

Insights into the electron transfer mechanisms of permanganate activation by carbon nanotube membrane for enhanced micropollutants degradation

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HIGHLIGHTS

- A CNT filter enabled effective KMnO_4 activation via facilitated electron transfer.
- Ultra-fast degradation of micropollutants were achieved in KMnO_4/CNT system.
- CNT mediated electron transfer process from electron-rich molecules to KMnO_4 .
- Electron transfer dominated organic degradation.

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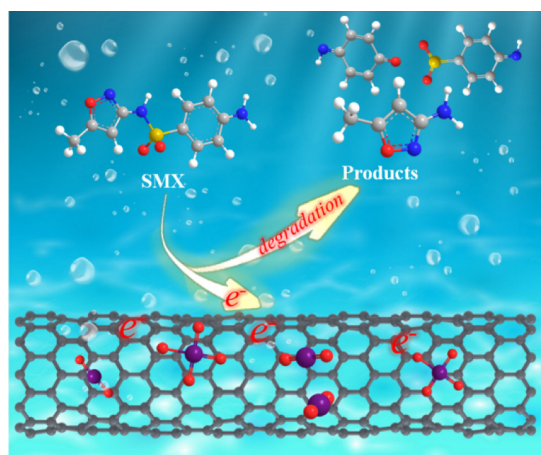
Carbon nanotubes

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GRAPHIC ABSTRACT



ABSTRACT

Numerous reagents have been proposed as electron sacrificers to induce the decomposition of permanganate (KMnO_4) by producing highly reactive Mn species for micropollutants degradation. However, this strategy can lead to low KMnO_4 utilization efficiency due to limitations associated with poor mass transport and high energy consumption. In the present study, we rationally designed a catalytic carbon nanotube (CNT) membrane for KMnO_4 activation toward enhanced degradation of micropollutants. The proposed flow-through system outperformed conventional batch reactor owing to the improved mass transfer via convection. Under optimal conditionals, a > 70% removal (equivalent to an oxidation flux of $2.43 \text{ mmol}/(\text{h}\cdot\text{m}^2)$) of $80 \mu\text{mol}/\text{L}$ sulfamethoxazole (SMX) solution can be achieved at single-pass mode. The experimental analysis and DFT studies verified that CNT could mediate direct electron transfer from organic molecules to KMnO_4 , resulting in a high utilization efficiency of KMnO_4 . Furthermore, the KMnO_4/CNT system had outstanding reusability and CNT could maintain a long-lasting reactivity, which served as a green strategy for the remediation of micropollutants in a sustainable manner. This study provides new insights into the electron transfer mechanisms and unveils the advantages of effective KMnO_4 utilization in the KMnO_4/CNT system for environmental remediation.

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1 Introduction

With the rapid growth of industrialization, toxic and non-biodegradable organic micropollutants of industrial origin are discharged into the environment directly or indirectly with a rise in environmental concerns regarding water pollution (Liu et al., 2009; Liu et al., 2018; Rasool et al., 2019). Different water treatment approaches, such as flocculation, bio-degradation, and physical adsorption, are not effective in dealing with organic contaminants (OCs) because of their chemical persistence, high toxicity and low concentration (Liu et al., 2009; Zheng et al., 2021; Shen et al., 2022). To meet the requirement of environmental remediation, advanced oxidation process has emerged as a viable technology for the abatement of different recalcitrant OCs in water due to their environmental friendliness and high efficiency (Zhou et al., 2019; Mustafa et al., 2021).

As a potential chemical oxidant, permanganate (KMnO_4) has been widely studied for water decontamination due to its high efficiency, cost-effectiveness and high stability (Guan et al., 2010; Ma et al., 2011; Chen et al., 2016; Shi et al., 2020). However, the poor stability and limited oxidation potential (1.68 V) of KMnO_4 restrict its applications in several circumstances (Bavasso et al., 2020). Hence, various innovative approaches have been attempted to boost the reactivity of KMnO_4 by producing highly reactive intermediate manganese species, such as ligands (Jiang et al., 2010), bisulfite (Cheng et al., 2019), electricity (Zhu et al., 2019a), ultraviolet irradiation (Hu et al., 2018). However, because of the addition of toxic and expensive chemicals and the occurrence of secondary pollution, these routes strongly impede the scientific progress of KMnO_4 oxidation toward practical applications.

In recent years, metal-free carbon material has emerged as an attractive additive to KMnO_4 oxidation due to its environmental friendliness. For instance, biochar was used as an electron sacrifier to enhance KMnO_4 oxidation by producing reactive manganese species (Tian et al., 2019). However, the reactivity of biochar is temporary due to the unattainable regeneration of reductive sites (Derakhshani and Naghizadeh, 2013). Moreover, the quenching effects of varying species strongly decreased the utilization efficiency of KMnO_4 . Alternatively, carbon nanotubes (CNT), an excellent electron transfer mediator, had been proved as a 'bridge' to facilitate the electron delivery from organic molecules (electron donor) to persulfate (electron acceptors) (Han et al., 2019; Hao et al., 2019; Shao et al., 2020). This may lead to oxidative decomposition of OCs, rather than converting from KMnO_4 to reactive manganese species (Peng et al., 2021; Wang et al., 2021).

With regard to mass transport limitation, conventional chemical oxidation methods mainly employ batch-mixed

mode, and the mass transport between the reactants and catalyst surface is often diffusion-limited. Alternatively, a flow-through system may serve as a crucial strategy to avoid this limitation *via* convection-regulated mass transport. Compared with conventional batch system, there are quantitative enhancements in the overall reaction kinetics realized by the flow-through system (Liu et al., 2020). Importantly, these one-dimensional CNT can be readily assembled into three-dimensional and porous networks, allowing for flow-through operation.

Herein, we design and establish a flow-through KMnO_4 /CNT system, where CNT was chosen as an electron mediator to simultaneously improve the oxidation ability and utilization efficiency of KMnO_4 . Sulfamethoxazole (SMX) was selected as the target compound due to its chemical persistence and ubiquitous distribution. The effects of several operational parameters were determined and optimized. The reaction mechanism was systematically explored according to a combination of experimental data and density functional theory (DFT) calculations by identifying reactive Mn species and analyzing the direct electron transfer from organic molecules to KMnO_4 . Besides, we comparatively investigated the utilization efficiency of KMnO_4 in different systems. Furthermore, ultraperformance liquid chromatography–quadrupole time-of-flight premier mass spectrometry (UPLC-QTOF-MS/MS) was used to unveil the SMX degradation pathway. Toxicity estimation software was employed to assess the toxicity of the intermediates based on quantitative structure–activity relationship (QSAR) prediction. Overall, this study reveals that CNT mediates electron transfer from SMX to KMnO_4 and presents the rational design of a robust, effective, and viable flow-through KMnO_4 /CNT system toward green environmental remediation.

2 Materials and methods

2.1 Reagents and chemicals

Details on the reagents and chemicals used in this study are summarized in Text S1. All chemical reagents were employed without further purification. Ultrapure water (resistivity $\geq 18.2 \text{ M}\Omega\cdot\text{cm}$) was utilized throughout the experiments.

2.2 Equipment and experimental procedures

The filtration trials were performed in a commercial Whatman polycarbonate filtration casing at ambient temperature ($25 \pm 1^\circ\text{C}$). Details on the fabrication of CNT filters were available in Text S2. To eliminate any contributions from physical adsorption, all experiments were carried out after achieving filter adsorption satura-

tion by flowing through 80 $\mu\text{mol/L}$ of SMX solution for 3 h at 0.5 mL/min, as controlled by an Ismatec ISM833C peristaltic pump (Switzerland). In a typical experiment, freshly prepared SMX solution with certain amount of KMnO_4 was passed through the filtration system at a flow rate of 0.5 mL/min in a single-pass filtration mode. Solution pH was adjusted using 1.0 mmol/L HCl and/or NaOH. At pre-determined time intervals, 1.0 mL samples were collected. The liquid samples were immediately filtered through a polytetrafluoroethylene membrane (PTFE, 0.22 μm) and mixed with 50 μL hydroxylamine hydrochloride (1.0 mol/L) to terminate any radical reactions. Oxidation flux was used to quantify the filtration performance, and the calculation equation was described in Text S3. All experiments were repeated in triplicate to ensure reproducibility.

2.3 Characterization

The crystal phase of pristine CNT filter specimens was determined by X-ray diffractometer (XRD, Rigaku D/max-2550PC, $\lambda = 1.5406 \text{ \AA}$). X-ray photoelectron spectroscopy (XPS, ESCALAB 250Xi, Thermo Fisher Scientific Escalab 250Xi) was applied to examine the variation on the chemical states on the CNT surface. Any generated reactive oxygen species (ROS) were identified using an EMXnano spectrometer (Bruker, Germany) with 5,5-dimethyl-1-pyrrolidine-N-oxide (DMPO) or 2,2,6,6-tetramethylpiperidine (TEMP). Chronoamperometry was conducted in 50 mmol/L Na_2SO_4 solution using a CHI 660E electrochemical workstation in a typical three-electrode system (e.g., a CNT working electrode, a saturated Ag/AgCl reference electrode and a Ti sheet counter electrode).

2.4 Analytical methods and calculations

High-performance liquid chromatography (HPLC, Shimadzu LC-16) coupled to a ShimNex UP C18 column

(5 μm , 4.6 mm \times 150 mm) was applied to determine the concentration of micropollutants. Details on the analytical protocols and operational parameters were summarized in Table S1. The concentration of KMnO_4 was determined using a UV2600 spectrophotometer (Shimadzu) at 525 nm. Sodium pyrophosphate (PP) was applied to determine the content of $\text{Mn(III)}_{\text{aq}}$ by forming a stabilized complex with $\text{Mn(III)}_{\text{aq}}$ (i.e., PP-Mn(III)) characterized by a 258-nm absorbance peak (Sun et al., 2015). The characteristic absorption peaks of Mn(VI) and Mn(V) were centered at 610 and 660 nm, respectively (Simandi et al., 1984). The SMX intermediates were quantified by a liquid chromatography-mass spectrometry (LC-MS, Agilent LC1290-QQQ-6470). Theoretical calculations of SMX degradation according to the density functional theory (DFT) were conducted via Gaussian 16 suite of programs. Details are available in Text S4. The quantitative structure activity relationship (QSAR) method was executed to analyze the mutagenicity, bioaccumulation factor, acute toxicity and developmental toxicity of SMX and its intermediates *via* Toxicity Estimation Software Tool (T.E.S.T., US, EPA). Details were available in Text S5.

3 Results and discussion

3.1 Ultrafast water decontamination by KMnO_4/CNT system

The catalytic performance of KMnO_4/CNT flow-through system was indirectly assessed by organic oxidative flux. To investigate the catalytic activity of CNT filter toward KMnO_4 , SMX was chosen as a model sulfonamide antibiotic. KMnO_4 , as a strong oxidizing agent, may be able to oxidize SMX directly. The results revealed a moderate SMX oxidation ability at an elevated concentration of KMnO_4 in the KMnO_4 -alone system (Fig. 1(a)). For example, negligible SMX was removed at a KMnO_4

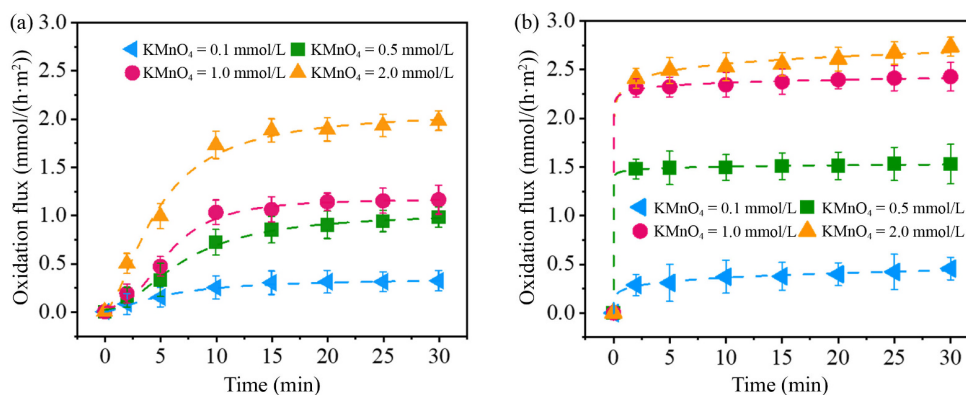


Fig. 1 SMX oxidation flux under different KMnO_4 concentrations in the (a) KMnO_4 alone system and (b) CNT/ KMnO_4 system. Experimental conditions: $[\text{SMX}]_0 = 80 \mu\text{mol/L}$, flow rate = 0.5 mL/min, and pH = 6.2.

concentration of 0.1 mmol/L, while a 1.98 mmol/(h·m²) SMX oxidation flux was observed at a KMnO₄ concentration of 2.0 mmol/L. As displayed in Fig. 1(b), the SMX oxidation flux increased from 0.46 to 2.43 mmol/(h·m²) with the KMnO₄ concentration increasing from 0.1 to 1.0 mmol/L in the KMnO₄/CNT system. This may be attributed to an increased number of electron acceptors with increasing KMnO₄ concentrations, and thus, more electrons are transferred from SMX to KMnO₄ with the mediation of CNT (Peng et al., 2021). It is worth noting that the SMX oxidation flux only increased from 2.43 to 2.74 mmol/(h·m²) when KMnO₄ concentration further increased from 1.0 to 2.0 mmol/L, which might be due to the limitation of CNT dosage, i.e., the dosage of CNT limited the electron transport capacity of the system (Peng et al., 2021). To be specific, the SMX oxidation flux was 2.43 mmol/(h·m²) in the KMnO₄/CNT system, which was significantly higher compared to the KMnO₄ alone system (1.16 mmol/(h·m²)) at a KMnO₄ concentration of 1.0 mmol/L. More importantly, the SMX oxidation can be profoundly improved by coupling with CNT, demonstrating the excellent reactivity of CNT toward the activation of KMnO₄.

3.2 Mechanistic investigation

The underlying mechanism of KMnO₄/CNT was explored by identifying the reactive species and exploring the potential role of electron transfer pathway.

3.2.1 Permanganate consumption in different processes

As shown in Fig. S1, 100 μmol/L KMnO₄ could be directly consumed by raw CNT in the absence of SMX. Although raw CNT possess a low electron-donating ability, it can give up electrons toward the KMnO₄ degradation (Zhang et al., 2022). This indicates that KMnO₄ may interact with CNT, but the reaction rate is rather slow. Besides, the utilization of KMnO₄ with SMX was evidently greatly than that without SMX. For example, only 30 μmol/L KMnO₄ without SMX was utilized in the KMnO₄/CNT system, whereas the utilization of KMnO₄ with 80 μmol/L SMX increased sharply to 340 μmol/L. These results further demonstrate that the reaction of KMnO₄ with CNT is sluggish and after addition of SMX into the KMnO₄/CNT system, the oxidation ability of KMnO₄ is significantly promoted. KMnO₄ degradation is largely dependent on the electron-donating effect of SMX via an electron-transfer pathway, which is contradictory to that of KMnO₄ activation by electron sacrificers. Hence, the interaction of KMnO₄ with electron donors may reduce the utilization efficiency of KMnO₄. Alternatively, direct electron-transfer mechanisms of the KMnO₄/CNT system could avoid KMnO₄ degradation, leading to an increase in KMnO₄ utilization efficiency.

3.2.2 Evaluating the contribution of diverse reactive species in the KMnO₄/CNT system

In the KMnO₄-based system, certain intermediates, such as manganese divalent ion (Mn(II)_{aq}), trivalent manganese (Mn(III)_{aq}), manganite (Mn(VI)_{aq}), hypomanganate (Mn(V)_{aq}) and manganese oxide (Mn(IV)_s), could be produced. Therefore, reactive Mn species (e.g., Mn(III), Mn(IV), Mn(V) and Mn(VI)) are often considered as the primary oxidative species for the degradation of contaminants (Chen et al., 2020). High-valence Mn species, such as Mn(V) and Mn(VI), may be formed and promotes the oxidation of contaminants during the KMnO₄ consumption (Gao et al., 2019). Methyl phenyl sulfoxide (PMSO) was chosen as scavengers for high metal-oxo species (e.g. Mn(V) and Mn(VI)) (Chen et al., 2020). Results suggested that the transformation kinetics of PMSO to PMSO₂ was similar regardless of the presence of CNT or not (Fig. S2). This indicated that a negligible contribution on the SMX degradation from high-valent Mn species (i.e., Mn(VI) and Mn(V)) in the KMnO₄/CNT system. Meanwhile, as shown in Fig. 2(a), the spike of PMSO cannot influence the SMX oxidation flux in the KMnO₄/CNT system. Besides, no characteristic peaks associated with Mn(V) and Mn(VI) were identified (Fig. 2(b)). Thus, excluding the formation and contribution of Mn(V) and Mn(VI) in this system (Chen et al., 2020).

Previous literature reported that MnO₂ with a moderate redox potential ($E_h = 0.464$ V) could directly oxidize OCs or promote the removal of OCs via oxidation and adsorption (Gao et al., 2012; Sun et al., 2019; Pan et al., 2021; Wang et al., 2021). To analyze these Mn intermediates on CNT surface, the surface characteristics of CNT pre- and post-reaction were assessed by XPS technique. As shown in Fig. 3(a), compared to the raw CNT, Mn elemental peak was determined on the surface of used CNT. In parallel, as shown in Fig. 3(b), Mn 2p_{3/2} photoelectron spectrum of used CNT was observed at the binding energy 642.0 eV and 643.8 eV, which can be assigned to Mn(IV). Additionally, the high-resolution O 1s spectrum of the raw CNT (Fig. 3(c)) located at 532.2 eV can be assigned to the surface of hydroxyl species (Yang et al., 2019). Compared with the raw CNT, a new peak (at 529.9 eV) appeared in the O 1s core level for the used CNT, which belonged to adsorbed Mn-O bonds on the CNT surface (Du et al., 2020). The results revealed that MnO₂ was formed and adsorbed on CNT surface during the reaction. To analyze the reactivity of Mn(IV), *ex situ* synthesized MnO₂ colloid was added dropwise into the KMnO₄/CNT system to assess the effects of MnO₂ on the SMX oxidation kinetics. Of note, this *ex situ* spiked MnO₂ colloid was according to the maximum consumption of KMnO₄ in the KMnO₄/CNT system (340 μmol/L, Fig. S1). As displayed in Fig. 2(c), the MnO₂/CNT system only showed a slight SMX oxidation flux of 0.42 mmol/(h·m²). Besides, no evident differences in

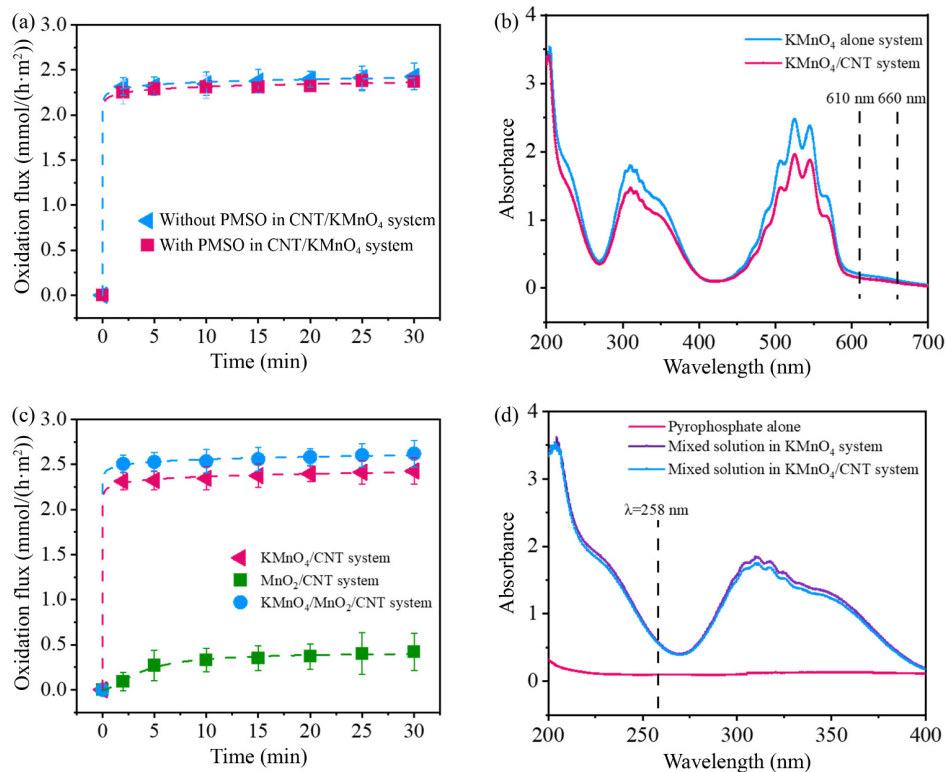


Fig. 2 (a) The effect of PMSO on SMX oxidation in the CNT/KMnO₄ system, (b) UV-visible spectrum of different solution samples, (c) SMX oxidation in various systems, (d) UV-visible spectrum of different solution samples in the presence of 1.0 mmol/L pyrophosphate. Experimental conditions: [SMX]₀ = 80 μmol/L, [KMnO₄]₀ = 1.0 mmol/L, flow rate = 0.5 mL/min, and pH = 6.2.

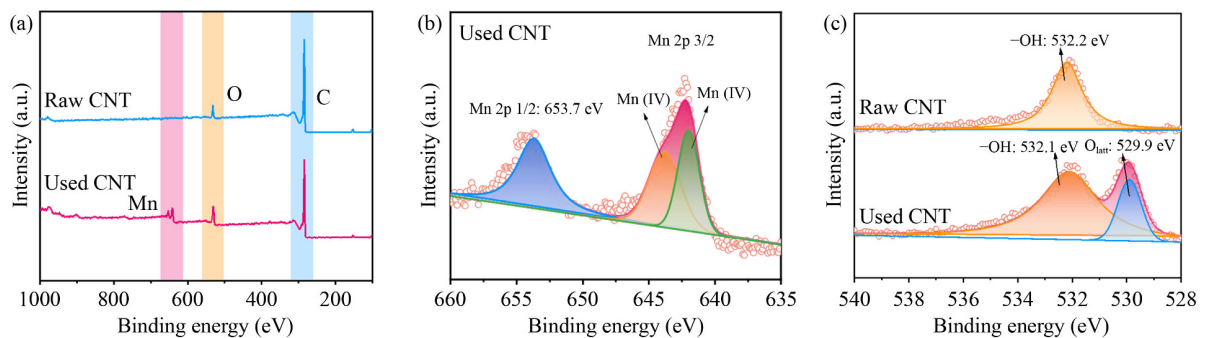


Fig. 3 XPS spectra of CNT: (a) survey spectrum, core-level spectra of (b) Mn 2p and (c) O 1s.

terms of SMX oxidation was observed between the KMnO₄/CNT and KMnO₄/MnO₂/CNT system (Fig. 2(c), 2.42 mmol/(h·m²) vs. 2.62 mmol/(h·m²), respectively), suggesting that the contribution from the *in situ* formed MnO₂ on SMX oxidation is rather limited.

It has been reported that Mn(III) obtained from KMnO₄ possess moderate oxidative ability ($E_h = 1.51$ V) toward OCs (Kostka et al., 1995; Zhu et al., 2019b). Therefore, to identify whether Mn(III) was involved in the system, the colorimetric determination of Mn(III)-pyrophosphate complex at a wavelength of 258 nm was detected (Zhu et al., 2019a; Zhu et al., 2019b). However, as displayed in Fig. 2(d), no obvious peak associated with Mn(III)-

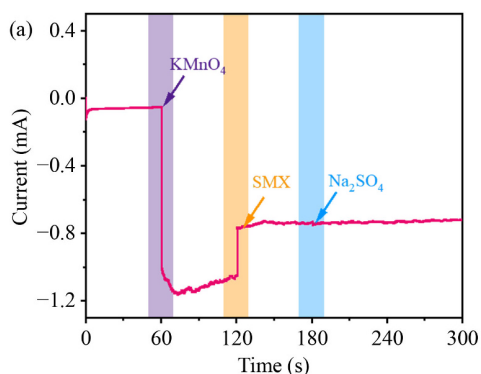
pyrophosphate complex was emerged. Thus, this excluded the possibility of Mn(III) production and its role in SMX decomposition.

Electron paramagnetic resonance spectrometer (EPR) analysis was conducted with the spin trapping agents TEMP and DMPO targeting ¹O₂ and HO[•], respectively. As displayed in Figs. S3a and S3b, the characteristic adduct signals were undetectable, thus excluding the possibility of ROS production in the KMnO₄/CNT system. The above discussion collectively indicated that the limited contribution from reactive Mn and oxygen species toward the degradation of SMX. Therefore, we hypothesize that other reactive mechanism may be

involved in SMX degradation in the KMnO_4/CNT system.

3.2.3 CNT mediates electron transfer from SMX to permanganate

In recent years, numerous researchers demonstrated that nanoscale carbon materials could mediate electron transfer from OCs to persulfate is the main pathway for OCs oxidation in persulfate-based oxidation systems (Ren et al., 2020; Zhou et al., 2021; Guo et al., 2022; Wu et al., 2022; Yu et al., 2023). Electrochemical characterization was an excellent technique to analyze the electrochemical characteristics of the electrode materials. For example, chronoamperometry tests had been employed to monitor the electron transfer process from the OCs to persulfate, and current density markedly elevated in the co-presence of persulfate and organic substrate on the carbon electrodes (Lee et al., 2016; Tang et al., 2018; Zhu et al., 2018; Jin et al., 2022). As a typical carbon-based material, CNT can also regulate the electron shuttle from SMX to KMnO_4 in the KMnO_4/CNT system. As displayed in Fig. 4(a), chronoamperometry was further performed to record the charge migration between KMnO_4 and SMX. When KMnO_4 was added into the system, the occurrence of a negative current showed the rapid electron transfer from CNT to KMnO_4 . Afterward, the current intensities were further strengthened with the addition of SMX. Hence, the adsorption/electron transfer represents the most likely mechanism of action, i.e., the SMX was first adsorbed by CNT, and then directly electron transfer to KMnO_4 with CNT employed as an electron channel (Fig. 4(b)). These results indicated the critical role of the electron transfer-mediated nonradical oxidation pathway in the KMnO_4/CNT system. Under the CNT mediation, KMnO_4 could directly oxidize the OCs via the electron transfer pathway, and ultimately transform into MnO_2 .



To further confirm this hypothesis, DFT calculations were conducted to evaluate the highest occupied molecular orbital (HOMO)-least unoccupied molecular orbital (LUMO) gaps of CNT, SMX, and permanganate. In general, HOMO indicates that electrons are more likely to escape, which can be easily attacked (Pearson, 1986; Casida et al., 1998). As shown in Table S2 and Fig. 5, the HOMO of SMX (-6.108563 eV) was more negative than the LUMO of CNT (-3.008827 eV), suggesting that the electrons of SMX could migrate from SMX to CNT (Han et al., 2019; Shao et al., 2020). Moreover, the electrons transported from the LUMO to HOMO of CNT because of the least energy gap (0.523347 eV) (Yun et al., 2018; Chen et al., 2020). Furthermore, the electrons transferred from the HOMO of CNT (-3.532174 eV) to the LUMO of permanganate (-2.585211 eV) for the decomposition of permanganate (Shao et al., 2020). DFT calculations indicated that CNT-mediated electron transfer from SMX to KMnO_4 was influenced by energy gaps. By integrating the above electrochemical analysis and theoretical data, we concluded that CNT could function as an electron mediator to accelerate electron transfer from SMX to KMnO_4 , thus leading to SMX oxidation and KMnO_4 decomposition.

3.3 Proposed SMX degradation pathways and toxicity assessment of the intermediates

The intermediates formed during SMX decomposition were identified using the HPLC-MS technique (Table S3 and Fig. S4). Three plausible SMX decomposition pathways are proposed according to the product identification. As shown in Fig. S5, in pathway I, the C-S bond can be hydrolyzed in the reaction process, leading to the formation of TP1 ($m/z = 110.0606$) and TP2 ($m/z = 99.0558$). In pathway II, N11 of SMX can also be oxidized to form TP3 ($m/z = 284.0348$), which is the classical oxidation product of SMX. Next, TP4 ($m/z = 300.0296$) can be formed by TP3 transformation. In pathway III, TP5 ($m/z = 503.0816$) may be a self-

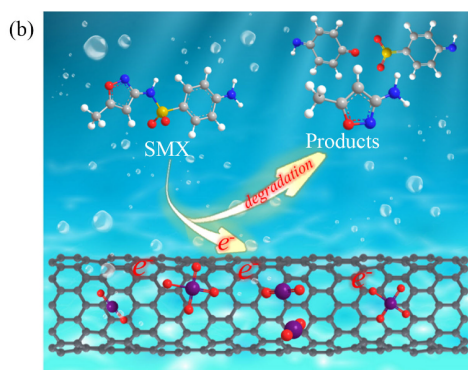


Fig. 4 (a) Chronoamperometry in 50 mmol/L Na_2SO_4 electrolyte solution at Pt electrodes and (b) proposed mechanisms of SMX oxidation in the CNT/KMnO_4 system.

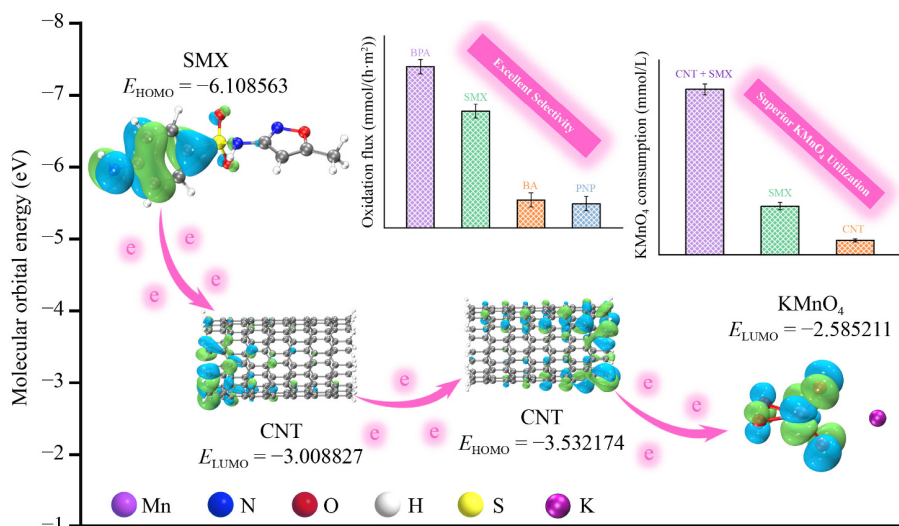


Fig. 5 The potential difference in molecular orbitals drives electron transmission via the structural defects of CNT.

coupling product formed by the coupling reaction of amino groups of two SMX molecules due to N11 of SMX. Finally, the above intermediates can be further transformed into small-molecule products, including TP2 ($m/z = 99.0558$), TP6 ($m/z = 155.003$) and TP7 ($m/z = 108.0451$)

The toxicity estimation of SMX intermediates is also a manifestation of the application potential of the CNT/ KMnO_4 system. As displayed in Fig. S6a, all the intermediates have negative mutagenicity, including SMX. In addition, for oral rat LD_{50} and bioaccumulation factor, all the intermediates have lower toxicity than SMX (Figs. S6b and S6c). The result reveals that the KMnO_4/CNT system significantly decreases the bioaccumulation factor and acute toxicity of intermediates compared to the original SMX. For developmental toxicity, SMX is 0.85, some intermediates have lower toxicity than SMX (Fig. S6d). However, only one toxic intermediate is produced, especially the developmental toxicity of TP4 is 0.95. In all, the above results show that the CNT/ KMnO_4 system

cannot only accomplish the oxidation of SMX but also decreases the toxicity of SMX.

3.4 Stability evaluation of the KMnO_4/CNT system

Heterogeneous oxidation processes usually require robust catalysts for the practical applications. To demonstrate the exceptional stability of the KMnO_4/CNT system, the SMX degradation experiments were performed under optimal conditions for six continuous running cycles (Fig. 6(a)). As displayed in Fig. 6(b), there was almost no decrease in the SMX oxidation flux, and the mean oxidation flux was $2.42 \text{ mmol}/(\text{h}\cdot\text{m}^2)$, indicating a good stability of the proposed system. This is mainly attributed to the outstanding capacity of catalytic and electronic transfer of CNT. Besides, XRD patterns of the raw and used CNT in the KMnO_4/CNT system are presented in Fig. S7. The characteristic peaks of raw CNT were detected at $2\theta = 25.80^\circ$, 40.08° and 53.03° , which can be indexed to the (110), (112), and (220) planes of crystal-

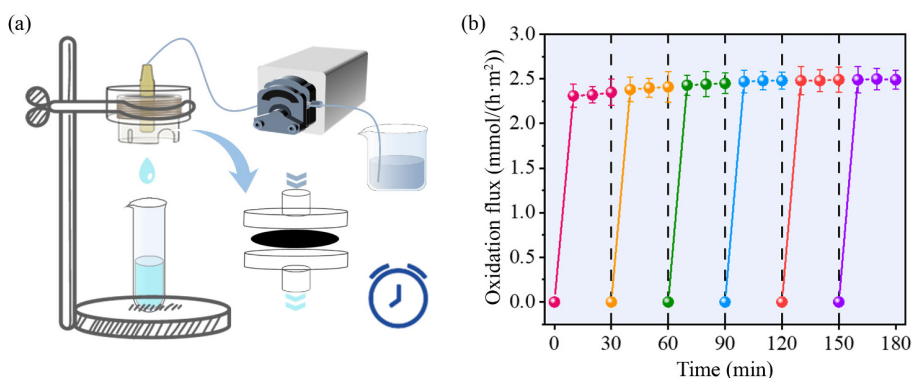


Fig. 6 (a) The schematic illustration of the KMnO_4/CNT system, and (b) the stability analysis for SMX oxidation flux. Experimental conditions: $[\text{SMX}]_0 = 80 \mu\text{mol}/\text{L}$, $[\text{KMnO}_4]_0 = 1.0 \text{ mmol}/\text{L}$, flow rate = $0.5 \text{ mL}/\text{min}$, and $\text{pH} = 6.2$.

line CNT (JCPDS 89-8489). Meanwhile, no additional peaks were detected in the used CNT, indicating that CNT maintained a high level of stability after the reaction.

The speciation of KMnO_4 was closely correlated with the solution pH. Thus, the pH effect on the efficacy of the KMnO_4/CNT system was examined. As shown in Fig. S8, the SMX oxidation kinetics under acidic pH conditions was much higher than that of alkaline conditions. The oxidation flux of SMX decreased from 2.90 to 1.00 $\text{mmol}/(\text{h}\cdot\text{m}^2)$ in the KMnO_4/CNT system as the solution pH increased from 2.8 to 9.4. Moreover, the presence of CNT evidently contributed to the SMX oxidation as well, as evidenced by the elevated SMX removal in the KMnO_4/CNT system, as compared with that of KMnO_4 alone system. Hence, the acidic conditions are more favorable for the SMX oxidation in the KMnO_4/CNT system.

Wastewater is consisted of various coexisting ions, which deteriorates the efficiency of OCs degradation. Therefore, this study investigated whether common inorganic anions can affect the efficacy of the KMnO_4/CNT system. As displayed in Fig. S9, the KMnO_4/CNT system exhibited little impact on SMX oxidation after the addition of various inorganic anions (Cl^- , SO_4^{2-} , and NO_3^-), indicating its strong resistance to environmental interference.

Humic acid (HA), a typical and natural organic matter (NOM) in real waters, was selected to assess the effect of NOM on the KMnO_4/CNT system. HA has been shown to exert a positive effect the oxidation of OCs in the KMnO_4 system (Jiang et al., 2010; Sun et al., 2013; Xu et al., 2017; Gao et al., 2018). Interestingly, HA with low concentrations have little effect on SMX oxidation (Fig. S9). However, HA with a background concentration of 5.0 mg/L can inhibit SMX degradation, owing to the direct reaction of HA with KMnO_4 . It has been reported that HA exhibits a reducing capacity, as it consists of reductive functional groups. As an electron donor that competes for KMnO_4 with SMX, HA can inhibit electron transfer from SMX to KMnO_4 (Peng et al., 2021).

To further verify the applicability of the KMnO_4/CNT system for OCs removal, the degradation of different refractory micropollutants that have wide pharmaceutical and industrial applications. As displayed in Fig. S10, benzoic acid and p-nitrophenol (electron-deficient contaminants) were difficult to undergo the electron-transfer reaction; while bisphenol A and SMX (electron-donating contaminants) were easy to undergo electron transfer reaction. These results indicated that the KMnO_4/CNT system could selectively and efficiently remove OCs with electron-donating. Taken together, our results demonstrate a novel flow-through system that enables robust, energy-effective, and high-efficiency water decontamination.

4 Conclusions

In this study, a flow-through KMnO_4/CNT system was designed for rapid and selective oxidation of OCs by enhancing the oxidation capacity of KMnO_4 . The system demonstrated high performance and stability with efficient KMnO_4 utilization in environmental remediation. Extensive experimental studies and theoretical calculations revealed the electron-transfer mechanisms of the KMnO_4 -based technologies. The excellent efficacy of the system was maintained in the presence of common anions, NOM, and an array of refractory micropollutants. The results of toxicity evaluation showed that the intermediates from SMX degradation had reduced toxicity. Hence, the KMnO_4/CNT system can serve as an efficient, cost-effective and sustainable advanced technology toward the remediation of OCs from water.

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