

# Abatement of mixed volatile organic compounds in a catalytic hybrid surface/packed-bed discharge plasma reactor

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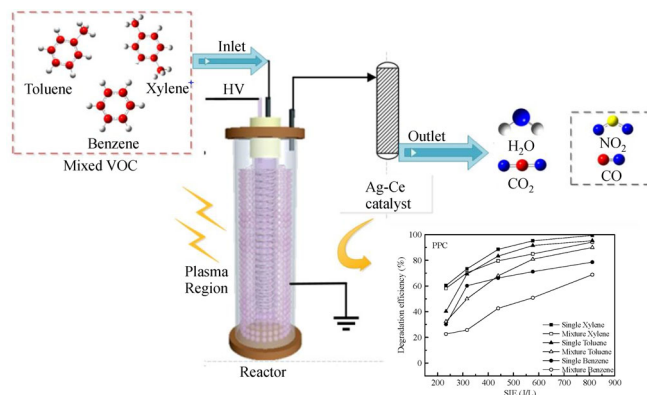
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## HIGHLIGHTS

- Mixed VOCs were successfully degraded by HSPBD reactor with Ag-Ce/ $\gamma$ -Al<sub>2</sub>O<sub>3</sub> catalyst at room temperature.
- The removal performance of single-component and mixed VOCs were compared in both NTP and PPC processes.
- The single-component and mixed VOCs decomposition products after plasma-catalysis treatment were analyzed.
- There existed an optimal gas humid to achieve the highest mixed VOCs removal efficiency.

## GRAPHIC ABSTRACT



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## ABSTRACT

In this study, post plasma-catalysis degradation of mixed volatile organic compounds (benzene, toluene, and xylene) has been performed in a hybrid surface/packed-bed discharge plasma reactor with Ag-Ce/ $\gamma$ -Al<sub>2</sub>O<sub>3</sub> catalyst at room temperature. The effect of relative air humidity on mixed VOCs degradation has also been investigated in both plasma-only and PPC systems. In comparison to the plasma-only system, a significant improvement can be observed in the degradation performance of mixed VOCs in PPC system with Ag-Ce/ $\gamma$ -Al<sub>2</sub>O<sub>3</sub> catalyst. In PPC system, 68% benzene, 89% toluene, and 94% xylene were degraded at 800 J·L<sup>-1</sup>, respectively, which were 25%, 11%, and 9% higher than those in plasma-only system. This result can be attributed to the high catalytic activity of Ag-Ce/ $\gamma$ -Al<sub>2</sub>O<sub>3</sub> catalyst to effectively decompose O<sub>3</sub> and lead to generating more reactive species which are capable of destructing the VOCs molecules completely. Moreover, the presence of Ag-Ce/ $\gamma$ -Al<sub>2</sub>O<sub>3</sub> catalyst in plasma significantly decreased the emission of discharge byproducts (NO<sub>x</sub> and O<sub>3</sub>) and promoted the mineralization of mixed VOCs towards CO<sub>2</sub>. Adding a small amount of water vapor into PPC system enhanced the degradation efficiencies of mixed VOCs, however, further increasing water vapor had a negative impact on the degradation efficiencies, which was primarily attributed to the quenching of energetic electrons by water vapor in plasma and the competitive adsorption of water vapor on the catalyst surface. Meanwhile, the catalysts before and after discharge were characterized by the Brunauer-Emment-Teller and X-ray photoelectron spectroscopy.

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## 1 Introduction

Volatile organic compounds (VOCs) in the atmosphere are

one of the major environmental problems due to serious hazards to ecological system and adverse effects on public health. Non-thermal plasma (NTP) technology has been widely studied as a promising method for the abatement of dilute VOCs at ambient conditions since the past 20 years [1,2]. NTP process is featured by non-equilibrium, fast ignition, and low energy cost. Electrons generated in a

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sufficient strong electric field collide with molecules, producing photons, oxide ions, free radicals, high-energy electrons and metastable molecules. These species can initiate a series of chemical reactions, such as excitation, ionization and dissociation, leading to decompose VOCs efficiently [3,4]. Various types of the NTP processes have been investigated for the removal of VOCs, including pulsed streamer corona discharge [5,6], gliding arc discharge [7], ferroelectric pellet-packed bed discharge [8,9], and surface discharge [10]. In our previous study, an innovative plasma reactor, hybrid surface/packed-bed discharge (HSPBD) plasma reactor, has been proposed for benzene degradation, which showed better performance in benzene degradation efficiency, mineralization efficiency and energy yield than surface discharge reactor or packed-bed discharge reactor alone [11].

However, this NTP process alone has some inherent drawbacks such as low energy yield, poor CO<sub>2</sub> selectivity and undesirable byproducts. A synergy between NTP and heterogeneous catalysis is expected to be the best choice to overcome these technological drawbacks. Noble metal catalysts and transition metal catalysts have been studied deeply to promote VOCs degradation. Among them, Ag-based catalysts were reported to exhibit a high oxidizing capacity and good physical stability [12,13]. Cerium oxide (CeO<sub>2</sub>) is known as a heterogeneous oxidation catalyst for the VOCs degradation reactions due to its excellent oxygen storage capacity (OSC) on the basis of the redox behavior between Ce<sup>3+</sup> and Ce<sup>4+</sup>. It can also act as a source of bulk oxygen species due to its labile oxygen vacancies and high oxygen mobility [14]. In recent years, compared to monometallic catalysts, bimetallic catalysts have gained increasing interest in the fields of combustion-catalysis and plasma-catalysis due to their multiple functionalities and excellent catalytic activity. The addition of Ce to Ag can facilitate the surface lattice of catalyst and increase the formation of surface adsorbed oxygen (O<sub>ads</sub>), which plays a key role in the plasma-catalytic reactions and significantly improved pollutants degradation [15]. Nevertheless, very limited work has been carried out using the supported Ag-Ce catalysts in plasma-catalytic oxidation of mixture VOCs.

In general, there are two different types depending on the catalyst location: by being exposed to the active plasma volume (In Plasma Catalysis, IPC) [16,17] or by placing the catalyst downstream of the discharge zone (Post Plasma Catalysis, PPC) [18–20]. It has been noticed that although IPC and PPC processes both have good performance of the degradation of VOCs, IPC process can change the retention time of pollutants in the plasma reactor. Einaga et al. [21] reported that the degradation efficiency of benzene was higher 48% in PPC system (i.e. a DBD reactor and MnO<sub>2</sub> in series) than in NTP at the specific input energy of 180 J·L<sup>-1</sup> and O<sub>3</sub> could vanish in PPC system. Zhu et al. [19] demonstrated that the maximum methanol removal efficiency of 95.4% using

Mn<sub>50</sub>Ce<sub>50</sub> oxide catalyst can be achieved in PPC system at a discharge power of 15 W and a gas flow rate of 1 L·min<sup>-1</sup>. In our previous research, it is found that PPC process was more effective at decomposing O<sub>3</sub> as well as destroying benzene than IPC system [22]. In addition, PPC system can also reduce the coke deposition to avoid the catalyst deactivation [23].

For industrial production, especially coating and chemical industries, various kinds of aromatic VOCs co-exists extensively and their emission has been limited strictly. The interaction between different kinds of VOCs in plasma-catalysis system is also fairly complex, and the synergistic mechanism study for mixed VOCs degradation is rather limited. Besides, from the view of industry application, gas relative humidity is also an important factor affecting the synergistic effect between plasma and catalysis as well as the energy utilization concerning VOCs degradation since VOCs emitted from industrial source generally contain water vapor.

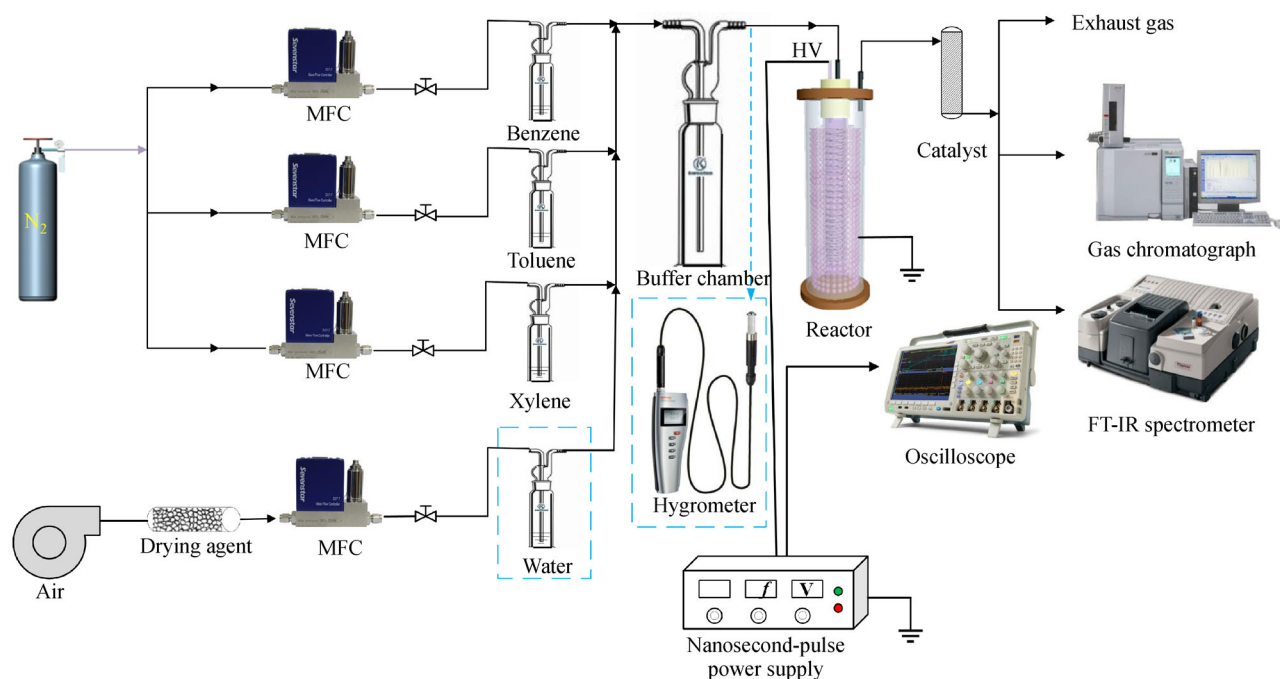
In the present work, benzene, toluene and xylene were chosen as target pollutants and a PPC system was constructed by introducing the Ag-Ce/γ-Al<sub>2</sub>O<sub>3</sub> catalyst downstream the discharge zone of the HSPBD plasma reactor. To understand the synergistic effects between plasma and catalysis, the degradation efficiencies, energy yields and CO<sub>2</sub> selectivities of the different single-component VOCs and mixed VOCs in both NTP and PPC processes were systematically discussed. FTIR was used to detect the decomposition products of single-component VOCs and mixed VOCs in NTP and PPC systems. In addition, the effects of the water vapor on the degradation efficiencies of mixed VOCs and O<sub>3</sub> concentration were also investigated.

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## 2 Experimental

### 2.1 Experimental setup and apparatus

The schematic diagram of the experimental setup is shown in Fig. 1. It consists of a bipolar pulsed power supply (P60D-III, Institute of Special Power of DLUT, China), a reaction gas supply system, a hybrid surface/packed-bed discharge (HSPBD) plasma reactor, a post plasma catalyst reactor, an analytical system and a relative humidity regulating system. Benzene, toluene and xylene were put in water baths and evaporated by passing a pure nitrogen. The flow rates of nitrogen were controlled by three mass flow controllers (MFCs) (Sevenstar, China, 0–100 mL·min<sup>-1</sup>) to make sure the concentrations of VOCs accuracy. According to the different boiling point and saturated vapor pressure of three kinds of VOCs, the flow rates of nitrogen were 5.5, 7.1, and 9.2 mL·min<sup>-1</sup> to obtain 100 ppm benzene, 100 ppm toluene and 100 ppm xylene, respectively. Three kinds of VOCs vapors were mixed with air completely and were fed into the reactor and then the



**Fig. 1** Schematic diagram of the experimental setup

catalyst zone at standard temperature and pressure. The gas flow rate was fixed at  $0.5 \text{ L} \cdot \text{min}^{-1}$  by using the mass flow controller (Sevenstar, China,  $0\text{--}3 \text{ L} \cdot \text{min}^{-1}$ ). The concentrations of VOCs at the inlet and outlet of reactor were analyzed by a gas chromatograph (Shimadzu GC-2010, Japan) equipped with a flame ionization detector (FID) and a 2 m Haysep-D packed column (GDX-102). The temperature program of the GC oven was the constant temperature  $100^\circ\text{C}$ . The inlet temperature and FID temperature were  $65^\circ\text{C}$  and  $250^\circ\text{C}$ , respectively. This condition was applied successfully to isolate and determine the benzene, toluene and xylene. Meanwhile, an online Fourier transform infrared spectroscopy (FTIR, Nicolet 6700, USA), equipped with a DTGS detector and a 2.4 m gas cell with KBr windows, was used to analyze the gaseous products. The  $\text{O}_3$  concentration was measured by the iodometric method [24]. The concentrations of NO and  $\text{NO}_2$  were measured by a flue gas analyzer (Testo 350, Germany). The electrical parameters were monitored with a four-channel digital oscilloscope (Tektronix TDS2024, USA), a voltage probe (Tektronix P6015A, USA) and a current probe (Tektronix P6021, USA). The relative humidity was adjusted by the drying agent and evaporable water content and tested by hygrometer (Rotronic 8303 Bassersdorf, Germany). The concentration of single-component VOC was 300 ppm, while mixed VOCs were comprised of 100 ppm benzene, 100 ppm toluene, and 100 ppm xylene.

## 2.2 Reactor configuration

A detailed description of the geometry of the plasma

reactor has been published elsewhere [11]. Briefly, a 280 mm-long Plexiglas cylinder (i.d. 35 mm) and a 250 mm-long quartz tube (i.d. 14 mm) were used as the dielectric barrier in a coaxial configuration. The inner surface of Plexiglas cylinder was wrapped by an aluminum-foil (190 mm length) as the ground electrode. The inner discharge electrode was a roll of stainless-steel coil (1 mm diameter) around the inner surface of the quartz tube connected to the bipolar pulsed power. The open area between the Plexiglas cylinder and the quartz tube was filled with glass beads ( $3 \pm 0.3 \text{ mm}$ ). The residence time of the VOC components was 30 s. When a nanosecond pulsed voltage is applied between high-voltage electrode and ground electrode, surface discharge starts from the peripheral edges of the coil electrode and stretches out along the surface of dielectric barrier (the quartz tube), namely, surface discharge zone. At the same time, many microdischarges are generated at the contact points between the glass beads, namely, packed-bed discharge zone. The surface discharge and the packed-bed discharge occur concurrently in the reactor, thus, this type of plasma is considered as “hybrid plasmas.” In addition, 3 g of  $\text{Ag-Ce}/\gamma\text{-Al}_2\text{O}_3$  catalyst was introduced after the discharge zone and supported by Pyrex glass, which has an inner diameter of 1.5 cm and an effective length of 8.0 cm. The space velocity of the catalyst is  $2062 \text{ h}^{-1}$ .

## 2.3 Catalyst preparation and characterization

$\text{Ag-Ce}/\gamma\text{-Al}_2\text{O}_3$  (13.5 wt.% Ag and 1.5 wt.% Ce) was prepared by the impregnation method in this study based on previous experimental results [22]. The  $\gamma\text{-Al}_2\text{O}_3$  pellets

of 2.5–3.5 mm in diameter as the support was first oxidized at 723 K for 2 h, followed by impregnation with silver nitrate and cerium nitrate aqueous solutions. Then it was dried at 378 K followed by calcination in air at 773 K for 5 h in the muffle furnace to obtain the catalyst. When the catalyst Ag-Ce/ $\gamma$ -Al<sub>2</sub>O<sub>3</sub> was introduced in the discharge plasma system, the discharge was started after the outlet concentration was equal to the inlet concentration, which implied that the adsorption-desorption equilibrium was achieved over the catalyst surface to avoid the adsorption of catalyst to reduce the concentration of pollutants in degradation process. To compare the physico-chemical properties of the catalysts, the catalyst Ag/ $\gamma$ -Al<sub>2</sub>O<sub>3</sub> (13.5 wt.% Ag) and Ce/ $\gamma$ -Al<sub>2</sub>O<sub>3</sub> (1.5 wt.% Ce) were prepared in the same manner.

The Brunauer-Emment-Teller (BET) measurement was used to characterize the specific surface areas ( $S_{BET}$ ) of the catalysts by carrying out a N<sub>2</sub> adsorption-desorption experiment at 77 K (Quantachrome Autosorb-1, America). The BET method was used to determine the total pore volume and the average pore diameter of the catalysts. X-ray photoelectron spectroscopy (XPS) spectra was recorded using a ThermoVG ESCALAN250 spectrometer fitted with Al-K $\alpha$  source. High resolution (0.05 eV) was performed for photoelectron peaks at the pass energy of 20 eV to identify the chemical states of all the elements. All the binding energies were calibrated using the C1s neutral carbon peak at 284.65 eV as an internal standard. XPS data were processed by using the XPS Peak 4.1 software.

## 2.4 Methods

The energy (J) of a single pulse was determined by the applied voltage and current:

$$E_p = \int_0^t U_t \times i_t dt. \quad (1)$$

The average discharged power:

$$P = E_p \times f, \quad (2)$$

where  $U_t$  is the pulse voltage (V),  $i_t$  is the pulse current (A),  $t$  is time (s),  $E_p$  represents for energy delivered per pulse (J),  $f$  is the pulse frequency (Hz),  $P$  is discharge power (W).

The  $SIE$  was defined as the discharge power divided by the gas flow rate, which could be calculated with the following equations:

$$SIE(J/L) = \frac{P}{V} \times 60. \quad (3)$$

The energy yield ( $EY$ ) as a measure of the energy efficiency was calculated as follows:

$$\eta_i = \frac{C_{inlet(i)} - C_{outlet(i)}}{C_{inlet(i)}} \times 100\%, \quad (4)$$

$$EY = \frac{3.6 \sum C_{inlet(i)} \eta_i M_i}{22.4 SIE}. \quad (5)$$

The selectivity of CO<sub>2</sub> was defined as follow:

$$S_{CO_2} = \frac{C_{outlet(i)[CO_2]}}{\sum n_i (C_{inlet(i)} - C_{outlet(i)})} \times 100\%, \quad (6)$$

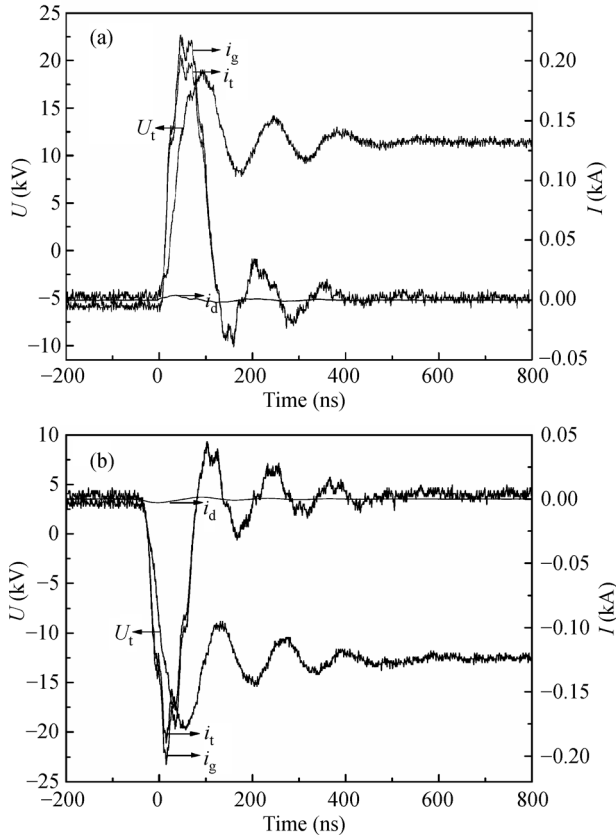
where  $V$  is gas flow rate (L·min<sup>-1</sup>);  $C_{inlet(i)}$  and  $C_{outlet(i)}$  is the inlet concentration and outlet concentration of benzene, toluene or xylene (ppm), respectively;  $\eta_i$  is the conversion of benzene, toluene or xylene; and  $M_i$  corresponds to the molar mass of VOCs (g·mol<sup>-1</sup>); 22.4 is the molar volume of gas (L·mol<sup>-1</sup>) at the ambient condition; 60 and 3.6 are the unit conversion coefficients;  $C_{outlet(i)[CO_2]}$  is the outlet concentration of CO<sub>2</sub>,  $i$  means one of the single-component or mixed VOCs;  $n_i$  is the number of carbon atoms in the molecules of benzene, toluene or xylene.

## 3 Results and discussion

### 3.1 Electric characterization of the discharge

The typical waveforms of applied voltage ( $U_t$ ) and current ( $i_t$ ) are shown in Fig. 2. When the HSPBD reactor was driven by bipolar nanosecond pulse power, the energy was injected into the discharge reactor in a very short moment (about 250 ns) and the peak value of current reached 200A at peak voltage 20 kV, so it can supply a very high instantaneous power  $4 \times 10^6$  W. The duration time of a single pulse discharge was nearly 100 ns, which was much less than the time for gas heating or ionization instability (about  $10^5$ – $10^6$  ns) [25], therefore the input energy could generate a tremendous amount of high energy electrons and active species. In addition, an equivalent circuit of air gap in a HSPBD can be simulated using a variable resistor connected with a capacitance ( $C_g$ ) in parallel, and then in series with a plane-parallel capacitor ( $C_d$ ) [26]. According to the equivalent circuit, air gas current ( $i_g$ ) (i.e. discharge current) and displacement current ( $i_d$ ) were calculated by the formulae as follow.  $C_g$  and  $C_d$  were calculated by the simplified structure model [26]. The applied current ( $i_t$ ) was the superposition of air gas current ( $i_g$ ) and displacement current ( $i_d$ ). It can be seen that gas current ( $i_g$ ) takes up a great proportion in applied current, indicating that the input energy mainly used for the gas discharge to produce energetic electrons rather than energy consumption.

In addition, instantaneous discharge current presents multiple peaks in the primary discharge of the applied voltage for both polarity. For positive pulse, a first current peak occurs during the rise time of the voltage, which is due to the accumulation of the electrical charges produced by the discharge initiated close to the inner discharge



**Fig. 2** Discharge voltage and current waveforms, gap current ( $i_g$ ) and displacement current ( $i_d$ ) of (a) the positive pulse and (b) the negative pulse

electrode where the electric field is intense enough. When the potential difference between the inner discharge electrode and the charged surface is adequate, a new discharge transmits on the dielectric layer, which results in the other discharge [27].

$$i_g = \left(1 + \frac{C_g}{C_d}\right) i_t - C_g \frac{dU_t}{dt}, \quad (7)$$

$$i_d = \frac{C_g C_d}{C_g + C_d} \frac{dU_t}{dt}. \quad (8)$$

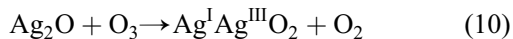
### 3.2 Catalyst characterization

Table 1 shows the surface area, total pore volume and average pore diameter of the support and the catalysts before and after discharge in the PPC process. Ag and Ce doping of the  $\gamma$ - $\text{Al}_2\text{O}_3$  cause a significant reduction in the surface area of catalysts. The  $S_{\text{BET}}$  of support  $\gamma$ - $\text{Al}_2\text{O}_3$  is  $204.41 \text{ m}^2 \cdot \text{g}^{-1}$ , while the  $S_{\text{BET}}$  of the Ag-Ce/ $\gamma$ - $\text{Al}_2\text{O}_3$  catalyst is about  $160 \text{ m}^2 \cdot \text{g}^{-1}$ . The pore volume of the catalysts is slightly lower than that of the support. The reduction in  $S_{\text{BET}}$  and pore volume for the Ag-Ce/ $\gamma$ - $\text{Al}_2\text{O}_3$  catalysts can be attributed to the coverage of  $\gamma$ - $\text{Al}_2\text{O}_3$  by Ag and Ce or partial occlusion of pores in the catalyst. In addition, the average pore diameters of catalysts are slightly larger than the support. Similar findings were reported by Zhu et al. using  $\gamma$ - $\text{Al}_2\text{O}_3$  supported single metal catalysts for the removal of acetone [28]. The average pore diameter (9.39 nm) of the Ag-Ce/ $\gamma$ - $\text{Al}_2\text{O}_3$  catalyst is enhanced compared to that Ag/ $\gamma$ - $\text{Al}_2\text{O}_3$  (8.89 nm) and Ce/ $\gamma$ - $\text{Al}_2\text{O}_3$  (8.76 nm), which could provide greater oxygen storage space for oxidation of VOCs. In addition, plasma exposure can result in the surface area of catalyst increase and the total pore volume decrease, indicating that ultrafine particles with higher specific surface area and crystal lattice with more vacancies have been formed. These physical changes could induce a higher catalytic activity, partially explaining the synergetic effect of plasma catalytic system [29].

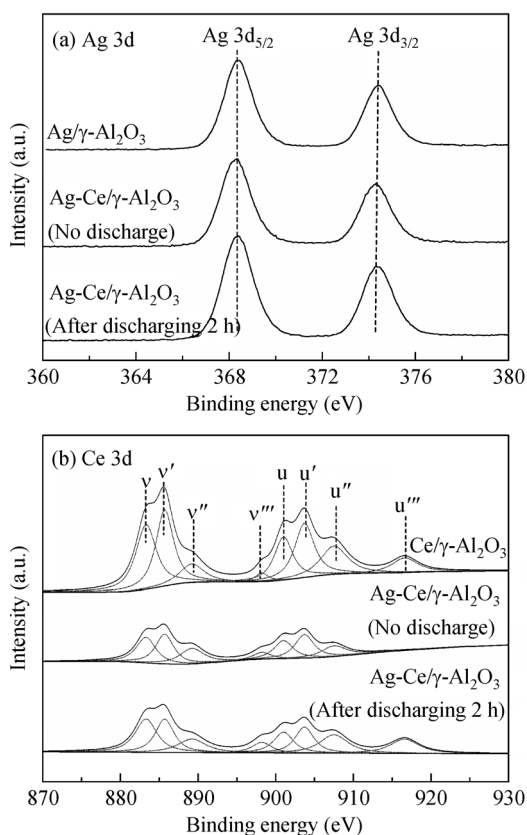
To determine the chemical form of the atoms, the XPS spectra of Ag/ $\gamma$ - $\text{Al}_2\text{O}_3$ , Ce/ $\gamma$ - $\text{Al}_2\text{O}_3$ , Ag-Ce/ $\gamma$ - $\text{Al}_2\text{O}_3$  and Ag-Ce/ $\gamma$ - $\text{Al}_2\text{O}_3$  (after discharging 2 h) are shown in Fig. 3. As shown in Fig. 3(a), the XPS spectra of Ag 3d of all catalysts are symmetric and similar, which consisted of two individual characteristic peaks Ag 3d<sub>5/2</sub> and Ag 3d<sub>3/2</sub> at binding energy of 367.7 and 373.8 eV correspond to Ag<sup>+</sup> and Ag<sup>0</sup>, respectively. Elemental Ag<sup>0</sup> can provide more atomic oxygen (Eq. (9)–(11)) [17], resulting in the adsorbed VOCs and intermediates on the catalyst surface more easily oxidized. In addition, it was observed that the peak intensity of the Ag 3d<sub>5/2</sub> of Ag-Ce/ $\gamma$ - $\text{Al}_2\text{O}_3$  catalysts after discharge increases, indicating that the plasma promotes a higher degree of dispersion of metallic silver on the support.

**Table 1** Physicochemical characteristics of the support and catalysts

Sample	Surface area ( $\text{m}^2 \cdot \text{g}^{-1}$ )	Total pore volume ( $\text{cm}^3 \cdot \text{g}^{-1}$ )	Average pore diameter (nm)
$\gamma$ - $\text{Al}_2\text{O}_3$	204.41	0.4382	8.38
Ag/ $\gamma$ - $\text{Al}_2\text{O}_3$	167.99	0.3732	8.89
Ce/ $\gamma$ - $\text{Al}_2\text{O}_3$	197.08	0.4319	8.76
Ag-Ce/ $\gamma$ - $\text{Al}_2\text{O}_3$	162.22	0.3808	9.39
Ag-Ce/ $\gamma$ - $\text{Al}_2\text{O}_3$ (After discharging 2 h)	163.33	0.3720	9.11



The XPS spectra of Ce 3d of all samples are showed in Fig. 3(b), where u and v indicate the spin-orbit splitting of states  $3d_{3/2}$  and  $3d_{5/2}$ , respectively. XPS spectra can be divided into eight peaks by fitting the curves. The peaks labeled as v' (885.7eV) and u' (903.7eV) are assigned to of  $\text{Ce}^{3+}$ , and the rest to  $\text{Ce}^{4+}$  [30]. As shown in Table 2, the concentration of  $\text{Ce}^{4+}$  are more than 60%, indicating that  $\text{CeO}_2$  is the main chemical state of element Ce for the catalysts. It is found that the effect of the discharge plasma on the morphologies of the catalysts changed greatly and the concentration of  $\text{CeO}_2$  was increased, indicating a



**Fig. 3** XPS spectra of Ag/ $\gamma$ - $\text{Al}_2\text{O}_3$ , Ce/ $\gamma$ - $\text{Al}_2\text{O}_3$ , Ag-Ce/ $\gamma$ - $\text{Al}_2\text{O}_3$  and Ag-Ce/ $\gamma$ - $\text{Al}_2\text{O}_3$  (after discharging 2 h) (a) Ag 3d (b) Ce 3d

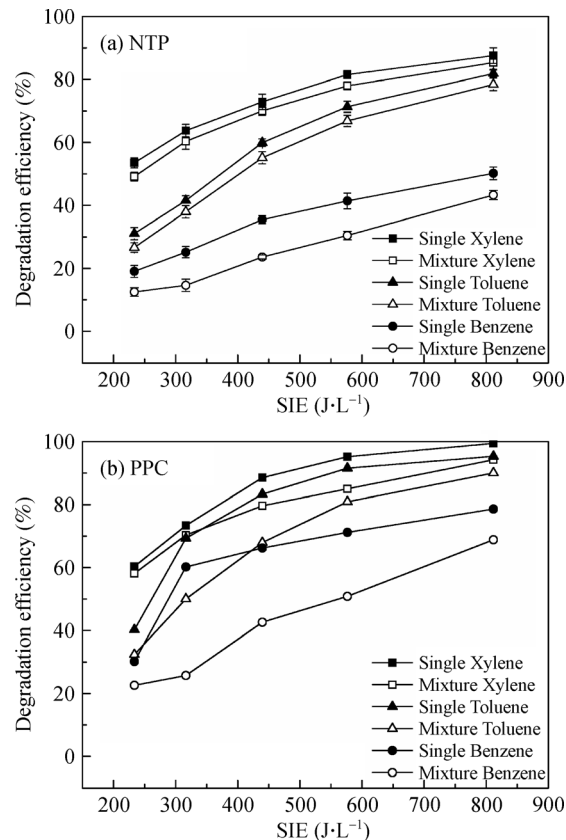
**Table 2** Surface Ce element states of different catalysts

Sample	Surface $\text{Ce}^{3+}$ (%)	Surface $\text{Ce}^{4+}$ (%)
Ce/ $\gamma$ - $\text{Al}_2\text{O}_3$	38.47	61.63
Ag-Ce/ $\gamma$ - $\text{Al}_2\text{O}_3$	37.99	62.01
Ag-Ce/ $\gamma$ - $\text{Al}_2\text{O}_3$ (After discharging 2 h)	29.56	70.44

strong interaction between Ag and Ce after discharging to promote the formation of  $\text{CeO}_2$ .

### 3.3 Comparison of single-component and mixed VOCs degradation in NTP and PPC processes

Figure 4 shows the degradation efficiencies of single-component VOCs and mixed VOCs as functions of SIE in NTP and PPC processes. The degradation efficiencies were measured after the discharge reached the steady-state, which means adsorption-desorption equilibrium of VOCs over the surface of catalyst to avoid the VOCs degradation by catalyst adsorption in PPC process. The result shows that the order of the degradation efficiencies was xylene>toluene>benzene in both single-component and mixed VOCs regardless of with or without catalyst. In PPC process, 98% of single-component xylene was degraded at  $800 \text{ J}\cdot\text{L}^{-1}$ , however, the corresponding degradation efficiencies of single-component toluene and benzene were 92% and 74%, respectively. This result is related to the stability of the molecular structure and the size of the molecular. The bond energies of the  $\text{C}_6\text{H}_5\text{-CH}_3$ ,  $\text{C}_6\text{H}_5\text{CH}_2\text{-H}$  in toluene and xylene are 4.4 and 3.7 eV, respectively, which are significantly lower than the  $\pi$  bond in a benzene ring (5.0–5.3 eV) and  $\text{C}_6\text{H}_5\text{-H}$  bonds (4.9 eV) in benzene molecule [31]. On the other hand, the molecular size also



**Fig. 4** The degradation efficiencies of single-component VOCs and mixed VOCs in NTP and PPC processes

influences the probability of collision because most reactions are initiated by inelastic collisions between aromatics and electrons and activated molecules, such as O and O<sub>3</sub> [31]. As a result, xylene is the most easily degradation component.

It can be observed that at the same *SIE*, the VOCs degradation efficiencies are higher in PPC process than in NTP process, suggesting that Ag-Ce/ $\gamma$ -Al<sub>2</sub>O<sub>3</sub> has good catalytic performance for three kinds of pollutants. For example, the degradation efficiencies of mixed VOCs of benzene, toluene and xylene were 68%, 89% and 94% in PPC process at 800 J·L<sup>-1</sup>, however, those were only 43%, 78% and 85% in NTP process. In PPC process, the active sites of the catalyst surface can decompose O<sub>3</sub> to generate the highly active species such as atomic oxygen and peroxide, which can decompose VOCs molecules more efficiently because the oxidation capacity of oxygen atom is much higher than O<sub>3</sub> [32]. It is also seen that the degradation efficiencies of the mixed VOCs are slightly lower than the single-component VOCs. The degradation efficiencies of benzene, toluene and xylene decreased by 7%, 3% and 2% respectively when they mixed together in NTP process at the *SIE* of 800 J·L<sup>-1</sup>. This result might be attributed to the fact that more complex degradation intermediates generated in mixed VOCs consume more energy at the same *SIE*.

To study the catalytic properties of the bimetal catalyst for mixed contaminants further, the relations between outlet concentration and time were examined, as shown in Fig. 5. In the stage of catalytic adsorption, it can be seen that the outlet concentration of the three pollutants would return to the initial 100 ppm after a period of time in the absence of the plasma, indicating that the catalyst alone could not decompose VOCs at room temperature. The adsorption time of three pollutants in the mixture was

different, which indicated that the adsorption capacities of the three kinds of pollutants were different. The order of adsorption capacity of the three pollutants was xylene>toluene>benzene. However, the benzene concentration decreased rapidly and was adsorbed preferentially on the catalyst surface in the first five minutes of the adsorption cycle. The adsorption rate of benzene to the catalyst was the highest, which explained that compared with the NTP process, the degradation efficiency of benzene in the PPC process increased most.

#### 3.4 Energy yields concerning single-component and mixed VOCs degradation in NTP and PPC processes

Figure 6 presents the energy yields concerning the decomposition of three kinds of single-component and mixed VOCs as a function of *SIE*. The energy yields increased first to the maximum and then decreased with the increase of *SIE*. This might suggest that input energy was utilized to degrade VOCs at lower *SIE* due to the more residual pollutants and the higher collision probability with the pollutant molecules at lower conversion. On the other hand, the concentration of VOCs decreased with the increase of VOCs degradation and hence the collision probability between the pollutant molecules and energetic species was lower at higher *SIE*. Meanwhile, a significant part of the input energy was converted into photons, heat, and used for byproducts formation at higher *SIE*. The order of the energy yield is xylene>toluene>mixture>benzene in both NTP and PPC processes. Comparing with the plasma only, the PPC process greatly improved the energy yields of VOCs. The energy yields of benzene, toluene and xylene were 2.3, 4.4 and 5.5 g·kWh<sup>-1</sup> in NTP process, respectively, which increased to 3.4, 5.0 and 6.2 g·kWh<sup>-1</sup> in PPC process at the *SIE* of 800 J·L<sup>-1</sup>.

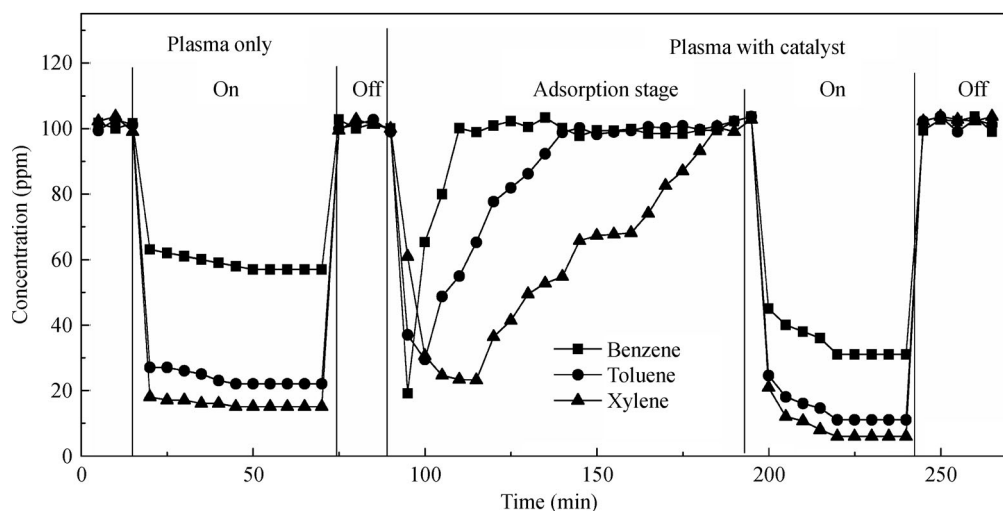
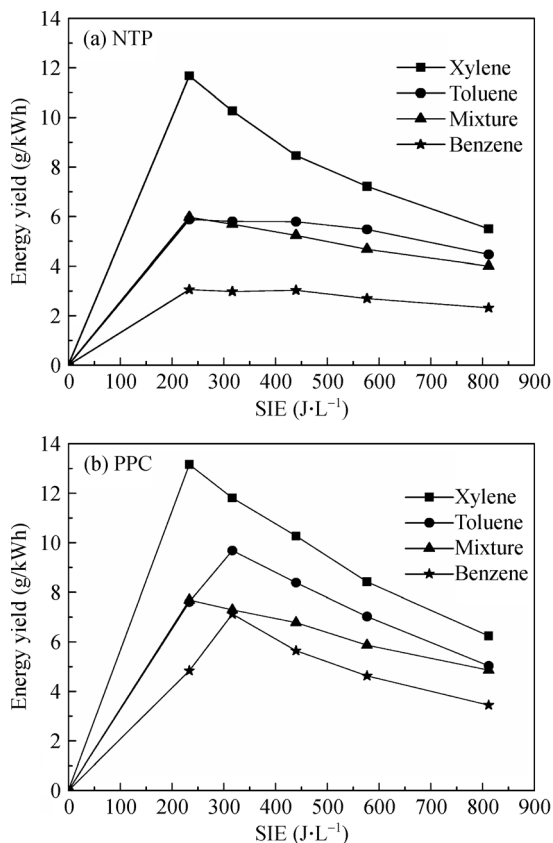


Fig. 5 The variation of the concentration of mixed VOCs with/without catalyst



**Fig. 6** Energy yield concerning single-component and mixed VOCs as a function of *SIE* in NTP and PPC systems

### 3.5 Behavior of products formation

#### 3.5.1 Products analysis

The FTIR was used to analyze the decomposition and discharge products in gas phase after NTP and PPC treatments at the *SIE* of  $800 \text{ J}\cdot\text{L}^{-1}$ , as shown in Fig. 7. The main gas products were C-containing products ( $\text{CO}_2$ , CO and HCOOH), discharge byproducts ( $\text{N}_2\text{O}$ ,  $\text{O}_3$ ) and  $\text{H}_2\text{O}$ . No other hydrocarbons except reactants and HCOOH were detected, suggesting that any other minor gaseous byproducts may be produced but they were at the sub-ppm level. As for xylene degradation, the absorption peak intensity of formic acid was stronger than benzene and toluene in NTP process, suggesting that formic acid can be produced by the approach, that is, the alkane fragments after benzene ring fracture and the methyl of benzene ring both can be further hydroxylated by the plasma consisted active substances  $\cdot\text{OH}$ ,  $\cdot\text{O}$  and  $\text{O}_3$ . Compared to NTP process, the absorption peak intensities of CO, HCOOH and  $\text{O}_3$  decreased, while the absorption peak intensities of  $\text{CO}_2$  and  $\text{H}_2\text{O}$  increased in PPC process. These results indicate that the Ag-Ce/ $\gamma$ - $\text{Al}_2\text{O}_3$  after the plasma region plays an important role in the further oxidation of VOCs and by-products. VOC molecules and the intermediates

can be adsorbed on the catalyst surface, and then converted to nontoxic  $\text{CO}_2$  and  $\text{H}_2\text{O}$  via the active oxygen species of the catalyst surface.  $\text{N}_2\text{O}$  formed via reaction of NO with N [33], and its absorption peak intensity was lower in the existence of the catalyst due to competing reactions of NO with the active surface oxygen species. In addition, the absorption peak intensity of  $\text{H}_2\text{O}$  was stronger in PPC process than in NTP process, indicating that more VOC molecules and intermediates were completely oxidized in PPC process. The brown deposits on the tube wall were dissolved in methylene chloride, extracted by Rotary evaporator, and analyzed by GC-MS. The degradation products of the single-component are relatively simple, such as, phenol, benzaldehyde, benzoic acid, hydroquinone and heptanoic acid, while the deposit of the mixed VOCs contains some large molecular weight substance, such as bis(2-ethylhexyl) phthalate, phenanthrene, 3-ethenyl-3-methylcyclopentane, 1, 2-benzenediazonium and butyl phthalate.

#### 3.5.2 $\text{CO}_2$ selectivities

$\text{CO}_2$  selectivity can directly reflect the mineralization degree of pollutants, thus quantitative analysis data for  $\text{CO}_2$  is shown in Fig. 8 as a function of the *SIE*. It is seen that  $\text{CO}_2$  selectivities in both NTP and PPC processes increased with the increase of *SIE* for all kinds of VOCs. For example, when the *SIE* increased from  $234 \text{ J}\cdot\text{L}^{-1}$  to  $800 \text{ J}\cdot\text{L}^{-1}$ ,  $\text{CO}_2$  selectivity of benzene increased from 29% to 72% in PPC process. This observation suggests that higher discharge power accelerates VOCs toward total oxidation. Compared to NTP process,  $\text{O}_3$  can be decomposed to active oxygen atoms on catalyst surface active sites in PPC process, which can further oxidize the intermediates to increase the concentration of  $\text{CO}_2$ . On the other hand, the residence time of intermediates can be extended effectively due to the catalyst adsorption characteristics, and thus the mineralization of intermediate products was accelerated. The  $\text{CO}_2$  selectivity of different VOCs follows this sequence: benzene>mixture>toluene>xylene. Although the degradation efficiency of benzene was lower than of toluene or xylene, the intermediate products of benzene were more easily to be completely oxidized to the final products ( $\text{CO}_2$  and  $\text{H}_2\text{O}$ ) by energetic electrons and active particles. However, when methyl groups in toluene and xylene molecules fractured, they can be further decomposed into small molecular hydrocarbon organic products, which can also synthesize stable organic compounds such as biphenyl type macromolecule products.

#### 3.5.3 The formation of $\text{NO}_x$

Formation of nitrogen oxides (NO and  $\text{NO}_2$ ) is inevitable byproducts during the discharge, which are must be



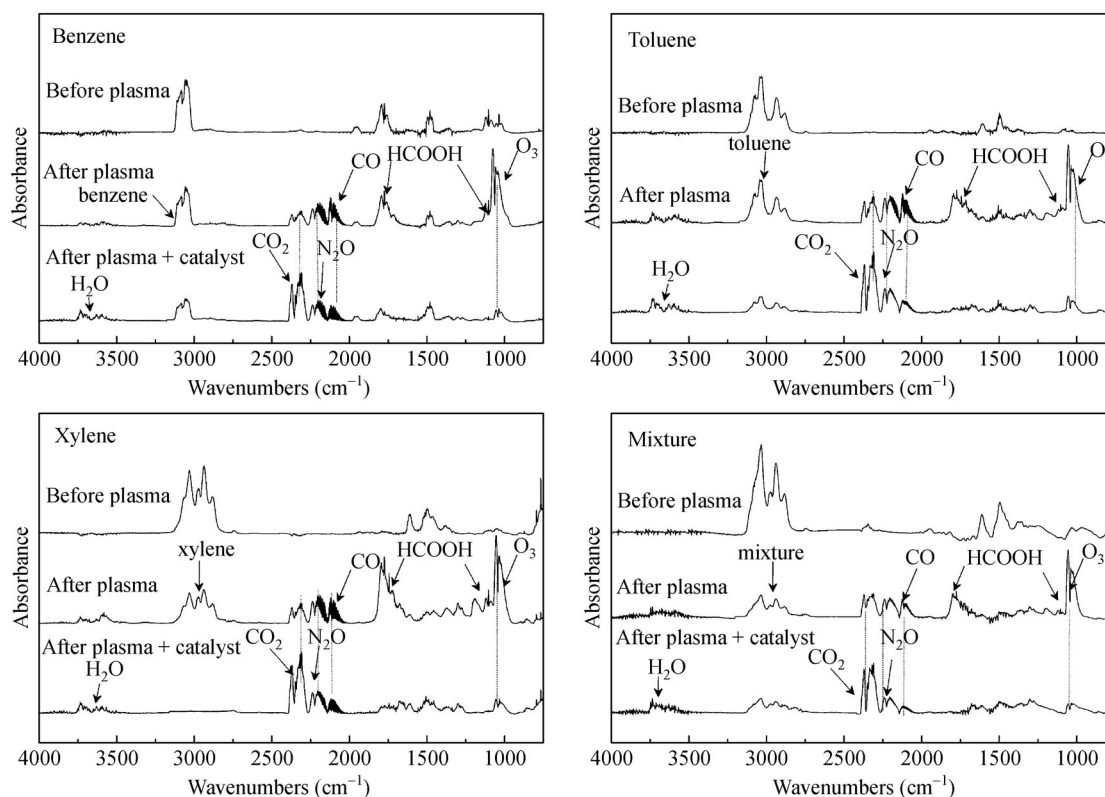
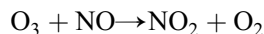


Fig. 7 FTIR spectra of gaseous products in hybrid surface/packed-bed discharge plasma reactor

characterized in the treatment process. Figure 9 shows  $\text{NO}_2$  concentration as a function of the *SIE* in NTP and PPC processes. The major  $\text{NO}_x$  species monitored by the flue gas analyzer was  $\text{NO}_2$ , and there was no  $\text{NO}$  detected in both NTP and PPC processes because the HSPBD reactor is good source of  $\text{O}_3$ , which can oxidize efficiently  $\text{NO}$  toward  $\text{NO}_2$  as Eq. (12):



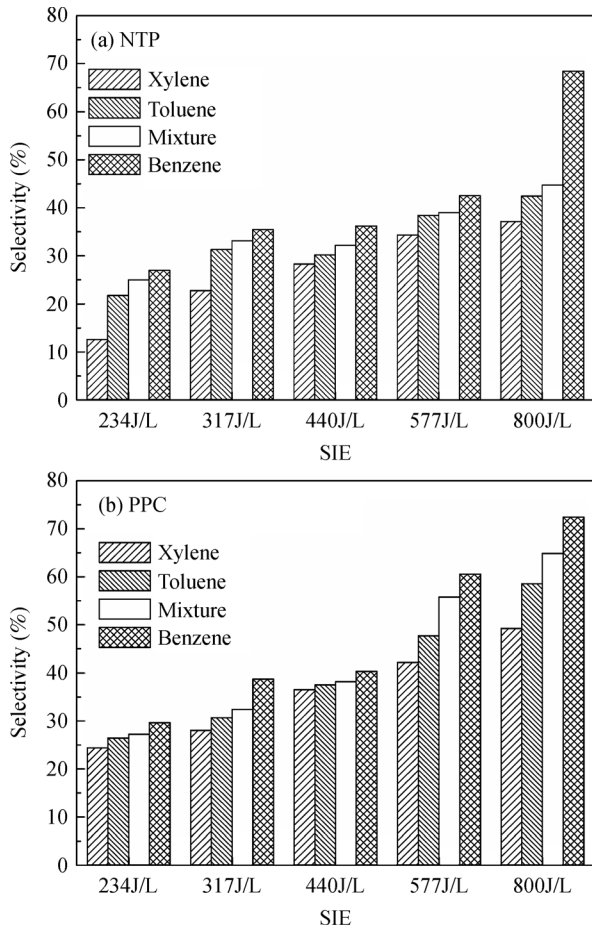
$$k = 1.8 \times 10^{-14} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1} \quad (12)$$

It is seen from Fig. 9 that the concentration of  $\text{NO}_2$  of single-component and mixed VOCs is approximately the same at the fixed *SIE* in both NTP and PPC processes. On the other hand, the  $\text{NO}_2$  concentration is much lower in PPC process than in NTP process at the same *SIE*. For at the *SIE* of  $800 \text{ J} \cdot \text{L}^{-1}$ , the amount of  $\text{NO}_2$  was 748 ppm in NTP process when mixed components were degraded, while only 170 ppm in PPC process. This could be attributed to the adsorption of  $\text{NO}_2$  on the catalyst surface and then conversion to  $\text{NO}_3^-$  by active oxygen species. Similar findings also have been reported in previous works. Durme et al. [34] has detected  $\text{NO}_3^-$  on the catalyst surface by ion chromatography method during the decomposition of high-concentration  $\text{O}_3$  generated by plasma discharge.

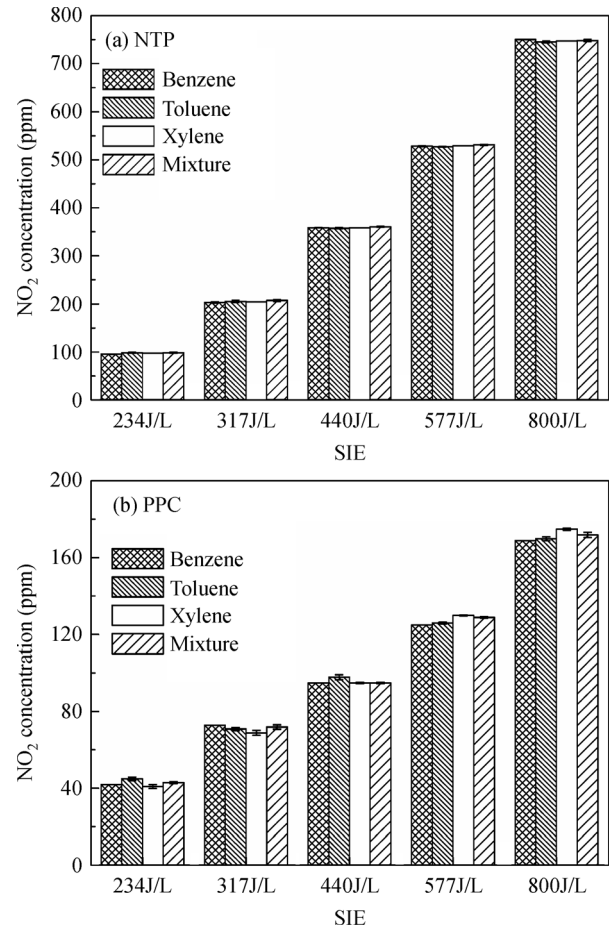
### 3.6 Effects of the relative humidity

#### 3.6.1 Effects of the relative humidity on degradation efficiencies

Most of the industrial processes produce exhaust gases containing water vapor, so the effects of water vapor on the decomposition of VOCs should be carefully investigated. Figure 10 shows the effects of relative humidity on the decomposition of mixed VOCs in both NTP and PPC processes at 23 kV at a constant flow rate of  $0.5 \text{ L} \cdot \text{min}^{-1}$ . It was found that the degradation efficiencies first increased at lower relative humidity, then decreased with the further increased relative humidity. In PPC process, the degradation efficiencies of benzene, toluene and xylene enhanced from 65%, 84.5% and 96.6% to 68%, 86.5% and 98% as the relative humidity increased from 0 to 40%, while they decreased significantly to 54.8%, 74% and 88.3% at 100% relative humidity, respectively. This is because increasing the water vapor in gas phase can produce more energetic species such as  $\cdot\text{H}$  and  $\cdot\text{OH}$  (Eqs. (13)–(15)) to improve the degradation efficiencies. Among these species,  $\cdot\text{OH}$  is more active than other particles, such as oxygen and peroxide (Eqs. (16)–(20)) [35,36]. However, under the high relative humidity condition, the energetic electrons were attached by excess water molecules and thus the

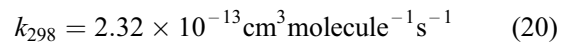
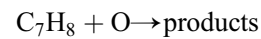
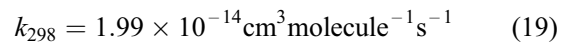
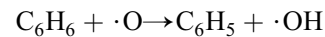
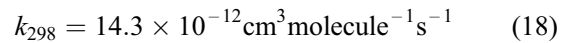
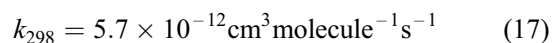
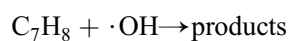
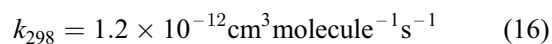
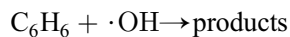
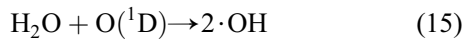
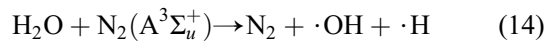
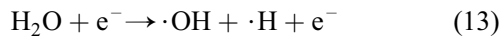


**Fig. 8** CO<sub>2</sub> selectivities as a function of the *SIE* in NTP and PPC processes



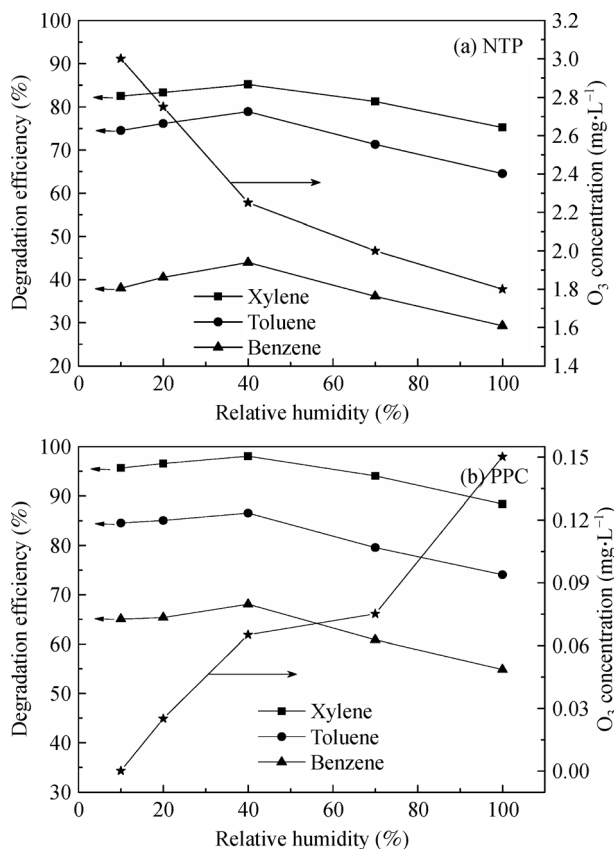
**Fig. 9** NO<sub>2</sub> concentration as a function of the *SIE* in NTP and PPC processes

electron density in the discharge zone was reduced. Secondly, excess water molecules can occupy the active sites on the catalyst surface, which inhibited O<sub>3</sub> decomposition, and reduced the amount of VOCs adsorption on the surface of the catalyst.



### 3.6.2 Effect of the relative humidity on O<sub>3</sub> formation

To understand the role of O<sub>3</sub> during the oxidation of mixture of VOCs, its concentration at the outlet was measured. Figure 10 also shows the O<sub>3</sub> outlet concentration as a function of relative humidity at 23 kV in NTP and PPC processes. The result shows that the O<sub>3</sub> concentration decreases as the relative humidity increased in NTP process, while increases in PPC process. In NTP process, O<sub>3</sub> outlet concentration was 3.0 mg·L<sup>-1</sup> at the relative



**Fig. 10** Effects of the relative humidity on degradation efficiencies and O<sub>3</sub> concentration of mixed VOCs in NTP and PPC processes at 23 kV

humidity of 10%, which decreased to 1.8 mg·L<sup>-1</sup> at 100% relative humidity. The lower of the O<sub>3</sub> concentration related to the electronegativity of water molecules, which limited the electron density and quenched activated chemical species, leading to generate fewer active particles. On the other hand, the discharge generated ·O reacted with H<sub>2</sub>O, leading to the number of ·O decline dramatically and inhibit the generation of O<sub>3</sub>.

Compare to NTP process, the introduction of the catalyst reduced the emissions of O<sub>3</sub> concentration due to O<sub>3</sub> catalytic decomposition on the Ag-Ce/γ-Al<sub>2</sub>O<sub>3</sub> catalyst surface. In PPC process, O<sub>3</sub> concentration was 0 under 10% humid condition, while O<sub>3</sub> concentration increased to 0.15 mg·L<sup>-1</sup> at the relative humidity of 100%. This result can be explained that the competitive adsorption between O<sub>3</sub> and H<sub>2</sub>O molecules was occurred under high relative humidity condition, which limited the absorption of O<sub>3</sub> on the catalytic surface.

## 4 Conclusions

This study mainly focused on the degradation of mixed VOCs (benzene, toluene, and xylene) in PPC configuration

using HSPBD plasma and Ag-Ce/γ-Al<sub>2</sub>O<sub>3</sub> catalyst. The degradation efficiencies of three-component mixed VOCs were in the following order: xylene>toluene>benzene. This result can be attributed to the stability of the molecular structure and the size of the molecular. Compared with plasma-only process, the combined application of HSPBD plasma and Ag-Ce/γ-Al<sub>2</sub>O<sub>3</sub> catalyst can lead to a significant enhancement in the degradation efficiencies of mixed VOCs. In PPC process, the degradation efficiencies of benzene, toluene, and xylene reached 68%, 89%, and 94% at the *SIE* of 800 J·L<sup>-1</sup>, respectively, which were 25%, 11%, and 9% higher than those in plasma-only system. The CO<sub>2</sub> selectivities greatly improved when compared to the plasma-only system, thereby minimizing the formation of organic intermediates. Moreover, the emission of discharge byproducts (such as NO<sub>x</sub> and O<sub>3</sub>) was also inhibited in PPC system. Under low relative humidity (relative humidity < 40%) condition, the presence of water vapor has a significant positive impact on mixed VOCs degradation due to the generation of more energetic species. However, an obvious decrease can be observed in mixed VOCs degradation at high relative humidity, which can be ascribed to the quenching of active species and competitive adsorption of H<sub>2</sub>O on the surface active sites of the catalyst. The optimal relative gas relative humidity is 40%.

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