

# Characterization of Copper-Resistant Rhizosphere Bacteria from *Avena sativa* and *Plantago lanceolata* for Copper Bioreduction and Biosorption

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**Abstract** Copper is a toxic heavy metal widely used to microbial control especially in agriculture. Consequently, high concentrations of copper residues remain in soils selecting copper-resistant organisms. In vineyards, copper is routinely used for fungi control. This work was undertaken to study copper resistance by rhizosphere microorganisms from two plants (*Avena sativa* L. and *Plantago lanceolata* L.) common in vineyard soils. Eleven rhizosphere microorganisms were isolated, and four displayed high resistance to copper. The isolates were identified by 16S rRNA gene sequence analysis as *Pseudomonas putida* (A1), *Stenotrophomonas maltophilia* (A2) and *Acinetobacter* sp. (A6), isolated from *Avena sativa* rhizosphere, and *Acinetobacter* sp. (T5), isolated from *Plantago lanceolata* rhizosphere. The isolates displayed high copper resistance in the temperature range from 25°C to 35°C and pH in the range from 5.0 to 9.0. *Pseudomonas putida* A1 resisted as much as 1,000 mg L<sup>-1</sup> of copper. The isolates showed similar behavior on copper removal from liquid medium, with a bioremoval rate of 30% at 500 mg L<sup>-1</sup> after

24 h of growth. Speciation of copper revealed high copper biotransformation, reducing Cu(II) to Cu(I), capacity. Results indicate that our isolates are potential agents for copper bioremoval and bacterial stimulation of copper biosorption by *Avena sativa* and *Plantago lanceolata*.

**Keywords** Copper contamination · Rhizosphere bacteria · *Avena sativa* · *Plantago lanceolata* · Vineyards · Copper resistance

## Introduction

Copper (Cu(II)) is an essential micronutrient to living organisms. However, at high concentrations in the environment, copper is hazardous to living systems [1]. Vineyard soils generally contain high levels of copper. This occurrence is due to frequent sprays used to control leaf diseases in horticulture [2]. Such contamination results in the selection of copper-resistant microorganisms and plants.

In vineyard areas of South of Brazil, oatmeal crops are widely used to control soil erosion and to improve physical and chemical soil properties. Oatmeal plants however display signs of copper toxicity [3]. Phytoremediation is one approach to bioremediation of metals. The efficiency of phytoremediation is however not very high [4]. Nonetheless, rhizosphere microorganisms from contaminated areas appear to be more efficient in plant growth-promoting under contaminated sites [5–7]. Variation of speciation of copper in the maize rhizosphere was reported by Tao et al. [8], and it was related to microbial activities in the rhizosphere zone.

Copper-resistant microorganisms can play a major role in copper removal from the environment [9]. Bioreduction of Cu(II) to Cu(I) increases its bioavailability to microbial cells [10, 11] as well as plants [9, 12]. Thus, copper-

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resistant microorganisms can be used in phytoremediation to promote phytoaccumulation of heavy metals. Moreover, copper-resistant microorganisms adapted in the rhizosphere environment can be better suited for bioaugmentation and can promote phytoremediation, such as the *Avena sativa* plants exhibited after inoculation of the rhizosphere bacteria isolated and characterized [3].

Selection of metal-resistant microorganisms with high bioremoval capacity is a major step in heavy metal bioremediation and phytoremediation. However, environmental parameters such as pH, temperature and contaminant concentration can interfere in the microbial activity [13]. In this study, it was isolated, and selected highly copper-resistant bacteria from *Avena sativa* and *Plantago lanceolata* rhizosphere soils are collected from vineyard areas; it was characterized as the isolates for copper bioreduction and biosorption. Selected isolates were then subjected to molecular characterization by 16S rRNA gene sequence analysis. The copper speciation by rhizosphere microorganisms was also examined.

## Materials and Methods

### Soil Samples

*Avena sativa* and *Plantago lanceolata* soil samples were collected from two copper-contaminated vineyard areas from Southern Brazil, in the EMBRAPA experimental station, Bento Gonçalves, RS, Brazil (29°09'53.92" S and 51°31'39.40" W). These soils were classified as Inceptisol and Mollisol. Table 1 presents the physical–chemical parameters: pH, cation exchange capacity, organic matter, clay, copper dissolved in acid (Cu) and copper soluble in water (Cu(H<sub>2</sub>O)). Copper soluble in water was extracted from 10 g of soil in 20 mL of deionized water for 60 min, with shaking. Copper was determined by atomic absorption spectrophotometer.

### Enrichment and Isolation of Cu(II)-Resistant Bacteria

Enrichment of Cu(II)-resistant bacteria was set up using 100 mL of nutrient broth in 250-mL Erlenmeyer flasks to which we added 500-mg mL<sup>-1</sup> Cu(II) using CuSO<sub>4</sub>·5H<sub>2</sub>O,

subsequently mentioned as Cu(II). The medium was adjusted to pH 7.0 using 1-M NaOH or 0.1-M HCl. The nutrient broth medium with copper was sterilized by autoclaving at 121°C for 20 min. The two different rhizosphere soils were independently used to inoculate (1% w/v) sterile nutrient broth amended with Cu(II) and incubated for 24 h, with orbital shaking (150 rpm) at 30°C. Subsequently, 1 mL of the enrichment culture was transferred to 99 mL of sterile nutrient broth amended with Cu(II) and incubated for 24 h, with shaking (150 rpm) at 30°C. The enrichment procedure was repeated for a third round. Cu(II)-resistant bacteria were thereafter purified by repeated streaking of the third-round cultures on nutrient agar plates containing Cu(II) (500 mg L<sup>-1</sup>). Cu(II)-resistant bacterial strains isolated from enrichment consortia were evaluated to Cu(II) resistance and biosorption.

### Analysis of Copper and Bacterial Mass

Total copper was analyzed using an atomic absorption spectrometer. Aliquots of culture supernatant (200-μL aliquots) were diluted 20 times and injected into the atomic absorption spectrometer. Copper remaining was calculated as the difference of the total copper added to the medium and the total copper removed from the medium after different microbial treatments. Copper reduction was quantified by measuring monovalent copper complex with 1-mM neocuproine hydrochloride [14]. Cultures were centrifuged (2,500 rpm, 10 min). For copper reduction, 1-mL aliquot of the cell-free culture supernatant was mixed with 2 mL of 1-mM neocuproine to complex Cu(I) and incubated for 30 min at 30°C in a 13 × 100 mm glass culture tubes. Thereafter, the absorbance was read using a spectrophotometer at λ<sub>454</sub> [14]. Copper remaining and copper reduction were calculated from copper (I or II) standard curves (25–500 mg L<sup>-1</sup> of copper) prepared in the nutrient broth medium. Copper resistance was determined by measuring bacterial biomass development using at λ<sub>600</sub>.

### Effect of Copper Concentration

The effects of different concentrations of Cu(II) were determined using nutrient broth amended with 250, 500,

**Table 1** Chemical–physical properties of rhizosphere soils from *Avena sativa* (Mollisol) and *Plantago lanceolata* (Inceptisol)

Soils	pH	CEC <sup>a</sup>	OM <sup>b</sup>	Clay	Cu	Cu (H <sub>2</sub> O)
	(1:1)	(cmol <sub>c</sub> kg <sup>-1</sup> )	(g kg <sup>-1</sup> )		(mg kg <sup>-1</sup> )	
Mollisol	6.9	16.5	64.0	22	238	10
Inceptisol	6.5	19.5	54.0	17	442	16

<sup>a</sup> CEC cation exchange capacity

<sup>b</sup> OM organic matter

750, 1,000 and 1,250 mg L<sup>-1</sup> of Cu(II). The inoculum was grown at 30°C for 24 h with shaking (150 rpm) and thereafter adjusted to 0.85 OD<sub>600</sub> units with sterile salt solution (0.85%). The sterile Cu(II) media were independently inoculated with 100-μL aliquot of each inoculum and incubated at 30°C for 24 h with shaking (150 rpm).

#### Effect of Temperature and pH

Temperature and pH effects on bacterial growth were examined using nutrient broth amended with Cu(II) (500 mg L<sup>-1</sup>). For the temperature effect, cultures were incubated at 25, 30, 35 and 40°C. In the pH experiment, sterilized nutrient broth amended with Cu(II) (500 mg L<sup>-1</sup>) was adjusted to pH 3.0, 5.0, 7.0, 9.0 and 11.0 by addition of predetermined amounts of sterilized 1-M NaOH or 0.1-M HCl. Cultures were inoculated with 100 μL of inoculum (OD<sub>600</sub>=0.85). Inoculum preparation and biomass determination were as described previously.

#### Time Course of Copper Biosorption and Reduction

The time course of divalent copper removal by biosorption and bioreduction was examined as follows. Nutrient broth amended with 500 mg L<sup>-1</sup> of Cu(II) was used. The medium was inoculated with 100 μL of each isolate (OD<sub>600</sub>=0.85). Cultures were incubated at 30°C with shaking (150 rpm) and analyzed at different time intervals (0, 6, 24, 30 and 48 h). Biomass, copper reduced and copper removed were then analyzed, and copper bioremoval was calculated as discussed previously.

#### DNA-Based Identification of Isolates

Isolates were identified by 16S ribosomal RNA gene sequence analysis as follows. The isolates were grown by streaking on nutrient agar with incubation at 30°C for 24 h. DNA of isolates was extracted from colony-forming units pooled from the nutrient agar plate using Promega Wizard Genomic DNA Purification Kit (Promega, Madison, WI) with slight modification. Briefly, cells were re-suspended in 300 μL of nucleic acid lyses solution, incubated at 80°C for 15 min and allowed to cool at room temperature. RNase solution (1.5 μL) was added and incubated at 37°C for 60 min. Protein precipitation solution (100 μL) was added and incubated on ice for 5 min. Following centrifugation, the supernatant was transferred to a tube, and cold ice 95% ethanol was added. The precipitate was recovered by centrifugation. The pellet was washed with 70% ethanol at ambient temperature and re-suspended in sterile nuclease-free distilled water. Two primers corresponding to *E. coli* positions 27 F (5'-AGATTTGATCMTGGCTCAG-3') and 1492R (5'-TACGGYTACCTTGTTACGAC TT-3') were

used for PCR amplification of the 16S ribosomal RNA [15]. The PCR reaction mixture consisted of 12.5 μL of PCR master mix (Promega, Madison, WI), genomic DNA template (0.5 μL), primer 27 F (2.5 μL=12.5 pmol) and primer 1492R (2.5 μL=12.5 pmol) and made up to 25-μL final volume with nuclease-free water. The 16S rRNA gene was amplified using a 35-cycle PCR (initial denaturation, 95°C for 5 min; subsequent denaturation, 95°C for 0.5 min; annealing temperature, 50°C for 1 min; extension temperature, 72°C for 1 min and final extension, 72°C for 5 min). The PCR amplification products were analyzed by electrophoresis on a 1% agarose gel. Millipore Montage PCR filter units (Millipore, Billerica, MA) were used to remove primers, salts and unincorporated dNTPs according to the manufacturer's instructions. DNA cycle sequencing was performed using BigDye terminator kit (Applied Biosystems, Foster City, CA) with sequencing primer 519r (5'-GWATTACCGCGGCKGCTG-3') in independent reactions at the Institute of Integrative Genome Biology (IIGB) of the University of California, Riverside, CA.

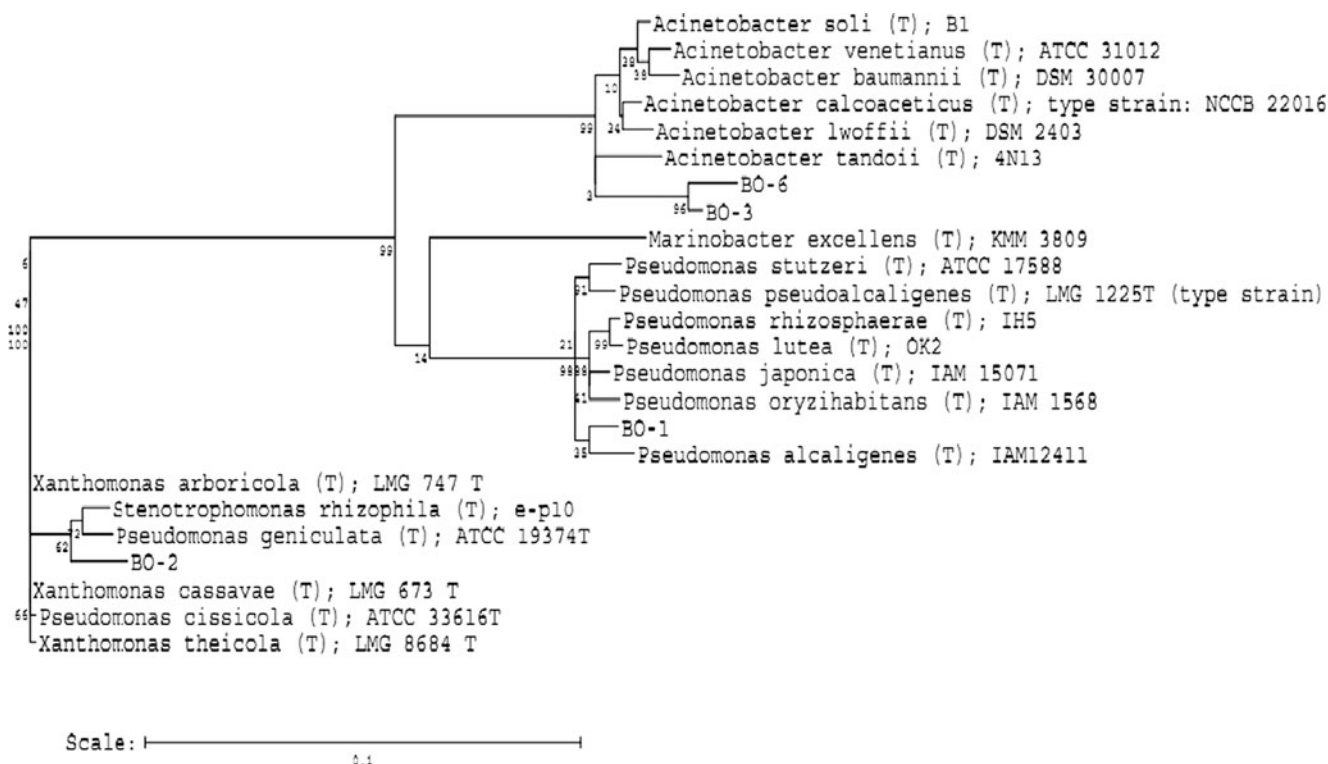
#### DNA Sequence Similarity and Phylogenetic Analysis

GenBank BLAST (N) was used for homology searches. The 16S rRNA gene sequence was deposited to the GenBank database. Phylogenetic analysis was conducted using RDP release 10 software [16].

## Results

#### Isolate Identification

A total of nine copper-resistant bacteria, six from *Avena sativa* (A) rhizosphere and three from *Plantago lanceolata* (T) rhizosphere, were isolated. These isolates resistant to high copper concentration (500 mg L<sup>-1</sup>) were further studied. The four selected isolates were identified by 16S rRNA, similarity searches were analyzed with my RDP software (Fig. 1), and the sources, the number of the 16S rDNA nucleotides, the GenBank Submission and the GenBank Match are presented in Table 2. The isolates were identified as *Pseudomonas putida* (A1) (BO-1 in the tree), *Stenotrophomonas maltophilia* (A2) (BO-2 in the tree) and *Acinetobacter* sp. (A6) (BO-3 in the tree) and were isolated from *Avena sativa* rhizosphere soil zone, whereas one isolate, *Acinetobacter* sp. (T5) (BO-6 in the tree), was isolated from *Plantago lanceolata* rhizosphere (Fig. 1). Isolates A1 and A2 were 98% similar to the GenBank BLAST analysis match, while isolates A6 and T5 were 96% and 94% similar to GenBank BLAST analysis match, respectively (Table 2).



**Fig. 1** Phylogenetic tree showing evolutionary distance of the selected isolates from the rhizosphere zone of two plants grown in the copper-contaminated areas based on 16S rRNA gene sequence. The identification of the isolates in the tree are isolate A1 (*Pseudomonas putida*) is

BO-1, isolate A2 (*Stenotrophomonas maltophilia*) is BO-2, isolate A6 (*Acinetobacter* sp.) is BO-3, and isolate T5 (*Acinetobacter* sp.) is BO-6. The numbers in the nodes are the bootstrap of 100 replicates. The scale is the evolutionary distance value

### Effect of Copper Concentration

The effect of copper concentration on the resistance profile of the four bacteria isolated from rhizosphere of *Avena sativa* and *Plantago lanceolata* plants was measured (Fig. 2). In general, all isolates showed growth decrease with increase in copper concentration. Isolates A1 and A6 showed similar resistance pattern at different copper concentrations and resisted up to 750 mg L<sup>-1</sup> in liquid medium. On the other hand, isolate A2 showed superior growth at all copper concentrations followed by T5, isolated from *Avena sativa* and *Plantago lanceolata* rhizosphere zone, respectively. Furthermore, A2 and T5 were the only isolates that tolerated 1,000 mg L<sup>-1</sup> of copper in liquid medium. However, none of the isolates survived at 1250 mg L<sup>-1</sup> of copper concentration.

### Effect of Temperature

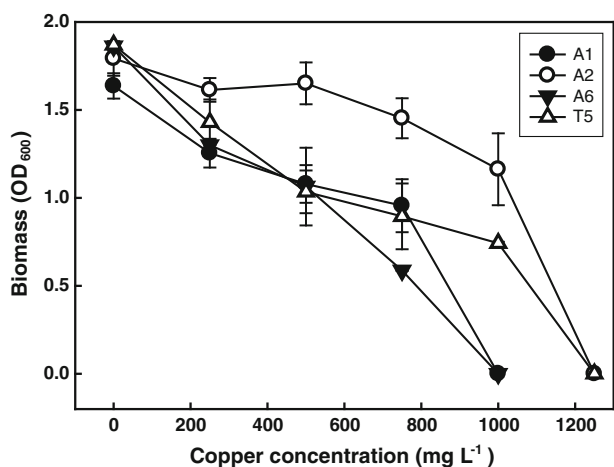
The effect of temperature on the copper resistance profile of bacteria isolated from rhizospheres of *Avena sativa* and *Plantago lanceolata* plants was studied (Fig. 3). Isolates A2, A6 and T5 were grown maximally between 25°C and 30°C in copper-amended medium. Contrarily, growth of isolate A1 in copper-amended medium was optimal at 35°C, with as high as 1.4 OD<sub>600</sub> units recorded. None of the isolates survived at 40°C in copper-amended medium.

### Effect of pH

The effects of pH on Cu(II)-resistant bacterial isolates are presented in Fig. 4. The optimum initial pH for biomass production by all isolates was at pH 7.0. There was no

**Table 2** DNA-based identification of four isolates from two rhizosphere plants grown in copper-contaminated vineyard soils

Isolates	Rhizosphere source	16S rDNA nucleotides	GenBank submission	GenBank match	Identity (%)
A1	<i>Avena sativa</i>	489	JN648685	AF509331.1	<i>Pseudomonas putida</i> (98)
A2	<i>Avena sativa</i>	479	JN648686	EF620472.1	<i>Stenotrophomonas maltophilia</i> (98)
A6	<i>Avena sativa</i>	506	JN648687	Z93454.1	<i>Acinetobacter</i> sp. (96)
T5	<i>Plantago lanceolata</i>	507	JN648688	FM213386.2	<i>Acinetobacter</i> sp. (94)

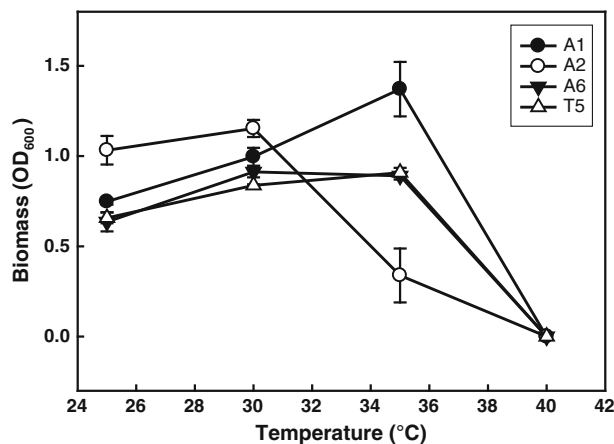


**Fig. 2** Copper resistance profile of bacteria isolated from rhizosphere of *Plantago lanceolata* (T5) and *Avena sativa* (A1, A2 and A6) at different concentrations of copper after 24 h with shaking (150 rpm) at 30°C. The error bars are standard errors of the mean

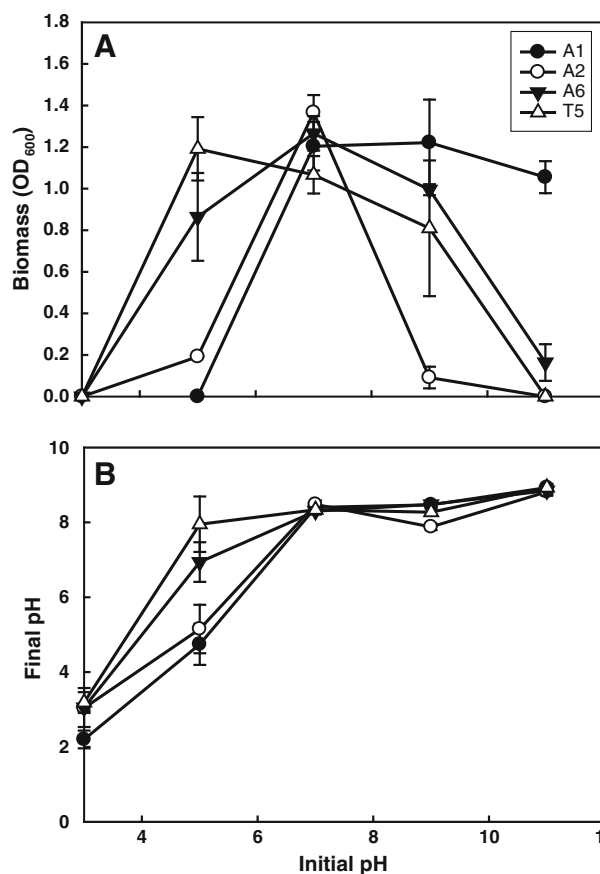
growth of the isolates at pH 3.0. Isolate A1 grew substantially at initial pH 11.0 (1.1 OD<sub>600</sub> units), but pH 11.0 was generally inhibitory to all the isolates. Slight growth (0.2 OD<sub>600</sub> units) of isolate A6 was observed at initial pH 11.0. Isolate A2 was sensitive to pH changes, with high biomass production only at pH 7.0 (1.4 OD<sub>600</sub> units). Isolates A6 and T5 were similar in their growth profiles at different pH (Fig. 4a). Growth of the most of the isolates induced significant changes in pH (Fig. 4b).

**Time Course of Copper Resistance and Copper Bioremoval**

Time course of Cu(II) resistance and Cu(II) bioremoval by selected rhizosphere bacterial isolates was evaluated (Fig. 5). Isolates A6 and T5 showed the highest biomass production after 9 h of incubation (1.1 and 1.2 OD<sub>600</sub> units,



**Fig. 3** Effect of temperature on growth of bacteria isolated from rhizosphere of *Plantago lanceolata* plants (T5) and *Avena sativa* (A1, A2 and A6) at high concentration of copper (500 mg L<sup>-1</sup>). The error bars are standard errors of the mean

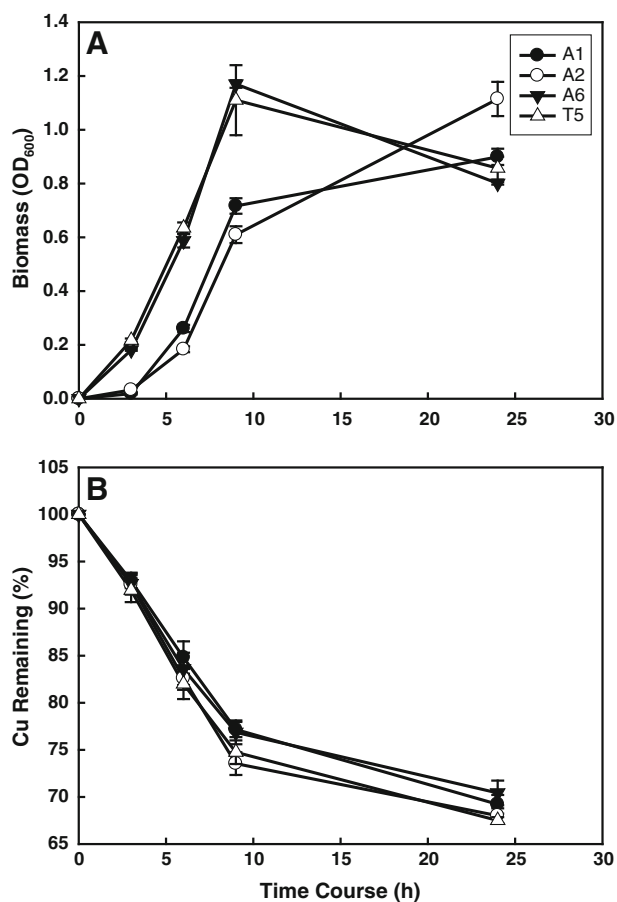


**Fig. 4** Effect of initial pH on growth (a) of bacterial isolates A1, A2, A6 and T5 at high copper concentration (500 mg L<sup>-1</sup>). Final pH values after 24 h of incubation at 30°C, with shaking (150 rpm) (b). The error bars are standard errors of the mean

respectively). Isolates A2 and A1 showed the highest biomass production after 24 h of incubation (1.2 and 0.9 OD<sub>600</sub> units, respectively). In both groups, the exponential phase of biomass production occurred after 9 h of incubation at 30°C with orbital shaking (150 rpm) (Fig. 5a). Copper bioremoval capacities of the four isolates were similar. The highest rate of Cu(II) bioremoval in liquid medium was in the first 9 h of incubation (25% removal, which is 125 mg L<sup>-1</sup> of copper removed). However, the isolates showed high Cu(II) bioremoval capacity (30% of the total amended copper or 150 mg L<sup>-1</sup> removed) from liquid medium with high copper concentration (500 mg L<sup>-1</sup>) in 24 h of incubation (Fig. 5b).

**Dynamics of Copper Speciation**

Figure 6a presents the dynamics of Cu(II) biosorption and bioreduction in medium amended with 500 mg L<sup>-1</sup> of copper. *Pseudomonas* species A1 displayed peak biomass production at 6 h of incubation (0.7 OD<sub>600</sub> units), and it was stable until 48 h of incubation. Cu(II) bioremoval by isolate A1 increased until 48 h of incubation, with more



**Fig. 5** Time course of biomass development (a) and copper bioremoval (b) in cultures of rhizosphere bacterial isolates A1, A2, A6 and T5 at high copper concentration ( $500 \text{ mg L}^{-1}$ ). The error bars are standard errors of the mean

than  $150 \text{ mg L}^{-1}$  of copper removed. Consequently, copper remaining in the medium was less than  $350 \text{ mg L}^{-1}$ , and more than  $240 \text{ mg L}^{-1}$  was reduced to Cu(I). This indicates that only  $110 \text{ mg L}^{-1}$  of the total Cu(II) added ( $500 \text{ mg L}^{-1}$ ) remained after 48 h of incubation.

Isolate A2 (*Stenotrophomonas maltophilia*) culture reached  $0.65 \text{ OD}_{600}$  units after 6 h of incubation, and biomass production thereafter remained stable after 48 h of incubation. Cu(II) bioremoval increased, and more than  $165 \text{ mg L}^{-1}$  of copper removal was observed after 48 h of incubation. Cu(II) remaining was less than  $335 \text{ mg L}^{-1}$ , where more than  $100 \text{ mg L}^{-1}$  of Cu(II) was reduced to Cu(I). This indicates that  $235 \text{ mg L}^{-1}$  of the  $500\text{-mg L}^{-1}$  Cu(II) added remained after 48 h of incubation of isolate A2 (Fig. 6b).

In isolate A6 (*Acinetobacter* sp.), culture biomass production rose to  $0.65 \text{ OD}_{600}$  units after 6 h and increased to  $0.75 \text{ OD}_{600}$  units after 48 h of incubation (Fig. 6c). Cu(II) removal increased and was greater than  $160 \text{ mg L}^{-1}$  of copper from the liquid medium amended with  $500 \text{ mg L}^{-1}$  of copper after 48 h of incubation. Consequently, copper remaining in the medium was  $340 \text{ mg L}^{-1}$ , of which

$160 \text{ mg L}^{-1}$  of Cu(II) was reduced to Cu(I). Thus, Cu(II) remaining after 48 h of incubation was  $265 \text{ mg L}^{-1}$  of Cu(II), originally added to the culture medium.

*Acinetobacter* sp. T5 isolated from *Plantago lanceolata* rhizosphere grew exponentially, reaching  $0.40 \text{ OD}_{600}$  units after 6 h, which increased to  $0.45 \text{ OD}_{600}$  units after 36 h (Fig. 6d). Copper bioremoval increased until 36 h of incubation, and more than  $190 \text{ mg L}^{-1}$  of copper bioremoval from liquid medium amended with  $500 \text{ mg L}^{-1}$  of copper was recorded at 36 h of incubation. Accordingly, copper remaining in the medium was less than  $310 \text{ mg L}^{-1}$  after 36 h of incubation. About this concentration, more than  $190 \text{ mg L}^{-1}$  of Cu(II) was reduced to Cu(I), indicating that  $120 \text{ mg L}^{-1}$  of Cu(II) originally added remained after 36 h of incubation.

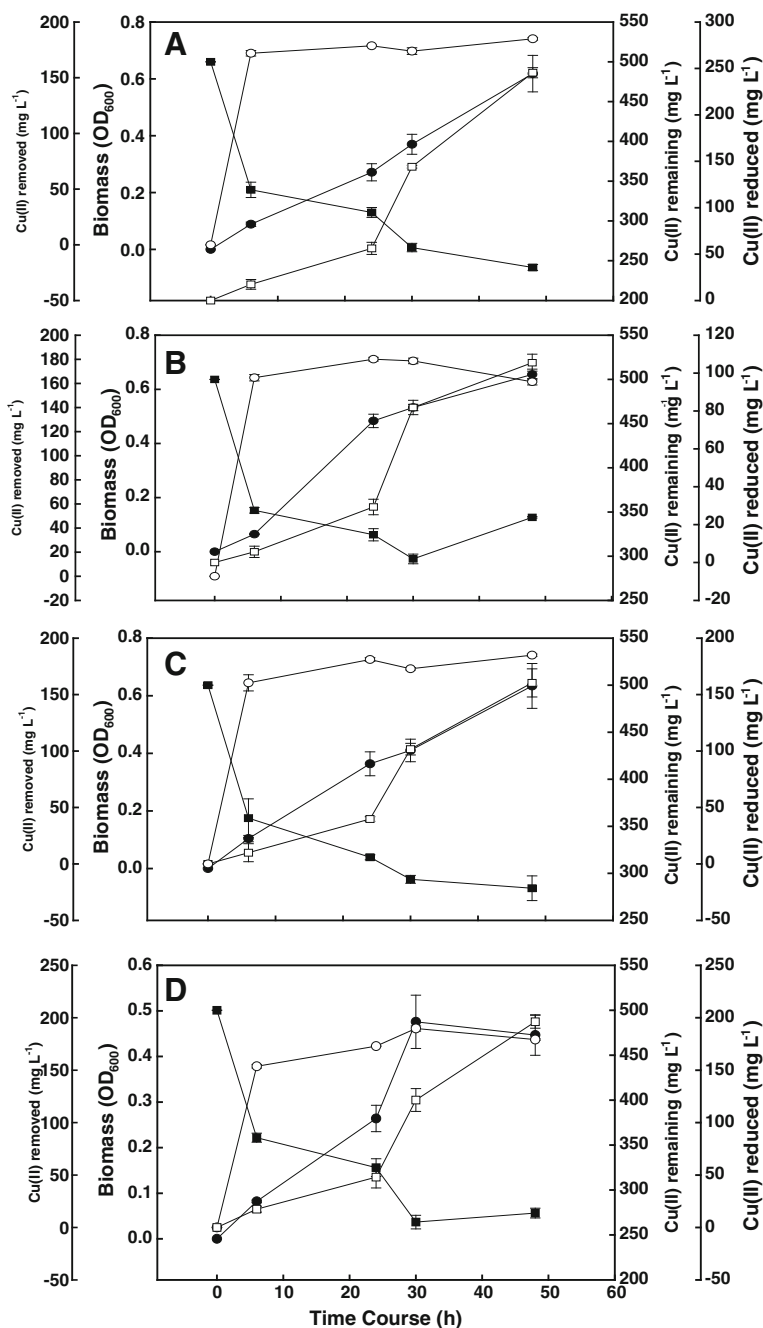
## Discussion

Copper pollution is widespread in the environment, and intensive research on the use of metal-resistant bacteria to remove metals from polluted sites is necessary. This study characterized four highly copper-resistant bacteria from the *Avena sativa* and *Plantago lanceolata* rhizospheres. However, some species of rhizosphere bacteria such as *Pseudomonas* and *Acinetobacter* were found in copper-resistant plants grown in copper-contaminated soils [5], although the copper uptake was not evaluated. *Stenotrophomonas maltophilia* was reported to be resistant to heavy metals [17], and Clausen [18], Chen et al. [19], Uslu and Tanyol [20], Chen et al. [21] reported metal resistance in *Pseudomonas putida*. *Acinetobacter calcoaceticus* strains were explored for the bioremediation of sites contaminated with metals such as chromium [18, 22], copper and arsenic [18] and sites contaminated with aromatic compounds [23]. However, copper resistance, bioreduction and biosorption capacity displayed by two of our rhizosphere bacterial isolates T5 and A6 are much higher than previously reported.

The development of biomass by bacteria was substantial in liquid medium containing copper concentrations greater than  $500 \text{ mg mL}^{-1}$  of Cu(II). Tolerance and sorption of divalent copper by our isolates were higher than observed in *Pseudomonas putida* CZ1 tolerant to only low concentrations of copper and zinc, in a range between 20 and  $25 \text{ mg L}^{-1}$  [19]. Interestingly, our isolate *Pseudomonas putida* A1 removed approximately  $150 \text{ mg L}^{-1}$  of copper after 24 h of growth in medium containing  $500 \text{ mg L}^{-1}$  of copper.

Factors such as pH, temperature and concentrations of pollutants influence microbial metabolism as enzyme activity [24]. Such factors either promote or inhibit enzyme activity [11, 25] and biosorption [13, 19, 26, 27]. In the culture of *Pseudomonas fluorescens*, the presence of  $20 \text{ mg L}^{-1}$  of copper reduced the level of  $\text{O}_2$ , which may interfere with the

**Fig. 6** Time course and relationship of biomass development, copper biosorption and bioreduction in culture of *Pseudomonas putida* A1 (**a**), *Stenotrophomonas maltophilia* A2 (**b**), *Acinetobacter calcoaceticus* A6 (**c**) and *Acinetobacter calcoaceticus* T5 (**d**) at high copper concentration ( $500 \text{ mg L}^{-1}$ ). Cultures were incubated for 48 h at  $30^\circ\text{C}$  with orbital shaking (150 rpm). Biomass (circle), Cu (II) removed (filled circle), Cu (II) remaining (filled square) and Cu(II) reduced (square) were evaluated. The error bars are standard errors of the mean



aerobic metabolism of the cell [28]. Temperature is an important factor, especially in relation to the speed of chemical and biochemical reactions. Optimal temperatures for growth of our isolates were between  $25^\circ\text{C}$  and  $30^\circ\text{C}$ . A similar optimal temperature for removal of copper was observed in a study with *Pseudomonas putida* [20]. In our study, isolate A1 identified as *Pseudomonas putida* optimally removed copper at  $35^\circ\text{C}$ . The genera *Pseudomonas* are important for bioremediation of contaminated sites [9, 18, 20, 21, 29].

The operon *cop* has a system for genetic regulation that is expressed in function of the copper concentration in the cell [24]. The gene *copA* captures the copper when the

concentration is low and transport it to the cytoplasm, and *copB* removes the copper from inside cells when there is toxic concentrations; the gene *copY* is the repressor, and *copZ* is the activator, which regulate the copper concentration in the cells [30]. Also, the environment conditions can increase and/or decrease the copper bioavailability for the microorganisms [29]. The optimum pH for enzyme activities is important for any metabolism; any pH lower than the optimum would protonate the ionized residue, while any pH higher than the optimum would ionize the protonated residue, leading to decreased activity [24]. Most enzymes have an optimum temperature, which may be

related to the type of organism from which the enzyme was isolated. Some organisms grow well near the room temperature, and so, their enzymes are most active at a temperature around 30–40°C [24]. Besides, the copper reductase activity is involved in copper reduction that has been exhibited as one of the most important pathways for copper biosorption [29].

Understanding copper speciation is important to predict copper availability in the environment [10]. Matrix pH is considered as a key factor in the variability of metal species in the environment. Thus, it can be used to determine copper species found in the environment, revealing copper bioavailability and potential toxicity in contaminated sites [31]. Changing pH from 5.0 to 3.0 decreased biosorption of Cu(II) and Mn(II) in a study with *Pseudomonas* species and *Bacillus* species [32]. All isolates in our study increased culture pH values to 7.5 and 8.5 after 24 or 48 h of incubation. Culture pH influenced the growth of the isolates, and it was high at neutral pH. Nonetheless, isolates T5 as well as A6 grew significantly at pH 5.0. This might be related to the genetic similarity of the isolates because both microorganisms were identified as *Acinetobacter* sp. The pH increase can be attributed to the exchange of Cu(II) complexed or bound with other organic compounds (COO<sup>-</sup>) existing in the medium to the H<sup>+</sup> ion, decreasing the concentration of free hydrogen, thus increasing the pH [33].

The highest sorption of copper by the isolates in this study was 190 mg L<sup>-1</sup>, achieved by isolate T5. The quantities removed in this study are much higher than 32 mg L<sup>-1</sup> removed by *Pseudomonas putida* [21] and 16 mg L<sup>-1</sup> of copper removed by *Acinetobacter calcoaceticus* [18]. *Bacillus licheniformis* removed 59 mg L<sup>-1</sup> of copper [18]. Copper bioremoval capacities of our isolates are hence remarkable.

In summary, our isolates displayed high capacity for copper resistance, reduction and uptake in aqueous media. The capacity of the rhizosphere bacteria to reduce copper as well as copper uptake are important, where the reduction of Cu(II) to Cu(I) increases copper bioavailability in the rhizosphere zone. Cu(II)-resistant bacteria isolated from the rhizosphere region of *Avena sativa* and *Plantago lanceolata* were characterized. The results indicate that our isolates are potential agents for Cu(II) bioremoval and bacterial stimulation of Cu(II) biosorption by *Avena sativa* and *Plantago lanceolata*. Furthermore, our isolates demonstrated high ability to copper biotransformation in the liquid medium, being important agents to copper reactions in the rhizosphere zone, when it was bioaugmented by the inoculation of the *Avena sativa* plants.

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