# **Ultrasonic Pretreatment of Cow Dung for Anaerobic Digestion: Efect on Methane Production and Microbial Community**

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## **Abstract**

Cow dung contains high concentrations of organic matter, which can be used for methane production by anaerobic digestion. However, the refractory lignocellulose in cow dung often hiders methane production through anaerobic digestion. Ultrasonic pretreatment is an efcient method to enhance lignocellulose degradation and methane production, but the mechanism is not well studied. In this work, the infuence of ultrasonic pretreatment on the anaerobic digestion of cow dung was studied. The impact of ultrasonic pretreatment on the dissolution rate of organic matter (including soluble chemical oxygen demand (sCOD) and carbohydrates) was investigated by modifying the ultrasonic power, length, and impulse time. The dissolution rate of sCOD increased by 96% with an ultrasonic power of 325 W by comparing the treatment without ultrasonic pretreatment. The optimized dissolution rates of sCOD (10,915 $\pm$ 112 mg/L) and carbohydrate (942 $\pm$ 12 mg/L) were achieved at 325 W ultrasonic power, 30 min ultrasonic time with impulse time of 2 s close and 1 s open. Under the above condition, after 26 days of anaerobic digestion, the cumulative methane production in the treatment with ultrasonic pretreatment was 851 mL, which was 1.36 times higher than that of the control without ultrasonic pretreatment (360 mL). In comparison with the treatment without ultrasonic pretreatment, the bacteria of *Actinaobacteria* phyla, which could degrade unselective organic substances, was signifcantly increased (by 44%) in the treatment with ultrasonic pretreatment. We conclude that ultrasonic pretreatment has a high potential to enhance methane production by anaerobic digestion of cow dung.

**Keywords** Ultrasonic pretreatment · Dissolved organic matter · Cow dung · Impulse time · Microbial community

# **Introduction**

In recent years, how to economically deal with livestock waste has become very important in many parts of the world. If treated with inappropriate disposal methods, livestock manure such as cow dung (CD) may cause serious

#### **Highlights**

The dissolution rate of sCOD and carbohydrates increased by 96% and 120% after ultrasonic pretreatment.

Cumulative methane production was improved by 136% when compared to the control.

The bacterial community was signifcantly changed after ultrasonic pretreatment.

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environmental problems such as pathogens in ground and surface water, odor, airborne ammonia, greenhouse gases, spills, etc. [\[1](#page-7-0)]. Generally, a mature cow can excrete around 15 kg of CD per day, and about 35 million tons of CD are produced annually in China [[2](#page-7-1), [3](#page-7-2)]. Recently, large volumes of CD generated from feedlot farming have increased annually, and most of them are disposed into landflls or piled up around farms without sufficient treatment  $[4]$  $[4]$ . The CD contains degradable organic materials, including a high content of lignin and lignocellulose fbers, about 40–50% of the total solids  $(TS)$  [[5\]](#page-7-4). Anaerobic digestion  $(AD)$  is an efficient method widely used for bioenergy production from sewage sludge, animal manure, agricultural residue, industrial sludge, and energy crops in developing and developed countries [[5,](#page-7-4) [6](#page-7-5)]. The process of AD includes four major steps: hydrolysis, acidogenesis, acetogenesis, and methanogenesis. In the frst step, hydrolysis is the rate-limiting step of the overall process due to the dissolution of insoluble particulate matter and the biological decomposition of organic polymers with complex structures (such as lignin and cellulose) [[5](#page-7-4)[–7](#page-7-6)].



The impact of the dissolution rate of organic matter was studied based on diferent ultrasonic parameters.

Hence, to accelerate the anaerobic digestion of CD, an efficient pretreatment method is highly recommended.

A few pretreatment methods have been developed for methane production from AD of the recalcitrant fraction of manure. These methods can be divided into four classes: physical, chemical, biological, and combined pretreatment [\[8](#page-7-7)]. Thermal pretreatment is a type of physical pretreatment in which the lignocellulosic biomass is subjected to heat at a certain temperature and pressure, and it can promote organic matter hydrolysis and breakdown during AD [[9](#page-7-8)[–11](#page-7-9)]. Chemical pretreatment is a widely used method to alter the physical and chemical characteristics of lignocellulose biomass [\[11](#page-7-9)]. Biological pretreatment is used to break down the complex structures in a substrate with enzymes such as cellulose and protein [\[12](#page-7-10)]. These pretreatments alter the lignocellulose structures by destroying the chemical bonds, which may improve biogas production [[13](#page-7-11)]. Some pretreatments studied for biogas production are summarized in Table [1.](#page-1-0) However, there are certain limitations in the use of these pretreatments. For instance, thermal pretreatment needs high energy requirements, chemical pretreatment produces secondary pollution, and biological pretreatment needs to monitor microorganism growth [\[13](#page-7-11)].

Ultrasonic pretreatment is a type of physical pretreatment that is simple, time saving, environmentally friendly, and has no chemical addition [[20](#page-8-0), [21](#page-8-1)]. The ultrasonic treatment is performed through the cavitation efect, including mechanical, thermal, and sonochemical effects [[22\]](#page-8-2). Previous research has established that ultrasonic technology can damage the fber structure, break the hydrogen bonds between fber molecules, and also increase the inner surface area of fiber  $[23, 24]$  $[23, 24]$  $[23, 24]$  $[23, 24]$ . Quiroga et al.  $[25]$  $[25]$  evaluated the effect of ultrasonic pretreatment in methane production when codigesting CD with food waste and sewage sludge. The results showed that after sonication treatment of CD and sludge, the

<span id="page-1-0"></span>**Table 1** Diferent pretreatments on biogas production

methane yield  $(0.85 \text{ L CH}_{4}/\text{L day})$  for mesophilic showed an increase of 31% with respect to the non-sonicated waste  $(0.74$  L CH<sub>4</sub>/L day). Ormaechea et al. [\[26](#page-8-6)] studied the pretreatment by ultrasound (sonication energy of 1040 kJ/kg TS) for a mixture of cattle manure, food waste, and raw glycerine. They obtained specifc methane production of 290 L  $CH<sub>A</sub>/kg$  VS without pretreatment and 520 L CH<sub>A</sub>/kg VS with pretreatment.

Many studies have reported ultrasonic pretreatment with CD can improve biogas production. However, they just focused on the ultrasonic power and time. Few researchers discussed the infuence of ultrasonic impulse time on anaerobic digestion. Therefore, the main objectives of this study are to (1) evaluate the efect of diferent ultrasonic pretreatment conditions (power, time length, and impulse time) on CD; (2) analyze the correlation of methane production and characteristics of substrate (pH, ammonia nitrogen, and soluble organic matters) during AD process; and (3) explore the changes of microbial community after AD with and without ultrasonic pretreatment of CD.

# **Materials and Methods**

## **Substrate and Inoculum**

The dry CD was collected from a cow farm in Linyi, Shandong Province, China. The dry CD was cut into small pieces, milled, and then mixed with tap water to the TS of 10% before further use. Anaerobic seed sludge was collected from an anaerobic digester treating waste-activated sludge from the Haibo River sewage treatment plant, Qingdao. The CD contained about  $10.69 \pm 0.87\%$ ,  $8.76 \pm 0.58\%$ , and  $14.16 \pm 1.03\%$  of TS, VS (volatile solid), and C/N (carbon to nitrogen ratio). The initial characteristics of anaerobic seed



sludge were TS 5.11 (0.53%), VS 3.82 (0.39%), and C/N 7.21 (0.64%). The substrates and inoculum were individually homogenized and stored in a refrigerator at 4 °C before further use.

## **Ultrasonic Pretreatment**

Ultrasonication was performed using an Ultrasonic cell pulverizer (SCIENTZ, China). The sonication frequency of the device was 20 kHz, and the maximum power input was 650 W when using a 6-mm probe. A volume of 40 mL of 10% CD added in a centrifuge tube (50 mL) was manually dosed in the ultrasonic horn. During sonication, the temperature was controlled at around 30 ℃. The pretreatment study was split into two approaches. In the frst approach based on the ultrasonic power (W) study, the prepared samples were kept at diferent applied ultrasonic powers (65, 130, 195, 260, and 325 W) [[27](#page-8-8)] at an ultrasonic time of 10 min and impulse time of 2 s opened, 1 s closed. After getting the suitable power, the second approach based on time study, the prepared samples were kept at certain times (5, 10, 15, 20, and 30 min) [[28\]](#page-8-9). At the suitable ultrasonic power and length, the third approach based on impulse time study, the samples were kept at diferent impulse times (1 s opened 1 s closed, 2 s opened 1 s closed, 2 s opened 2 s closed, 3 s opened 2 s closed, and 3 s opened 3 s closed) [[29](#page-8-10)]. The control treatment without ultrasonic pretreatment was prepared in the same way.

#### **Anaerobic Biodegradability Tests**

A batch reactor was performed to test the rate and extent of anaerobic biodegradability of untreated (control) and ultrasonic pretreated CD. All batch tests were performed at 37 ℃ in 500-mL anaerobic reactors with a working volume of 400 mL. The reactors were set with a CD-to-ASS ratio of 1 based on VS content [\[13](#page-7-11)]. All of the reactors were purged with nitrogen for at least 3 min to remove oxygen from the reactor space in order to maintain anaerobic conditions. Each reactor was connected with an inverted bottle, containing a solution of 1.5 N sodium hydroxide (NaOH), through a silicon S-tube to ensure the scrubbing of  $CO<sub>2</sub>$  and  $H<sub>2</sub>S$  from the biogas generated from the reactor [\[30](#page-8-11)]. The  $CH<sub>4</sub>$  was gathered in aluminum foil bag. The pH value of the culture was not controlled. The experiment was conducted in triplicate for 26 days.

### **Analytical Methods**

**Scanning Electron Microscopy (SEM)** The surface morphology of the adsorbents was determined by scanning electron microscopy (SEM) (SU8010, Hitachi, Japan). The raw

substrate samples were oven dried at 105 ℃ for 24 h and then analyzed after cooling to room temperature [\[17](#page-7-15)].

**pH and Lignocellulose** The pH was determined with a pH meter (Thermo, USA). The cellulose and lignin were studied by using the titration method.

**Solubility Index** Samples from the reactor were taken and the volume of methane was measured every 2 days. Before analysis, the samples were centrifuged at 10,000 rpm for 5 min and the supernatant was fltered using 0.45-μm flters. The sample's TS, VS, and soluble chemical oxygen demand (sCOD) were obtained using standard methods [[31](#page-8-12)]. The total ammonia nitrogen (TAN) content and carbohydrate were studied using the spectrophotometer (HACH, DR/2800) [\[32](#page-8-13)].

The free ammonia nitrogen (FAN) concentration was calculated based on the following equation [\[33](#page-8-14)]:

$$
FAN = \frac{TAN}{1 + \frac{10^{-pH}}{K_a}}
$$

where TAN is the total ammonia nitrogen,  $K_a$  is a dissociation constant that refects on temperature with values  $1.29 \times 10^{-9}$  for 37 °C, and pH is equal to the pH of the substance.

**Methane Production** The produced methane was collected using an aluminum foil bag and the volume of methane was measured by using a graduated 100-mL plastic syringes.

#### **Statistical Analysis**

The analysis of the solubility index was performed in triplicate sets. The average and standard deviation of the duplicates were calculated and shown using GraphPad Prism v. 7.0 (GraphPad Software, Inc.). Also, an analysis of variance and the Tukey test were performed, with a 95% confdence interval  $(p < 0.05)$ .

# **Results and Discussions**

#### **Pretreatment Efect on Chemical Characteristics**

The effect of solution characteristics with various ultrasonic powers was studied under the operation condition of the ultrasonic length of 10 min with impulse time of 2 s opened and 1 s closed. The efect of ultrasonic power with constant time on CD in ultrasonic pretreatment is shown in Fig. [1a](#page-3-0). It was observed that the increase in ultrasonic power could improve the solubilization rate, which was measured by sCOD and carbohydrate. The increase



<span id="page-3-0"></span>**Fig. 1** Efect of **a** ultrasonic power, **b** ultrasonic length, and **c** ultrasonic impulse time on sCOD and carbohydrate

of soluble organic matter indicated that the polymer was broken and the intracellular constituents were released [[22\]](#page-8-2). The maximum value was achieved by the ultrasonic power of 325 W. The maximum value of sCOD and carbohydrate reached 10,362 mg/L and 760 mg/L (94% increase in sCOD and 77% increase in carbohydrate). Qi et al. [\[21\]](#page-8-1) proved the similar conclusion that the sCOD increased in ultrasound power and peaked at 200 W-20 min, which was 66% greater than the control sCOD. Kisielewska et al. [[34](#page-8-15)] reported that ultrasound pretreatment of *Sida hermaphreodita* made the maximum COD solubilization by 21% increased with energy input (Es) ranging from 200 to 550 kJ/kg.

The effect of solution characteristics with various ultrasonic lengths was studied under the operation condition of ultrasonic power of 130 W with impulse time of 2 s opened and 1 s closed. Figure [1](#page-3-0)b illustrates the efect of ultrasonic time on CD measured by sCOD and carbohydrates. As seen in Fig. [3](#page-4-0), disintegration increased after 10 min of ultrasound and then remained stable of ultrasonic length from 10 to 20 min, and increased at the ultrasonic length of 30 min. It was observed that the highest soluble organic matters, including sCOD and carbohydrate, can be achieved by 30 min of ultrasonic (96% increase in sCOD and 120% increase in carbohydrate). Apul et al. [[28](#page-8-9)] indicated that soluble COD achieved the highest value by 15 min of ultrasonic pretreatment and increased nearly 5 times. Longer ultrasonic pretreatment time led to more complex reactions and resulted in fuctuations of soluble organic matter. The stabilization of sCOD and carbohydrate at the ultrasonic time from 10 to 20 min is most probably due to the entrapment of organics into foc structure since the released organics and polymers can have focculation during the ultrasonic pretreatment [[35](#page-8-16)].

The effect of solution characteristics with various ultrasonic impulse times was studied under the operation condition of ultrasonic power of 130 W and ultrasonic time of [1](#page-3-0)0 min. Figure 1c evaluates the effect of sCOD and carbohydrates with diferent ultrasonic impulse times. It can be seen that at the impulse time of continuous time: 2 s and pulsive time: 2 s, sCOD and carbohydrate decreased sharply. While the sCOD and carbohydrate reached the highest level, increasing 58% and 67%, respectively, with the continuous time of 2 s and pulsive time of 1 s. This phenomenon indicated that smaller bubbles dissolve into the liquid more rapidly as the pulse-off time becomes longer, while an appropriate pulse-off time could increase the number of bubbles, which were adequate for sonochemical reactions [\[36](#page-8-17)].

Considering the energy input and the effect of solubilization, we use the condition of the ultrasonic power of 325 W, ultrasonic length of 30 min with the ultrasonic continuous time of 2 s, and pulsive time of 1 s to study the anaerobic digestion process.

### **SEM Analysis**

The micromorphology of the untreated (control) and ultrasound-pretreated CD was studied by scanning electron microscopy (SEM). SEM images of control and ultrasound pretreated CD (Fig. [2\)](#page-4-1) showed signifcant changes in microscopic morphology. The control sample had a rigid, compacted smooth surface with some batches, while the surface of the ultrasoundpretreated sample had more fssures and larger hollows. This was because the micro-jet produced by the cavitation efect of ultrasonic impacted the shear substances and destroyed the interior transistor of the substances [\[37](#page-8-18)]. Thus, the results

<span id="page-4-1"></span>

<span id="page-4-0"></span>**Fig. 3 a** Cumulative and daily methane production during anaerobic digestion; **b** TAN and FAN in the anaerobic digestion process; **c** sCOD and carbohydrate during anaerobic digestion; **d** the change of pH during anaerobic digestion

concluded that the ultrasonic pretreatment damaged the physical structure of CD, which could increase the accessible surface and release more biodegradable substances.

# **The Infuence of Pretreatment on Anaerobic Digestion**

#### **Efect of Ultrasonic Pretreatment on Methane Production**

The quantity of biogas produced from CD over 26 days solid retention time (SRT) is shown in Fig. [3](#page-4-0)a. The daily methane yields of pretreatment exhibited two peaks (day 6 and day 12). In contrast, the group of control only has one peak (day 12). The maximum daily methane production was found on day 2, which was 150 and 222 mL in control and ultrasonic pretreated treatment, respectively. That may be because the dissolved organic matter and easily degradable substances existed and were digested by methanogens at the beginning of AD. The daily methane yield of ultrasonic pretreated CD reached the frst peak on day 6. This result indicates that ultrasonic pretreated CD was more easily accessible to hydrolytic bacteria at the early stage of digestion [[38](#page-8-19)].

The organic acids gained from macromolecule organic matter digested by hydrolytic acidifcation resulted in the inhibition of methanogen activity and decline in the daily menthane yield [[39](#page-8-20)]. The cumulative methane production of pretreatment increased by 136% compared with the control. The result demonstrated that ultrasonic pretreatment could increase the daily methane peak and improve the total methane production of AD. This result was consistent with Braguglia et al. [[40\]](#page-8-21) and Zou et al. [\[37](#page-8-18)]. The reason for this might be that ultrasonic pretreatment dissolves the soluble substances and strengthen the activity of methanogens.

#### **Efect of Ultrasonic Pretreatment on TAN and FAN**

The variation of TAN contents was observed to evaluate the impact of ultrasonic pretreatment on CD with AD. TAN was made up of ionized ammonium nitrogen  $(NH_4^+)$  and unionized free ammonium nitrogen (FAN/NH<sub>3</sub>). Unionized NH<sub>3</sub> is more toxic than ionized form due to its uncharged nature and solubility in lipids, which may cause the unionized  $NH<sub>3</sub>$ to pass the biological cell membranes easily and can afect the stability of the microbial cells during the methane production process [[41](#page-8-22)]. Figure [3](#page-4-0)b illustrates the evolution of TAN and FAN concentration following the AD time. The transitions of TAN of the group of control and pretreatment were dominated by an upward trend universally in the frst 14 days as the nitrogenous organics, such as protein, were degraded by hydrolytic bacteria. Then there was a brief dip in evolution, which may be because of the rapid metabolism of TAN for methanogens growth. After that, the stages in control and ultrasonic pretreatment treatment remained stable at 450 and 390 mg/L, respectively. Ultrasonic pretreatment can enhance the activity of the methanogens, which utilize more ammonium for growth metabolism and cause the TAN concentrations to decrease after stability. The variation trends of FAN and pH were similar, and pH afected FAN concentration in the AD process [[42](#page-8-23)].

### **Efect of Ultrasonic Pretreatment on sCOD and Carbohydrate**

Ultrasonic pretreatment signifcantly infuenced the sCOD and carbohydrates during the AD process. The variation of sCOD and carbohydrate is shown in Fig. [3c](#page-4-0). The sCOD and carbohydrate concentration decreased at the beginning, which was diferent from the result of Pan et al. [\[39](#page-8-20)], the sCOD was increased during the frst 10 days, and then decreased. This phenomenon indicated that the methanogenic rate was higher than the hydrolysis rate during the initial stages of AD. The dissolved organic matter was consumed higher than that produced by the microorganism. Ultrasonic pretreatment can efectively improve the consumption of biodegradable organic matter and enhance the activity of methanogenic microbic. The sCOD removal rate of pretreatment (58%) was higher than that of the control (36%). The same result was observed in the removal rates of carbohydrates, which were 47% and 35% in the treatment with ultrasonic pretreatment and control, respectively.

#### **Efect of Ultrasonic Pretreatment on pH**

One of the most important factors that directly afect biogas production during AD process is the pH value. The variation of pH during the AD process is shown in Fig. [3d](#page-4-0). The pH value showed a rapid drop during the frst 4 days for both control and pretreatment reactors. This phenomenon was attributed to the production of volatile fatty acids resulting from the digestible organic matter decomposition. Then, due to the consumption of volatile fatty acids by methanogens, the pH value increased gradually from day 4 to day 18. The optimum pH range during AD is estimated between 6.5 and 7.2, which is preferable for methanogenic archaea [\[43](#page-8-24)]. The pH of group pretreatment is lower than the group control after stable, which may be because ultrasonic pretreatment made the organic matter easier degradation and produced more volatile fatty acids.

# **Efect of Ultrasonic Pretreatment on the Removal Rate of Organic Matter**

The removal rates of VS, cellulose, and lignin are represented in Fig. [4.](#page-6-0) The VS removal rate was increased after ultrasonic pretreatment (21.8%) compared with the control (18.4%). This result indicated the substance was easily biodegraded and used after ultrasonic pretreatment. The increasing VS removal rate could be interpreted as ultrasonic pretreatment could destroy the cell structure of organic matter, and hydrolysis, the rate-limiting step of digestion, was overcome by sonication [[28\]](#page-8-9). Ultrasonic pretreatment of CD results in the alteration of the surface structure and production of oxidizing radicals that chemically attack the lignocellulosic matrix [[23](#page-8-3)]. The cellulose and lignin removal rates with ultrasonic pretreatment were 30.1% and 10.4%. Without pretreatment, the removal rate of cellulose and lignin was 29.6% and 7.5%, respectively. After anaerobic digestion, the removal rate of cellulose with ultrasonic pretreatment was increased with no signifcant change compared with the control. However, the lignin removal rate with ultrasonic pretreatment has a major diference compared with the control. Because through pretreatment, lignin is dissolved in the liquid [[44](#page-8-25)]. The hydrolysis of cellulose yields fermentable sugars such as glucose, xylose, arabinose, mannose, and galactose as by-products. These fermentable sugars can be further used as a carbon source for energy production [\[45](#page-8-26)].



<span id="page-6-0"></span>**Fig. 4** The removal rate of VS, cellulose, and lignin of control and ultrasonic pretreatment after anaerobic digestion

## **Microbic Community**

In the process of AD, microorganisms play an important role, and the stability of the microbial community can be used as an indicator of system stability. To explore which specifc microbic species were predominant under ultrasonic pretreatment, we used the barplot graph to compare the average relative distribution of microbial communities at the phylum level of bacteria and the genus level of archaea, observed with and without ultrasonic pretreatment (Fig. [5](#page-6-1)). The relative abundance of microbic community in the culture with ultrasonic pretreatment is higher than control. Figure [5a](#page-6-1) shows the relative abundance of bacterial community at the phylum level. Several phyla were found ubiquitous, and three phyla (*Proteobacteria*, *Acidobacteria*, and *Actinobacteria*) were found to be dominant*. Proteobacteria* are known for utilizing glucose, acetate, propionate, and butyrate [\[46](#page-8-27)]. *Acidobacteria* encodes a wide repertoire of carbohydrate-active enzymes involved in the breakdown, utilization, and biosynthesis of diverse carbohydrates [\[47](#page-8-28)]. *Actinobacteria* have the main functions, including degradation/decomposition of all sorts of organic substances, for example, cellulose, polysaccharides, protein fats, organic acids, and so on [\[48](#page-8-29)]. The relative abundance of *Proteobacteria* was decreased after AD, which can be explained by the biodegradable substance utilized by the microorganism. The *Acidobacteria* abundance had no noticeable changes after ultrasonic pretreatment, while the relative abundance of *Actinobacteria* was increased after pretreatment (control: 6.38%, pretreatment: 9.16%). The explanation for the increasing of *Actinobacteria* was that ultrasonic pretreatment could enhance the hydrolysis of macromolecular organic matter and the utilization of the hydrolysate.

The relative abundance of the archaeal community at the genus level is shown in Fig. [5](#page-6-1)b. *Methanothrix*, a type of acetoclastic methanogen (which utilizes acetate as the substrate to produce methane), was the most abundant. The next most abundant specie was diferent, *Methanobrevibacter* for P-0 and C-0, which is hydrogenotrophic methanogens, utilizes  $H_2/CO_2$  as the substrate to produce methane, while *Methznolinea* for C-26 and P-26, which is both hydrogenotrophic and acetoclastic methanogen, utilizes both  $H<sub>2</sub>/CO<sub>2</sub>$ and acetate as the substrate to produce methane. The third most abundant specie was *Methanomassiliicoccus*, which



<span id="page-6-1"></span>**Fig. 5** Relative abundance of predominant bacterial phylum (**a**) and archaeal genus (**b**)

is hydrogenotrophic methanogen [[49\]](#page-8-30). And it also shows that the relative abundance of *Methanothrix* decreased and *Methznolinea* increased from day 0 to day 26, both in control and ultrasonic pretreatment. The reason for this result was that the pH in the reactor increased, and the acetate substance decreased. For day 26, the abundance of *Methanothrix* in the group of ultrasonic pretreated (44%) was higher than that of the control (38%). And also, the next most abundant specie was *Methznolinea*, with 14% detected in the control culture and 17% detected in the pretreatment culture, respectively. That may be because the soluble organic matter increased, and nutrients were provided to microbes in ultrasonic pretreatment culture.

# **Conclusion**

Ultrasonic pretreatment can increase the solubilization rate of sCOD and carbohydrates in the anaerobic digestion of CD. Ultrasonic pretreatment signifcantly improves methane production by increasing the methanogen activity to utilize ammonium. During anaerobic digestion, the ultrasonic pretreatment can also change the bacterial community, where *Actinobacteria* phyla were signifcantly increased. In the future, the studies should perform the energy balance and energy beneft when using pretreatment to increase methane production.

**Author Contribution** All the authors contributed to the conception and design. Yuxing Xu: frst draft of the manuscript and data analysis; Hao Chang: experimental operation and data collection; Chen Yang: experimental operation; Changqing Liu: funding acquisition, supervision, and reviewing; Yihua Xiao: funding acquisition and reviewing.

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**Data Availability** The datasets are available from the corresponding author upon reasonable request.

#### **Declarations**

**Competing Interests** The authors declare no competing interests.

# **References**

- <span id="page-7-0"></span>1. Harikishan S, Sung S (2003) Cattle waste treatment and class A biosolid production using temperature-phased anaerobic digester. Adv Environ Res 7:701–706. [https://doi.org/10.1016/S1093-](https://doi.org/10.1016/S1093-0191(02)00034-5) [0191\(02\)00034-5](https://doi.org/10.1016/S1093-0191(02)00034-5)
- <span id="page-7-1"></span>2. Li K, Yang Z, Zhang Y et al (2022) Efect of pretreated cow dung fber on mechanical and shrinkage properties of cementitious

composites. J Clean Prod 348:131374. [https://doi.org/10.1016/j.](https://doi.org/10.1016/j.jclepro.2022.131374) [jclepro.2022.131374](https://doi.org/10.1016/j.jclepro.2022.131374)

- <span id="page-7-2"></span>3. Shen R, Chen R, Yao Z et al (2021) Engineering and microbial characteristics of innovative lab and pilot continuous dry anaerobic co-digestion system fed with cow dung and corn straw. Bioresour Technol 342:126073. [https://doi.org/10.1016/j.biortech.2021.](https://doi.org/10.1016/j.biortech.2021.126073) [126073](https://doi.org/10.1016/j.biortech.2021.126073)
- <span id="page-7-3"></span>4. Liu Y, Wang T, Xing Z et al (2022) Anaerobic co-digestion of Chinese cabbage waste and cow manure at mesophilic and thermophilic temperatures : digestion performance, microbial community, and biogas slurry fertility. Bioresour Technol 363:127976. <https://doi.org/10.1016/j.biortech.2022.127976>
- <span id="page-7-4"></span>5. Alfa IM, Dahunsi SO, Iorhemen OT et al (2014) Comparative evaluation of biogas production from poultry droppings, cow dung and lemon grass. Bioresour Technol 157:270–277. [https://doi.org/](https://doi.org/10.1016/j.biortech.2014.01.108) [10.1016/j.biortech.2014.01.108](https://doi.org/10.1016/j.biortech.2014.01.108)
- <span id="page-7-5"></span>6. Liu L, Zhang T, Wan H et al (2015) Anaerobic co-digestion of animal manure and wheat straw for optimized biogas production by the addition of magnetite and zeolite. Energy Convers Manag 97:132–139.<https://doi.org/10.1016/j.enconman.2015.03.049>
- <span id="page-7-6"></span>7. Veluchamy C, Raju VW, Kalamdhad AS (2017) Prerequisite – an electrohydrolysis pretreatment for anaerobic digestion of lignocellulose waste material. Bioresour Technol 235:274–280. [https://](https://doi.org/10.1016/j.biortech.2017.03.137) [doi.org/10.1016/j.biortech.2017.03.137](https://doi.org/10.1016/j.biortech.2017.03.137)
- <span id="page-7-7"></span>8. Meegoda JN, Li B, Patel K, Wang LB (2018) A review of the processes, parameters, and optimization of anaerobic digestion. Int J Environ Res Public Health 15(10):2224. [https://doi.org/10.](https://doi.org/10.3390/ijerph15102224) [3390/ijerph15102224](https://doi.org/10.3390/ijerph15102224)
- <span id="page-7-8"></span>9. Atelge MR, Atabani AE, Banu JR et al (2020) A critical review of pretreatment technologies to enhance anaerobic digestion and energy recovery. Fuel 270:117494. [https://doi.org/10.1016/j.fuel.](https://doi.org/10.1016/j.fuel.2020.117494) [2020.117494](https://doi.org/10.1016/j.fuel.2020.117494)
- 10. Menardo S, Airoldi G, Balsari P (2012) The efect of particle size and thermal pre-treatment on the methane yield of four agricultural by-products. Bioresour Technol 104:708–714. [https://doi.](https://doi.org/10.1016/j.biortech.2011.10.061) [org/10.1016/j.biortech.2011.10.061](https://doi.org/10.1016/j.biortech.2011.10.061)
- <span id="page-7-9"></span>11. Zhou S, Zhang Y, Dong Y (2012) Pretreatment for biogas production by anaerobic fermentation of mixed corn stover and cow dung. Energy 46:644–648. [https://doi.org/10.1016/j.energy.2012.](https://doi.org/10.1016/j.energy.2012.07.017) [07.017](https://doi.org/10.1016/j.energy.2012.07.017)
- <span id="page-7-10"></span>12. Oh Y, Shih I, Tzeng Y, Wang S (2000) Protease produced by Pseudomonas aeruginosa K-187 and its application in the deproteinization of shrimp and crab shell wastes. Enzyme Microb Technol 27:3–10. [https://doi.org/10.1016/S0141-0229\(99\)00172-6](https://doi.org/10.1016/S0141-0229(99)00172-6)
- <span id="page-7-11"></span>13. Rajput AA, Zeshan VC (2018) Efect of thermal pretreatment on chemical composition, physical structure and biogas production kinetics of wheat straw. J Environ Manage 221:45–52. [https://doi.](https://doi.org/10.1016/j.jenvman.2018.05.011) [org/10.1016/j.jenvman.2018.05.011](https://doi.org/10.1016/j.jenvman.2018.05.011)
- <span id="page-7-12"></span>14. McVoitte WPA, Clark OG (2019) The effects of temperature and duration of thermal pretreatment on the solid-state anaerobic digestion of dairy cow manure. Heliyon 5:e02140. [https://doi.](https://doi.org/10.1016/j.heliyon.2019.e02140) [org/10.1016/j.heliyon.2019.e02140](https://doi.org/10.1016/j.heliyon.2019.e02140)
- <span id="page-7-13"></span>15. Wang C, Shao Z, Qiu L (2021) The solid-state physicochemical properties and biogas production of the anaerobic digestion of corn straw pretreated by microwave irradiation. R Soc Chem 11:3575–3584.<https://doi.org/10.1039/d0ra09867a>
- <span id="page-7-14"></span>16. Veluchamy C, Raju VW, Kalamdhad AS (2018) Electrohydrolysis pretreatment for enhanced methane production from lignocellulose waste pulp and paper mill sludge and its kinetics. Bioresour Technol 252:52–58. <https://doi.org/10.1016/j.biortech.2017.12.093>
- <span id="page-7-15"></span>17. Jafar M, Pang Y, Yuan H et al (2016) Wheat straw pretreatment with KOH for enhancing biomethane production and fertilizer value in anaerobic digestion. Chinese J Chem Eng 24:404–409. <https://doi.org/10.1016/j.cjche.2015.11.005>
- <span id="page-7-16"></span>18. Breton-deval L, Méndez-acosta HO, González-álvarez V et al (2018) Agave tequilana bagasse for methane production in batch

and sequencing batch reactors : acid catalyst effect, batch optimization and stability of the semi-continuous process. J Environ Manage 224:156–163.<https://doi.org/10.1016/j.jenvman.2018.07.053>

- <span id="page-8-7"></span>19. Shen F, Li H, Wu X et al (2018) Efect of organic loading rate on anaerobic co-digestion of rice straw and pig manure with or without biological pretreatment. Bioresour Technol 250:155–162. [https://](https://doi.org/10.1016/j.biortech.2017.11.037) [doi.org/10.1016/j.biortech.2017.11.037](https://doi.org/10.1016/j.biortech.2017.11.037)
- <span id="page-8-0"></span>20. Zou S, Wang H, Wang X et al (2016) Application of experimental design techniques in the optimization of the ultrasonic pretreatment time and enhancement of methane production in anaerobic co-digestion. Appl Energy 179:191–202. [https://doi.org/10.1016/j.apenergy.](https://doi.org/10.1016/j.apenergy.2016.06.120) [2016.06.120](https://doi.org/10.1016/j.apenergy.2016.06.120)
- <span id="page-8-1"></span>21. Qi N, Zhao X, Zhang L et al (2021) Performance assessment on anaerobic co-digestion of Cannabis ruderalis and blackwater: ultrasonic pretreatment and kinetic analysis. Resour Conserv Recycl 169:105506.<https://doi.org/10.1016/j.resconrec.2021.105506>
- <span id="page-8-2"></span>22. Chen W, Gao X, Xu H et al (2017) Infuence of extracellular polymeric substances (EPS) treated by combined ultrasound pretreatment and chemical re-focculation on water treatment sludge settling performance. Chemosphere 170:196–206. [https://doi.org/10.1016/j.](https://doi.org/10.1016/j.chemosphere.2016.12.004) [chemosphere.2016.12.004](https://doi.org/10.1016/j.chemosphere.2016.12.004)
- <span id="page-8-3"></span>23. Hassan SS, Williams GA, Jaiswal AK (2018) Emerging technologies for the pretreatment of lignocellulosic biomass. Bioresour Technol 262:310–318.<https://doi.org/10.1016/j.biortech.2018.04.099>
- <span id="page-8-4"></span>24. Chu X, Cheng Q, Xu Y et al (2021) Anaerobic digestion of corn straw pretreated by ultrasonic combined with aerobic hydrolysis. Bioresour Technol 341:125826
- <span id="page-8-5"></span>25. Quiroga G, Castrillón L, Marañón E et al (2014) Effect of ultrasound pre-treatment in the anaerobic co-digestion of cattle manure with food waste and sludge. Bioresour Technol 154:74–79. [https://doi.](https://doi.org/10.1016/j.biortech.2013.11.096) [org/10.1016/j.biortech.2013.11.096](https://doi.org/10.1016/j.biortech.2013.11.096)
- <span id="page-8-6"></span>26. Ormaechea P, Castrillón-pelaez L, Marañón E et al (2017) Infuence of the ultrasound pretreatment on anaerobic digestion of cattle manure, food waste and crude glycerine. Environ Technol 38:682– 686. <https://doi.org/10.1080/09593330.2016.1208278>
- <span id="page-8-8"></span>27. Wang H, Yang SC, Cai W et al (2019) Enhanced organic matter and nutrient release from waste activated sludge using ultrasound and surfactant synergetic pre-treatment. Bioresour Technol Reports 6:32–38.<https://doi.org/10.1016/j.biteb.2019.01.017>
- <span id="page-8-9"></span>28. Apul OG, Sanin FD (2010) Ultrasonic pretreatment and subsequent anaerobic digestion under diferent operational conditions. Bioresour Technol 101:8984–8992.<https://doi.org/10.1016/j.biortech.2010.06.128>
- <span id="page-8-10"></span>29. Qu R, Tang M, Wang Y, et al (2020) TEMPO-oxidized cellulose fbers from wheat straw: efect of ultrasonic pretreatment and concentration on structure and rheological properties of suspensions. Carbohydr Polym 117386.<https://doi.org/10.1016/j.carbpol.2020.117386>
- <span id="page-8-11"></span>30. Dhamodharan K, Kumar V, Kalamdhad AS (2015) Efect of diferent livestock dungs as inoculum on food waste anaerobic digestion and its kinetics. Bioresour Technol 180:237–241. [https://doi.org/10.](https://doi.org/10.1016/j.biortech.2014.12.066) [1016/j.biortech.2014.12.066](https://doi.org/10.1016/j.biortech.2014.12.066)
- <span id="page-8-12"></span>31. Association APH (2012) Standard methods for the examination of water and wastewater. DC, US, Washington
- <span id="page-8-13"></span>32. Xue Y, Liu H, Chen S et al (2015) Efects of thermal hydrolysis on organic matter solubilization and anaerobic digestion of high solid sludge. Chem Eng J 264:174–180. [https://doi.org/10.1016/j.](https://doi.org/10.1016/j.cej.2014.11.005) [cej.2014.11.005](https://doi.org/10.1016/j.cej.2014.11.005)
- <span id="page-8-14"></span>33. Angelidaki I (2013) Efect of ammonium and acetate on methanogenic pathway and methanogenic community composition. FEMS Microbiol Ecol 83:38–48.<https://doi.org/10.1111/j.1574-6941.2012.01456.x>
- <span id="page-8-15"></span>34. Kisielewska M, Rusanowska P, Dudek M et al (2020) Evaluation of ultrasound pretreatment for enhanced anaerobic digestion of Sida hermaphrodita. Bioenergy Res 13:824–832. [https://doi.org/10.1007/](https://doi.org/10.1007/s12155-020-10108-9) [s12155-020-10108-9](https://doi.org/10.1007/s12155-020-10108-9)
- <span id="page-8-16"></span>35. Laurent J, Pierra M, Casells M et al (2009) Activated sludge properties after ultrasonic and thermal treatments and their potential infuence on dewaterability. J Residuals Sci Technol 6:19–26
- <span id="page-8-17"></span>36. Tuziuti T, Yasui K, Lee J et al (2008) Mechanism of enhancement of sonochemical-reaction efficiency by pulsed ultrasound. J Phys Chem A 112:4875–4878.<https://doi.org/10.1021/jp802640x>
- <span id="page-8-18"></span>37. Zou S, Wang X, Chen Y et al (2016) Enhancement of biogas production in anaerobic co-digestion by ultrasonic pretreatment. Energy Convers Manag 112:226–235. [https://doi.org/10.1016/j.enconman.](https://doi.org/10.1016/j.enconman.2015.12.087) [2015.12.087](https://doi.org/10.1016/j.enconman.2015.12.087)
- <span id="page-8-19"></span>38. Zhu J, Wan C, Li Y (2010) Enhanced solid-state anaerobic digestion of corn stover by alkaline pretreatment. Bioresour Technol 101:7523–7528. <https://doi.org/10.1016/j.biortech.2010.04.060>
- <span id="page-8-20"></span>39. Pan J, Ma J, Liu X et al (2019) Efects of diferent types of biochar on the anaerobic digestion of chicken manure. Bioresour Technol 275:258–265.<https://doi.org/10.1016/j.biortech.2018.12.068>
- <span id="page-8-21"></span>40. Braguglia CM, Gagliano MC, Rossetti S (2012) High frequency ultrasound pretreatment for sludge anaerobic digestion: effect on floc structure and microbial population. Bioresour Technol 110:43–49. <https://doi.org/10.1016/j.biortech.2012.01.074>
- <span id="page-8-22"></span>41. Ngo T, Shahsavari E, Shah K et al (2022) Improving bioenergy production in anaerobic digestion systems utilising chicken manure via pyrolysed biochar additives: a review. Fuel 316:123374. [https://](https://doi.org/10.1016/j.fuel.2022.123374) [doi.org/10.1016/j.fuel.2022.123374](https://doi.org/10.1016/j.fuel.2022.123374)
- <span id="page-8-23"></span>42. Wang H, Fotidis IA, Angelidaki I (2015) Ammonia efect on hydrogenotrophic methanogens and syntrophic acetate-oxidizing bacteria. FEMS Microbiol Ecol 91:1–8. [https://doi.org/10.1093/femsec/](https://doi.org/10.1093/femsec/fiv130)  $fiv130$
- <span id="page-8-24"></span>43. Lahbab A, Djaafri M, Kalloum S et al (2021) Co-digestion of vegetable peel with cow dung without external inoculum for biogas production : experimental and a new modelling test in a batch mode. Fuel 306:121627. <https://doi.org/10.1016/j.fuel.2021.121627>
- <span id="page-8-25"></span>44. Ren N, Wang A, Cao G et al (2009) Bioconversion of lignocellulosic biomass to hydrogen: potential and challenges. Biotechnol Adv 27:1051–1060.<https://doi.org/10.1016/j.biotechadv.2009.05.007>
- <span id="page-8-26"></span>45. Prapinagsorn W, Sittijunda S, Reungsang A (2018) Co-digestion of napier grass and its silage with cow dung for bio-hydrogen and methane production by two-stage anaerobic digestion process. Energies 11:.<https://doi.org/10.3390/en11010047>
- <span id="page-8-27"></span>46. Li Y, Achinas S, Zhao J et al (2020) Co-digestion of cow and sheep manure: performance evaluation and relative microbial activity. Renew Energy 153:553–563. [https://doi.org/10.1016/j.renene.2020.](https://doi.org/10.1016/j.renene.2020.02.041) [02.041](https://doi.org/10.1016/j.renene.2020.02.041)
- <span id="page-8-28"></span>47. Dedysh SN, Sinninghe Damsté JS (2018) Acidobacteria. Encycl Life Sci 1–10.<https://doi.org/10.1002/9780470015902.a0027685>
- <span id="page-8-29"></span>48. Anadan R, Dhanasekaran D. MGP (2016) An Introduction to Actinobacteria. Basics Biotechnol Appl 3–38. [https://doi.org/10.5772/](https://doi.org/10.5772/62329) [62329](https://doi.org/10.5772/62329)
- <span id="page-8-30"></span>49. Wang P, Peng H, Adhikari S et al (2020) Enhancement of biogas production from wastewater sludge via anaerobic digestion assisted with biochar amendment. Bioresour Technol 309:123368. [https://](https://doi.org/10.1016/j.biortech.2020.123368) [doi.org/10.1016/j.biortech.2020.123368](https://doi.org/10.1016/j.biortech.2020.123368)

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