



Uncovering Anaerobic Hydrocarbon Biodegradation Pathways in Oil Sands Tailings from Two Different Tailings Ponds via Metabolite and Functional Gene Analyses

Mohd Faizd Mohamad Shahimin^{1,2} · Tariq Siddique¹

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Abstract

Oil sands tailings, a slurry of alkaline water, silt, clay, unrecovered bitumen, and residual hydrocarbons generated during bitumen extraction, are contained in ponds. Indigenous microbes metabolize hydrocarbons and emit greenhouse gases from the tailings. Metabolism of hydrocarbons in tailings ponds of two operators, namely, Canadian Natural Upgrading Limited (CNUL) and Canadian Natural Resources Limited (CNRL), has not been comprehensively investigated. Previous reports have revealed sequential and preferential hydrocarbon degradation of alkanes in primary cultures established from CNUL and CNRL tailings amended separately with mixtures of hydrocarbons (*n*-alkanes, *iso*-alkanes, paraffinic solvent, or naphtha). In this study, activation pathway of hydrocarbon biodegradation in these primary cultures was investigated. The functional gene analysis revealed that fumarate addition was potentially the primary activation pathway of alkanes in all cultures. However, the metabolite analysis only detected transient succinylated 2-methylpentane and 2-methylbutane metabolites during initial methanogenic biodegradation of *iso*-alkanes and paraffinic solvent in all CNUL and CNRL cultures amended with *iso*-alkanes and paraffinic solvent. Under sulfidogenic conditions (prepared only with CNUL tailings amended with *iso*-alkanes), succinylated 2-methylpentane persisted throughout incubation period of ~1100 days, implying dead-end nature of the metabolite. Though no metabolite was detected in *n*-alkanes- and naphtha-amended cultures during incubation, *assA/masD* genes related to *Peptococcaceae* were amplified in all CNUL and CNRL primary cultures. The findings of this present study suggest that microbial communities in different tailings ponds can biodegrade hydrocarbons through fumarate addition as activation pathway under methanogenic and sulfidogenic conditions.

Keywords Anaerobic metabolic pathway · Oil sands tailings · Alkylsuccinate synthase · Hydrocarbon biodegradation · Fumarate addition

✉ Mohd Faizd Mohamad Shahimin
faizd.shahimin@gmail.com

Tariq Siddique
tariq.siddique@ualberta.ca

¹ Department of Renewable Resources, University of Alberta, Edmonton, AB T6G 2G7, Canada

² Present Address: Faculty of Chemical Engineering & Technology, Universiti Malaysia Perlis, Aras 2, Blok S2, UniCITI Alam Campus, 02100 Padang Besar, Perlis, Malaysia

Introduction

Alkanes constitute major proportion of crude oil and its derived products, which can impact many environments, naturally or anthropogenically. Alkanes are inert in nature and have high C-H bond dissociation energy [1]. Under aerobic condition, aerobic microorganisms utilize highly reactive oxygen species to overcome the high C-H bond dissociation energies of alkanes via mono and/or dioxygenase reactions [2]. However, under anaerobic conditions, anaerobic microorganisms must activate alkanes via alternative mechanisms. For the last two decades, fumarate addition has been implicated as the main hydrocarbon activation pathway in many anaerobic cultures amended with (1) *n*-alkanes [3–7], (2) *iso*-alkanes [8, 9], (3) monoaromatics [10–13], and (4) polyaromatics [13–16]. Other anaerobic hydrocarbon activation pathways including carboxylation and hydroxylation [17] have also been implicated and highly debated but not widely reported.

Oil sands tailings ponds are engineered environments that are impacted by fugitive hydrocarbons containing a large proportion of alkanes (*n*-alkanes, *iso*-alkanes, and cycloalkanes). Extraction solvents/diluents used in the process of bitumen extraction from oil sands ores are recovered but their residual concentrations in generated waste tailings escape to oil sands tailings ponds. Indigenous methanogenic microbial communities in oil sands tailings [18] metabolize a suite of alkanes (*n*- and *iso*-alkanes) and some monoaromatics [19–21] contributing to methane (CH₄) emissions from tailings ponds [22]. Most previous studies investigated microbial degradation of hydrocarbons in the primary cultures established using fluid fine tailings (FFT) collected from Syncrude Canada Ltd. Mildred Lake Settling Basin (MLSB). These primary cultures metabolized a wide range of alkanes to CH₄. The metabolite and functional gene analyses performed on these primary cultures revealed possibility of fumarate-addition as the primary activation pathway of *n*- and *iso*-alkanes biodegradation in MLSB [8, 9].

The biodegradation pathways of alkanes in tailings ponds operated by Canadian Natural Upgrading Limited (CNUL) and Canadian Natural Resources Ltd. (CNRL) remain relatively unexplored. Previous studies [23–27] have focused on the biodegradation of specific hydrocarbon diluents used in bitumen extraction processes at these sites. While these studies characterized the microbial communities involved in degradation, the exact metabolic pathways for hydrocarbon activation are still unknown.

CNUL uses primarily paraffinic solvents (C₅–C₆; *n*- and *iso*-alkanes), whereas CNRL employs a more complex diluent such as naphtha (*n*-, *iso*-, and cycloalkanes, and alkylbenzenes primarily in the C₆–C₁₀ range). Despite identifying key microbial players, understanding the specific pathways by which these hydrocarbons are activated remains as a knowledge gap.

To address this gap, the present study is aimed to conduct functional gene and metabolite analyses on the previously studied primary cultures from CNUL and CNRL. By analyzing these cultures' metabolites and functional genes, the primary activation pathways employed by the indigenous microbial communities in these tailings ponds will be elucidated. This research will significantly contribute to existing understanding of alkane activation mechanisms and anaerobic hydrocarbon biodegradation processes in distinct tailings pond environments.

Materials and Methods

Description of Primary Cultures Established

The primary cultures (in triplicate) were established by incubating 50 mL of FFT collected from CNUL's Muskeg River Mine tailings pond or CNRL's Horizon tailings pond FFT with 50 mL of methanogenic medium or sulfate-reducing medium in a 158-mL serum bottle. Heat-killed (autoclaved at 121 °C and 20 psi for 4 consecutive days) and unamended microcosms were also prepared in duplicates to consider abiotic degradation and the production of CH₄ resulting from residual indigenous substrates in the FFTs, respectively. All primary cultures analyzed here has been described in detail in previous studies (Table 1). All cultures were incubated at room temperature (~20 °C) statically in the dark until CH₄ production or sulfate-reduction plateaued. The CH₄ concentration was measured by analyzing the headspace of all microcosms using gas chromatography equipped with flame ionization detector (GCFID) whereas sulfate concentration in all sulfate-reducing microcosms was analyzed using ion chromatography following method as described in a recent report [27]. The spiked hydrocarbon concentrations in all cultures were examined using purge-trap gas chromatography with flame ionization detection (GCFID), following the methodology outlined in Siddique et al. [20]. Additional PONA (paraffins, olefins, naphthenes, aromatics, and unknown components) analysis was performed on microcosms amended with paraffinic solvent and naphtha as detailed in the previous reports [23, 24]. These primary cultures are discussed briefly in Supplementary Information.

Determination of Metabolites During Hydrocarbon Biodegradation

The cultures from all studies were sampled for metabolite analysis at two time points: first sampling was performed during the exponential production of CH₄ or reduction of sulfate, and second sampling was done after CH₄ production or sulfate reduction plateaued [28]. Two to four milliliters of cultures were drawn from all replicates of each treatment and pooled in 20-mL EPA vials (Fisher Scientific). One microgram of 4-fluoro-1-naphthoic acid (4FIN; CAS#573-03-5; Alfa Aesar) was then added to all samples as a surrogate standard [29]. Using concentrated hydrochloric acid (CAS#7647-01-0; Sigma-Aldrich), all samples were acidified to pH < 2 [30]. The acidified samples were extracted three times with 15 mL of ethyl acetate (CAS#141-78-6; Fisher Scientific) and dried overnight in a fume hood. The dried samples were dissolved in 1.5 mL ethyl acetate and transferred into 2-mL GC vials. The solution was concentrated to ~100 µL under N₂ flow and derivatized with 100 µL of *N,O*-bis(trimethylsilyl)trifluoroacetamide (BSTFA; CAS#25561-30-2; Thermo Scientific) at 70 °C for ~90 min. The derivatized samples were diluted to 500 µL with ethyl acetate before analysis by gas chromatography-mass spectrometry (GC-MS). One microliter of derivatized sample was analyzed by TRACE 1300 gas chromatograph equipped with a TraceGold TG-5MS GC column (30 m by 0.25 mm internal diameter; Thermo Scientific) in splitless mode. The initial temperature of the oven was held at 65 °C for 5 min, increased at 5 °C min⁻¹ to 280 °C and then held at 280 °C for 15 min. Mass spectra of metabolites were obtained using ISQ LT Single Quadrupole mass spectrometer (Thermo Scientific), and the data was acquired in the scan mode from 50 to 600 mass units.

Table 1 Description of primary cultures used for metabolites and functional genes analyses

| MFT | Treatment name | Description | References |
|------|----------------|---|-------------------------------------|
| CNUL | 2-Alkanes | MFT amended with two <i>n</i> -alkanes, C5 and C6 (~400 ppm each), under methanogenic conditions | Mohamad Shahimin et al. 2016 |
| | 4-Alkanes | MFT amended with four <i>n</i> -alkanes, C5, C6, C8, and C10 (~400 ppm each), under methanogenic conditions | Mohamad Shahimin et al. 2016 |
| Par | Par | MFT amended with CNUL paraffinic solvent at ~1500 ppm under methanogenic conditions | Mohamad Shahimin and Siddique 2017a |
| | Naph | MFT amended with CNRL Naphtha at ~2000 ppm under methanogenic conditions | Mohamad Shahimin and Siddique 2017b |
| | M-3I | MFT amended with three <i>iso</i> -alkanes, 2-methylbutane, 2-methylpentane, and 3-methylpentane (~400 ppm each), under methanogenic conditions | Mohamad Shahimin et al. 2021 |
| | S-3I | MFT amended with three <i>iso</i> -alkanes, 2-methylbutane, 2-methylpentane, and 3-methylpentane (~400 ppm each), under sulfidogenic conditions | Mohamad Shahimin and Siddique 2023 |
| CNRL | 2-Alkanes | MFT amended with two <i>n</i> -alkanes, C5 and C6 (~400 ppm each), under methanogenic conditions | Mohamad Shahimin et al. 2016 |
| | 4-Alkanes | MFT amended with four <i>n</i> -alkanes, C5, C6, C8, and C10 (~400 ppm each), under methanogenic conditions | Mohamad Shahimin et al. 2016 |
| Par | Par | MFT amended with CNUL paraffinic solvent at ~1500 ppm under methanogenic conditions | Mohamad Shahimin and Siddique 2017a |
| | Naph | MFT amended with CNRL Naphtha at ~2000 ppm under methanogenic conditions | Mohamad Shahimin and Siddique 2017b |
| | M-3I | MFT amended with three <i>iso</i> -alkanes, 2-methylbutane, 2-methylpentane, and 3-methylpentane (~400 ppm each) under methanogenic conditions | Mohamad Shahimin et al. 2021 |

Nucleic Acid Extraction and Functional Gene Amplification

Triplicate 300 μL volumes of primary culture were collected from each of three replicate primary cultures described above at the end of incubation for functional genes (*assA/masD*) analysis. The total DNA was extracted from the cultures using the protocol described in detail in previous report [25]. The recovered DNA from each replicate of a single treatment was pooled before PCR amplification. The *assA/masD* gene was amplified in triplicate from pooled DNA using primers 1432F (5'-CCNACCACNAAGCAYGG-3') and 1936R (5'-TCRTCATTNCCC CAYTTNGG-3') [28, 31]. The temperature program for amplification in a S1000™ Thermal Cycler (BIO RAD) was as follows: 95 °C for 3 min followed by 40 cycles of 96 °C for 45 s, 52 °C for 30 s, 72 °C for 45 s, and a final extension step at 72 °C for 10 min. The PCR reaction (25 μL) contained 12.5 μL AccuStart II PCR ToughMix (Quanta Biosciences, Gaithersburg, MD), 1 μL extracted DNA (~20 ng/ μL), 2.5 μL (10 μM) of each primer, and 6.5 μL of sterile nuclease-free water. Negative controls containing only PCR reagents and nuclease-free water were also included with every set of samples to maintain quality control. The triplicate amplifications from single treatment were pooled, and the amplicons' size was confirmed via 1% agarose gel electrophoresis. All amplicons with the right size were excised and purified using Qiagen's QIAquick Gel Extraction kit according to the manufacturer's procedure.

Construction of *assA/masD* Gene Clone Libraries and Phylogenetic Analysis

The gel purified amplicons were quantified on NanoDrop-1000 spectrophotometer v3.3 before used for cloning with Invitrogen's TA Cloning Kit (pCR™2.1-TOPO® vector and DH5 α -T1^R *E. coli* competent cells) according to manufacturer's procedure. All clones were first checked for inserts via PCR amplification with insert-specific primers and resolved by gel electrophoresis. The amplicons were sequenced using T7 primers and BigDye Terminator v3.1 cycle sequencing kit (Applied Biosystem) on an ABI 3730 sequencer (Applied Biosystem, Foster City, CA) in Molecular Biology Services Unit (MBSU) at the University of Alberta. Initial analysis showed that the *assA/masD* genes in the samples had minimal diversity; therefore, we performed restriction analysis on all amplicons by *Msp*I digestion for restriction fragment pattern similarities to reduce the number of clones for sequencing [32]. The sequences from *assA/masD* clone libraries were trimmed to remove vector sequences. The trimmed sequences were then compared against GenBank Database using BLASTX algorithm to determine closest *assA/masD*-derived amino acid sequences. The *assA/masD* gene sequences were aligned with reference sequences from GenBank using MUSCLE [33] and manually edited. The sequences were then clustered into OTUs and translated using Geneious R8 (Biomatters Ltd, New Zealand). Maximum likelihood tree with WAG model and 100 bootstrap replicates was constructed using PhyML [34]. All cloned sequences were deposited in GenBank under the accession numbers KU840850–KU840905.

Results

Detection of Succinylated Metabolites in Primary Cultures

During incubation, different primary cultures started biodegrading the amended hydrocarbons at different times (Supplementary Figs. A–F). However, two sampling time points were selected for all cultures to determine metabolites: first, during the

active biodegradation of the amended hydrocarbons and, second, at the end of incubation to determine if the metabolites detected (if any) persisted after prolonged incubation (sampling times from all cultures are indicated in Table 2). In all methanogenic CNUL and CNRL primary cultures amended with Naph, 2-Alkanes, and 4-Alkanes, no metabolite was detected at both sampling points. However, in methanogenic samples collected from CNUL Par, CNRL Par, CNRL M-3I, and CNUL M-3I, succinylated metabolites were detected at the first sampling point but the same metabolites were not detected at second sampling point. The succinylated metabolites, however, were detected in both sampling points in samples collected from sulfidogenic CNUL S-3I (Table 2). In all samples where succinylated metabolites were detected, a cluster of two GC-MS peaks at 25.9 and 26.1 min were detected with mass spectra profiles identical to mass spectra expected from (2-methylpentyl)succinic acid (Fig. 1A, B). The mass spectra of putative double-derivatized trimethylsilyl (di-TMS) ester of (2-methylpentyl)succinic acid comprised an (M-15)⁺ ion with m/z 331 and other key ion fragments (m/z 73, 147, 174, 217, and 262) (Fig. 1B), which were similar to previously reported spectra of derivatized fumarate-added 2-methylpentane (or 1,3-dimethylbutylsuccinic acid) [9].

Additionally, we also detected another cluster of two peaks at 23.9 and 24.1 min exclusively at the first sampling point from CNUL M-3I (Fig. 1A, C). The mass spectral profiles of both peaks were similar, and the profiles were consistent with derivatized fumarate-added 2-methylbutane with diagnostic (M-15)⁺ ion fragments of m/z 317 and other key fumarate-added ion fragments (m/z 73, 147, 172, 217, and 262) (Fig. 1A, C). The key ion fragments were the same as the ion fragments associated with fumarate addition metabolites reported in previous studies [8, 35–38]. These peaks were not detected in the second sampling point from CNUL M-3I.

assA/masD Gene Detection During Anaerobic Biodegradation of Alkanes

Because alkylsuccinate metabolite was detected in some of the primary cultures and the fact that the amendments used in all the treatments were completely or mostly consisted of alkanes, functional genes encoding catalytic subunits of alkylsuccinate synthase (*assA*), also known as methylalkylsuccinate synthase (*masD*), were chosen to be investigated. The PCR was performed to amplify partial *assA/masD* genes in all the primary cultures to support the detection of postulated succinylated metabolites and to ascertain if fumarate addition was the potential activation pathway in cultures where activated metabolites were not detected.

assA/masD-like genes that are considered as key functional genes involved in activation of alkanes via fumarate addition pathway were found in the total DNA extracted from all the amended primary cultures. However, no amplification of *assA/masD*-like genes was observed from unamended cultures, which might indicate low abundance of key bacteria carrying the functional genes in alkane-deficient cultures. The restriction analysis revealed only one gene fragmentation pattern in all amended primary cultures except CNUL and CNRL Naph, which exhibited two and three unique fragmentation patterns, respectively (reflected in the number of OTUs present in each treatment; Fig. 2), indicating low diversity of *assA/masD*-like genes in the cultures. This observation coincided with observations from previous studies [8, 9] which also reported low diversity of *assA/masD*-like genes in FFT-derived enrichment cultures incubated with various *n*- and *iso*-alkanes amendments.

Clone libraries generated from all CNUL and CNRL amended primary cultures revealed that all the cloned sequences were related to *assA/masD* genes (Fig. 2).

Table 2 Silylated putative alkylsuccinate metabolites detected in extraction from methanogenic and sulfate-reducing primary cultures

| Parent compound | Retention times (min) | Selected ions (<i>m/z</i>) | CNRL MFT | | | CNUL MFT | | |
|-----------------|-----------------------|------------------------------|----------|------|-----|----------|------|------|
| | | | Par | M-3I | Day | Par | M-3I | S-3I |
| 2-Methylbutane | 23.9, 24.1 | 73, 147, 172, 217, 262, 317 | - | - | - | - | - | - |
| 2-Methylpentane | 25.9, 26.1 | 73, 147, 172, 217, 262, 331 | + | - | + | - | - | + |

M-3I, methanogenic microcosms amended with three *iso*-alkane mixture

S-3I, sulfate-reducing microcosms amended with three *iso*-alkane mixture

-, peak not detected; +, peak detected. Detection limit of alkylsuccinate metabolites on GC-MS was not determined in this study, although previous study [45] has reported detection limit in the nanomolar range for alkylsuccinates

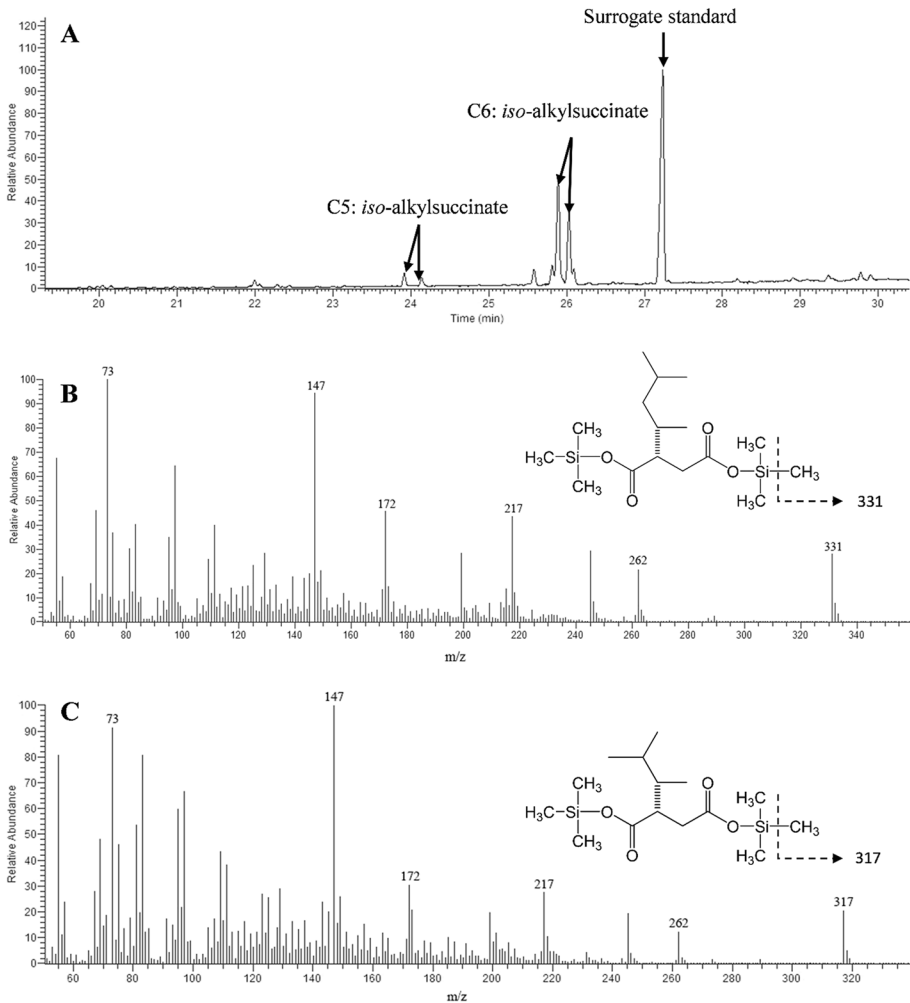


Fig. 1 Postulated TMS-derivatized metabolites extracted from CNUL M-3I detected using GC–MS. **A** Retention time of the postulated TMS-derivatized C5 and C6 *iso*-alkylsuccinate. Surrogate standard: 4-fluoro-1-naphthoic acid added in all samples prior to metabolite extraction procedure. **B** Mass spectrum of putative TMS-derivatized 2-methylpentylsuccinic acid. **C** Mass spectrum of putative TMS-derivatized 2-methylbutylsuccinic acid. Inferred putative structures and diagnostic ($M-15$)⁺ ion fragments of m/z 317 and 331 are shown in the insets. Chromatograms (**A**, **B**) were used as representative chromatograms for metabolites detected in CNUL S-3I, CNRL M-3I, CNUL Par, and CNRL Par

Subsequent analysis performed at the protein level of the *assA/masD* genes to construct maximum-likelihood tree revealed that the translated *assA/masD* fragments were closely related to a putative methylalkylsuccinate subunit (MasD) from an uncultured bacterium recovered from gas seepage-impacted and pristine cold marine sediments and putative AssA from an uncultured *Peptococcaceae* recovered from a methanogenic short-chain alkane-degrading enrichment culture (SCADC) [9, 39] [9]. All AssA/MasD sequences had high similarity at protein level ($\geq 98\%$ identity) to each other and to AssA related

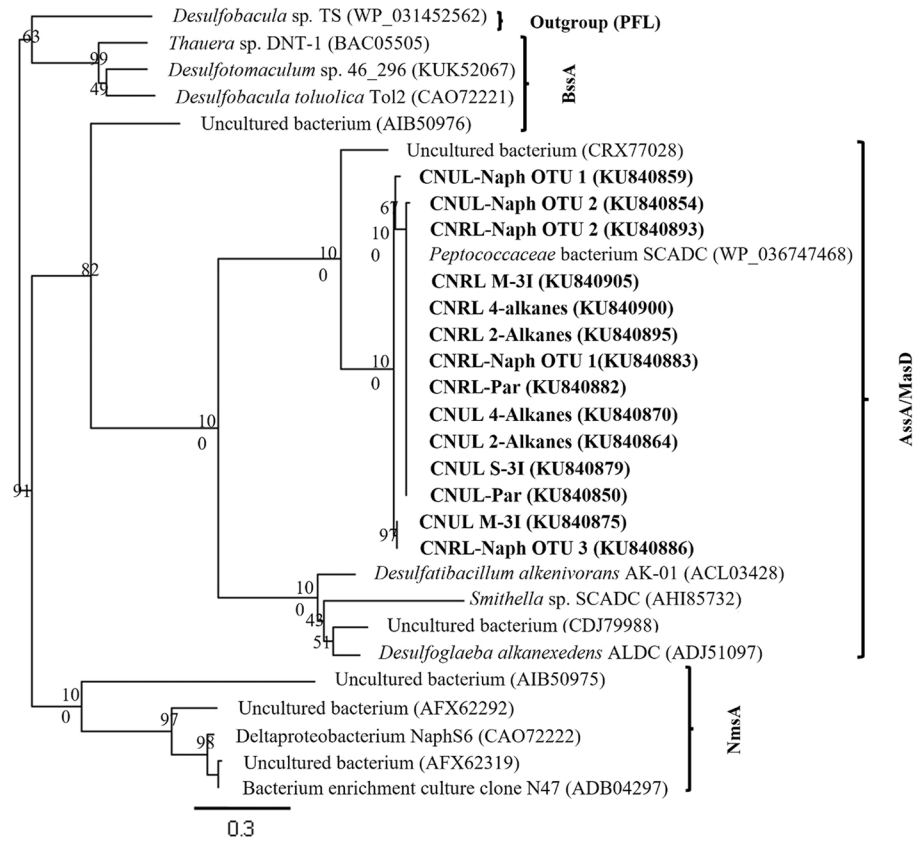


Fig. 2 Maximum likelihood tree showing the affiliation of translated sequences coding for the alpha subunit of alkylsuccinate synthase (*assA*) to selected reference sequences. *AssA*-like sequences from current study (bold) was aligned with closely related *AssA*, *BssA*, and *NmsA* reference sequences recovered from NCBI nr-database through BLASTX searches. Maximum likelihood tree was constructed using PhyML with WAG model and 100 bootstrap replicates. Pyruvate formate lyase sequence from *Desulfobacula* sp. TS was used as the outgroup. The GenBank accession numbers are indicated in parentheses

to *Peptococcaceae* SCADC except CNUL-Naph OTU 1, CNUL M-3I, and CNRL-Naph OTU 3 which had ~94–95% identity to *AssA* from *Peptococcaceae* SCADC (Fig. 2).

Discussion

In recent years, there has been increasing evidence indicating that hydrocarbons are primarily activated via fumarate activation pathway under anaerobic conditions. Fumarate addition has been demonstrated under various reducing conditions, and the corresponding succinylated metabolites have been detected in various hydrocarbon-impacted environments and hydrocarbon-degrading enrichment and pure cultures [17, 40–44]. Fumarate addition has also been demonstrated in MLSB FFT-derived enrichment cultures grown on various alkanes [8, 9]. However, no similar study on hydrocarbon activation pathway has been reported in FFTs from relatively “younger” oil sands tailings ponds

managed by CNUL and CNRL, which employed different oil sands extraction process and treated their tailings differently for solid consolidation enhancement before being deposited into the tailings ponds. Initial studies on biodegradation of different hydrocarbon mixtures in CNUL and CNRL FFT [23–25] revealed the enrichment of different microbial taxa during the biodegradation process and distinct biodegradation pattern in CNUL and CNRL FFT-associated microbial communities. These differences beg the question whether hydrocarbons in CNUL and CNRL tailings ponds were also activated differently in the biodegradation process.

To identify the metabolites resulting from activation of the amended alkanes and to discover if the detected metabolites persist in the cultures during extended incubation, culture samples retrieved from two sampling points (sampling time indicated in Table 2) from all primary cultures were analyzed. Succinylated metabolites were only detected in CNUL M-3I, CNUL S-3I, CNRL M-3I, CNUL Par, and CNRL Par (Table 2) but not in the other primary cultures. The putative metabolites were only detected during active biodegradation of *iso*-alkanes; however, the metabolites appeared to be transient because the metabolites were not detected at second sampling point excepting CNUL S-3I. Similarly, in previous reports which examined methanogenic biodegradation of alkanes, succinylated metabolites were only transiently detected in cultures grown on *iso*-alkanes [8, 9] but not in cultures amended with *n*-alkanes [9, 45]. This phenomenon could be attributed to the recalcitrance nature of *iso*-alkanes over *n*-alkanes, which have been reported previously [23, 24], where *iso*-alkanes were shown to have lower conversion rate to CH₄ than *n*-alkanes. Thus, although under methanogenic condition the succinylated *iso*-alkanes were only transient products, the succinylated *iso*-alkanes were not further oxidized as rapidly as succinylated *n*-alkanes, allowing the metabolic intermediates to transiently accumulate and, thus, be detected during the exponential *iso*-alkanes degradation phase. The reason for the unsuccessful detection of any metabolites from methanogenic cultures amended with *n*-alkanes, therefore, may be explained by (1) non-accumulating metabolites resulting from the close link of alkane activation with metabolites turnover and (2) instrumentation limitation in detecting low yield metabolites.

Nonetheless, *iso*-alkanes also constitute a major portion of naphtha (~39 wt% [24]). However, naphtha comprised a wide range of *iso*-alkanes (C₆–C₁₀) as opposed to paraffinic solvent, which only has three major *iso*-alkane components constituting ~58 wt% of paraffinic solvent's total mass [23]. Therefore, the concentration of individual *iso*-alkanes in naphtha is low compared to the concentration of individual *iso*-alkanes in paraffinic solvent; hence, the concentration of any metabolites that might arise from the oxidation of the *iso*-alkanes in naphtha amended cultures may be too low to be detected by GC–MS. Interestingly under sulfate-reducing conditions, the succinylated *iso*-alkane, 1,3-dimethylbutylsuccinic acid persisted throughout ~1100 days incubation even though sulfate concentration has plateaued since day ~900 (Supplementary Fig. F). Accumulation of 1,3-dimethylbutylsuccinic acid in CNUL S-3I might be attributed to the absence or low growth of microbes carrying the enzymes appropriate for oxidation of 1,3-dimethylbutylsuccinic acid. The accumulation of 1,3-dimethylbutylsuccinic acid might have also resulted in inhibition of further degradation of 2-methylpentane in CNUL S-3I despite the fact that high concentration of sulfate was still present in the culture (Supplementary Fig. F). The derivatized metabolites from 2-methylbutane and 2-methylpentane exhibited a cluster of two GC–MS peaks. These peaks might have represented diastereomers formed by enzymatic mechanisms as described by [46], which have also been reported in previous studies examining anaerobic biodegradation of *iso*-alkanes [8, 9, 47].

Activation of alkanes via fumarate addition pathway is carried out by glycyl radical alkylsuccinate or methylalkylsuccinate synthase (Ass or Mas, respectively). To determine the presence of genes encoding Ass/Mas and diversity of the genes in all the cultures, amplification of *assA/masD* genes, which encode for the catalytic subunit of Ass/Mas, was performed using degenerate *assA/masD* primers. *assA/masD* genes were amplified from all CNUL and CNRL amended primary cultures, indicating that microbial communities in all the primary cultures have the potential for activating alkanes via fumarate addition pathway despite that the expected alkylsuccinate metabolites were not detected in all the cultures, possibly due to quick further transformation. The diversity of the translated *assA/masD* genes in all the cultures, however, is low and has high similarity to *Peptococcaceae* SCADC (Fig. 2). Incidentally, all the primary cultures exhibited enrichment of *Peptococcaceae* (~36–77% sequence reads) during biodegradation of the amended alkanes [23–25], and similar enrichment of *Peptococcaceae* has also been reported in *n*- and *iso*-alkane-degrading cultures from FFT [8, 9, 21, 25–27, 39, 48], providing further evidence for implication of *Peptococcaceae* as important degraders of alkanes in oil sands tailings ponds.

Conclusion

In summary, this study significantly advances the overall understanding of hydrocarbon activation pathways under methanogenic and sulfidogenic conditions, particularly highlighting the pivotal role played by the fumarate addition pathway. Specifically, this study underscores the complexities involved in studying hydrocarbon metabolism due to the transient nature of metabolites during biodegradation process. Moreover, the findings underscore the significance of *Peptococcaceae* in the degradation of hydrocarbons within oil sands tailings ponds. The metabolite and functional gene profiling data, combined with results from previous studies [8, 9, 49], strongly indicate the prevalence of the fumarate addition pathway across various oil sands tailings ponds, despite differences in extraction techniques and tailings management practices among operators. Nevertheless, the potential existence of alternative metabolic routes, such as carboxylation and/or hydroxylation for hydrocarbon activation within these environments, is acknowledged.

Future investigations should delve deeper into elucidating the governing mechanisms behind these pathways and explore the factors influencing the persistence of specific metabolites. Such endeavors will undoubtedly propel our understanding of anaerobic hydrocarbon biodegradation in intricate environments to new frontiers, further enriching our knowledge base and potentially informing improved environmental management strategies, not only for remediation oil sands tailings ponds but also other hydrocarbon-impacted environments.

Supplementary Information The online version contains supplementary material available at <https://doi.org/10.1007/s12010-024-04855-0>.

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Author Contribution Conceptualization: TS and MFMS, data curation: MFMS, formal analysis: MFMS, funding acquisition: TS, investigation: MFMS, methodology: MFMS, project administration: MFMS, resources: TS, supervision: TS, validation: TS, writing original draft: MFMS, and review and editing: MFMS and TS.

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Data Availability All data generated or analyzed during this study are included in this published article and its supplementary information files.

Declarations

Ethics Approval Not applicable.

Consent to Participate Not applicable.

Consent for Publication Not applicable.

Competing Interests The authors declare no competing interests.

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