

# A Review on the Genetics of Aliphatic and Aromatic Hydrocarbon Degradation

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**Abstract** Because of the high diversity of hydrocarbons, degradation of each class of these compounds is activated by a specific enzyme. However, most of other downstream enzymes necessary for complete degradation of hydrocarbons maybe common between different hydrocarbons. The genes encoding proteins for degradation of hydrocarbons, including the proteins required for the uptake of these molecules, the specific enzyme used for the initial activation of the molecules and other necessary degrading enzymes are usually arranged as an operon. Although the corresponding genes in many phylogenetic groups of microbial species show different levels of diversity in terms of the gene sequence, the organisation of the genes in the genome or on plasmids and the activation mode (inductive or constitutive), some organisms show identical hydrocarbon-degrading genes, probably as a result of horizontal gene transfer between microorganisms.

Keywords Genetics · Hydrocarbons · Alkanes · Aromatics

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# Introduction

Due to the high diversity of the molecular structures of hydrocarbons, each class of compound has to be degraded by a specific enzyme [1]. Through the catabolic process, hydrocarbon molecules are first taken up by specific microbes, and then converted to simple organic molecules [70, 78]. Based on the microbial species and the community in which the degrader species lives, these simple organic molecules may then be used by the same organism or may be released into the environment and further catabolised by other microbial members of the community [27]. In terms of molecular investigations, such synergistic relationships suggest the probable presence of corresponding degradation genes in more than one member of the microbial community [7]. This review describes the genes involved in the degradation of different types of aliphatic and aromatic hydrocarbons in both aerobic and anaerobic conditions.

All hydrocarbon oxygenases, including both monooxygenases and aromatic-ringhydroxylating dioxygenases, are classified into a large family of iron-sulphur-flavoproteins containing enzymes capable of transferring one or two electrons to their substrates [113, 115]. Despite the existence of some monoxygenases that are structurally homomultimer ( $\alpha_6$ ) and physiologically related to each other [38], the aromatic-ring-hydroxylating dioxygenases usually have a heterohexamer ( $\alpha_3\beta_3$ ) structure. Based on the sequence diversity of their  $\alpha$  subunits, these aromatic hydroxylating dioxygenases are sub-classified into four groups: toluene/benzene oxygenase, toluate/benzoate oxygenase, naphthalene oxygenase and biphenyl oxygenase [47, 93]. Furthermore, based on their native substrates, the aromatic ring hydroxylating dioxygenases are classified into four groups: the toluene/biphenyl family (specialised for initial oxidative attack to benzene, toluene, chlorobenzenes, isopropylbenzene and biphenyl), the naphthalene family (for activation of naphthalene, phenanthrene, nitrotoluene and nitrobenzene), the benzoate family (for catabolism of toluate, benzoate, anthranilate, isopropylbenzoate, trichlorophenoxyacetate, 2-chlorobenzoate) and finally the phthalate family (for the initial catabolism of aromatic acids such as phenoxybenzoate, p-toluene sulfonate, phthalate, vanillate, 3-chlorobenzoate) [41, 42]. In addition to these classified enzymes, several oxygenase enzymes have been identified which are specific for different substrates, such as salicylate, ohalobenzoate, 3-phenylpropionate, dibenzodioxin, aniline and dehydroabeitate [28, 76].

The full range of genes required for hydrocarbon degradation and their genetic organisation is not fully understood in many organisms. Although most of the genetic elements involved in the catabolism of aliphatic hydrocarbons are ordered in inducible operon structures with several coregulated genes present in the same transcription unit, in some cases, these operons are expressed constitutively [15, 120]. Both inducible and constitutive genes can also be located on chromosomes either in the form of an operon unit or as separate genes [19, 88]. Based on the catabolic genes present and their homology in endonuclease restriction patterns, DNA rearrangements and electrophoretic mobility, plasmids containing hydrocarbon-degrading genes are divided into three groups: the plasmids containing alkane degradation genes (like the OCT plasmid) [88], the plasmids containing naphthalene and salicylate degradation genes (such as the NAH plasmids) [75] and finally, the plasmids containing toluene- and xylene-oxidising genes (like the TOL plasmids) [75].

#### The Alkane Degradation Genes

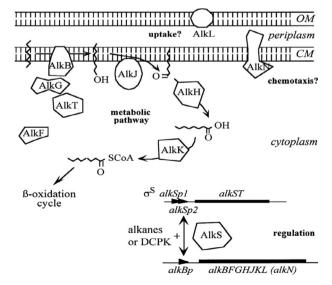
As an outline, biodegradation of hydrocarbons starts with oxidation of the substrate molecules by an electron-carrier-dependent monooxygenase system, producing the corresponding alcohol [60]. After conversion of the hydrocarbons to their corresponding alcohol, the alcohols are further oxidised and broken down to smaller molecules that finally are utilised via the central catabolic reactions of the cells (reaction 1) [89].

Reaction 1: R-CH2OH 
$$\frac{\text{Alcohol}}{\text{Dehydrogenase}} > \text{R-CH2O} \frac{\text{Aldehyde}}{\text{Dehydrogenase}} > \text{R-COOH}$$

Based on the chain length of the aliphatic hydrocarbons utilised, *n*-alkane utilising organisms are classified into three groups: methanotrophs, gaseous alkane-utilising (C2 to C4) microorganisms and finally liquid alkane-catabolising (C5 to C20) microorganisms [89, 112]. Furthermore, based on the molecular structure and the supporting electron transport system, monooxygenases are classified into rubredoxin-dependent enzymes and (bacterial and fungal) cytochrome P450-containing monooxygenases [29, 90]. The rubredoxin-dependent enzymes are composed of a rubredoxin reductase, a rubredoxin and an alkane hydroxylase [45]. In most bacteria, the rubredoxin-dependent monooxygenases are encoded by the gene *alkB* while some bacteria, such as *Acinetobacter* sp., express the enzyme encoded by the *alkM* gene [114].

*Pseudomonas putida* GP01, for instance, uses a monooxygenase to convert *n*-alkanes (C6– C10) into their fatty acids (Fig. 1) [88]. The *alk* gene organisation in this strain is located on OCT plasmid and encodes the enzymes required for degradation of C5–C13 *n*-alkanes [88]. In *P. putida*, it is postulated that the *alkL gene* is involved in import of *n*-alkanes into the bacterial cells [37]. Furthermore, it has been shown that changes in the configuration of an outermembrane protein encoded by *blc* in *Alcanivorax borkumensis* Sk2 can lead to the transport of short-length chain hydrocarbons into the cells [55]. Furthermore, it is believed that the longlength chain fatty acid transporter proteins (*FadL*) in many bacteria participates in the transportation of long-length chain hydrocarbons into the cell [6]. The alkanes are initially oxidised by a trimer alkane hydroxylase (a complex of alkane monooxygenase, rubredoxin and rubredoxin reductase encoded by *alkB*, *alkG* and *alkK*, respectively), which are integrated into the inner cell membrane of the bacterium via the product of *alkB* [88]. The resulting alcohol is further oxidised by the products of *alkJ* and *alkH* (respectively for an alcohol

Fig. 1 The position and role of alkane-degrading proteins in *Pseudomonas putida. alkB* alkane hydroxylase, *alkF* rubredoxin 1, *alkG* rubredoxin 2, *alkH* aldehyde dehydrogenase, *alkJ* alcohol dehydrogenase, *alkL* outer-membrane protein, *alkN* chemotactic transducer of *Pseudomonas aeruginosa*, *alkT* rubredoxin reductase, *alkS* positive regulator [116]



dehydrogenase and aldehyde dehydrogenase) into an aldehyde and acid, respectively, which is activated by addition of a CoA to the acid (through the action of the product of *alkK*). These genes are organised as *alkBFGHJKL* on the OCT plasmid and are controlled by the action of the products of another operon (alkST) located 40 kb away from the first operon [40]. The same operon structure (alkBFGHJKL) exists in P. putida P1, but alkST has been moved to a position upstream of the operon and also the *alkL* and *alkN* genes are not separated by an insertion sequence (IS) [116]. In Acinetobacter sp. strain ADP1, the three subunits of alkane hydroxylase (specialised for C6–C12 alkanes), alkane monooxygenase, rubredoxin and rubredoxin reductase, are encoded by *alkM*, *rubA* and *rubB*, respectively [109]. With the exception of Rhodococcus ervithropolis, none of the rubredoxin reductase genes are situated near a hydroxylase gene probably because of the involvement of the rubredoxin reductases in other metabolic pathways as well [102]. This well-organised aliphatic hydrocarbon-degrading gene cluster is not always observed in all aliphatic hydrocarbon-degrading bacteria. Acinetobacter HOI-N, for instance, is a hydrocarbon-degrading bacterium that contains a set of aliphatic hydrocarbon-degrading enzymes located at three separate loci on the chromosome: the gene encoding the alkane hydroxylase is located at a considerable distance from the genes specified for the alcohol dehydrogenase as well as aldehyde dehydrogenase [5, 85].

The alkane hydroxylating enzymes involved in initial activation of long-chain length aliphatic hydrocarbons (>C18) are evolutionary distinct from the previously mentioned enzymes involved in hydroxylation of short-length chain hydrocarbons. For instance, *AlmA*, encoded by *alma*, is an alkane monooxygenase belonging to the flavin-binding family in *Acinetobacter* sp. DSM17874, *Alcanivorax* and many other bacteria which involves in initial activation of the hydrocarbons with bigger than C32 [119]. Furthermore, *LadA* is a thermophilic alkane hydroxylase, belonging to flavin-dependent oxygenase, obtained from *Geobacillus thermodenitrificans* NG80-2 with the ability to hydroxylate C15–C36 alkanes [59].

The genes involved in degradation of alkane hydrocarbons are downregulated by two regulatory systems, cytochrome ubiquinol oxidase (*Cyo*) and the global regulatory protein Crc [18] to ensure expression of these genes just in certain physiological conditions. The *Cyo* gene product is known as a global regulatory factor able to regulate carbon metabolism and respiration. This factor suppresses the expression of alkane-degrading genes in the presence of easily metabolised carbon sources [18]. The *Crc* gene product is a RNA-binding protein with the ability to stop the mRNA translation via binding to the 5<sup>/</sup> end of the mRNAs responsible for production of both the regulatory factor *alkS* and alkane-degrading proteins [40].

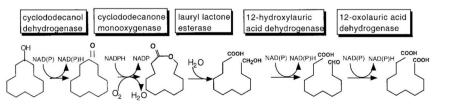
### Cycloaliphatic Compounds

The cycloaliphatic hydrocarbons, like cyclopentane, methylcyclopentane and cyclohexane are degraded by a large range of bacteria [16]. *Acinetobacter* strain SE19, for instance, uses six chromosomal catabolic genes, arranged as *chnBER* ORF and *chnADC* ORF for the degradation of cycloaliphatic hydrocarbons [16]. Through this reaction, cyclohexanone monooxygenase (encoded by *chnB*) and NAD(P)H-dependent aldehyde dehydrogenases (encoded by *chnE*) convert cyclohexane into cyclohexanone, which is further oxidised by the products of the *chnADC* ORFs. The *chnADC*, which is located in opposite direction of the *chnBER* ORF, encodes cyclohexanol dehydrogenase, 6-hydroxyhexanoic acid dehydrogenase

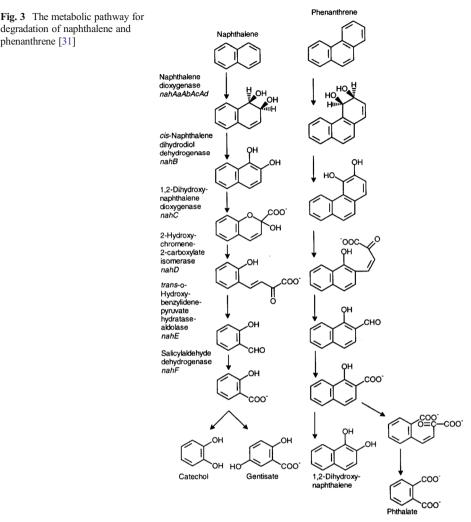
and caprolactone hydrolase, respectively, by *chnB*, *chnE* and *chnR*, to produce an end product of oxohexanoic acid. In the same way (Fig. 2), *Rhodococcus ruber* SC1 uses a cyclododecane monooxygenase and a NAD(P)H-dependent aldehyde dehydrogenases to oxidise cyclododecane first to cyclododecanol and then to cyclododecanone [51]. A cyclododecanone monooxygenase, encoded by *cddA* oxidises it into a lactone oxacyclotridecan-2-one (lauryl lactone), which is first hydroxylated by lauryl lactone esterase (encoded by *cddB*) to 12hydroxydodecanoic acid and then is oxidised twice by two dehydrogenases (12-hydroxylauric acid dehydrogenase and 12-oxolauric acid dehydrogenase, encoded respectively by *cddC* and *cddD*, to make a 12-oxolauric acid and finally a DDDA (dodecanedioic acid) [51]. Although different bacteria utilise the same genes for the degradation of cyclododecane, their gene organisation may be different. In *R. ruber* SC1, the gene cluster is arranged as *cddABCDXY* with two-space ORFs between *cddABCD* and *cddXY*, while the gene order in *chn* cluster of *Acinetobacter* sp. strain SE19 is random [51].

# The Plasmids Containing Naphthalene and Salicylate Degradation Genes

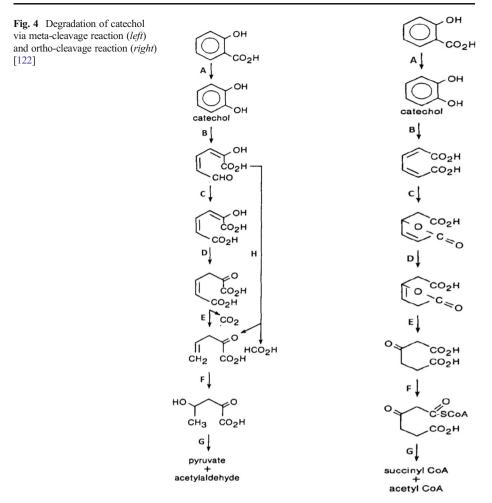
Several aromatic-degrading bacteria are able to convert mono/multiple cyclic aromatic hydrocarbons into salicylate, which undergoes a meta-cleavage to present the products to tricarboxylic acid cycle (TCA) [43]. As a prototype dioxygenase enzyme, a (Rieske-type two-iron twosulphur centre containing) naphthalene dioxygenase (NOD; encoded by *nahAaAbAcAd*) inserts two oxygen atoms into the aromatic ring of a broad range of aromatic hydrocarbons, such as naphthalene, phenanthrene and anthracene, converting them to corresponding dihydrodiols, such as cis-naphthalene dihydrodiol and cis-phenanthrene dihydrodiol, respectively [9, 44]. Next, a *cis*-dihydrodiol dehydrogenase (encoded by *nahB*) dehydrogenates the dihydrodiols to make 1,2-dihydroxynaphthalene, which is subjected to meta-cleavage by 1,2dihydroxynaphthalene dioxygenase (nahC) to form 2-hydroxychromene-2-carboxylic acid (Fig. 3). After an enzymatic *cis* to *trans* isomerisation (by an isomerase encoded by *nahD*), the side-chain at the trans-unsaturated bond of the trans-o-hydroxybenzylidenepyruvate product is cleaved by a hydratase-aldolase (encoded by *nahE*) to produce a salicylaldehyde. The product is finally dehydrogenated by NAD-dependent salicylaldehyde dehydrogenase to salicylate (encoded by *nahF*). Depending on the bacterial strain, the covalent bond of the aromatic ring of salicylate is cleaved between two adjacent carbon atoms with hydroxyl groups (meta-cleavage) or between a carbon with a hydroxyl group and its adjacent carbon with a carboxyl group (ortho-cleavage) [97]. In most cases, like P. putida PpG7 (containing NAH7 plasmid) and P. putida R1 (containing SAL1 plasmid), bacteria use a meta-cleavage reaction on salicylate (Fig. 4) [80] in which bacteria salicylate hydroxylase (nahG) convert salicylate



cyclododecanol cyclododecanone lauryl lactone 12-hydroxylauric acid 12-oxolauric acid dodecanedioic acid Fig. 2 The metabolic pathway for degradation of cyclododecanol by *Rhodococcus ruber* SC1 [51]



into catechol. The product is oxidised by catechol oxygenase (*nahH*) to 2-hydroxymuconic semialdehyde. From here this intermediate can pass two different ways: in one way, the molecule is directly hydrolysed by a hydroxymuconic semialdehyde hydrolase (*nahN*) into 2-Oxo-4-pentenoic acid, while through a second pathway, the product is acted on by 2-hydroxymuconic semialdehyde dehydrogenase (*nahI*) and 4-oxalocrotonate isomerase (*nahJ*) to produce 2-hydroxymuconic acid and 4-oxalocrotonic acid before conversion by 4-oxalocrotonate decarboxylase (*nahK*) into 2-Oxo-4-pentenoic acid. This intermediate is the substrate for 2-Oxo-4-pentenoate hydratase (*nahL*) and is converted to 4-hydroxy-2-oxovaleric acid which is broken by 2-Oxo-4-hydroxypentanoate aldolase (*nahM*) into pyruvic acid and acetaldehyde. The acetaldehyde is converted by Acetaldehyde dehydrogenase (*nahO*) into acetyl-CoA. In the ortho-cleavage pathway, on the other hand, bacteria use three subsequent enzymes, 2-oxo-4-hydtoxypentanoate aldolase (*nahM*), catechol 1,2-oxygenase (*carA*) and *cis*-muconate lactonising enzyme (*catB*) to convert catechol into succinate and acetyl-CoA (Fig. 4) [117].



Through the gentisate pathway, bacteria employ an alternative pathway in which gentisate (2,5-dihydroxybenzoate) is subjected to a ring cleavage by gentisate 1,2-dioxygenas (*BagI*) to produce a maleylpyrovate [3, 64]. The product can directly hydrolyse into pyruvate and malate or may go into another process in which the product is first isomerised by an isomerase (*bagKL*) into fumarylpyruvate before hydrolysing (*bagK*) to pyruvate and fumarate [65]. Several other bacteria, such as *Salmonella typhimurium* and *Pseudomonas alkaligenese* are also able to use this system to degrade other substrates, such as 3-hydroxybenzoate and xylenol, respectively, through conversion to gentisate as an intermediate [21, 30].

Although most of the genes responsible for degradation of naphthalene identified in different bacteria show 99–100 % homology with their counterparts in other strains, the location (plasmid or chromosome) and organisation of these gene clusters may be different in each strain (Fig. 5) [62]. However, there is a lower similarity between those genes identified in mycobacterial species and those in other bacteria probably due to the origin of the genes being from different sources or, less likely due to a greater rate of genetic changes in the mycobacterial genes [50]. NAH plasmids are a group of highly homologous plasmids, which

<b>Fig. 5</b> The gene organisation in										$^{\circ}$		
different strains of <i>Pseudomonas</i> sp. [31]	Consensus	nah	Aa	Ab	Ac	Ad	В	F	С	Q	E	D
	P. putida NCIB 9816-4	nah	Aa	Ab	Ac	Ad						
	P. putida G7	nah	Aa	Ab	Ac	Ad			С	Q	Ε	D
	P. putida BS202	nah	A1	A2	A3	A4	В	F	С		Е	D
	Pseudomonas sp. C18	dox		Α	В	D	Е	F	G	н	1	J
	P. aeruginosa PaK1	pah	Α	A2	AЗ	A4	В	F	С	Q	Е	D
	P. putida OUS82	pah	Aa	Ab	Ac	Ad	В	F				
	P. putida NCIB 9816	ndo		Α	В	С						
	P. fluorescens ATCC 17483	ndo		C1	C2	C3						
	P. putida ATCC 17484	ndo		C1	C2	C3						

carry naphthalene catabolic genes. These plasmids can be distinguished by their restriction endonuclease digestion patterns [10]. The gene sequences of all identified NAH plasmids, such as NAH7 in *P. putida* PpG7, pNL1 in *Novosphingobium aromaticivorans* F199, pND6-1 in *Pseudomonas* sp. strain ND6 and pWW60-1 in *P. putida* NCIB9816 are highly conserved, with 90–100 % homology in the gene sequences [61, 75, 80]. NAH7 in *P. putida* PpG1carries two separate operons of which the NAH operon is specialised for conversion of PAHs, including naphthalene, anthracene and phenanthrene, to salicylate (*nahAaAbAcAdBFCED*) and the Sal operon is used for the catabolism of salicylate to catechol and further to TCA cycle intermediates (*nahGTHINLOMKJ*) [75]. In addition, *nahX* (with an unknown function) and *nahY* (a chemotaxic transducer protein) are located downstream of the *nahJ* [75, 81]. The product of the *nahY* gene acts as a methyl-accepting chemotaxis protein for cell attraction towards naphthalene via flagella-dependent movement [81]. A *nahR* gene located between these two operons positively regulates the expression of both of the operons [33, 75].

# The Plasmids Containing Toluene- and Xylene-Oxidising Genes

Through several pathways, bacteria insert one or more hydroxyl groups into aromatic rings to form a catechol (Fig. 6), which is later cleaved for further catabolism. In one of these pathways, the toluene-degrading genes in *P. putida*, located on the TOL plasmid, degrade this molecule into benzoic acid, cis-benzoate dihydrodiol and finally to catechol that in turn is cleaved for further oxidation processes (Fig. 6; pathway A). Conversion of toluene into benzoate is performed by xylA, benzylalcohol dehydrogenase (xylB) and benzaldehyde dehydrogenase (xylC), while the next process, oxidation of toluate to catechol, is carried out by the products of the xylD, xylE, xylF and xylG genes. The first group of enzymes for the production of benzoate (encoded by xylCAB) is located on a plasmid, while the genes responsible for conversion of toluate into catechol can be found on both plasmids and chromosomes (Fig. 7) [48, 57]. Catechol and its derivatives are cleaved via one of two meta-cleavage activities. In one pathway, 2-hydroxymuconic semialdehyde is directly converted via hydroxymuconic semialdehyde hydrolase (HMSH; encoded by xylF) into 2-oxopent-4-enoate or its derivatives [36]. In the second pathway, 2-hydroxymuconic semialdehyde is first oxidised by 2hydroxymuconic semialdehyde dehydrogenase (HMSD; encoded by xylG) to its corresponding dioates before isomerisation (encoded by xvlI) to 2-oxopent-4-enoate (Fig. 8) [36]. This last product is finally hydrolysed (by the xy/K-encoded enzyme) to 4-hydroxy-2-oxovalerat before cleavage by 4-hydroxy-2-oxovalerate aldolase (encoded by xvlJ) into pyruvate and

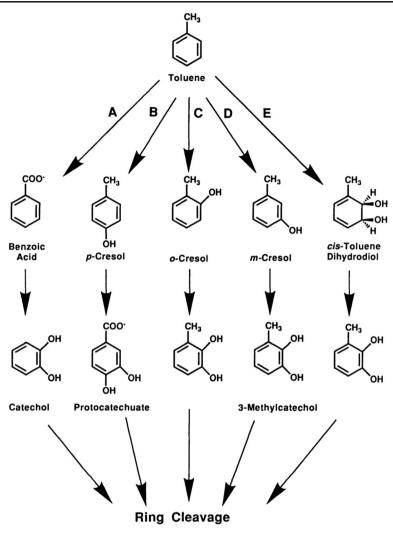


Fig. 6 Five possible pathways for degradation of toluene [124]

propionaldehyde [36]. The gene cluster on pTOL in *P. putida* is ordered as *xyl XYZLTEGFJQKIH* where *xylZ* and *xylL* encodes for 1,2-dioxygenase and 1,2-dihydroxycyclohexa-3,5-diene-carboxylate dehydrogenase, respectively and other genes downstream to *xylL* involve in the lower catabolic pathway [48].

Different strains of bacteria harbour quite similar operons in terms of DNA sequence and gene organisation. *Sphingomonas yanoikuyae* B1 harbours a *xylXYEFGJQKIHT* operon where in addition to internal gene rearrangements, the *xylL and xylZ* genes have moved to a separate place on the genome [49]. Furthermore, the operon *bphR1bphA1A2(orf3)bphA3A4BCX0X1X2X3D* in *Burkholderia* sp. strain LB400 and *Pseudomonas pseudoalcaligenes* KF707, responsible for degradation of biphenyls to pyruvate and acetaldehyde, is highly homologous to the operons for degradation of toluene (Fig. 9) [25]. The corresponding operon in *Pseudomonas* KKS102 is broken into a

benzaldehyde dehydrogenase [98]

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XYIXYZ

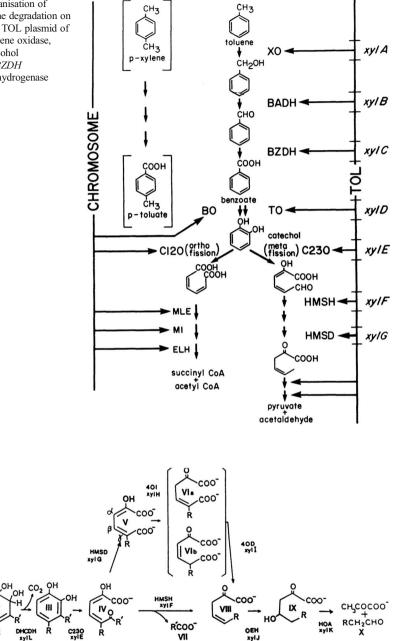
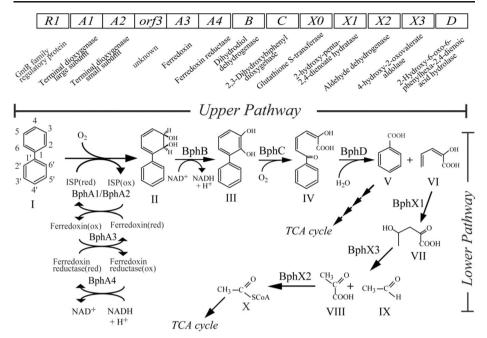


Fig. 8 The metabolic pathway for the degradation of toluate and benzoate via the meta-cleavage process. The 2hydroxymuconic semialdehyde intermediate can oxidise via two different reactions. It may be oxidised directly by hydroxymuconic semialdehyde hydrolase (HMSH; into 2-oxopent-4-enoate or may be first converted by HMSD (2-hydroxymuconic semialdehyde dehydrogenase into a dioates before isomerisation to 2-oxopent-4enoate. TO toluate 1,2-dioxygenase, DHCDH 1,2-dihydroxycyclohexa-3,5-diene-carboxylate dehydrogenase, C230 catechol 2,3-dioxygenase, 40D 4-oxalocrotonate decarboxylase, OEH 2-oxopent-4-enoate hydratase, HOA 4-hydroxy-2-oxovalerate aldolase. xylE to xylZ are names of the genes specified for the degrading enzymes [36]

vII

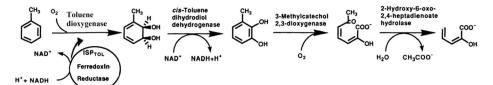
xylJ



**Fig. 9** The degradation pathway of biphenyl into acetyl-CoA and the order of the corresponding genes (*bph* operon) in *Pseudomonas pseudoalcaligenes* KF707. The biphenyl dioxygenase (consisting of two subunits of the terminal dioxygenase and ferrodoxin encoded respectively by bphAlA2A3A4) inserts two oxygen atoms into biphenyls (I) to yield dihydrodiols (II). The product is further oxidised by dihydrodiol dehydrogenase (encoded by bphB) to 2,3-dihydroxybiphenyl (III), which undergoes a cleavage at the meta-position by the product of the *bphC* gene (2,3-dihydroxybiphenyl dioxygenase) to yield 2-hydroxy-6-oxo-6-phenylhexa- 2,4-dienoate (IV). This molecule is broken by 2-hydroxy-6-oxo-6-phenylhexa-2,4-dienoic acid hydrolase (encoded by bphD) into benzoic acid (V) and 2-hydroxypenta-2,4-dienoate (VI). A hydratase, encoded by (bphX1) converts the 2-hydroxypenta-2,4-dienoate to 4-hydroxy-2-oxovalerate (VII), which is cleaved by 4-hydroxy-2-oxovalerate aldolase (encoded by bphX3) into pyruvate (VIII) and acetaldehyde (XI). The acetaldehyde is then activated by acetaldehyde dehydrogenase (encoded by bphX2) and glutathione-S-transferase (encoded by bphX0) through the addition of CoA (X) before entering the TCA cycle. *ISP* iron-sulphur protein [25]

*bphEGF*(ORF4)*A1A2A3BCD*(ORF1)*A4* operon, which allows the catabolic reaction of biphenyls to proceed to 2-hydroxypenta-2,4-dienoate and benzoic acid, and the gene cluster *bphEGF* located 4 kbp upstream of the first operon, which encode for hydratase, aldolase and dehydrogenase and convert these intermediate products into pyruvate and acetaldehyde [77]. Finally, the gene cluster for the degradation of biphenyl in *Rhodococcus* sp. RHA1 is distributed between several linear plasmids, referred to as RHA1, RHA2 and RHA3. Most of the genes for initiating the catabolism of biphenyl (*bphA1A2A3A4CB*) are located on RHA1, while the *bhpDEF* cluster is placed on pRHL2 [108].

Through a completely different pathway in *P. putida* Fl (Figs. 6d and 10), a multimeric enzyme referred to as toluene dioxygenase (encoded by tod *C1C2BA*) converts toluene and many other aromatics into (+)-*cis*-1(S),2(R)-dihydroxy-3-methylcyclohexa-3,5-diene [17, 100]. This reaction is driven by the protons and electrons originating from NADH that are passed through an electron transport system composed of a reductase (encoded by *todC1* and *todC2*). A NAD<sup>+</sup>-dependent-*cis*-toluene dihydrodiol dehydrogenase (encoded by *todD*) oxidises the dihydrodiol to form 3-methylcatechol, which is cleaved twice by 3-methylcatechol-2,3-dioxygenase



**Fig. 10** Conversion of toluene into 2-hydroxypenta-2,4-dienoate by *P. putida* F1. The electron originating from NADH, and  $H^+$  is passed through an electron chain system, respectively, consisting of reductase, ferrodoxin and an iron-sulphur protein *(ISP)* to catabolise toluene into toluene dihydrodiol dehydrogenase [100]

(encoded by todE) to 2-hydroxy-6-oxo-2,4-heptadienoate and then by 2-hydroxy-6-oxo-2,4-hepladlenoale hydrolase (encoded by todF) to 2-hydroxypenta-2,4-dienoate and acetate. All of the toluene-degrading genes are ordered as todFClC2BADE gene cluster.

### Nitroaromatic Compounds

Nitroaromatic compounds (NACs) are synthetic molecules broadly utilised in different industries as plastics, pharmaceuticals, precursors for dyes, explosives and pesticides [35]. Although there are many different types of nitroaromatics, 2,4,6-trinitrotoluene (TNT), dinitrotoluenes and nitrotoluenes are the most abundant environmental pollutants [8, 24]. Based on the gene capacity and type of the original nitroaromatic compound, microorganisms use oxidative and/or reductive degrading pathways to convert these NACs completely to CO<sub>2</sub> and H<sub>2</sub>O or partially to an organic compound [4]. While aerobic bacteria use both the catabolic systems, anaerobic bacteria are able to use only the reductive degrading mechanism to catabolise NACs [4].

The oxidative reactions are triggered through the reaction of a mono/di-oxigenase enzyme, releasing a nitrite and dihydroxy aromatic compounds. The substrate specificity and the intermediate and final products are unique based on the substrate, the type of oxygenases used in the reaction and the organisms involved in the degradation. The monooxygenase systems, for instance, are able to react with different substrates, including 2-nitrophenol (P. putida B2) [123], 4-nitrophenol (Moraxella sp.) [103] and 4-methyl-5-nitrocatechol (Pseudomonas sp. strain DNT) [34], 2-nirrotoluene (from Acidovorax sp. JS42) [56], nitrobenzene Comamonas sp. strain JS765 [73], 3-nitrobenzoate [52], 1,3-dinitrobenzene (Burkholderia cepacia R34) [46], 2chloronitrobenzene (Pseudomonas stutzeri strain ZWLR2-1) [63] and 2,4-dinicrotoluene [104]. The monooxygenases belonging to the two-component flavin-diffusible monooxygenase (TC-FDM) family in Moraxella sp., Pseudomonas sp. strain ENV2030, Rhodococcus sp. strain PN1, Rhodococcus opacus SAO101 and many other bacteria, oxidise 4-nitrophenol in expense of two NADPH and a molecular oxygen to hydroquinone and releases a nitrite molecule [79]. The members of this family can be divided into two homology groups: the phenol 2monooxygenase and phenol 4-monooxygenase groups. While the members of the first group (such as phenol monooxygenase (*PheA*), nitrophenol monooxygenase (*NphA1*) and 4hydroxyphenylacetate monooxygenase (HpaB), hydroxylate the ortho group of phenols, members of the second group, including 2,4,6-trichlorophenol monooxygenases (TcpA), 2,4,5trichlorophenol monooxygenase (TftD), PNP monooxygenase (NpcA) and 4-chlorophenol monooxygenase (*CphC*-I), hydroxylate their para position. Following release of nitrite in both bacteria, the products are directed into normal cell metabolism that leads to production of maleylacetic acid and further to  $\beta$ -ketoadipate. The *npd* gene cluster in *Arthrobacter* sp. JS443, responsible for catabolism of *p*-nitrophenol, consists of three genes, *npdB* (hydroxyquinol 1,2dioxygenase), *npdA1* (*p*-nitrophenol monooxygenase) and *npdA2* (*p*-nitrophenol hydroxylase) [79]. This cluster is 85 % similar to the *cph I* gene cluster found in *A. chlorophenolicus* A6 (Fig. 11) [79]. However, while *cph* gene cluster are regulated by products of two genes (cphR and cphS), these regulatory genes are combined into a single gene, called as *npdR* [74, 79].

The multicomponent NAC dioxygenase system, such as NBDO, NDO, DNTDO, 3-NTDO, 2NTDO and TNT dioxygenase, consists of a terminal iron-sulphur oxygenase, an iron-sulphur ferredoxin and a flavoprotein reductase to substitute a hydroxyl group to the ring by a nitrite [101]. NAC dioxygenases are usually non-specific enzymes able to react with several NACs, such as nitrotoluenes, dinitrobenzenes and nitrophenols and non-nitrogen aromatic hydrocarbons as well. These dioxygenases use the electrons transferred by two other components to the system to add two oxygen atoms into the ring to produce catechol intermediates [101]. The dioxygenase genes show highly similarity with the sequence and gene structure of the naphthalene, which belongs to Rieske non-heme iron dioxygenases [22]. The genes responsible for expression of these subunits are organised in a sequence of a reductase (*mntAa*), followed by one or two other ORF, the ferredoxin subunits, the large subunit of oxygenase (MntAc) and its small subunit (MntAd). These genes are under control of a regulator (mntR) located at upstream of the functional dioxygenase gene. Although the two ORFs between genes reductase and ferredoxin in many strains are inactive, in several cases, such as DNTDO identified in Burkholderia sp. strain DNT and NDO identified in Ralstonia sp., are responsible for expression of two subunits of salicylate-5-hydroxylase, which accept the electron released

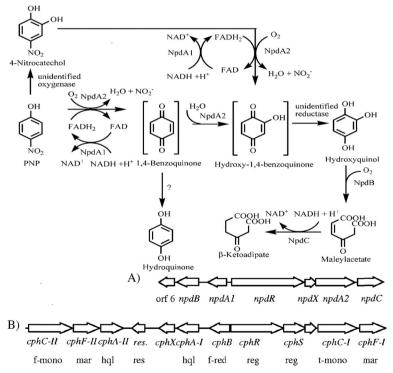
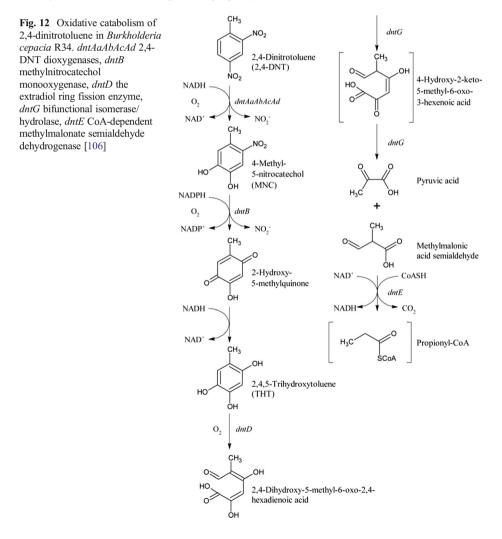


Fig. 11 The metabolic pathway of *p*-nitrophenol (*PNP*) and the *npd* gene cluster in *Arthrobacter* sp. JS443 (a) and *Arthrobacter chlorophenolicus* A6 (b); *f-mono* flavoprotein monooxygenase, *mar* maleylacetate reductase, *hql* hydroxyquinol 1,2-dioxygenase, *res* resolvase pseudogene, *f-red* flavin reductase, *reg* transcriptional regulator, *t-mono* TC-FDM [79]

from reductase and ferredoxin. As a prototype NAC, 2,4-dinitrotoluene (2,4-DNT) (Figs. 12 and 13) is initially oxidised by a tetramer dioxygenase (encoded by *dntAaAbAcAd*) to release  $NO_2^-$ , converting it to 4-methyl-5-nitrocatechol (MNC). While in *Acidovorax* sp. strain JS42 this operon is regulated by a transcriptional activator (*NtdR*) located immediately upstream of the operon [82], there is a gap between the *dntAaAbAcAd* and its upstream regulator (*dntR*) in *B. cepacia* R34 [106]. This intermediate is affected by the MNC monooxygenase (encoded by *dntB*) to produce 2-hydroxy-5-methylquinone (HMQ), which is further oxidised by HMQ reductase (encoded by *dntC*) to 2,4,5-trihydroxytoluene (THT). THT should undergo extradiol ring fission by THT oxygenase (encoded by *dntD*) to produce 2,4-dihydroxy-5-methyl-6-oxo-2,4-hexadiennoic acid (DMOH). This intermediate is converted by bifunctional DMOH isomerase (encoded by *dntG*) first to 4-hydroxy-2-keto-5-methyl-6-oxo-3-hexenoic acid and further to pyruvic acid and methlmalonic acid semialdehyde. The last product is bound by a CoA-dependent methylmalonate semialdehyde dehydrogenase (encoded by *dntE*) to CoA-SH in conjunction with NAD<sup>+</sup> to produce propionyl-CoA.



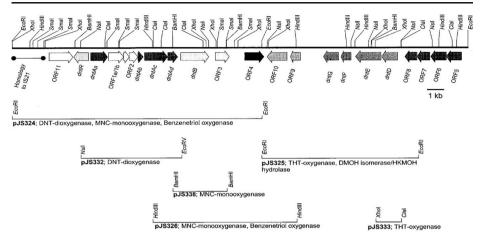


Fig. 13 Physical map of region encoding 2,4-DNT pathway genes in Burkholderia cepacia R34 [46]

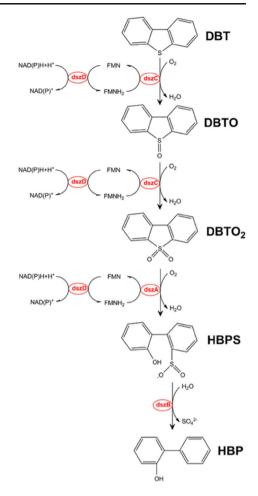
# Dibenzothiophene

Since a carbon group of dibenzothiophene (DBT) and its derivatives is substituted with sulphur, their degradation requires an initial set of reactions to remove this sulphur [67]. The genes involved in desulfurization of dibenzothiophene (DszA, DszB and DszC) are located on an operon. These genes remove the sulphur through a four step reaction referred to as the 4S desulfurization pathway [84, 99] (Fig. 14). The two first reactions are catalysed by the action of dibenzothiophene monooxdase (DszC) in which DBT is converted initially to dibenzothiophenesulfoxide and later to dibenzothiophene sulfone (DBTO<sub>2</sub>) [2, 68]. The product of DszA (DBTO<sub>2</sub>) monooxygenase) converts this intermediate to 2-(2-hydroxyphenyl)-benzene sulfonate (HBPS) [2, 68], which then is more metabolised by HBPS-desulfinase (DszB) releases sulphur from the compound and leaves 2-hydroxybiphenyl (HBP). The first three steps are O2-dependent oxidative reactions and require an electron and hydrogen transportation system, composed of FMNH<sub>2</sub> and NAD(P)H as a reductant [2, 68]. Oxidation of NAD(P)H to NAD(P) is catalysed by action of a flavin reductase, which is encoded by DszD [2, 68]. This later genes is located at a separate locus from dszABC [2, 68]. Following desulfurisation of these compounds, the product is a simple aromatic compound which is metabolised by the aromatic degradation system as mentioned before. Although there is not enough evidence for the regulation of these genes, it has been shown that expression of the operon is under the control of a repressor and is limited in the presence of different readily bioavailable sulphur sources, such as SO<sub>4</sub><sup>2-</sup>, casamino acids, methanesulfonic, taurine, cysteine and methionine [71].

# Anaerobic Degradation of Hydrocarbons

Several phylogenetically and physiologically distinct microorganisms degrade hydrocarbons through anaerobic metabolic pathways utilising the reduction of unusual electron acceptors, such as sulphate, thiosulfate, nitrate, nitrite, nitrous oxide, metal ions and carbonate, or using anoxygenic phototrophic reactions involving the donation of electrons and hydrogen for substrate catabolism activities [32]. These anaerobic bacteria degrade hydrocarbons via five

Fig. 14 The dibenzothiophene desulfurizing pathway in Gordonia sp.; dszA dibenzothiophene sulfone (DBTO<sub>2</sub>) monooxygenase, dszB 2-(2-hydroxyphenyl)benzene sulfonate desulfinase. dszC dibenzothiophene monooxdase, dszD NAD(P)H/ FMN oxidoreductase, DBT dibenzothiophene, DBTO dibenzothiophene-sulfoxide, DBTO<sub>2</sub> dibenzothiophene sulfone, HBPS 2-(2-hydroxyphenyl)benzene sulfonate, HBP 2hydroxybiphenyl (taken from http://2012.igem.org/ File:12SJTU desulpathway1.png)



different pathways: (a) addition of a fumarate to methylene or methyl groups of hydrocarbons [118], (b) oxygen-independent hydroxylation of 2nd or 3rd terminal C-atoms (to make secondary or tertiary alcohols) [107], (c) carboxylation of unsubstituted carbon atoms of aromatics [11], (d) hydration of the double and triple bond of alkenes and alkynes [66] and (e) reverse methanogenesis [111].

Many anaerobic bacteria, including denitrifying microorganisms, sulphate-reducing bacteria, methanogenic consortia and metal-reducing (Mn(IV), Fe(III)) bacteria, are able to activate hydrocarbons via terminal or sub-terminal addition of a carbonic group, such as fumarate, to a carbon atom of the hydrocarbon [96]. Conversion of toluene into (R)-benzylsuccinate is a common example in which a trimer benzylsuccinate synthase (BSS) enzyme (encoded by *bbsABC*) adds a fumarate to the substrate (Fig. 15) [118]. Then, after addition of a CoA to the product via the action of succinyl-CoA/benzylsuccinate CoA-transferase (encoded by *bbsEF*), the benzylsuccinyl-CoA undergoes an oxidation step, a hydration and another oxidation respectively by benzylsuccinyl-CoA dehydrogenase (encoded by *bbsCD*) before cleavage (by benzoylsuccinyl-CoA thiolase (encoded by *BbsB*) into a benzyl-CoA and

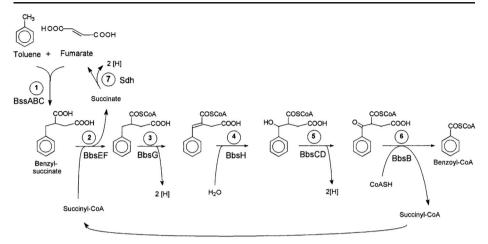
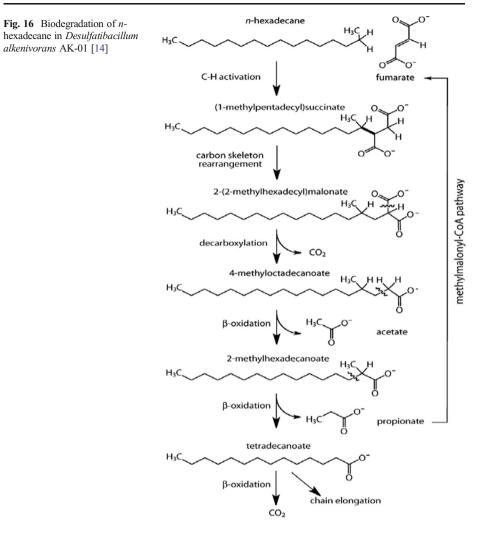


Fig. 15 The catabolic pathway for anaerobic toluene degradation. *1*, Benzylsuccinate synthase, encoded by BssABC; *2*, succinyl-CoA/benzylsuccinate CoA-transferase, encoded by BbsEF; *3*, benzylsuccinyl-CoA dehydrogenase, encoded by BbsG; *4*, phenylitaconyl-CoA hydratase, encoded by BbsH; *5*, 3-hydroxyacyl-CoA dehydrogenase, encoded by BbsCD; *6*, benzoylsuccinyl-CoA thiolase, encoded by BbsB; *7*, succinate dehydrogenase, encoded by *Sdh* [58]

a succinate. All of these genes are clustered as a *bbsABCDEFGHI* operon (*bbs* is abbreviation of beta-*oxidation* of benzylsuccinate) and is controlled by a regulatory factor (TdiSR) that activates this operon in the absence of  $O_2$  and the presence of toluene [39]. The naphthyl-2-methyl-succinate synthase in SRBs, encoded by *nmsABC*, specific to transfer a fumarate to naphthalene is similar to the corresponding enzyme involved in the biodegradation of toluene [94]. Furthermore, the smallest subunit of alkylsuccinate synthases, involved in addition of a fumarate to alkanes through anaerobic degradation, is highly similar to the corresponding subunit of benzylsuccinate synthetase, encoded by *bssC* [20].

*n*-alkane and cycloalkanes are metabolised with similar strategy by sulphate-reducing microorganisms in which a fumarate addition step activates these substrates to yield alkyl-succinates and cycloalkylsuccinate derivatives, respectively [72, 86]. The metabolism of cyclohexane by a nitrate-reducing microorganism to cyclohexylsuccinate [72] and of ethylcyclopentane by a sulphate-reducing organism to cyclopentylsuccinate [86] have been reported before. In the case of the anaerobic degradation of hexadecane by this mechanism (Fig. 16), alkylsuccinate synthetase (encoded by *assABC*) binds a fumarate to the substrate to produce 1-methylpentadecyclisuccinate. This intermediate is converted by methylmalonyl-CoA mutase (encoded y mcm) to 2-(2-methylhexadecyl)malonate, which is decarboxylated by the activity of a carboxyl transferase to produce 4-methyloctadecanoate. This last product is directed into the beta-oxidation pathway for further catabolism [14]. In the genome of *Desulfatibacillum alkenivorans* AK-01, these genes are located at two different loci (*assA1* and *assA2*), and there is no similarity between them [14] (Fig. 17).

Several denitrifying bacteria use an oxygen-independent hydroxylation process for the degradation of some aromatics in which a trimer ( $\alpha$ ,  $\beta$  and  $\gamma$ ) molybdenum-containing ethylbenzene dehydrogenase (EBDH; encoded by *ebdABC*) hydroxylates the terminal carbon of this molecule to produce (*S*)-1-phenylethanol [83, 107]. The (*S*)-1-phenylethanol is converted into acetophenone and then to benzoylacetate and finally to benzoylacetyl-CoA as a result of reactions involving NAD-dependent (*S*)-1-phenylethanol dehydrogenase,



acetophenone carboxylase (APC) and benzoylacetate-CoA ligase (BAL), respectively [83, 107]. The gene cluster for these three subunits of ethylbenzene dehydrogenase, along with the genes encoding (S)-1-phenylethanol dehydrogenase (*ped*) and a chaperone-like protein (encoded by *ebdD*) (necessary for transferring molybdenum into ethylbenzene dehydrogenase) are located on one operon, whereas the genes encoding APC (subunits A, B, C, D and E) and BAL are present on another operon [83, 107].

Addition of a carboxyl group to substrates is an alternative reaction utilised for the anaerobic catabolism of hydrocarbons by several sulphate and nitrate-reducing bacteria [95]. For instance, after conversion of propylene to acetone by *Xanthobacter autotrophicus* strain Py2, the acetone is carboxylated by acetone carboxylase at the expense of a CO<sub>2</sub> and an ATP to produce acetoacetate [11]. This multimeric ( $\alpha 2\beta 2\gamma 2$ ) enzyme is encoded by three genes: *acxA*, *acxB* and *acxC* which encode the  $\beta$ ,  $\alpha$  and  $\gamma$  subunits, respectively, which are clustered

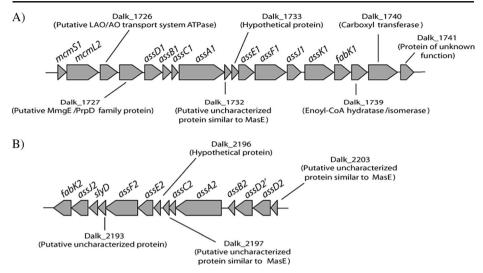


Fig. 17 Gene organisation of the enzymes involved in hexadecane degradation in anaerobic conditions (in the D. alkenivorans AK-01 genome). **a** Alkylsuccinate synthase locus 1, **b** locus 2; mcmS1 and mcmS2 small and large subunits of methylmalonyl-CoA mutase, assD1 alkylsuccinate synthase (I) glycyl radical activating enzyme, assB1 alkylsuccinate synthase beta subunit, assC1 alkylsuccinate synthase gamma subunit, assD1 alkylsuccinate synthase alpha subunit, assE1 a chaperone protein, assF1 an uncharacterised protein, assJ1 encyl-CoA hydratase/isomerase, assK1 AMP-dependent synthetase and ligase, fabK1 encyl-(acyl-carrier protein) reductase II, fabK2 putative encyl-(acyl-carrier protein) reductase II, assJ2 encyl-CoA hydratase/isomerase, slyD putative peptidyl-prolyl cis-trans isomerase (fkbp-type), assF2 putative uncharacterised protein, assE2 chaperone protein, assC2 gamma subunit of alkylsuccinate synthase (II), A2 alpha subunit of alkylsuccinate synthase (II), B2 beta subunit of alkylsuccinate synthase (II), assD2 and assD2 alkylsuccinate synthase (II) glycyl radicalactivating enzyme [14]

in an operon as acxABC and its expression is regulated by a gene (AcxR) upstream of the acxABC cluster [11].

$$CH_3$$
-COCH<sub>3</sub> + CO<sub>2</sub> + ATP  $\rightarrow$  CH<sub>3</sub>COCH<sub>2</sub>COO + H<sup>+</sup> + AMP + 2Pi

Some microorganisms are able to anaerobically degrade alkenes and alkynes through the addition of H<sub>2</sub>O to the unsaturated bond, producing the corresponding alcohols [66]. Using a monomeric thermostable acetylene hydratase (encoded by *AH* gene), for instance, *Pelobacter acetylenicus* adds a H<sub>2</sub>O molecule to acetylene converting it to an acetaldehyde [66, 110]. In the degradation of  $\beta$ -myrcene (7-methyl-3-methylen-1,6-octadien) by *Castellaniella defragrans*, linalool dehydratase/isomerase (LDI; encoded by *ldi* gene) acts as a dualfunction enzyme that desaturates linalool to myrcene before isomerisation to geraniol [12]. The molecule is then oxidised by two dehydrogenases, referred to geraniol dehydrogenase (GeDH) and geranial dehydrogenase (GaDH) (encoded by *geoA* and *geoB*, respectively) into geranial and geranic acid [12].

Several methanotrophic microorganisms use methyl-coenzyme M reductase (Mcr) for the initial activation of methane, which then binds to coenzyme B (CoBSH) via methyl-coenzyme M (CoMSH) to produce a complex of CoM-S-S-CoB-heterodisulfide and methane [111]. This enzyme consists of two of each of the  $\alpha$ ,  $\beta$  and  $\gamma$  subunits, one nickel atom and a tetrapyrrole cofactor (normally F<sub>430</sub> factor or 17(2)-methylthio-F430) [91, 111]. The encoding genes for these subunits, *mcrBGA*, and two other additional genes with unidentified roles (*mcrC* and *mcrD*) are located on one operon [91, 111]. The *mcrA* gene is used as a marker to track methanogens and anaerobic methanotrophic microorganisms [91, 111].

### Anaerobic Degradation of NACs

Several bacteria, such as sulphate-reducing bacteria and Clostridia, are able to degrade nitroaromatic compounds in anaerobic conditions. The anaerobic metabolism of these compounds is performed through the reduction of the nitro group. However, it has to be mentioned that this mechanism is not exclusive to anaerobic bacteria and many other bacteria, such as *Enterobacter* sp., *Nocardiodes* sp. and *Rhodococcus* sp., are able to metabolise NACs using similar reductive pathways. The reductive pathways can be proceed by two different processes depending on the microbial gene capacity. *Enterobacter cloacae* PB2 expresses a monomeric flavoenzyme, referred to as PETN reductase (PETNr), which enables this microorganism to reduce triply nitrated aromatic compounds, such as picric acid and TNT. Furthermore, presence of *PnrA* gene, encoding for an NADPH-dependent nitroreductase, enable several microorganisms to transform a variety of NACs, such as 3,5-dinitroaniline, 3- and 4-nitrobenzoate, 3-nitrotoluene, 2,4-DNT, TNT and 3,5-dinitrobenzamide.

Hydride transferases are a second reducing system in which the aromatic ring of some NACs, such as TNT and picric acid loss their nitrite. The initial reduction of TNT by PETNr, for instance, leads to production of hydride- and dihydride-Meisenheimer TNT complexes (H-TNT and 2H-TNT), which is further reduced through an unknown mechanism to release nitrite. *Nocardiodes simplex* FJ2-1A and *Rhodococcus erythropolis* are two bacteria with the ability to degrade picric acid using this mechanism. In these examples, a hydride is initially added to the aromatic ring by the activity of a  $F_{420}$ -dependent hydride transferase (*NpdI*) and  $F_{420}$  reductase (*NpdG*) to produce a hydride Meisenheimer complex. While in *N. simplex* the second hydride is added by the same enzyme, this step on *R. erythropolis* is performed by the product of *NpdC/NpdG*. The dihydride-Meisenheimer complex is later undergone a tautomerisation performed by the product of *NpdH* to release nitrite and various products.

### Genetics of Microbial Adaptation to High Hydrocarbon Concentrations

A limitation of mineral nutrients is a common problem in oil contaminated marine environments due to an imbalanced C/P/N ratio [70]. Microorganisms can adapt themselves to these limitations using either their own gene products or through the creation of a cooperative relationship with other microbes to decrease this stress [69]. Due to such adaptation processes, the community of hydrocarbon-degrading microorganisms may show a sudden increase in cell mass after an initial temporary drop in the total number of microorganisms [70]. In addition to the genes responsible for degradation of hydrocarbons, many bacteria adapt their physiology to the shortage of mineral nutrition through increased expression of their existing ion transporters [92, 121]. *A. borkumensis* strain SK2, for instance, adapts itself to various mineral deficiencies through the induction of the genes responsible for different transport proteins such as *narKGHJI* cluster and *nrtCB-nasDTS* cluster (for reduction and uptake of nitrogen), *amt* (for uptake of ammonium), *phoBR* and *phoUpstBACS* gene cluster (for uptake of phosphate), *znuAB* (for uptake of zinc), *modABC* (for uptake of molybdite), *mgtE* (for uptake of magnesium) and CorA-like MIT(for uptake of cobalt) [92].

In addition, one of the critical factors for biodegradation of hydrocarbons is the ability of the degrading microorganisms to be resistant to high concentrations of hydrocarbons, especially when the cells are suddenly exposed to large amounts of the compounds [26, 52]. A common resistance strategy to toxic solvents is to intensify the cell membrane density using isomerisation of *cis*-unsaturated fatty acids to their *trans* forms [69]. Furthermore, penetration

of the solvents into cells induces expression of several chaperons to refold the proteins denatured by the reagents [52]. Some bacteria tolerate high concentrations of solvents through expression of efflux pumps on their cell membrane, enabling them to export the toxic solvents out of the cells [26]. *P. putida* S12, for instance, is a highly resistant strain to organic compounds due to the expression of a solvent resistance pump (encoded by *srpABC*), which discharges several types of solvents, hydrocarbons included [105].

Furthermore, many bacteria use chemotactic strategies to improve their resistance to different toxic compounds, hydrocarbons included [23, 53]. The integral cell membrane associated proteins encoded by *alkN* [23] and *nahY* in *P. putida* [53], for instance, interact respectively with alkanes and naphthalene. These proteins act as a methyl-accepting chemotaxis factor that triggers a cell signalling pathway, regulating the flagella motor and resulting in cell attraction towards naphthalene [23, 53]. Interestingly, both chemotactic activities and solvent resistance abilities are gene-dose dependent, and the amounts of the corresponding genes and the level of gene activity in cells determine the level of cell susceptibility to a solvent [54]. Based on a study, performed by *Lacal* [54] on *P. putida* DOT-T1E, the presence of two alleles of chemoreceptor *mcpT* gene genes is enough to enable the cell to response strongly to different aromatic hydrocarbons. They showed that the level of *mcpT* gene expression in this strain was directly under control of the substrates, and there was an inverse relationship between the amounts of toluene and the level of gene methylation. The increases in the level of *McpT* methylation induce the activity of flagella motor using the autophosphorylation of *CheA* [54].

The behaviour of bacteria in the case of exposure to a carbon source, which is potentially toxic for the cells, depends to the reaction of signal transduction proteins. Although the signalling pathway leads to triggering a degradation shunt in the cells, some bacteria prefer to migrate away from these compounds. This decision making in *P. putida* DOT-T1E in the case of exposure to toluene depends on the activity of *TodS/TodT* two-component system and *TtgV* in the cells. Since affinity of the sensor kinase *TodS* to toluene is double in comparison with *TodV*, this signalling protein is activated in lower concentrations of the substrate, inducing the TOD pathway [13]. However, binding of toluene to *todV* in higher concentrations of toluene upregulates expression of *TtgGHI* efflux pump to enhance the cell's resistance to the toxicity of this compound [87].

# Conclusion

Although there are some slight differentiations between many hydrocarbon-degrading genes in different phylogenetic groups of microbial species, the homology of DNA sequences and organisation of these genes as well as intensive overlapping of the activity of their products indicates that horizontal gene transfers have occurred between these groups. Slight differences in the corresponding hydrocarbon-degrading genes can change the ability of microbes to degrade hydrocarbons, including altering the time of expression (due to presence of inducible or constitutive promoters), the level of expression (due to the activity of the promoter) and the enzymatic activity of the product (due to sequence of amino acids and the protein configuration).

This information assists environmental microbiologists and biotechnologists to choose suitable/ stronger hydrocarbon-degrading genes and the hydrocarbon resistance genes in order to create efficient genetic-engineered microorganisms (GEM). Furthermore, such information can be used for the selection of appropriate (non-genetically engineered) microbial consortia with higher hydrocarbon-degrading ability for use as an inoculum in the bioremediation of contaminated sites. Acknowledgments The authors would like to appreciate Australian Government, University of South Australia, University of Newcastle and Cooperative Research Centre for Contamination Assessment and Remediation of the Environment (CRC-CARE) for funding towards this research.

**Conflicts of Interest** There is no conflict of interest.

### References

- Abbasian, F., Lockington, R., Mallavarapu, M., & Naidu, R. (2015). A comprehensive review of aliphatic hydrocarbon biodegradation by bacteria. *Appl Biochem Biotech*, 176, 670–699.
- Abin-Fuentes, A., Mohamed, M. E.-S., Wang, D. I., & Prather, K. L. (2013). Exploring the mechanism of biocatalyst inhibition in microbial desulfurization. *Applied and Environmental Microbiology*, 79, 7807–7817.
- Agrawal, P., Latha, S., & Mahadevan, A. (1997). Utilization of phenylalanine and phenylacetic acid by Pseudomonas solanacearum. Appl Biochem Biotech, 61, 379–391.
- Arora, P. K., Srivastava, A., & Singh, V. P. (2014). Bacterial degradation of nitrophenols and their derivatives. *Journal of Hazardous Materials*, 266, 42–59.
- Asperger, O., & Kleher, H. (1991). Metabolism of alkanes by ac in etobacter. *The Biology of Acinetobacter: Taxonomy, Clinical Importance, Molecular Biology, Physiology, Industrial Relevance*, 57, 323.
- 6. B, V. d. (2005). The FadL family: unusual transporters for unusual substrates. Curr Opin Struc Biol, 15, 401–407.
- Beškoski, V. P., Gojgić-Cvijović, G., Milić, J., Ilić, M., Miletić, S., Šolević, T., & Vrvić, M. M. (2011). Bioremediation of a soil contaminated by mazut (heavy residual fuel oil)—a field experiment. *Chemosphere*, 83, 34–40.
- Biń, A. K., Machniewski, P., Sakowicz, R., Ostrowska, J., & Zieliński, J. (2001). Degradation of nitroaromatics (MNT, DNT and TNT) by AOPs. *Ozone Science and Engineering*, 23, 343–349.
- 9. Boronin, A., & Kosheleva, I. (2010). Handbook of hydrocarbon and lipid microbiology (pp. 1155–1163). Springer.
- Boronin, A. M., & Kosheleva, I. A. (2014). Current environmental issues and challenges (pp. 159–168). Springer.
- Broberg, C. A., & Clark, D. D. (2010). Shotgun proteomics of *Xanthobacter autotrophicus* Py2 reveals proteins specific to growth on propylene. *Archives of Microbiology*, 192, 945–957.
- Brodkorb, D., Gottschall, M., Marmulla, R., Lüddeke, F., & Harder, J. (2010). Linalool dehydrataseisomerase, a bifunctional enzyme in the anaerobic degradation of monoterpenes. *The Journal of Biological Chemistry*, 285, 30436–30442.
- Busch, A., Lacal, J., Silva-Jímenez, H., Krell, T., & Ramos, J. L. (2010). Catabolite repression of the TodS/ TodT two-component system and effector-dependent transphosphorylation of TodT as the basis for toluene dioxygenase catabolic pathway control. *J Biotech*, *192*, 4246–4250.
- Callaghan, A., Morris, B., Pereira, I., McInerney, M., Austin, R. N., Groves, J. T., Kukor, J., Suflita, J., Young, L., & Zylstra, G. (2012). The genome sequence of *Desulfatibacillum alkenivorans* AK-01: a blueprint for anaerobic alkane oxidation. *Environmental Microbiology*, 14, 101–113.
- Cappelletti, M., Fedi, S., Frascari, D., Ohtake, H., Turner, R., & Zannoni, D. (2011). Analyses of both the alkB gene transcriptional start site and alkB promoter-inducing properties of *Rhodococcus* sp. strain BCP1 grown on *n*-alkanes. *Applied and Environmental Microbiology*, 77, 1619–1627.
- Cheng, Q., Thomas, S., & Rouviere, P. (2002). Biological conversion of cyclic alkanes and cyclic alcohols into dicarboxylic acids: biochemical and molecular basis. *Appl Microbiol Biotech*, 58, 704–711.
- Choi, E. N., Cho, M. C., Kim, Y., Kim, C.-K., & Lee, K. (2003). Expansion of growth substrate range in *Pseudomonas putida* F1 by mutations in both cymR and todS, which recruit a ring-fission hydrolase CmtE and induce the tod catabolic operon, respectively. *Microbiology*, 149, 795–805.
- Dinamarca, M. A., Aranda-Olmedo, I., Puyet, A., & Rojo, F. (2003). Expression of the *Pseudomonas putida* OCT plasmid alkane degradation pathway is modulated by two different global control signals: evidence from continuous cultures. *J Biotech*, *185*, 4772–4778.
- dos Santos, V. M., Yakimov, M. M., Timmis, K. N. and Golyshin, P. N. (2008) Genomic insights into oil biodegradation in marine systems. Microbial Biodegradation: Genomic and Molecular Biology.
- Ehrenreich, P., Behrends, A., Harder, J., & Widdel, F. (2000). Anaerobic oxidation of alkanes by newly isolated denitrifying bacteria. *Archives of Microbiology*, 173, 58–64.
- Feng, Y., Khoo, H. E., & Poh, C. L. (1999). Purification and characterization of gentisate 1, 2-dioxygenases from pseudomonas alcaligenes NCIB 9867 and *Pseudomonas putida* NCIB 9869. *Applied and Environmental Microbiology*, 65, 946–950.

- Ferraro, D. J., Gakhar, L., & Ramaswamy, S. (2005). Rieske business: structure–function of rieske nonheme oxygenases. *Biochem Biophys Res Comm*, 338, 175–190.
- Fondi, M., Rizzi, E., Emiliani, G., Orlandini, V., Berna, L., Papaleo, M. C., Perrin, E., Maida, I., Corti, G., & De Bellis, G. (2013). The genome sequence of the hydrocarbon-degrading acinetobacter venetianus VE-C3. *Research in Microbiology*, 164, 439–449.
- Funk, S., Crawford, D., Crawford, R., Mead, G., & Davis-Hoover, W. (1995). Full-scale anaerobic bioremediation of trinitrotoluene (TNT) contaminated soil. *Appl Biochem Biotech*, 51, 625–633.
- Furukawa, K., Suenaga, H., & Goto, M. (2004). Biphenyl dioxygenases: functional versatilities and directed evolution. *J Biotech*, 186, 5189–5196.
- 26. Gallegos, M.-T., Molina-Henares, A. J., Zhang, X., Terán, W., Bernal, P., Alguel, Y., Guazzaroni, M.-E., Krell, T., Segura, A. and Ramos, J.-L. (2008) Genomic insights into solvent tolerance and pumps that extrude toxic chemicals. In Microbial biodegradation: genomics and molecular biology.
- 27. Gargouri, B., Karray, F., Mhiri, N., Aloui, F. and Sayadi, S. (2013) Bioremediation of petroleum hydrocarbons-contaminated soil by bacterial consortium isolated from an industrial wastewater treatment plant. Journal of Chemical Technology and Biotechnology.
- Gibson, D. T., & Parales, R. E. (2000). Aromatic hydrocarbon dioxygenases in environmental biotechnology. *Current Opinion in Biotechnology*, 11, 236–243.
- Girhard, M., Klaus, T., Khatri, Y., Bernhardt, R., & Urlacher, V. B. (2010). Characterization of the versatile monooxygenase CYP109B1 from *Bacillus subtilis. Appl Microbiol Biotech*, 87, 595–607.
- Goetz, F. E., & Harmuth, L. J. (1992). Gentisate pathway in Salmonella typhimurium: metabolism of mhydroxybenzoate and gentisate. FEMS Microbiology Letters, 97, 45–49.
- Goyal, A., & Zylstra, G. (1997). Genetics of naphthalene and phenanthrene degradation by Comamonas testosteroni. *Journal of Industrial Microbiology and Biotechnology*, 19, 401–407.
- Grossi, V., Cravo-Laureau, C., Guyoneaud, R., Ranchou-Peyruse, A., & Hirschler-Réa, A. (2008). Metabolism of *n*-alkanes and *n*-alkenes by anaerobic bacteria: a summary. *Organic Geochemistry*, 39, 1197–1203.
- Habe, H., & Omori, T. (2003). Genetics of polycyclic aromatic hydrocarbon metabolism in diverse aerobic bacteria. *Bioscience, Biotechnology, and Biochemistry*, 67, 225–243.
- Haigler, B. E., Nishino, S. F., & Spain, J. C. (1994). Biodegradation of 4-methyl-5-nitrocatechol by *Pseudomonas* sp. strain DNT. J Biotech, 176, 3433–3437.
- Halecky, M., Karlova, P., Paca, J., Stiborova, M., Kozliak, E. I., Bajpai, R., & Sedlacek, I. (2013). Biodegradation of a mixture of mononitrophenols in a packed-bed aerobic reactor. *Journal of Environmental Science and Health, Part A*, 48, 989–999.
- Harayama, S., & Timmis, K. N. (2012). Catabolism of aromatic hydrocarbons by pseudomonas, genetics of bacterial diversity. Elsevier B.V.: Academic Press Limited, pp. 151–174.
- Hearn, E. M., Patel, D. R., Lepore, B. W., Indic, M., & B., A. (2009). Transmembrane passage of hydrophobic compounds through a protein channel wall. *Nature*, 458, 367–370.
- Herman, P. L., Behrens, M., Chakraborty, S., Chrastil, B. M., Barycki, J., & Weeks, D. P. (2005). A threecomponent dicamba O-demethylase from pseudomonas maltophilia, strain DI-6 gene isolation, characterization, and heterologous expression. *Journal of Biological Chemistry*, 280, 24759–24767.
- Hermuth, K., Leuthner, B., & Heider, J. (2002). Operon structure and expression of the genes for benzylsuccinate synthase in *Thauera aromatica* strain K172. Archives of Microbiology, 177, 132–138.
- Hernández-Arranz, S., Moreno, R., & Rojo, F. (2013). The translational repressor Crc controls the *Pseudomonas putida* benzoate and alkane catabolic pathways using a multi-tier regulation strategy. *Environmental Microbiology*, 15, 227–241.
- Iwai, S., Chai, B., Sul, W. J., Cole, J. R., Hashsham, S. A., & Tiedje, J. M. (2010). Gene-targeted-metagenomics reveals extensive diversity of aromatic dioxygenase genes in the environment. *The ISME Journal*, 4, 279–285.
- Iwai, S., Johnson, T. A., Chai, B., Hashsham, S. A., & Tiedje, J. M. (2011). Comparison of the specificities and efficacies of primers for aromatic dioxygenase gene analysis of environmental samples. *Applied and Environmental Microbiology*, 77, 3551–3557.
- Izmalkova, T. Y., Sazonova, O. I., Nagornih, M. O., Sokolov, S. L., Kosheleva, I. A., & Boronin, A. M. (2013). The organization of naphthalene degradation genes in *Pseudomonas putida* strain AK5. *Research in Microbiology*, 164, 244–253.
- Izmalkova, T. Y., Sazonova, O. I., Nagornih, M. O., Sokolov, S. L., Kosheleva, I. A., & Boronin, A. M. (2013). The organization of naphthalene degradation genes in *Pseudomonas putida* strain AK5. *Research in Microbiology*, 164, 244–253.
- Ji, Y., Mao, G., Wang, Y., & Bartlam, M. (2013). Structural insights into diversity and n-alkane biodegradation mechanisms of alkane hydroxylases. Frontiers in Microbiology, 4.
- Johnson, G. R., Jain, R. K., & C., S. J. (2002). Origins of the 2,4-dinitrotoluene pathway. J Biotech, 184, 4219–4232.

- Jouanneau, Y., Martin, F., Krivobok, S., & Willison, J. C. (2011). Ring-hydroxylating dioxygenases involved in PAH biodegradation: structure, function and biodiversity. In *Microbial bioremediation of non metals: current research* (pp. 149–175). Norflok, UK: Caister Academic Press.
- Jutkina, J., Hansen, L. H., Li, L., Heinaru, E., Vedler, E., Jõesaar, M., & Heinaru, A. (2013). Complete nucleotide sequence of the self-transmissible TOL plasmid pD2RT provides new insight into arrangement of toluene catabolic plasmids. *Plasmid*, 70, 393–405.
- Kim, E., & Zylstra, G. (1999). Functional analysis of genes involved in biphenyl, naphthalene, phenanthrene, and m-xylene degradation by Sphingomonas yanoikuyae B1. J Indus Microbiol Biotech, 23, 294–302.
- Kim, S.-J., Kweon, O., Jones, R. C., Edmondson, R. D., & Cerniglia, C. E. (2008). Genomic analysis of polycyclic aromatic hydrocarbon degradation in *Mycobacterium vanbaalenii* PYR-1. *Biodegradation*, 19, 859–881.
- Kostichka, K., Thomas, S. M., Gibson, K. J., Nagarajan, V., & Cheng, Q. (2001). Cloning and characterization of a gene cluster for cyclododecanone oxidation in *Rhodococcus ruber* SC1. *The Journal of Bacteriology*, 183, 6478–6486.
- 52. Krajewski, S. S., Joswig, M., Nagel, M. and Narberhaus, F. (2014) A tricistronic heat shock operon is important for stress tolerance of *Pseudomonas putida* and conserved in many environmental bacteria. Environ Microbiol.
- Krell, T., Lacal, J., Guazzaroni, M. E., Busch, A., Silva-Jiménez, H., Fillet, S., Reyes-Darías, J. A., Muñoz-Martínez, F., Rico-Jiménez, M., & García-Fontana, C. (2012). Responses of *Pseudomonas putida* to toxic aromatic carbon sources. *Journal of Biotechnology*, 160, 25–32.
- Lacal, J., Muñoz-Martínez, F., Reyes-Darías, J. A., Duque, E., Matilla, M., Segura, A., Calvo, J. J. O., Jímenez-Sánchez, C., Krell, T., & Ramos, J. L. (2011). Bacterial chemotaxis towards aromatic hydrocarbons in *Pseudomonas. Environmental Microbiology*, 13, 1733–1744.
- Lai, Q., Li, W., & Shao, Z. (2012). Complete genome sequence of *Alcanivorax dieselolei* type strain B5. J Biotech, 194, 6674–6674.
- Lee, K.-S., Parales, J. V., Friemann, R., & Parales, R. E. (2005). Active site residues controlling substrate specificity in 2-nitrotoluene dioxygenase from *Acidovorax* sp. strain JS42. *J Indus Microbiol Biotech*, 32, 465–473.
- Lehrbach, P., Ward, J., Meulien, P., & Broda, P. (1982). Physical mapping of TOL plasmids pWWO and pND2 and various R plasmid-TOL derivatives from *Pseudomonas* spp. *J Biotech*, *152*, 1280–1283.
- 58. Leuthner, B., & Heider, J. (2000). Anaerobic toluene catabolism of Thauera aromatica: the bbs operon codes for enzymes of  $\beta$  oxidation of the intermediate benzylsuccinate. *J Biotech*, *182*, 272–277.
- Li, L., Liu, X., Yang, W., Xu, F., Wang, W., Feng, L., Bartlam, M., Wang, L., & Rao, Z. (2008). Crystal structure of long-chain alkane monooxygenase (LadA) in complex with coenzyme FMN: unveiling the long-chain alkane hydroxylase. *J Mol Bio*, 376, 453–465.
- Li, L., Liu, X., Yang, W., Xu, F., Wang, W., Wang, L., Feng, L., Bartlam, M., & Rao, Z. (2008). Crystal structure of long-chain alkane monooxygenase (LadA) in complex with coenzyme FMN: unveiling the long-chain alkane hydroxylase. *J Mol Bio*, 376, 453–465.
- Li, S., Zhao, H., Li, Y., Niu, S., & Cai, B. (2012). Complete genome sequence of the naphthalene-degrading Pseudomonas putida strain ND6. J Biotech, 194, 5154–5155.
- Li, W., Shi, J., Wang, X., Han, Y., Tong, W., Ma, L., Liu, B., & Cai, B. (2004). Complete nucleotide sequence and organization of the naphthalene catabolic plasmid pND6-1 from pseudomonas sp. strain ND6. *Gene*, 336, 231–240.
- 63. Liu, H., Wang, S.-J., Zhang, J.-J., Dai, H., Tang, H., & Zhou, N.-Y. (2011). Patchwork assembly of nag-like nitroarene dioxygenase genes and the 3-chlorocatechol degradation cluster for evolution of the 2chloronitrobenzene catabolism pathway in *Pseudomonas stutzeri* ZWLR2-1. *Applied and Environmental Microbiology*, 77, 4547–4552.
- Liu, T.-T., Xu, Y., Liu, H., Luo, S., Yin, Y.-J., Liu, S.-J., & Zhou, N.-Y. (2011). Functional characterization of a gene cluster involved in gentisate catabolism in *Rhodococcus* sp. strain NCIMB 12038. *Appl Microbiol Biotech*, 90, 671–678.
- Liu, T.-T., & Zhou, N.-Y. (2012). Novel L-cysteine-dependent maleylpyruvate isomerase in the gentisate pathway of *Paenibacillus* sp. strain NyZ101. *J Biotech*, 194, 3987–3994.
- 66. Liu, Y.-F., Liao, R.-Z., Ding, W.-J., Yu, J.-G., & Liu, R.-Z. (2011). Theoretical investigation of the first-shell mechanism of acetylene hydration catalyzed by a biomimetic tungsten complex. *Journal of Biological inorganic Chemistry: JBIC: A Publication of the Society of Biological Inorganic Chemistry*, 16, 745–752.
- Lu, L., Cheng, S., Gao, J., Gao, G., & He, M.-Y. (2007). Deep oxidative desulfurization of fuels catalyzed by ionic liquid in the presence of H<sub>2</sub>O<sub>2</sub>. *Energy & Fuels*, 21, 383–384.
- Martinez, I., Santos, V. E., Alcon, A., & Garcia-Ochoa, F. (2015). Enhancement of the biodesulfurization capacity of *Pseudomonas putida* CECT5279 by co-substrate addition. *Process Biochemistry*, 50, 119–124.
- McCammick, E., Gomase, V., McGenity, T., Timson, D., & Hallsworth, J. (2010). Handbook of hydrocarbon and lipid microbiology (pp. 1451–1466). Springer.

- Megharaj, M., Ramakrishnan, B., Venkateswarlu, K., Sethunathan, N., & Naidu, R. (2011). Bioremediation approaches for organic pollutants: a critical perspective. *Environment International*, 37, 1362–1375.
- Gasemali, M., & Andrew, B. S. (2008). Biocatalytic desulfurization (BDS) of petrodiesel fuels. *Microbiol*, 154, 2169–2183.
- Musat, F., Wilkes, H., Behrends, A., Woebken, D., & Widdel, F. (2010). Microbial nitrate-dependent cyclohexane degradation coupled with anaerobic ammonium oxidation. *The ISME Journal*, 4, 1290–1301.
- Nishino, S. F., & Spain, J. C. (1995). Oxidative pathway for the biodegradation of nitrobenzene by Comamonas sp. strain JS765. Applied and Environmental Microbiology, 61, 2308–2313.
- Nordin, K., Unell, M., & Jansson, J. K. (2005). Novel 4-chlorophenol degradation gene cluster and degradation route via hydroxyquinol in *Arthrobacter chlorophenolicus* A6. *Applied and Environmental Microbiology*, 71, 6538–6544.
- Obayori, S., & Salam, L. B. (2010). Degradation of polycyclic aromatic hydrocarbons: role of plasmids. Scientific Research and Essays, 5, 4093–4106.
- Oberoi, A. S., Philip, L., & Bhallamudi, S. M. (2015). Biodegradation of various aromatic compounds by enriched bacterial cultures: part A—monocyclic and polycyclic aromatic hydrocarbons. *Appl Biochem Biotech*, 176, 1870–1888.
- Ohtsubo, Y., Nagata, Y., Kimbara, K., Takagi, M., & Ohta, A. (2000). Expression of the bph genes involved in biphenyl/PCB degradation in *Pseudomonas* sp. KKS102 induced by the biphenyl degradation intermediate, 2-hydroxy-6-oxo-6-phenylhexa-2, 4-dienoic acid. *Gene*, 256, 223–228.
- Oliveira, F. J. and De França, F. P. (2005) Increase in removal of polycyclic aromatic hydrocarbons during bioremediation of crude oil-contaminated sandy soil. Twenty-Sixth Symposium on Biotechnology for Fuels and Chemicals, pp. 593–603. Springer.
- Perry, L. L., & Zylstra, G. J. (2007). Cloning of a gene cluster involved in the catabolism of *p*-nitrophenol by *Arthrobacter* sp. strain JS443 and characterization of the *p*-nitrophenol monooxygenase. *J Biotech*, 189, 7563–7572.
- PhaLe, P. S., MahajaN, M. C., & VaidyaNathaN, C. S. (2013). Biodegradation of polycyclic aromatic hydrocarbons. *Journal of the Indian Institute of Science*, 77, 141.
- Pickrell, W. O., Rees, M. I., & Chung, S.-K. (2012). 1 next generation sequencing methodologies—an overview. Advances in Protein Chemistry and StructuralBiology, 89, 1.
- Rabinovitch-Deere, C. A., & Parales, R. E. (2012). Three types of taxis used in the response of *Acidovorax* sp. strain JS42 to 2-nitrotoluene. *Applied and Environmental Microbiology*, 78, 2306–2315.
- Rabus, R., Kube, M., Heider, J., Beck, A., Heitmann, K., Widdel, F., & Reinhardt, R. (2005). The genome sequence of an anaerobic aromatic-degrading denitrifying bacterium, strain EbN1. Archives of Microbiology, 183, 27–36.
- Raheb, J. (2011). The study of biodesulfurization activity in recombinant *E. coli* strain by cloning the *dsz* genes involve in 4S pathway. *Journal of Sciences, Islamic Republic of Iran*, 22, 213–219.
- Ratledge, C. (1984). Microbial conversions of alkanes and fatty acids. *Journal of the American Oil Chemists' Society*, 61, 447–453.
- Rios-Hernandez, L. A., Gieg, L. M., & Suflita, J. M. (2003). Biodegradation of an alicyclic hydrocarbon by a sulfate-reducing enrichment from a gas condensate-contaminated aquifer. *Applied and Environmental Microbiology*, 69, 434–443.
- Rodríguez-Herva, J. J., García, V., Hurtado, A., Segura, A., & Ramos, J. L. (2007). The ttgGHI solvent efflux pump operon of *Pseudomonas putida* DOT-T1E is located on a large self-transmissible plasmid. *Environmental Microbiology*, 9, 1550–1561.
- 88. Rojo, F. (2009). Degradation of alkanes by bacteria. Environmental Microbiology, 11, 2477-2490.
- 89. Rojo, F. (2010), in Handbook of Hydrocarbon and Lipid Microbiology, Springer, pp. 1141-1154.
- Sayavedra-Soto, L. A., Hamamura, N., Liu, C. W., Kimbrel, J. A., Chang, J. H., & Arp, D. J. (2011). The membrane-associated monooxygenase in the butane-oxidizing Gram-positive bacterium *Nocardioides* sp. strain CF8 is a novel member of the AMO/PMO family. *Environmental Microbiology Reports*, 3, 390–396.
- Scheller, S., Goenrich, M., Boecher, R., Thauer, R. K., & Jaun, B. (2010). The key nickel enzyme of methanogenesis catalyses the anaerobic oxidation of methane. *Nature*, 465, 606–608.
- Schneiker, S., dos Santos, V. A. M., Bartels, D., Bekel, T., Brecht, M., Buhrmester, J., Chernikova, T. N., Denaro, R., Ferrer, M., & Gertler, C. (2006). Genome sequence of the ubiquitous hydrocarbon-degrading marine bacterium *Alcanivorax borkumensis*. *Nature Biotechnology*, *24*, 997–1004.
- Schuler, L., Jouanneau, Y., Chadhain, S. M. N., Meyer, C., Pouli, M., Zylstra, G. J., Hols, P., & Agathos, S. N. (2009). Characterization of a ring-hydroxylating dioxygenase from phenanthrenedegrading *Sphingomonas* sp. strain LH128 able to oxidize benz[a]anthracene. *Appl Microbiol Biotech*, 83, 465–475.
- 94. Selesi, D., Jehmlich, N., von Bergen, M., Schmidt, F., Rattei, T., Tischler, P., Lueders, T., & Meckenstock, R. U. (2010). Combined genomic and proteomic approaches identify gene clusters

involved in anaerobic 2-methylnaphthalene degradation in the sulfate-reducing enrichment culture N47. *J Biotech*, 192, 295–306.

- Selesi, D., & Meckenstock, R. U. (2009). Anaerobic degradation of the aromatic hydrocarbon biphenyl by a sulfate-reducing enrichment culture. *FEMS Microbiology Ecology*, 68, 86–93.
- Selesi, D. e., & Meckenstock, R. U. (2009). Anaerobic degradation of the aromatic hydrocarbon biphenyl by a sulfate-reducing enrichment culture. *FEMS Microbiology Ecology*, 68, 86–93.
- 97. Seo, J.-S., Keum, Y.-S., Hu, Y., Lee, S.-E., & Li, Q. X. (2006). Phenanthrene degradation in *Arthrobacter* sp. P1-1: initial 1,2-, 3,4- and 9,10-dioxygenation, and meta-and ortho-cleavages of naphthalene-1,2-diol after its formation from naphthalene-1,2-dicarboxylic acid and hydroxyl naphthoic acids. *Chemosphere*, 65, 2388–2394.
- Shapiro, J., Charbit, A., Benson, S., Caruso, M., Laux, R., Meyer, R., & Banuett, F. (1981). Trends in the biology of fermentations for fuels and chemicals. Springer-Verlag: Springer, pp. 243–272.
- Shavandi, M., Soheili, M., Zareian, S., Akbari, N., & Khajeh, K. (2013). The gene cloning, overexpression, purification, and characterization of dibenzothiophene monooxygenase and desulfinase from *Gordonia* alkanivorans ripi90a. Journal of Petroleum Science and Technology, 3, 57–64.
- 100. Shindo, K., Nakamura, R., Osawa, A., Kagami, O., Kanoh, K., Furukawa, K., & Misawa, N. (2005). Biocatalytic synthesis of monocyclic arene-dihydrodiols and -diols by *Escherichia coli* cells expressing hybrid toluene/biphenyl dioxygenase and dihydrodiol dehydrogenase genes. *Journal of Molecular Catalysis B: Enzymatic*, 35, 134–141.
- 101. Singh, D., Kumari, A., & Ramanathan, G. (2014). 3-nitrotoluene dioxygenase from *Diaphorobacter* sp. strains: cloning, sequencing and evolutionary studies. *Biodegradation*, 25, 479–492.
- 102. Smits, T. H., Balada, S. B., Witholt, B., & van Beilen, J. B. (2002). Functional analysis of alkane hydroxylases from Gram-negative and Gram-positive bacteria. J Biotech, 184, 1733–1742.
- Spain, J. C., & Gibson, D. T. (1991). Pathway for biodegradation of *p*-nitrophenol in a Moraxella sp. Applied and Environmental Microbiology, 57, 812–819.
- Spanggord, R. J., Spain, J., Nishino, S., & Mortelmans, K. (1991). Biodegradation of 2, 4-dinitrotoluene by a *Pseudomonas* sp. *Applied and Environmental Microbiology*, 57, 3200–3205.
- 105. Sun, X., Zahir, Z., Lynch, K. H., & Dennis, J. J. (2011). An antirepressor, SrpR, is involved in transcriptional regulation of the SrpABC solvent tolerance efflux pump of *Pseudomonas putida* S12. *J Biotech*, 193, 2717–2725.
- Symons, Z. C., & Bruce, N. C. (2006). Bacterial pathways for degradation of nitroaromatics. *Natural Product Reports*, 23, 845–850.
- Szaleniec, M., Hagel, C., Menke, M., Nowak, P., Witko, M., & Heider, J. (2007). Kinetics and mechanism of oxygen-independent hydrocarbon hydroxylation by ethylbenzene dehydrogenase. *Biochemistry*, 46, 7637–7646.
- Takeda, H., Yamada, A., Miyauchi, K., Masai, E., & Fukuda, M. (2004). Characterization of transcriptional regulatory genes for biphenyl degradation in rhodococcus sp. strain RHA1. J Biotech, 186, 2134–2146.
- Tanase, A.-M., Ionescu, R., Chiciudean, I., Vassu, T., & Stoica, I. (2013). Characterization of hydrocarbon-degrading bacterial strains isolated from oil-polluted soil. *International Biodeterioration & Biodegradation*, 84, 150–154.
- Tenbrink, F., Schink, B., & Kroneck, P. M. H. (2011). Exploring the active site of the tungsten, iron-sulfur enzyme acetylene hydratase. *J Biotech*, 193, 1229–1236.
- Thauer, R. K. (2011). Anaerobic oxidation of methane with sulfate: on the reversibility of the reactions that are catalyzed by enzymes also involved in methanogenesis from CO<sub>2</sub>. *Current Opinion in Microbiology*, 14, 292–299.
- Tinberg, C. E., & Lippard, S. J. (2011). Dioxygen activation in soluble methane monooxygenase. Accounts of Chemical Research, 44, 280–288.
- 113. Tinberg, C. E., Song, W. J., Izzo, V., & Lippard, S. J. (2011). Multiple roles of component proteins in bacterial multicomponent monooxygenases: phenol hydroxylase and toluene/o-xylene monooxygenase from *Pseudomonas* sp. OX1. *Biochemistry*, 50, 1788–1798.
- Torres Pazmino, D., Winkler, M., Glieder, A., & Fraaije, M. (2010). Monooxygenases as biocatalysts: classification, mechanistic aspects and biotechnological applications. *Journal of Biotechnology*, 146, 9–24.
- 115. van Beilen, J. B., Funhoff, E. G., van Loon, A., Just, A., Kaysser, L., Bouza, M., Holtackers, R., Röthlisberger, M., Li, Z., & Witholt, B. (2006). Cytochrome P450 alkane hydroxylases of the CYP153 family are common in alkane-degrading eubacteria lacking integral membrane alkane hydroxylases. *Applied and Environmental Microbiology*, 72, 59–65.
- 116. van Beilen, J. B., Panke, S., Lucchini, S., Franchini, A. G., Röthlisberger, M., & Witholt, B. (2001). Analysis of *Pseudomonas putida* alkane-degradation gene clusters and flanking insertion sequences: evolution and regulation of the *alk* genes. *Microbiology*, 147, 1621–1630.
- 117. Van der Meer, J. (2008). A genomic view on the evolution of catabolic pathways and bacterial adaptation to xenobiotic (compounds. ed., ). Norfolk, UK: Caister Academic Press.
- 118. von Netzer, F., Pilloni, G., Kleindienst, S., Krüger, M., Knittel, K., Gründger, F., & Lueders, T. (2013). Enhanced gene detection assays for fumarate-adding enzymes allow uncovering of anaerobic hydrocarbon degraders in terrestrial and marine systems. *Applied and Environmental Microbiology*, 79, 543–552.

- Wentzel, A., Ellingsen, T. E., Kotlar, H.-K., Zotchev, S. B., & Throne-Holst, M. (2007). Bacterial metabolism of long-chain n-alkanes. *Appl Microbiol Biotech*, 76, 1209–1221.
- 120. Whyte, L., Smits, T., Labbe, D., Witholt, B., Greer, C., & Van Beilen, J. (2002). Gene cloning and characterization of multiple alkane hydroxylase systems in rhodococcus strains Q15 and NRRL B-16531. *Applied and Environmental Microbiology*, 68, 5933–5942.
- Yakimov, M. M., Timmis, K. N., & Golyshin, P. N. (2007). Obligate oil-degrading marine bacteria. *Current Opinion in Biotechnology*, 18, 257–266.
- Yen, K.-M., Serdar, C. M., & Gunsalus, I. C. (1988). Genetics of naphthalene catabolism in pseudomonads. Critical reviews in microbiology, 15, 247–268.
- Zeyer, J., Kocher, H., & Timmis, K. (1986). Influence of para-substituents on the oxidative metabolism of o-nitrophenols by *Pseudomonas putida* B2. *Applied and Environmental Microbiology*, 52, 334–339.
- 124. Zylstra, G. J., & Gibson, D. T. (1991). Genetic Engineering. New York: Springer US, Plenum Press, pp. 183–203.