based on density functional theory (DFT) and X-ray photoelectron spectroscopy (XPS) analysis.

### **2 Experimental**

#### **2.1 Preparation of graphene oxide microspheres**

Water-in-oil (W/O) emulsion drops were first generated

as templates for preparing the microspheres using a polydimethylsiloxane (PDMS)-based microfluidic device made by standard soft-lithography. The inner phase comprised an aqueous solution containing 20 wt.% poly-(ethylene glycol)diacrylate (PEG-DA,  $M_W$  = 750, Sigma-Aldrich), 0.5 wt.% 2-hydroxy-2methylpropiophenone as the photoinitiator (Sigma-Aldrich), and 0.5 wt.% double layer graphene oxide (US Research Nanomaterials). The outer phase was composed of a fluorinated oil, HFE 7500 (3 M), with 1 wt.% of Krytox-PEG surfactant to stabilize the emulsion drops against coalescence. The velocity of the two phases in the microfluidic device was controlled by syringe pumps (PHD 2000, Harvard Apparatus, USA). Monodisperse drops were generated by controlling the flow rates of the outer phase and inner phase at 1,000 and 200 μL/h, respectively. UV light (365 nm) was then used to irradiate the emulsion drops for photopolymerization to form microspheres, in which the GO material was embedded in the PEG-DA network. Subsequently, the microspheres were washed with isopropyl alcohol and deionized water three times, respectively. At the end, intact microspheres were obtained.

#### **2.2 Characterization**

Bright-field and fluorescence images were obtained with a Leica inverted fluorescence microscope equipped with a digital camera (Hamamatsu, C4742-95) and Simple PCI acquisition software (Compix). Optical and confocal microscopy images were obtained with a Leica TCS SP5 confocal microscope. The morphology and structure of the microspheres was characterized using a Hitachi S-800 scanning electron microscope (SEM). XPS spectra were acquired with a PHI Quantera SXM Scanning X-ray microprobe. The operating pressure was  $4.5 \times 10^{-12}$  Bar with an Al-K $\alpha$  X-ray anode radiation source was as the excitation source (*hν* = 1,486.7 eV). XPS peak software was used for curve fitting analysis of the C 1s spectra. The XPS spectra used Shirley-type background subtraction; the Gaussian–Lorentzian peak shape was used after performing background subtraction. The zeta potential of the membrane surfaces was measured by using a Zetasizer Nano ZS 90 series (Malvern, UK) instrument.

#### **2.3 Adsorption experiments**

The adsorption performance of the GO microspheres was examined by calculating the change in the concentration of PFOS (99%, AccuStandard, Inc., USA). A 25 mL aqueous solution was prepared with 100 ppb PFOS,  $0.005$  g GO microspheres, and  $1$  mM  $MgCl<sub>2</sub>$ (99%, Fisher Scientific) at  $pH = 7$ , 25 °C. The solution was vibrated on an orbital shaker at a speed of 130 rpm. A fixed volume (10 μL) of solution was withdrawn at set time intervals of 20 s, filtered through a  $0.22 \mu m$ membrane filter needle, and injected into the ultraperformance liquid chromatography-tandem mass spectrometer (UPLC-MS/MS) for analysis. The PFOS concentrations were measured via UPLC-MS/MS using an UPLC system (Waters Corp., USA) equipped with a C18 column (BEH, particle size 1.7 μm, 2.1 × 150 mm i.d.) and an MS system Quattro Premier XE tandem quadrupole mass spectrometer (Waters Corp., USA) with an electrospray ionization source. The mobile phase contained a binary mixture of (a) 10 mmol/L ammonium acetate in MilliQ water and (b) acetonitrile applied at a flow rate of 0.2 mL/min. The ratio of (a) to (b) was 3:7 and the injection volume was  $10 \mu L$ . The cone voltage and collision energy were 30 V and

11 eV respectively, and the selected ion monitoring (SIM) model was used to quantify the analytes (*m/z* = 499.05 for PFOS).

The PFOS removal efficiency (*R*) was defined as

$$
R\left(\% \right) = \left(1 - \frac{C_{\rm p}}{C_{\rm f}}\right) \times 100\tag{1}
$$

where  $C_f$  and  $C_p$  are the PFOS concentrations in the feed and residual, respectively.

#### **2.4 DFT calculation**

The DFT Becke three-parameter nonlocal exchange functional [47] was used to calculate the geometry of the molecules. The frequencies were used to confirm the optimized structures as ground states without an imaginary frequency.

# **3 Results and discussion**

#### **3.1 Preparation and characterization of droplets**

A cross-junction microfluidic device was used to generate monodisperse drops with the inner phase broken up by the outer phase (Fig.  $1(a)$ ). The drops



**Figure 1** (a) Formation of W/O emulsion drops at the cross-junction. (b) Optical micrograph of the monodisperse drops containing PEG-DA and graphene. (c) Optical micrograph of the deformed drops. (d1) Fluorescence microscope image of the drops. (d2) Brightfield microscope image of the drops at the same area as in (d1). (e1) Fluorescence microscope image of the polymerized microspheres. (e2) Bright-field microscope image of polymerized microspheres at the same area as in (e1). (f) Cryo-SEM images of a microsphere showing the porous surface structure. (g) Cryo-SEM image of the cross-section of a microsphere.

were collected in a chamber with two parallel glass slides and the drops assembled into a single-layer hexagonal packing structure due to the high monodispersity, as shown in Fig. 1(b). After removing the oil phase from the suspension, the droplets were transferred onto the glass substrate. The drops deformed into an hexagonal shape due to partial evaporation of the volatile HFE oil (Fig. 1(c)).

To characterize the uniformity of the structure, the inner fluid was doped with a fluorescent dye to observe the emulsions under confocal microscopy. Figures 1(d1) and 1(d2) demonstrate the excellent encapsulation efficiency based on encapsulation of the hydrophilic fluorescent dye, where the fluorescent dye was encapsulated inside the polymersome. After irradiating the drops with UV light (365 nm) in a chamber for 10 min, monodisperse GO microspheres were obtained, as shown in Figs. 1(e1) and 1(e2). The fluorescence intensity for the drops and microspheres could be quantified as 49.95 and 47.79, respectively, which further indicated that uniform microspheres were generated. Based on cryo-SEM visualization (Figs. 1(f) and 1(g)), the microspheres were porous, both on the surface and internally, which is advantageous for pollutant removal.

#### **3.2 Adsorption performance**

To verify the adsorption performance of the GO microspheres, we first prepared two common aqueous solutions containing sulforhodamine B as the model molecules, and added GO microspheres to one of them (Fig. 2(a)). For the sample with the added GO microspheres, after 10 min, the solution changed to almost colorless from its original red color (Fig. 2(b)), again clearly indicating that the developed microspheres can effectively adsorb free molecules in the aqueous environment. We also used sulforhodamine B to characterize the adsorption performance of the microspheres. The microspheres without fluorescent dye were dispersed in an aqueous solution of sulforhodamine B (Fig. 2(a)).

Over a period of about 2 min, the dye gradually diffused to the inside of the spheres, where the dye accumulated in the spheres due to the presence of GO, as shown in Figs. 2(d) and 2(e). This confirms the



**Figure 2** Effective absorption of fluorescent dye molecules in bulk by PEGDA-GO microspheres: (a) 0 and (b) 10 min after adding PEGDA-GO microspheres. Time-lapse confocal microscope images showing the process of dye diffusion into the microspheres at (c) 0, (d) 1 and (e) 2 min.

great capacity of the GO microspheres to adsorb free molecules in water due to their porous structure.

Furthermore, to demonstrate the application of the GO microspheres to removal of persistent organic pollutants, the GO microspheres, as a novel adsorbent, were applied to PFOS removal. Because common divalent cations are co-existent in real water samples, here, magnesium ions were used as a representative to illustrate the effect of the cation type. Two feed systems, one with 100 ppb PFOS only and the other with a mixture of 100 ppb PFOS and 1 mM  $Mg^{2+}$ , were examined. A control experiment was also performed by adding PEG-DA microspheres without encapsulated GO.

Figure 3 shows that the GO microspheres rapidly removed PFOS within 1 min, and equilibrium was reached within approximately 2 min with a removal efficiency of >95%. To the best of our knowledge, the adsorption speed and capacity of  $5,300 \text{ mg} \cdot \text{g}^{-1}$  s is much higher than previously reported values [3]. This superior performance may be ascribed to the merits of the GO structure for adsorption and the diffusion capacity. However, low PFOS removal efficiency was achieved by using the microspheres without GO. This confirms that GO plays an important role in the PFOS adsorption process. We attribute the marked enhancement of the PFOS removal efficiency in the initial 1 min to the fact that PFOS was first primarily adsorbed on the GO microspheres.



**Figure 3** Removal of PFOS by GO microspheres versus spheres without encapsulated GO.

Interestingly, it was also found that the presence of  $Mg^{2+}$  can improve the PFOS removal efficiency (98%) relative to the microspheres without  $Mg^{2+}$ . This might be because the microsphere surface is negatively charged (the zeta potential is  $-5$  mV), and therefore the  $Mg^{2+}$  ions improve the surface adsorption capacity. Moreover,  $Mg^{2+}$  can bridge and interact with the PFOS molecules, thereby increasing the adsorption capability.

#### **3.3 Mechanism analysis**

To further study the mechanism of the interaction between the  $Mg^{2+}$  and PFOS molecules on the microscopic scale, DFT was used to simulate the interaction of Mg2+ with the PFOS molecules. The simulated geometries and structural parameters of the Mg<sup>2+</sup> ionbound PFOS molecules are presented in Fig. 4 and Table 1.

Figures 4(a) and 4(b) respectively show the structures of H<sub>2</sub>O and Mg(H<sub>2</sub>O) $_6^{2+}$  as the stable species in solution. Frontier orbital analysis shows that the highest occupied molecular orbital (HOMO) of anionic PFOS is primarily localized on the sulfonate head group (Fig.  $4(c)$ ), and the  $-SO<sub>3</sub>$  groups are the active sites for binding with the positively charged  $Mg^{2+}$  ion. The  $Mg<sup>2+</sup>$  ion may interact with one or two PFOS anions to yield two bridging structures as follows

$$
CF_3(CF_2)_7SO_3^- + Mg(H_2O)_6^{2+} \to
$$
  
\n
$$
Mg[CF_3(CF_2)_7SO_3]^+(H_2O)_4 + 2H_2O
$$
 (2)



**Figure 4** (a) Structure of H<sub>2</sub>O, (b) structure of Mg(H<sub>2</sub>O)<sup>2+</sup>, (c) structure and HOMO of PFOS, (d) geometry of Mg<sup>2+</sup> ion bound to one PFOS molecule, (e) geometry of Mg<sup>2+</sup> ion bound to two PFOS molecules. Yellow, dark yellow, red, gray, and blue circles denote S, Mg, O, C, and F atoms, respectively. The isosurface plots of the HOMO orbitals were generated with an isodensity value of 0.02 a.u.



**Table 1** Molecule-chain length and select parameters calculated based on DFT



$$
2CF_3(CF_2)_{7}SO_3 + Mg(H_2O)_{6}^{2+}
$$
  
\n
$$
\rightarrow CF_3(CF_2)_{7}SO_3MgO_3S(CF_2)_{7}CF_3(H_2O)_{2} + 4H_2O
$$
 (3)

Based on DFT calculation and thermodynamic analysis, the above reactions result in formation of the  $CF_3(CF_2)_7SO_3Mg^*(H_2O)_4$  or  $CF_3(CF_2)_7SO_3MgO_3S(CF_2)_7CF_3(H_2O)_2$  configuration depending on the number of PFOS molecules bound to  $Mg^{2+}$  (Figs. 4(d) and 4(e)). Formation of the  $CF_3(CF_2)_7SO_3Mg_+(H_2O)_4$  structure enhanced the polarity of the molecule (*μ* increased from 22.9 to 31.2), which promoted interaction with the negatively charged GO microsphere surface.

The vibrational properties of the two states before and after PFOS adsorption were also investigated by XPS (Fig. 5) and Table 2.

Figures 5(a) and 5(c) show the C 1s XPS spectra of GO and the standard PFOS material; the corresponding binding energies are summarized in Table 2. When  $Mg^{2+}$  and PFOS were co-existent (Fig. 5(c)), the spectra of PFOS with  $MgCl<sub>2</sub>$  showed significant changes; the C–C peak was negatively shifted from 284.4 to 283.8 eV and a new peak (288.1 eV) was observed that has not been reported in previous literature. The disappearance of the  $CF_2$  peak and the shift of the  $CF_3$  peak to 285.1 eV indicate that the electron cloud density of



Figure 5 XPS spectra. (a) Graphene oxide, (b) standard PFOS material, (c) PFOS with MgCl<sub>2</sub>, (d) graphene oxide with MgCl<sub>2</sub>, (e) graphene oxide with PFOS,  $(f)$  graphene oxide with PFOS and MgCl<sub>2</sub>.

Substance	Binding energy (eV)	The types of bonds
Graphene oxide [48]	283.8	$C-C$
	285.9	$C$ -OH
	287.5	$C = O$
<b>PFOS</b> [49]	284.4	$C-C$
	298.7	$C-S$
	291.2	CF <sub>2</sub>
	292.9	CF <sub>3</sub>
$PFOS + MgCl2$	283.8	$C-C$
	285.1	CF <sub>x</sub>
	288.1	Unknown
Graphene oxide + MgCl <sub>2</sub>	284.2	$C-C$
	286.3	$C$ -OH
	287.2	$C = O$
Graphene oxide + <b>PFOS</b>	284.8	$C-C$
	287.0	$C=O$
	288.9	$-COO^-$
	291.4	CF <sub>2</sub>
	293.3	CF <sub>3</sub>
	296.0	Unknown
Graphene + PFOS	284.6	$C-C$
$+ MgCl2$	286.9	$C=O$
	287.5	$C-O-C$
	291.2	$CF_x$

**Table 2** Binding energy data

these two functional groups increased. Therefore, these phenomena prove that  $Mg^{2+}$  can chemically bond to the negative  $SO_3^-$  groups of the PFOS molecules. Adding  $MgCl<sub>2</sub>$  to GO (as shown in Fig. 5(d)) caused the C–C bond energies to shift from 283.8 to 284.3 eV. The disappearance of the C–OH peak (285.9 eV) in Fig. 5(a) after the adsorption process and the appearance of the peak at 286.3 eV confirm that  $Mg^{2+}$ is chemically bonded to the GO adsorbent. The C=O binding energy at 287.5 eV was negatively shifted to 287.2 eV, demonstrating that  $Mg^{2+}$  can increase the electron cloud density of GO. The  $Mg^{2+}$  ions can be adsorbed on the negative sites of the GO surface. Figure 5(e) demonstrates that the addition of PFOS to GO caused significant changes, including the appearance of the C 1s peak (288.9 and 296.0 eV) after

adsorption, confirming that the PFOS molecules are chemically bonded to the GO adsorbent. The disappearance of the C 1s peak at 285.9 eV and the negative shift of the C=O peak to 287.0 eV show that there are some interactions between the carboxyl group and the PFOS molecules. The positive shift of the  $CF_2$  and the  $CF_3$  peaks from 291.2 to 291.4 eV and 292.9 to 293.3 eV, respectively, confirm that GO contained adsorbed PFOS molecules and the electron cloud density decreased. When  $Mg^{2+}$  and PFOS were both added (Fig.  $5(f)$ ), the C–C bond energy was positively shifted from 283.8 to 284.6 eV and the C=O peak at 287.5 eV was negatively shifted to 286.9 eV. The disappearance of the C–OH peak at 285.9 eV might indicate transformation of these species to C–O–C groups with a binding energy of 287.5 eV, and the new peaks at 291.2 eV are attributed to CF*x*. Comparison of Figs. 5(d) and 5(e) shows that GO can adsorb both PFOS molecules and  $Mg^{2+}$  ions. Moreover, the adsorbed  $Mg^{2+}$  ions can attract the anionic  $SO_3^$ groups to the GO surface, thereby playing a bridging role in neutralizing the negative surface charge. The  $-SO_3-Mg^+$  groups on the molecules then undergo attractive interactions with the negatively charged GO surface and can be adsorbed on the GO surface. The presence of  $Mg^{2+}$  would increase the capacity for adsorption of PFOS owing to the bridging during the adsorption process (Fig. 6).

Thus, the adsorption of PFOS on the GO microsphere surface is enhanced accordingly with improved PFOS removal efficiency.



graphene oxide.

# **4 Conclusions**

Monodisperse hydrogel microspheres with encapsulated GO were developed as a novel adsorbent for water treatment. The microspheres not only efficiently adsorbed the dye molecules in water but also removed PFOS from water, and the efficiency of PFOS removal reached 98% in the presence of  $Mg^{2+}$  within 2 min. This is attributed to the interactions and bridging of the microspheres with  $Mg^{2+}$  and PFOS molecules, which is supported by XPS analysis and density functional theory calculation. This new strategy should facilitate extensive research on microfluidic technology and GO materials, and pave the way for potential applications of these materials in water treatment in the near future.

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