

Metal Biomonitoring and Comparative Assessment in Urine of Workers in Lead-Zinc and Steel-Iron Mining and Smelting

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Abstract

The exposure of heavy metals (lead (Pb), cadmium (Cd), copper (Cu), nickel (Ni), and metalloid arsenicals) and their effects on workers' health from a lead-zinc mine (145 workers) and a steel smelting plant (162 workers) was investigated. Information on subject characteristics was obtained through a questionnaire. We determined the urinary levels of Pb, Cd, Cu, Ni, and arsenicals (including inorganic arsenic (iAs), monomethylarsonic acid (MMA), and dimethylarsinic acid (DMA), as were 8-hydroxydeoxyguanosine (8-OHdG) and cystatin C. Lead-zinc mine foundry workers had significantly higher concentrations of urinary Pb, Cd, Cu, Ni, iAs, and MMA than did steel smelting plant workers. Individuals who had consumed seafood in the previous 3 days had higher concentrations of urinary Ni than did individuals who had not consumed seafood. The urinary Cd concentrations in the two groups of factory workers may have been affected by daily smoking. There was no significant difference in urinary 8-OHdG between workers from the lead-zinc mine foundry and the steel smelting plant. Urinary Pb and Cd had significant positive linear dose-dependent effects on 8-OHdG. Urinary cystatin C, a sensitive biological indicator reflecting early renal damage, was found at higher levels in lead-zinc mine workers than in steel smelting plant workers. Binary logistic regression analysis showed that age and urinary Cd were significantly associated with urinary cystatin C. These results indicated that workers from lead-zinc mines may be exposed to higher levels of heavy metals which could lead to greater risk of kidney damage.

Keywords Heavy metals · Lead-zinc mine · Occupational exposure · 8-OHdG · Cystatin C · Risk factors

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Introduction

Lead-zinc mines are rich in mineral metals such as lead (Pb), cadmium (Cd), copper (Cu), and nickel (Ni), which are widely used in the electronics, mechanical, and military industries. The emission of metals caused by mining and smelting in lead-zinc mines pollutes the surrounding atmosphere, water, and soil. This pollution is harmful not only to the occupational population but also to the residents living in the vicinity of the emissions. Environmental and occupational exposure to heavy metals induces several adverse health effects [1]. The production process in lead-zinc mines, including underground ore mining, transportation, crushing, grinding, roughing, and smelting, can release Pb, Cd, Cu, and Ni, as well as metalloid arsenic (As). These heavy metals are not readily degradable in the environment and easily accumulate in human bodies to a very high amount, and beyond a certain limit which can induce toxicity and adverse health effects [2].

Because of the potential risk to the occupational population working in lead-zinc mines, there are growing concerns about the adverse health effects associated with metal exposure encountered in the occupational environment [3, 4]. Several metals, such as Cd, Ni, and As, are known human carcinogens, and the carcinogenic potential of these heavy metals has been widely studied in humans and experimental animals [5, 6]. Lead poisoning can manifest in multiple systems in the body including in the central and peripheral nervous systems, erythropoietin system, cardiovascular system, musculoskeletal system, immune system, and/or reproductive system [7]. Rafal et al. found that workers with occupational exposure to Pb were far more likely to have cardiac injury than those without occupational exposure [8]. Cd was discovered in Germany in 1817 [9]. On the basis of a large number of epidemiological studies, Cd and its compounds were classified as class I carcinogens by the International Agency for Research on Cancer (IARC) in 1993 [10]. Cd causes several chronic toxicity symptoms in human body systems, including in the bones and kidney, as well as the cardiovascular system, respiratory system, reproductive system, and immune system [11]. Moreover, many studies have shown that lung, prostate, kidney, and liver cancers are associated with Cd. Occupational populations with excessive amounts of Cd in their working environment, as compared with individuals with no exposure, present with higher Cd concentrations in the body.

Non-ferrous metal smelting is often accompanied by As release [12]. As is a known human carcinogen that is widely distributed in the environment [13]. Exposure to As can lead not only to pigmentation and hyperkeratosis but also to damage in the liver, kidney, cardiovascular system, and central nervous system. Inorganic As (iAs) undergoes methylation metabolism after absorption into the human body, which is then distributed in various tissues and organs [14]. The metabolites of iAs include monomethylarsonic acid (MMA) and dimethylarsinic acid (DMA); the inorganic species and organic methylated metabolites of MMA and DMA are then excreted in the urine. The toxicities of these arsenicles are different and iAs and MMA are more toxic than DMA [13, 15]. Indeed, current evidence indicates that the unifying factor determining the toxicity and carcinogenicity of heavy metals is the generation of free radicals and oxidative stress, which in turn causes oxidative damage to DNA and enhance lipid peroxidation [16, 17]. Urinary 8-hydroxydeoxyguanosine (8-OHdG) is a widely accepted sensitive biomarker of oxidative DNA damage, whose concentration reflects general oxidative stress levels in the human body [18].

 Cu, Ni, and As. In addition, we chose cystatin C as an early renal damage biomarker to compare the degree of kidney damage in both groups. Finally, we also evaluated oxidative damage (urinary 8-OHdG) in the study population.

Materials and Methods

Subjects

This study was conducted in two foundries with a total of 307 workers: 145 exposed workers from a lead-zinc mine foundry and 162 workers from a steel smelting plant, both located in the northeastern part of China. All subjects provided informed consent to participate in this study, and the China Medical University Ethics Committee approved the research. A questionnaire was given to all subjects to obtain information about their age, weight, height, occupational history, disease history, dietary status (including consumption of seafood, meat, vegetables, and fruits in the 3 days preceding the start of the study), alcohol consumption, smoking habits, and other lifestyle parameters. Blood pressure was measured after the questionnaire was completed. All subjects with urinary creatinine beyond the range of 0.3-3.0 g/l were excluded according to the World Health Organization criteria for the acceptability of urine samples for biological monitoring [19]. Information on the study subjects is listed in Table 1. The lead-zinc mine foundry workers (average age 46.3 years) were older than the steel smelting plant workers (average age 40.1 years). The smoking rates in the two groups of foundry workers were high, and more than half of the workers were smokers. The mean employment years of the lead-zinc mine foundry workers (16.28 years) were higher than that of the smelting plant workers (7.04 years). There was no difference in BMI between the two groups of foundry workers. The amount of seafood consumption in the previous 3 days was lower in the lead-zinc mine foundry workers (30.69 g) than the smelting plant workers (131.08 g). Moreover, lead-zinc mine foundry workers consumed more alcohol than did smelting plant workers. Spot urine samples from all subjects were collected in PVC bottles and kept on ice; urine samples were stored at -20 °C, then transported to the China Medical University Laboratory, and stored at - 80 °C before being analyzed.

Determination of Urinary Heavy Metals

Determination of urinary Pb, Cd, Cu, and Ni was performed with a graphite furnace atomic absorption spectrophotometer (Z-2000, Hitachi, Japan). The urine samples and nitric acid were mixed at a ratio of 100:1 (ν/ν) and then digested. Standard solutions of Pb, Cd, Cu, and Ni (Meilun, Dalian, China) were diluted with normal human urine and used to prepare standard curves. The standard reference materials
 Table 1 Demographics of the study subjects

Characteristics	Lead-zinc mine foundry workers	Steel smelting plant workers
Number of subjects	145	162
Age (mean \pm SD)	$46.34 \pm 8.28*$	40.14 ± 10.19
Years of employment (mean \pm SD)	$16.28 \pm 12.07*$	7.04 ± 8.87
Smoking habits $(n, \%)$	107 (73.80%)*	108 (66.67%)
BMI (kg/m ²) (mean \pm SD)	24.79 ± 4.20	24.44 ± 4.13
Numbers of worker eating seafood in recent 3 days $(n, \%)$	36 (24.83%)*	102 (62.96%)
Amount of seafood consumption in recent 3 days (g) (mean \pm SD)	$30.69 \pm 84.41*$	131.08 ± 176.04
Numbers of worker drinking alcohol in recent 3 days $(n, \%)$	53 (36.55%)*	42 (25.93%)
Amount of alcohol consumption in recent 3 days (g) (mean ± SD)	$97.24 \pm 159.29*$	58.95 ± 134.60

*P < 0.05 compared with the steel smelting plant workers

purchased from Dalian Meilun Biotechnology Corporation (China) was used as quality control. We estimated the accuracy by comparing the difference between the measured values and available certified values with their uncertainty according to the calculation method reported by Linsinger [20]. For measurement of urinary metals, batches of 20 urine samples were analyzed and the standard curves were reanalyzed after each batch to ensure normal performance of the instrument. If the detected values of standard curves were not in agreement with available certified values, the instrument was recalibrated, and the 20 urine samples were re-tested again. The limits of detection (LOD) for these four metals were 0.050 μ g/L for Pb, 0.001 μ g/L for Cd, 0.009 μ g/L for Cu, and 0.01 μ g/L for Ni. Samples with concentrations below LOD were set to LOD/2 for each metal.

Determination of Arsenic Species

iAs, MMA, DMA, and trimethylated As (TMA) in urine were analyzed with an atomic absorption spectrophotometer (AA-6800, Shimadzu, Kyoto, Japan) with an As speciation pretreatment system (ASA-2sp, Shimadzu, Kyoto, Japan). Concentrations of As species were measured as described previously [21]. Quality control for arsenic determinations included the analysis of standard reference material of freezedried urine (SRM 2670a). The certified concentration values for arsenic were $480 \pm 100 \ \mu g/L$, and the values measured in this study were $474 \pm 20 \ \mu g/L$. The reliability of arsenic species separation was evaluated by analytical recoveries of added arsenic species. Spiking urine samples with 10 µg/L of iAs, MMA, and DMA resulted in recoveries of 81-92, 88-98, and 89-103% for iAs, MMA, and DMA, respectively. Because TMA was not a urinary iAs metabolite, we calculated total As (TAs) concentrations by summing the concentrations of iAs, MMA, and DMA.

Determination of Urinary Creatinine

The concentrations of urinary creatinine (Cr) were measured through the Jaffe reaction with a method reported previously to remove the influence of the effect of urine dilution on exposure biomarkers measured in spot samples [22]. The concentrations of urinary heavy metals (including urinary Pb, Cd, Cu, Ni, and As) were calibrated on the basis of urinary Cr and were reported as micrograms per gram Cr.

Determination of Urinary 8-OHdG

To determine the concentrations of 8-OHdG, the urine samples were centrifuged at 2000 rpm for 10 min and the supernatants were used for measurement. Urinary concentrations of 8-OHdG were determined using an Abcam ELISA kit (ab201734) according to the manufacturer's instructions. The concentrations of urinary 8-OHdG were calibrated on the basis of urinary Cr and reported as micrograms per gram Cr.

Determination of Urinary Cystatin C

The urine samples were centrifuged at 2000 rpm for 10 min before tested, and the supernatants were used for measurement with a cystatin C kit (Mlbio, Shanghai, China) according to the manufacturer's instructions, the levels of urinary cystatin C were further corrected by the concentrations of individual urinary Cr levels and reported as micrograms per gram Cr.

Statistical Analysis

Statistical analysis was performed in the SPSS statistical package (SPSS version 19.0). The distributions of the data were examined with one-sample Kolmogorov-Smirnov normality tests. Nonparametric tests were used if the data were not normally distributed or after transformation. The differences in frequencies were evaluated with chi-square tests. We compared the levels of urinary heavy metals and arsenicals between the two groups of factory workers. The dose-response effects of urinary heavy metals on the oxidative stress biomarker 8-OHdG and kidney damage biomarker cystatin C were separately estimated using bivariate correlation analysis. A binary logistic regression model was used to screen the factors influencing the levels of urinary heavy metals, including age, employment years, factories, BMI, smoking habits, alcohol consumption, and seafood intake in the previous 3 days. In addition, we analyzed factors influencing the levels of urinary 8-OHdG and cystatin C using a binary logistic regression model. Risk factors with significant differences were selected for binary logistic regression. The statistical significance threshold was defined as P < 0.05.

Results

Urinary Heavy Metal Levels

Urinary heavy metal levels of the workers in the two factories are shown in Table 2. Urinary Pb, Cu, and Ni levels of lead-zinc mine foundry workers were significantly higher than those of steel smelting plant workers (P < 0.05). Although there was no significant difference between the levels of urinary Cd of workers in the two factories (P > 0.05), Cd levels of the workers in the lead-zinc mine were higher than those of the steel smelting plant. The lead-zinc mine foundry workers had significantly higher concentrations of iAs and MMA than did the steel smelting plant workers (P < 0.05), and the concentrations of DMA and TAs were significantly lower than those of the steel smelting plant workers (P < 0.05). The steel smelting plant workers also had higher concentrations of DMA and TAs. Furthermore, we found that the steel smelting plant workers had higher amounts of seafood consumption (131.08 g) in the 3 days preceding urine sample collection than did the lead-zinc

mine foundry workers (30.69 g). Thus, seafood consumption was probably a major factor that increased DMA and TAs levels in the steel smelting plant workers, similar to many previous reports [23, 24].

Factors Influencing Urinary Heavy Metal Levels

Beyond the analysis comparing foundries, we also analyzed the influences of other potential factors, such as employment years, smoking habits, and seafood consumption in the previous 3 days, on urinary heavy metals. We stratified the employment years into three groups: 10 years or fewer, 10 to 20 years, and 20 years or more. We found that the lead-zinc mine foundry workers with 10 to 20 employment years had the highest urinary Pb. For smelting plant workers, there were no differences in urinary heavy metals across the three employment year groups. In this study, the daily smoking habits of workers in the two factories were also investigated. We divided the subjects into four groups: no smoking, <10, 10-20, and > 20 branches per day. Urinary Cd levels increased with increasing daily smoking levels. In lead-zinc mine foundry workers, those who had consumed seafood in the previous 3 days had lower levels of urinary Pb than did those who did not consume seafood. In steel smelting plant workers, levels of urinary Ni were higher in those who had consumed seafood than in those who had not consumed seafood (Table 3).

To screen the risk factors affecting the concentrations of urinary heavy metals, we performed binary logistic regression. Age, employment years, factories, BMI, smoking habits, alcohol consumption, and seafood consumption were included. The logistic regression analysis showed that working in leadzinc mine foundry was a risk factor for urinary levels of Pb and Ni of the workers compared to working in steel smelting plant. In addition, alcohol consumption was another risk factor for urinary levels of Ni. Age and smoking habits were the risk factors for urinary levels of Cd. BMI was a protective factor regarding Cd levels in the urine (Table 4).

	Lead-zinc mine foundry workers	Steel smelting plant workers
Number of subjects	145	162
Pb (µg/g Cr)	23.91 (4.29~58.06)*	0.025 (0.025~1.84)
Cd (µg/g Cr)	1.41 (0.28~3.24)	0.93 (0.17~3.15)
Cu (µg/g Cr)	10.33 (4.48~24.69)*	6.44 (0.0045~17.80)
Ni (µg/g Cr)	23.13 (0.005~128.40)*	1.18 (0.005~4.40)
iAs (µg/g Cr)	3.33 (2.19~4.65)*	2.49 (1.44~3.95)
MMA (µg/g Cr)	2.54 (1.76~4.02)*	2.02 (1.26~3.01)
DMA (µg/g Cr)	2.54 (1.76~4.02)*	2.02 (1.26~3.01)
TAs (µg/g Cr)	25.19 (18.67~37.17)*	33.77 (26.94~46.27)

*P < 0.05 compared with the steel smelting plant workers

Table 2 Urinary heavy metalconcentrations of the two factoryworkers. Median (range)

Table 3 Urinary	/ heavy 1	metal and arsenide concentr	ations in the different ch	aracteristics of the two factor	ry workers [median (range)]		
	Ν	Pb (µg/g Cr)	Cd (µg/g Cr)	Cu (µg/g Cr)	Ni (µg/g Cr)	iAs (µg/g Cr)	MMA (µg/g Cr)	DMA (µg/g Cr)
Lead-zinc mine fo	w drbd w	vorkers						
Employment year	s							
< 10	67	38.83 (7.58~63.96) ^a	1.36 (0.29~3.19)	8.88 (2.90~25.30)	20.28 (0.005~108.72)	2.85 (1.71~4.76)	2.54 (1.34~4.06)	19.00 (12.96~31.26)
10 - 20	20	27.0 (20.36~65.95) ^a	1.01 (0.0005~3.62)	9.68 (1.83~16.84)	6.10 (0.005~163.78)	3.51 (2.31~4.74)	2.46 (1.95~4.51)	17.49 (13.64~27.05)
> 20	58	12.88 (0.25~35.23)	1.50 (0.38~3.61)	14.29 (6.48~30.92)	27.29 (0.005~137.76)	3.49 (2.29~4.62)	2.53 (2.05~4.01)	18.77 (13.76~25.45)
Daily smoking an	10 nount (pe	er day)						
No smoking	38	16.73 (0.025~45.39)	0.75 (0.0005~1.78)	11.17 (3.89~19.59)	12.26 (0.005~130.65)	2.64 (1.65~4.00)	2.24 (1.36~3.02)	17.26 (13.01~24.89)
> 10	6	23.39 (2.06~83.82)	0.82 (0.17~3.31)	12.02 (5.90~25.68)	51.03 (0.005~134.67)	2.27 (1.26~4.27)	2.18 (1.82~3.73)	26.96 (8.54~38.67)
10-20	84	30.15 (8.96~61.28)	1.59 (0.53~3.62) ^b	9.55 (4.01~25.68)	20.00 (0.005~120.14)	3.47 (2.33~5.45)	2.62 (1.95~4.43)	19.17 (13.87~30.01)
> 20	14	27.58 (9.64~45.50)	2.55 (1.26~4.28) ^b	15.14 (6.40~34.63)	78.52 (6.10~154.44)	4.38 (2.37~4.66)	2.70 (1.87~4.43)	17.96 (16.37~24.05)
Seafood consump	tion in r	ecent 3 days						
Yes	36	11.16 (0.025~28.76)	1.66 (0.13~4.13)	10.04 (2.69~23.40)	23.34 (0.005~142.78)	3.52 (1.97~4.74)	2.40 (1.16~3.23)	20.61 (15.99~30.45)
No	109	30.86 (7.15~68.05) ^c	1.36 (0.32~3.21)	10.33 (4.84~25.56)	21.12 (0.005~121.25)	3.31 (2.19~4.57)	2.61 (1.92~4.29)	18.31 (12.74~26.88)
Steel smelting pla	nt work(STS						
Employment year	s							
< 10	129	0.025 (0.025~2.11)	0.81 (0.19~3.05)	7.16 (0.0045~17.93)	1.07 (0.005~4.61)	2.39 (1.42~3.88)	2.08 (1.28~3.12)	27.85 (21.57~38.65)
10 - 20	11	0.025 (0.025~1.94)	0.35 (0.0005~1.68)	$0.0045 (0.0045 \sim 11.01)$	0.005 (0.005~3.70)	3.32 (2.79~6.03)	2.28 (1.38~3.07)	27.36 (23.03~40.84)
> 20	22	0.025 (0.025~1.11)	1.91 (0.38~3.87)	5.79 (0.0045~21.34)	2.07 (0.005~5.28)	2.32 (1.27~4.07)	1.80 (0.77~2.79)	32.16 (25.50~50.08)
Daily smoking an	ount (pe	er day)						
No smoking	55	0.025 (0.025~1.52)	0.27 (0.0005~1.11)	3.28 (0.0045~21.98)	0.46 (0.005~3.62)	2.31 (1.41~4.31)	2.24 (1.39~3.32)	29.67 (23.24~38.10)
< 10	38	0.025 (0.025~1.77)	0.59 (0.0005~2.32)	$10.04\ (0.0045{\sim}18.48)$	1.98 (0.005~5.14)	2.24 (1.37~3.49)	1.72 (1.17~2.74)	26.04 (21.07~41.96)
10 - 20	56	0.025 (0.025~2.60)	1.60 (0.59~3.57) ^d	6.83 (0.0045~15.42)	0.99 (0.005~4.35)	2.60 (1.49~4.58)	2.16 (1.33~3.05)	30.35 (21.25~39.88)
> 20	13	$0.025\ (0.025{\sim}10.52)$	1.46 (0.85~4.40) ^d	0.0045 (0.0045~12.52)	0.06 (0.005~4.61)	2.53 (1.80~3.88)	1.56 (0.22~2.26)	32.23 (22.49~48.38)
Seafood consump	tion in r	ecent 3 days						
Yes	102	0.025 (0.025~1.75)	0.89 (0.14~2.89)	8.96 (0.0045~18.60)	1.73 (0.005~5.02) ^e	2.60 (1.72~3.86)	2.15 (1.20~2.99)	31.11 (22.52~42.51)
No	60	0.025 (0.025~3.17)	0.80 (0.17~2.68)	4.20 (0.0045~13.59)	0.005 (0.005~3.41)	2.46 (1.27~4.45)	1.94 (1.28~3.25)	25.97 (21.31~34.37)

 $^{a}p < 0.05$ compared with > 20 employment years group in lead-zinc mine foundry workers

p<0.05 compared with no smoking group in lead-zinc mine foundry workers

 $^{c}p < 0.05$ compared with seafood consumption in three recent days group in lead-zinc mine foundry workers

p < 0.05 compared with no smoking group in steel smelting plant workers

 $^{\circ}p < 0.05$ compared with no seafood consumption in three recent days group in steel smelting plant workers

Independent variable	Pb levels (µg/g Cr)		Cd levels (µg/g Cr)		Ni levels (µg/g Cr)	
	OR(95% CI)	р	OR(95% CI)	р	OR(95% CI)	р
Age	1.190 (0.867, 1.634)	0.282	1.563 (1.168, 2.091)	0.003	0.902 (0.682, 1.192)	0.468
Employment years	0.754 (0.516, 1.101)	0.144	1.112 (0.800, 1.545)	0.527	1.184 (0.861, 1.626)	0.298
Factories	11.927 (6.273, 22.676)	0.000	1.248 (0.708, 2.198)	0.444	2.918 (1.667, 5.108)	0.000
BMI	0.879 (0.641, 1.204)	0.422	0.608 (0.455, 0.811)	0.001	0.895 (0.680, 1.180)	0.432
Smoking habits	0.625 (0.343, 1.139)	0.125	3.404 (1.955, 5.925)	0.000	0.873 (0.515, 1.481)	0.615
Alcohol consumption	1.006 (0.546, 1.853)	0.985	1.258 (0.725, 2.183)	0.414	2.086 (1.220, 3.567)	0.007
Seafood consumption	0.811 (0.451, 1.461)	0.486	1.214 (0.701, 2.102)	0.490	1.443 (0.838, 2.484)	0.186

Table 4 The factors of influencing the level of urinary Pb, Cd, and Ni

The medians for the urinary Pb (1.94 μ g/g Cr), Cd (1.11 μ g/g Cr), and Ni (2.36 μ g/g Cr) were used in these models. The binary logistic regression models, with adjusted for age (20–30, 31–40, 41–50, and > 50 years old groups), employment years (< 10, 10–20, > 20 years groups), factories (lead-zinc mine foundry and steel smelting plant), BMI (< 23, 23–25, > 25 kg/m² groups), smoking habits (no smoking and smoking groups), alcohol consumption (no alcohol consumption and alcohol consumption groups), seafood consumption (no seafood consumption and seafood consumption groups) *OR* odd ratio

Concentrations of 8-OHdG and Cystatin C Biomarkers in the Urine

Present evidence indicates that 8-OHdG in urine is a widely accepted sensitive biomarker of oxidative DNA damage. Occupational exposure to heavy metals can trigger oxidative stress responses in the body, which can increase 8-OHdG levels. In this study, there was no significant difference in the concentrations of 8-OHdG between the two groups of factory workers (Table 5). Bivariate correlation analysis showed that 8-OHdG levels of workers were positively correlated with urinary Pb, Cd, iAs, MMA, and DMA concentrations (Spearman correlation coefficients were 0.223, 0.225, 0.267, 0.215, and 0.195, respectively) (P < 0.05) (Table 6). Binary logistic regression was used to screen the risk factors affecting the level of urinary 8-OHdG, and age, employment years, factories, BMI, smoking habits, alcohol consumption, and urinary heavy metals were included in the models. The factors that affected urinary 8-OHdG levels were urinary Pb (OR = 2.502, 95% CI 1.058–5.913) and Cd (OR = 2.473, 95% CI 1.125–5.437) (Table 7).

Cystatin C, a secretory low-molecular-weight protein (13 kDa), is abundant in body fluids and various tissues [25]. The kidney is an important target organ of Cd and Ni. Concentrations of cystatin C increase in urine in cases of renal

tubule dysfunction. The preliminary results of this study showed that lead-zinc mine foundry workers might be exposed to Cd and Ni; hence, we selected cystatin C as an early renal damage biomarker, which we used to compare the degree of kidney damage among the two groups of factory workers. The results showed that urinary cystatin C levels in the lead-zinc mine foundry workers were higher than those in the steel smelting plant workers (P < 0.05) (Table 5). Furthermore, to determine the effects of metals on cystatin C, bivariate correlation was performed. Table 6 shows that cystatin C levels in workers was positively correlated with urinary Pb, Cd, Ni, iAs, and MMA (Spearman correlation coefficients were 0.294, 0.176, 0.195, 0.268, and 0.137, respectively) (P < 0.05). Binary logistic regression showed that the risk factors affecting the concentration of urinary cystatin C were urinary Cd (OR = 1.992, 95% CI 1.031-3.849). In addition, urinary cystatin C was associated with age (Table 7).

Discussion

Heavy metal pollution health accidents occur frequently worldwide. Heavy metal pollution has become a major social problem that has received substantial research attention both at home and abroad. Occupational populations contact the

Table 5 Levels of urinary 8-OHdG and cystatin C of the twofactory workers

	Lead-zinc mine foundry workers	Steel smelting plant workers
Number of subjects	97	76
Urinary 8-OHdG (µg/g Cr) (median (range))	103.40 (70.03~157.17)	92.70 (66.27~148.10)
Number of subjects	116	97
Urinary cystatin C (μ g/g Cr) (mean \pm SD)	$1704.24 \pm 900.93 *$	1327.35 ± 878.45

*P < 0.05 compared with the steel smelting plant workers

Table 6Bivariate correlationbetween the biomarkers and theheavy metals of the two factoryworkers

	8-OHdG levels ($\mu g/g \ Cr$)		Cystatin C levels (µg/g Cr)		
	r	Р	r	Р	
iAs (µg/g Cr)	0.267	0.000	0.268	0.000	
MMA (µg/g Cr)	0.215	0.004	0.137	0.046	
DMA (µg/g Cr)	0.195	0.010	0.070	0.308	
Pb (µg/g Cr)	0.223	0.003	0.294	0.000	
Cd (µg/g Cr)	0.225	0.003	0.176	0.010	
Cu (µg/g Cr)	-0.280	0.715	-0.470	0.496	
Ni (µg/g Cr)	0.058	0.452	0.195	0.004	

The correlation coefficients of the biomarkers and heavy metals were spearman correlation coefficients

heavy metal elements directly, owing to a lack of strict mining production supervision and a lack of basic occupational protection knowledge among most mine workers. Occupational populations are thus likely to encounter occupational disease hazards.

This study included a total of 307 workers from two foundries in Liaoning Province. The five heavy metals analyzed in this study are the most commonly encountered toxic metals detected after occupational and dietary exposure. Urinary concentrations of Pb, Cu, Ni, and As are recognized as short-term indicators of recent exposure, while urinary excretion of Cd is proportional to the body burden and therefore represents an indicator of a lifetime exposure. Our results showed that the lead-zinc mine foundry workers' urinary Pb, Cd, Cu, and Ni were higher than those of the steel smelting plant workers, indicating that lead-zinc mine foundry workers may have a higher exposure risk of heavy metals. Usually, non-ferrous smelters exposed to inorganic arsenic in the workplace, inorganic As is methylated into MMA and DMA in the liver [12, 26]. The concentrations of urinary iAs and MMA were higher in the lead-zinc mine foundry workers, thus indicating that those workers were probably exposed to As. Many studies have found that seafood products contain a large number of As glucose and As lipids that can be metabolized as DMA [27, 28]. DMA levels in the body significantly increased after seafood consumption [29, 30]. Our study also

Independent variable	8-OHdG levels (µg/g Cr)		Cystatin C levels (µg/g	Cystatin C levels (µg/g Cr)	
	OR (95% CI)	р	OR (95% CI)	р	
Age	0.738 (0.487, 1.117)	0.151	1.755 (1.207, 2.553)	0.003	
Employment years	0.823 (0.513, 1.322)	0.421	0.701 (0.462, 1.064)	0.095	
Factories	0.948 (0.341, 2.633)	0.919	1.520 (0.672, 3.439)	0.314	
BMI	1.097 (0.742, 1.622)	0.642	1.211 (0.853, 1.719)	0.285	
Smoking habits	0.702 (0.317, 1.556)	0.384	0.971 (0.493, 1.913)	0.932	
Alcohol consumption	0.718 (0.316, 1.633)	0.430	1.208 (0.603, 2.422)	0.594	
iAs	2.091 (0.994, 4.402)	0.052	1.639 (0.841, 3.193)	0.146	
MMA	1.961 (0.954, 4.030)	0.067	1.361 (0.714, 2.592)	0.349	
DMA	1.549 (0.729, 3.288)	0.255	1.526 (0.796, 2.928)	0.203	
Pb	2.502 (1.058, 5.913)	0.037	1.667 (0.801, 3.472)	0.172	
Cd	2.473 (1.125, 5.437)	0.024	1.992 (1.031, 3.849)	0.040	
Cu	0.616 (0.305, 1.242)	0.176	0.964 (0.525, 1.772)	0.907	
Ni	1.306 (0.608, 2.808)	0.494	1.238 (0.640, 2.394)	0.525	

The medians for the urinary 8-OHdG (98.87 μ g/g Cr) and cystatin C (1305.29 μ g/g Cr) were used in these models. The binary logistic regression models for 8-OHdG, with adjusted for age (20–30, 31–40, 41–50, and > 50 years old groups), employment years (<10, 10–20, > 20 years groups), factories (lead-zinc mine foundry and steel smelting plant), BMI (<23, 23–25, > 25 kg/m² groups), smoking habits (no smoking and smoking groups), alcohol consumption (no alcohol consumption and alcohol consumption groups), seafood consumption (no seafood consumption groups), medians of iAs, MMA, DMA, Pb, Cd, Cu, and Ni. The binary logistic regression models for cystatin C, with adjusted for age, employment years, factories, BMI, smoking habits, alcohol consumption and seafood consumption, and medians of iAs, MMA, DMA, Pb, Cd, Cu, and Ni

Table 7 The factors ofinfluencing the level of urinary 8-OHdG and cystatin C

found that the steel smelting plant workers had higher urinary DMA than those in the lead-zinc mine foundry workers, possibly because the steel smelting plant workers had a higher amount of seafood consumption (131.08 g) in the 3 days preceding urine samples collection than did the lead-zinc mine foundry workers (30.69 g). Furthermore, among the steel smelting plant workers, those who ate seafood exhibited significantly higher urinary Ni levels than those that did not, thus suggesting that seafood products may be enriched in Ni. Usually, seafood is rich in protein, calcium, and zinc which can be replaced and antagonized with Pb after absorption in intestine and excreted with feces. In lead-zinc mine foundry, we observed that workers with no seafood intake have higher levels of urinary Pb than those with seafood intake which indicated that seafood consumption may be an advantageous factor for promoting Pb excretion. However, further investigations are needed to verify these results.

In order to screen for the risk factors affecting the concentrations of urinary heavy metals and arsenic, binary logistic regression analysis of Pb, Cd, Cu, and Ni was carried out. We chose age, employment years, factories, BMI, smoking habits, alcohol consumption, and seafood consumption as influencing factors. In this study, we have found that working in lead-zinc mine foundry was a risk factor for elevating urinary Pb and Ni. In addition, alcohol consumption was another risk factor for elevating urinary Ni. Alcohol consumption can lead to the accumulation of Pb and Cd in the body, but whether it can induce Ni accumulation has not been reported yet. The mechanism of the effect of alcohol consumption on the increase of urinary Ni level needs further confirmation. Age and smoking habits were the risk factors for urinary levels of Cd. BMI was a protective factor regarding Cd levels in the urine.

Smoking is another route of Cd exposure, each cigarette contains approximately 1-2 µg of Cd, and blood cadmium levels in smokers in Europe are three to four times higher than those in non-smokers [31]. In this study, urinary Cd level was elevated with higher daily smoking amounts in the two groups of factory workers, thus indicating that smoking may affect Cd levels in the human body, in agreement with previous reports. Cd can accumulate and remain in the human body for a long time, and can cause serious harm to various organs [32]. Many studies have reported that Cd induces damage to the kidneys, bones, respiratory system, cardiovascular system, reproductive system, and immune system [33–35]. The kidney not only is a site of Cd accumulation but also one of the target organs for Cd damage. Cystatin C has recently been used as a sensitive biomarker for early injury of renal function [36]. This study determined the levels of cystatin C in urine to compare the degree of early renal injury among the workers. There were higher levels of cystatin C in the workers in the leadzinc mine foundry, thus suggesting that lead-zinc mine foundry workers are more likely to have higher renal injury than steel smelting plant workers. Furthermore, we observed a positive linear dose-dependent effect on the urinary levels of cystatin C and urinary Pb, Cd, Ni, iAs, and MMA in the two groups of factory workers. Binary logistic regression analysis of cystatin C also supported the positive association urinary levels of Cd with urinary cystatin C.

Oxidative DNA damage is a widely accepted mechanism of heavy metal-induced carcinogenesis. Although the lead-zinc mine foundry workers had a higher degree of heavy metal exposure than steel smelting plant workers in this study, there was no significant difference in urinary 8-OHdG, thus suggesting that higher doses of heavy metals are required to produce oxidative damage. Oxidative DNA damage is an established mechanism for As-induced carcinogenesis [37]. Many reports carried out in various populations, such as glass production workers, semiconductors, healthy adults, as well as in children, consistently showed that subjects with highly As exposure had elevated the levels of 8-OHdG than those with low As exposure [38–41]. However, there were no significant associations that have been found between urinary Pb and Cd and 8-OHdG in the present population [42]. In this study, we observed that urinary Pb, Cd, and As had a positive linear dose-dependent effect on oxidative damage to DNA. These results suggested that the occupational exposure to heavy metals may significantly enhanced oxidative damage to DNA.

To our knowledge, there are few studies on the exposure of heavy metals in lead-zinc mine foundry workers. Therefore, this study investigated the levels of heavy metals and arsenide in the urine of two groups of foundry workers. Our results showed that lead-zinc mine foundry workers may have higher exposure to heavy metals, and these results also provide basic data for occupational protection. Further epidemiological studies are needed to develop functional environmental protection measures and to protect occupational populations' health.

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Compliance with Ethical Standards

Conflict of Interest The authors declare that they have no conflict of interest.

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