### **ORIGINAL ARTICLE**





# **A global analysis of the drivers of human pressure within protected areas at the national level**

**Christos Mammides1**

Received: 4 June 2019 / Accepted: 18 April 2020 / Published online: 30 April 2020 © Springer Japan KK, part of Springer Nature 2020

#### **Abstract**

Global conservation efforts rely largely on effective protected areas. Recently, it was shown that one-third of the world's protected land is under intense human pressure. However, this proportion varies substantially across countries. Some countries are more successful than others in keeping their protected areas relatively free from intense human pressure. In this study, I explore the possible reasons behind this pattern. I fnd that countries with lower human population densities, lower percentage of agricultural land, and a larger area tend to have a lower proportion of their protected land under intense human pressure. These three factors alone account for approximately two-thirds of the variation in intense human pressure within protected areas. Other factors include the percentage of protected land under strict protection status (i.e., IUCN Categories I and II) and the current amount of funds invested in conservation at the national level. However, these factors are less important and account for little of the variation in human pressure. Moreover, there is no relationship between the levels of human pressure within protected areas and the countries' economic development status and efectiveness of national governance. These fndings suggest that under the current conditions—and assuming no major reforms in national conservation policies and actions—countries with high population densities and extensive areas of agricultural land are likely to struggle to keep human pressure within protected areas at low levels, irrespective of their economic development level, national governance strength, and current investments in conservation. Worse still, future projected increases in human population densities and agricultural land will likely exacerbate the human pressure within protected areas; these increases will occur mostly in developing countries—many of which are located in biodiverse regions—making conservation in those regions more diffcult. To achieve their sustainability goals, countries must take actions to address the key drivers of human pressure within protected areas.

**Keywords** Aichi targets · Biodiversity conservation · Convention on biological diversity · Human footprint index · Protected planet · Sustainable development goals

Handled by Osamu Saito, Institute for Global Environmental strategies, Japan.

**Electronic supplementary material** The online version of this article [\(https://doi.org/10.1007/s11625-020-00809-7\)](https://doi.org/10.1007/s11625-020-00809-7) contains supplementary material, which is available to authorized users.

 $\boxtimes$  Christos Mammides cmammides@outlook.com

<sup>1</sup> Guangxi Key Laboratory of Forest Ecology and Conservation, College of Forestry, Guangxi University, Daxuedonglu 100, Nanning 530004, China

## **Introduction**

We live in an era in which humans are impacting the natural environment at an unparalleled scale (Gaston et al. [2008](#page-8-0); Corlett [2015\)](#page-7-0). Human activities—which now afect most of the planet (Ellis et al. [2010](#page-7-1); Venter et al. [2016a](#page-8-1); Watson and Venter [2019\)](#page-9-0)—result in substantial biodiversity loss (Butchart et al. [2010;](#page-7-2) Newbold et al. [2015\)](#page-8-2). Protected areas represent one of the most important conservation strategies for addressing this loss (Visconti et al. [2019\)](#page-9-1). According to the Strategic Plan for Biodiversity, adopted in 2010, the signatory countries to the Convention on Biological Diversity (CBD) have agreed to expand their network of protected areas to cover 17% of the planet's land area (Watson et al. [2014;](#page-9-2) Gray et al. [2016](#page-8-3)). Many countries are nearing

that target, while several others have already exceeded it although some are still lagging behind (Jones et al. [2018](#page-8-4)). Considering the pivotal role of protected areas in addressing biodiversity loss (Fuller et al. [2010;](#page-7-3) Watson et al. [2014](#page-9-2); Visconti et al. [2019](#page-9-1)), it is essential we understand the factors that afect the countries' capacity to mitigate human impact within their protected areas and to conserve natural environments (Watson et al. [2014\)](#page-9-2). Not surprisingly, this has been a major research topic in conservation science (Gaston et al. [2008;](#page-8-0) Geldmann et al. [2018](#page-8-5)). Most studies concur that protected areas tend to have a positive efect on biodiversity (Nagendra [2008;](#page-8-6) Joppa and Pfaff [2011;](#page-8-7) Coetzee et al. [2014](#page-7-4); Gray et al. [2016;](#page-8-3) Geldmann et al. [2018\)](#page-8-5). However, the extent to which they are able to do so varies substantially (Coetzee et al. [2014](#page-7-4); Spracklen et al. [2015](#page-8-8); Geldmann et al. [2019](#page-8-9); Anderson and Mammides [2020a\)](#page-7-5). Evidence suggests that many of the world's protected areas remain inefective in conserving biodiversity and in curbing human impact (Gaston et al. [2008;](#page-8-0) Coetzee et al. [2014;](#page-7-4) Watson et al. [2014](#page-9-2); Spracklen et al. [2015\)](#page-8-8).

In fact, it was shown recently that one-third of the protected land across the planet is under intense human pressure (Jones et al. [2018\)](#page-8-4). Researchers reached this conclusion by measuring the mean human footprint (Venter et al. [2016a](#page-8-1), [b\)](#page-8-1) within the world's protected areas. The human footprint is a composite index of the human pressure across the globe—at a resolution of  $1 \text{ km}^2$  (Venter et al. [2016a](#page-8-1)). It is based on eight types of human pressure, including human population densities, built-up areas, intensive agriculture, and infrastructure (Venter et al. [2016b\)](#page-9-3). Although the human footprint index is based on satellite-derived data—which may not always capture fner scale anthropogenic disturbance (Peres et al. [2006;](#page-8-10) Mammides [2018](#page-8-11))—and although it does not include other key pressures, such as pollution and invasive species, the index has been shown to be a good proxy of the human pressure on natural environments (Watson and Venter [2019](#page-9-0)). For example, increases in human footprint have been linked to increased animal extinction risk (Di Marco et al. [2018](#page-7-6)) and reduced animal movement (Tucker et al. [2018](#page-8-12)).

A glance at the list of countries and the proportion of their protected land under intense human pressure (Jones et al. [2018](#page-8-4)) reveals substantial variation. Inevitably, this variation leads to the following question: why are some countries more successful than others in keeping their protected areas relatively free from intense human pressure? It is a simple yet intriguing question that goes beyond mere academic interest. Given the fact that human pressure is still rising within many of the world's protected areas (Jones et al. [2018;](#page-8-4) Geldmann et al. [2019](#page-8-9); Anderson and Mammides  $2020a$ ), and that it is significantly affecting biodiversity (Tucker et al. [2018](#page-8-12); Di Marco et al. [2019;](#page-7-7) Watson and Venter [2019\)](#page-9-0), the answer to this question could provide us with insights important for conservation. Once we understand better the factors associated most with the levels of human pressure within protected areas, we can shape our national and international policies accordingly to address those factors. Protected areas are imperative to the protection of the world's biodiversity and are essential for attaining the UN Sustainable Development Goals (SDGs), particularly Goals 14 and 15, which concern the sustainable use of natural resources (Dudley et al. [2017](#page-7-8)).

One could hypothesize that the proportion of protected land under intense human pressure is higher in countries within certain geographic regions. In their study, Jones et al. ([2018](#page-8-4)) mention that protected areas under intense human pressure tend to be found in western Europe, southern Asia, and Africa. Other studies, too, have shown that environmental degradation varies across geographic regions (Nagendra [2008](#page-8-6); Geldmann et al. [2019](#page-8-9); Anderson and Mammides [2020a\)](#page-7-5). Deforestation rates, for instance, tend to be higher within protected areas in Asian countries compared to countries in Latin America (Spracklen et al. [2015](#page-8-8)). Agricultural expansion rates also vary amongst regions (Geldmann et al. [2019](#page-8-9); Anderson and Mammides [2020b](#page-7-9)).

These geographic variations are partly driven by dissimilarities in socioeconomic factors. For example, studies often cite poor national and local governance as one of the main reasons for why protected areas are sometimes inefective in mitigating human impact (Wright et al. [2007](#page-9-4)) and conserving biodiversity (Smith et al. [2003;](#page-8-13) Smith and Walpole [2005\)](#page-8-14). Countries with more efective governance should have a higher capacity to enforce national and international environmental policies (Smith et al. [2003\)](#page-8-13). Therefore, they should be able to regulate better the levels of human pressure within their protected areas (Watson et al. [2014\)](#page-9-2).

Besides governance, another commonly cited reason for increased human pressure within protected areas is lack of resources (Leader-Williams and Albon [1988](#page-8-15); Bruner et al. [2001\)](#page-7-10), particularly fnancial resources. Having adequate fnancial resources for conservation is important (Bruner et al. [2001](#page-7-10); McCarthy et al. [2012](#page-8-16); Miller et al. [2013](#page-8-17); Waldron et al. [2013\)](#page-9-5). Waldron et al. [\(2017\)](#page-9-6) showed that higher financial investments in conservation—at the national level—can be linked to lower rates of biodiversity loss within countries. It is reasonable to hypothesize that higher conservation funding translates into lower proportions of human pressure within protected areas; this could even be the mechanism behind the reduced biodiversity loss reported by Waldron et al. [2017.](#page-9-6)

High human population densities are a major determinant of anthropogenic disturbance in natural environments (McKee et al. [2004\)](#page-8-18). It is for that reason that human population densities are incorporated explicitly into the human footprint index (Venter et al. [2016b\)](#page-9-3). Human population densities are also the driving force behind a range of other factors associated with anthropogenic disturbance (Luck [2007\)](#page-8-19). For instance, higher human densities result in more built-up environments and more infrastructure and hence further human-induced environmental change (Cincotta et al. [2000\)](#page-7-11). It is, therefore, probable that countries with higher population densities have a higher proportion of their protected land under intense human pressure. It is essential to clarify here, though, that although this reasoning may seem circular at frst—i.e., to use human densities to explain the levels in human footprint, when human densities are part of the index—, in reality it is not, because here we are interested in understanding human pressure within protected areas specifically. One would expect countries—including densely populated countries—to be able to keep their protected areas mostly free from intense human pressure, especially considering the demonstrated negative impact of the human footprint on biodiversity (Di Marco et al. [2018](#page-7-6); Tucker et al. [2018](#page-8-12); Watson and Venter [2019\)](#page-9-0).

Agricultural land expansion is also a known major driver of anthropogenic disturbance in natural environments (Brussaard et al. [2010;](#page-7-12) Rockström et al. [2017](#page-8-20)) and has been shown to affect biodiversity negatively (Brussaard et al. [2010](#page-7-12); Karp et al. [2012\)](#page-8-21)—e.g., through habitat loss and fragmentation (Ramankutty et al. [2018\)](#page-8-22). Reports from global and regional assessments suggest that it is not uncommon to have large areas of agricultural land within protected areas (Geldmann et al. [2019](#page-8-9); Anderson and Mammides [2020b\)](#page-7-9), even within strictly protected nature reserves. Considering the current scale of the agricultural activities across the planet (Molotoks et al. [2018\)](#page-8-23), it is probable that countries with more agricultural land have more of their protected areas under intense human pressure. It is also probable that else being equal, countries with smaller land area have a higher proportion of their protected land under intense human pressure because of increased competition for land resources (Smith et al. [2010](#page-8-24)).

Lastly, it is possible that human pressure within protected areas is linked to their type of protection status (Nelson and Chomitz [2011\)](#page-8-25). There are several types of protected areas, and although in many of them human activities and use of natural resources are strictly prohibited, in others they are allowed (Dudley  $2008$ ) in an effort to minimize the social impacts on the local communities and the resulting conficts (West and Brockington [2006;](#page-9-7) West et al. [2006;](#page-9-8) Dudley et al. [2018;](#page-7-14) Mammides [2020\)](#page-8-26). The International Union for Conservation of Nature (IUCN) recognizes six categories of protected areas, which refect a gradient of permissible human uses (Dudley [2008](#page-7-13)). For example, protected areas belonging to IUCN Categories V and VI permit multiple sustainable nonindustrial uses (Dudley [2008\)](#page-7-13), which are often prohibited in strict nature reserves and national parks, i.e., IUCN Categories Ia and II (Chape et al. [2005](#page-7-15); Leroux et al. [2010\)](#page-8-27). It is possible that countries with more of their protected land under stricter status have lower proportions of protected areas under intense human pressure.

In this study, I explore the factors that infuence most the levels of human pressure within the world's protected areas at the national level (Jones et al. [2018](#page-8-4)). In particular, I answer the following questions:

- a. Is the proportion of protected land under intense human pressure related to the country's level of economic development, measured using its per capita gross domestic product (GDP)?
- b. Do effective national governance and higher conservation spending lead to lower levels of human pressure within protected areas?
- c. Do higher human population densities lead to higher levels of human pressure within protected areas?
- d. Do countries with a higher percentage of agricultural land or smaller land area have a higher proportion of protected land under intense human pressure?
- e. Does a higher percentage of protected land under strict protection status lead to lower levels of human pressure within protected areas?

# **Materials and methods**

#### **Data collection**

I calculated each country's proportion of protected land under intense human pressure using the data provided by Jones et al. ([2018\)](#page-8-4). The data represent the percentage of protected land within each country—in terms of its total land area—under low and high human pressure (i.e., a mean human footprint value of  $<$  4 and  $\geq$  4, respectively). The data are based on the most recent version of the human footprint index from the year 2009 (Venter et al. [2016b](#page-9-3)). The human footprint index is a composite measure of the human impact across the planet, calculated using the following eight types of human pressure: (1) built-up areas; (2) intensively farmed crop land; (3) pasture land; (4) human population density; (5) night-time lights; (6) railways; (7) roads; (8) navigable waterways (Venter et al. [2016a](#page-8-1); Jones et al. [2018](#page-8-4)). These eight types of human pressure are frst standardized and weighted according to their relative impact (Sanderson et al. [2002](#page-8-28); Venter et al. [2016a](#page-8-1)) and then summed to provide a cumulative score of human pressure ranging from 0–50 (Venter et al. [2016a](#page-8-1); Jones et al. [2018\)](#page-8-4).

To calculate each country's proportion of protected land under intense human pressure, I divided each country's percentage of protected land with a mean human footprint value of≥4, by its total percentage of land covered by protected areas (Jones et al. [2018](#page-8-4)). For the purposes of the analysis here—and following the protocol of the Environmental Sustainability Index (Schmiedeknecht [2013](#page-8-29))—I excluded countries with a land area of less than  $5000 \text{ km}^2$  ( $n = 17$ ) and I also excluded countries that had less than 3% of the area covered by protected areas  $(n=15)$ . In total, there were 130 countries for which there were complete data for all the variables included in the analysis (the list of variables is presented below).

The data on conservation spending and government effectiveness were obtained from Waldron et al. [\(2013](#page-9-5)) and cover the years 2001–2008. Conservation spending represents the total amount of funds spent for conservation within each country during the above-mentioned period (Waldron et al. [2013\)](#page-9-5). These were either funds coming from large international and foreign donors, such as the World Bank and the Global Environment Facility Trust Fund, or from the countries' own budgets (Waldron et al. [2013\)](#page-9-5). Importantly, many of these funds concern investments in protected areas (Waldron et al. [2013\)](#page-9-5) and hence are relevant to the analysis presented here, which explores the factors associated with the levels of human pressure within protected areas.

The data on national governance are essentially a measure of each country's efectiveness in "formulating and imple-menting policies" (Waldron et al. [2013](#page-9-5)). Admittedly, governance is a multidimensional and complicated concept and, therefore, difficult to capture with one metric (Eklund and Cabeza [2017\)](#page-7-16). It is possible that the results of the analysis presented here were infuenced, at least partly, by the metric chosen to measure national governance. To assess the extent to which this was the case, I ran the analysis again using the national governance metric included in the Environmental Sustainability Index (Schmiedeknecht [2013\)](#page-8-29). The index was prepared jointly by institutions led by the Yale Center for Environmental Law and Policy and the Center for International Earth Science Information Network (CIESIN) of Columbia University, and is comprised of a series of metrics that aim at measuring each country's level of environmental sustainability (Schmiedeknecht [2013\)](#page-8-29). Among the metrics included, one is labeled "environmental governance" and its purpose is to measure the governance efectiveness of the countries specifcally regarding environmental matters. Hence, in addition to indicators such as corruption and rule of law, the metric includes indicators related also to the environment, such as memberships to IUCN organizations and Local Agenda 21 initiatives. Yet, regardless of which national governance metric was used in the analysis, the results were similar; therefore, here I present the results based on the metric compiled by Waldron et al. ([2013\)](#page-9-5), which was available for a few more countries (resulting in a higher sample size). The results based on the governance metric extracted from the Environmental Sustainability Index can be found in the supporting information (Table S1).

The data on the countries' human population densities, extent of agricultural land, and economic development level

were obtained from the World Bank; using the "WDI" package in R, I downloaded the following three indicators, which are also available online at [https://data.worldbank.org/indic](https://data.worldbank.org/indicator) [ator:](https://data.worldbank.org/indicator) (a) total human population size ("SP.POP.TOTL"); (b) percent of agricultural land ("AG.LND.AGRI.ZS"); and (c) gross domestic product per capita ("NY.GDP.PCAP.CD"). Following the approach of Waldron et al. ([2013](#page-9-5)), I used each indicator's average value between the years 2001–2008. To calculate each country's percentage of protected land under strict protection status I used the World Database of Protected Areas (WDPA; October 2018 version) available at <https://protectedplanet.net>. In line with the methods of Jones et al. ([2018](#page-8-4)), I classified protected areas  $\geq$  5 km<sup>2</sup> according to their IUCN category as follows: (a) strictly protected areas (Categories I and II), (b) non-strictly protected areas (Categories III–VI), and (c) areas with no IUCN category reported. Then, for each country, I estimated the percentage of protected land (in  $km^2$ ) belonging to the first class; a higher percentage corresponded to more protected land under strict protection status.

#### **Data analysis**

To identify the factors that are associated most with the proportion of protected land under intense human pressure, I ran a generalized linear mixed model (with a binomial error distribution), which included the following seven variables for each country: (1) total land area  $(km^2)$ ; (2) human population density (inhabitants/ $km^2$ ); (3) percentage of land area covered by agricultural land  $(\%)$ ; (4) economic development level, measured using the gross domestic product (GDP) per capita (current US\$/inhabitants); (5) conservation spending (million US\$); (6) national governance effectiveness; and (7) percentage of protected land under strict protection status (%). I log-transformed total land area, human population density, GDP per capita, and conservation spending to meet the assumptions of linearity and homoscedasticity, i.e., constant variance (Zuur et al. [2010\)](#page-9-9). Since it was possible that some of the independent variables used in the analysis were correlated with each other—leading to biased estimates (Akinwande et al. [2015](#page-7-17))—I examined whether collinearity was an issue by calculating each variable's variance infation factor (Akinwande et al. [2015\)](#page-7-17), using the *vif* function in the "car" package (Fox and Weisberg [2011](#page-7-18)). I used the continent of each country as a random efect to account for the possibility that countries within the same geographic region have similar characteristics (Lira-Noriega and Soberón [2015](#page-8-30)). However, it accounted for zero variance, and therefore I proceeded with a simpler generalized linear model (i.e., one with no random effects).

I used information-theory (Burnham and Anderson [2002](#page-7-19); Hegyi and Garamszegi [2011](#page-8-31)) and model averaging (Symonds and Moussalli [2011](#page-8-32); Grueber et al. [2011](#page-8-33)) to assess the relative importance of each independent variable. Specifically, I used the "MuMin" package (Bartoń [2019](#page-7-20)) in R to average all models with a  $\triangle$ AICc of less than 2.  $\triangle$ AICc is the difference between the AICc value of each model and that of the best model, i.e., the one with lowest AICc value (Grueber et al. [2011](#page-8-33)). The rationale behind the model averaging approach is that models with a  $\triangle$ AICc of < 2 are also plausible (Burnham and Anderson [2002](#page-7-19)) and consequently contain information that should be included in the results (Burnham and Anderson [2002](#page-7-19); Grueber et al. [2011\)](#page-8-33); this is achieved by averaging the regression coefficients of the selected models. I used the full-average method—also called the zero method (Grueber et al. [2011\)](#page-8-33)—because it is more appropriate for studies such as this one in which the aim of the analysis is to assess the relative importance of each independent variable (Grueber et al. [2011](#page-8-33)). For each variable, I calculated its unstandardized and standardized regression coefficients. I used the "rsq" package (Zhang [2018\)](#page-9-10) in R to calculate the pseudo- $R^2$  value of the best model and the pseudo- $R^2$  value of the model which contained all the variables selected during the averaging process. For each variable in the latter model, I also calculated its partial pseudo-*R*<sup>2</sup> value (Zhang [2018](#page-9-10)) to assess its relative contribution to the overall variance explained.

#### **Results**

The median proportion of protected land under intense human pressure within each country was 0.63 (ranging from 0.03 to 1.00), showing that in most countries more than half of their protected land has been already afected by intense human pressure. The variables in the best model—i.e., the model with the lowest AICc value—explained approximately two-thirds of the variance in the proportion of protected land under intense human pressure  $(R^2 = 0.61)$ . The model consisted of the following three variables: (1) human population density; (2) percentage of agricultural land; (3) total land area. Countries with lower human population densities, lower percentage of agricultural land, and a larger area had a lower proportion of their protected land under intense human pressure (Fig. [1\)](#page-4-0). The highest variance inflation factor—when the full model was tested (i.e., with all seven variables)—was 4.5, suggesting no substantial collinearity (Defries et al. [2010;](#page-7-21) Akinwande et al. [2015\)](#page-7-17). There were ten models with a  $\Delta AICc$  of less than 2 (Table S2), showing that some of the other four factors were also related to the levels of human pressure within protected areas. Indeed, the fnal averaged model included all seven original variables  $(R^2=0.63;$  Table [1\)](#page-5-0).

The relationship between conservation spending and the proportion of protected land under intense human pressure was negative (Table [1\)](#page-5-0), meaning that countries that had received and spent more funds on conservation had a lower

<span id="page-4-0"></span>**Fig. 1** Relationship between the countries' proportion of protected land under intense human pressure and **a** human population density, **b** percentage of agricultural land, and **c** total land area. The fgures are based on the best model (i.e., the model with the lowest AICc value; pseudo- $R^2$ =0.61)



Variable	Standardized		Unstandardized		Sum of weights	Partial pseudo- $R^2$
	Coefficient	Std. Error	Coefficient	Std. Error		
Population density	0.82	0.28	0.68	0.22	1.00	0.282
Agricultural land	0.41	0.22	0.02	0.01	1.00	0.163
Land area	$-0.19$	0.25	$-0.13$	0.17	0.48	0.060
Protection status	$-0.08$	0.17	$-0.30$	0.63	0.46	0.025
Conservation spending	$-0.18$	0.23	$-0.11$	0.14	0.69	0.015
GDP per capita	0.04	0.13	0.03	0.09	0.16	0.002
National governance	0.04	0.13	0.06	0.17	0.18	0.001

<span id="page-5-0"></span>**Table 1** Results of the model-averaging process showing the factors associated most with the proportion of protected land under intense human pressure at the national level

The table shows the averaged standardized and unstandardized regression coefficients along with their corresponding standard errors. The sum of weights of each variable and its partial pseudo- $R^2$  value are also provided

proportion of human pressure within their protected areas. Similarly, the relationship between human pressure and the protection status was also negative; the higher the percentage of protected land under strict status (i.e., IUCN categories I and II) the lower the proportion of intense human pressure. However, the amount of variance explained by these two variables was little and the size of their effect was small (Table [1](#page-5-0)). Conversely, most of the variation in the proportion of protected land under intense human pressure was explained by human population density, followed by the percent of agricultural land (Table [1](#page-5-0)). The countries' level of economic development and efectiveness of national governance related only weakly to their proportion of protected land under intense human pressure (Table [1\)](#page-5-0).

# **Discussion**

Several key messages can be extracted from the above analysis. First, there was no strong relationship between the proportion of protected land under intense human pressure and the efectiveness of national governance. This was true even when using the national governance metric of the Environmental Sustainability Index (Schmiedeknecht [2013\)](#page-8-29), which is more relevant to environmental issues (Table S1). Considering that several studies have reported that successful conservation depends on efective national and local governance (Bruner et al. [2001](#page-7-10); Smith et al. [2003](#page-8-13); Wright et al. [2007](#page-9-4); Lira-Noriega and Soberón [2015](#page-8-30)), one would expect that countries with stronger institutions are better able to curb intense human pressure within protected areas (Smith et al. [2003\)](#page-8-13). I found no strong evidence for this pattern; in retrospect, this is understandable, considering that multiple countries that tend to score high on national governance (Waldron et al. [2013](#page-9-5)), including environmental governance (Schmiedeknecht [2013\)](#page-8-29)—such as Netherlands, Austria, Belgium, and Israel—have also some of the highest proportions of protected land under intense human pressure  $(>0.94)$ . Conversely, other countries that tend to score low on national governance—such as Niger and Gabon—have very low proportions of protected land under intense human pressure. Moreover, it is also possible that the lack of a strong relationship between governance and human pressure was driven partly by the fact that local governance is likely to be more important for curbing human pressure within protected areas; however, the efectiveness of the local governance may not have been captured successfully by the national metrics.

There was also no signifcant relationship between the human pressure within protected areas and the economic development level of the countries (as measured using their per capita gross domestic product). This suggests that it may be possible—at least to a certain extent—to decouple economic development and human pressure within protected areas. In other words, it appears possible for countries to achieve their economic development goals and yet maintain much of their protected land under relatively low human pressure. Examples of such countries include Finland, Canada, Australia, and the United States. All four of them have less than 15% of their protected land under intense human pressure (Jones et al. [2018\)](#page-8-4) and yet they have some of largest economies in the world. Conversely, countries such as Ghana, Bangladesh, and Tanzania have more than 80% of their protected land under intense human pressure, despite their lower economic development level. This fnding, however, must be interpreted cautiously, because it is not possible to decipher from the analysis presented here whether the low levels of human pressure within the protected areas of those developed countries is actually a recent outcome, i.e., achieved following their economic development. It may be that economic development and human pressure within protected areas follow an inverted-*U* shaped relationship, similar to what would be expected by the Environmental Kuznets Curve theory (Mills and Waite [2009](#page-8-34); Carson [2010](#page-7-22)). In such case, developing countries would have to frst tolerate elevated levels of human pressure within their protected areas before they develop the means and the conditions to reduce them (Carson [2010\)](#page-7-22). That said, there appears to be no evidence for a Kuznets curve when one looks at the relationship between the proportion of protected land under intense human pressure and each country's GDP per capita (Figure S1). Similarly, other researchers have found no such pattern when examining, for instance, the relationship between national deforestation rates and economic development levels (Mills and Waite [2009](#page-8-34)).

The percentage of protected land under strict protection status was only weakly related to the levels of intense human pressure within protected areas (Table [1](#page-5-0)). As expected, countries with more of their protected land in IUCN Categories I and II had lower proportions of it under intense human pressure. However, this factor explained only a small amount of the variation in human pressure (Table [1](#page-5-0)). By now, several studies have demonstrated that the IUCN categories of the protected areas infuence only minorly their efectiveness (Leroux et al. [2010](#page-8-27); Jones et al. [2018](#page-8-4); Leberger et al. [2019](#page-8-35)). These fndings tend to be consistent regardless of the scale of the analysis or the metric used, e.g., rates of deforestation, changes in land-cover, or declines in animal population sizes (Leroux et al. [2010](#page-8-27); Coetzee et al. [2014](#page-7-4); Jones et al. [2018\)](#page-8-4).

Conservation spending was also expectedly negatively related to the human pressure within protected areas; higher conservation spending was associated with less protected land under intense human pressure. The relationship, though, was also very weak; conservation spending accounted for only a small amount of the variation in human pressure within protected areas (Table [1\)](#page-5-0). As interesting as this finding may be, it is important not to interpret it as if investments in conservation provide no substantial benefts in general. This fnding is specifc to human pressure within protected areas—as measured using the human footprint index—and, therefore, does not represent the overall relationship between spending and conservation. In fact, Waldron et al. [\(2017\)](#page-9-6) have already shown that conservation spending has a positive effect on biodiversity, since on average it reduces loss by approximately 29% per country (Waldron et al. [2017](#page-9-6)). However, conservation spending—at least at the current levels—does not seem to be signifcantly associated with lower levels of human pressure within protected areas. That said, it is probable that larger investments are needed (Miller et al. [2013\)](#page-8-17) to have a substantial effect on the levels of human pressure within the protected areas.

The three most important factors, which showed the strongest association with human pressure within protected areas, were human population densities, percent of agricultural land, and total land area of the country (Table [1](#page-5-0)). Countries with a higher population density, a larger percentage of agricultural land, and a smaller area had a higher proportion of their protected areas under intense human pressure (Fig. [1](#page-4-0)). This fnding is important, because it suggests that under a business-as-usual scenario, countries with those characteristics are likely to struggle to keep their protected areas relatively free from intense human pressure—regardless of their economic development level, efectiveness of national governance, and current investments in conservation. In addition, this fnding is important, because the population density of many countries—especially developing countries, many of which are located in biodiverse regions—is projected to increase during the next few decades (Cincotta et al. [2000;](#page-7-11) United Nations [2017](#page-8-36)). To complicate matters even further, the global demand for agricultural products will also increase—partly due to the growing human populations—resulting in further agricultural land expansion and intensifcation (Brussaard et al. [2010](#page-7-12); Tscharntke et al. [2012](#page-8-37); Ramankutty et al. [2018](#page-8-22))—in many cases within some of the same countries (Molotoks et al. [2018](#page-8-23)). The combination of the above-mentioned two drivers (i.e., the increase of human population densities and demand for agricultural products) will likely make it even more difficult for the governments of those countries to conserve their biodiversity within protected areas.

A possible concern regarding the above conclusions is that the results of the analysis refect the patterns in human pressure but not necessarily the patterns in biodiversity loss; human activities are not invariably incompatible with biodiversity conservation. Although this is true to a large extent, undoubtedly there is a strong negative relationship between the two (Newbold et al. [2016;](#page-8-38) Gray et al. [2016](#page-8-3)). For instance, studies have demonstrated that increases in human pressure—measured using the human footprint index—result in higher species extinction risks (Di Marco et al. [2018](#page-7-6), [2019](#page-7-7)) and reduced animal movement (Tucker et al. [2018\)](#page-8-12). Consequently, it is unlikely that countries will be able to conserve biodiversity successfully, and achieve their sustainability goals, without frst addressing the levels of human pressure within their protected areas (Jones et al. [2018](#page-8-4)).

Two caveats, though, to consider when interpreting the results of this study. First, the analysis does not capture the variation in human pressure within the countries. Therefore, it does not necessarily refect the factors that afect the efectiveness of protected areas at the local level (Leverington et al. [2010\)](#page-8-39). Yet, since conservation policies are often designed and implemented at the national level, understanding the factors that infuence human pressure at this level is undoubtedly essential. Second, the index used to measure human pressure (Jones et al. [2018](#page-8-4)), i.e., the human foot-print index (Venter et al. [2016b\)](#page-9-3), does not take into account several other types of human pressure, which also have a negative efect on biodiversity (Jones et al. [2018](#page-8-4)), e.g., pollution, overexploitation, and invasive species (Jones et al. [2018\)](#page-8-4). Although these other types are likely to correlate with the pressures already included in the index (Jones et al. [2018](#page-8-4); Bowler et al. [2020](#page-7-23)), it is not possible to know to what extent the factors identifed here relate to those pressures. For example, it may be that national governance infuences strongly the levels of pollution within protected areas.

## **Conclusion**

Although studies often cite poor governance and lack of resources as two of the major reasons behind the inefectiveness of protected areas, the fndings presented here suggest that human population densities—and to a lesser degree the extent of agriculture—are probably more important in determining intense human pressure within protected areas. Countries—especially smaller area countries—aiming at reducing human pressure within their protected areas need to address high human population densities and to invest in efficient agricultural methods and policies, so that less land is required to meet the rising global demand for agricultural products (Ramankutty et al. [2018\)](#page-8-22). Protected areas represent one of the most important conservation strategies for addressing the current rates of global biodiversity loss (Visconti et al. [2019](#page-9-1)) and for attaining the Sustainable Development Goals (Dudley et al. [2017\)](#page-7-8); hence, it is imperative we understand and address the factors driving the levels of human pressure within the world's protected areas.

**Acknowledgements** I am thankful to all the researchers who have made available the data used in this study. Specifcally, I would like to thank Kendall R. Jones and his colleagues for the data on the human pressure within protected areas and Anthony Waldron and his colleagues for the data on conservation spending and governance. I am also thankful to the researchers at the World Bank, the United Nations Environment World Conservation Monitoring Center, the Yale Center for Environmental Law and Policy, and the Center for International Earth Science Information Network (CIESIN) of Columbia University for their commitment in compiling the related datasets and for making them accessible to researchers across the globe. Lastly, I would like to thank Francesco Martini and two anonymous reviewers for their highly constructive and useful feedback.

## **References**

- <span id="page-7-17"></span>Akinwande MO, Dikko HG, Samson A (2015) Variance infation factor: as a condition for the inclusion of suppressor variable(s) in regression analysis. Open J Stat. [https://doi.org/10.4236/](https://doi.org/10.4236/ojs.2015.57075) [ojs.2015.57075](https://doi.org/10.4236/ojs.2015.57075)
- <span id="page-7-5"></span>Anderson E, Mammides C (2020a) The role of protected areas in mitigating human impact in the world's last wilderness areas. Ambio 49:434–441.<https://doi.org/10.1007/s13280-019-01213-x>
- <span id="page-7-9"></span>Anderson E, Mammides C (2020b) Changes in land-cover within high nature value farmlands inside and outside Natura 2000 sites in Europe: a preliminary assessment. Ambio. [https://doi.](https://doi.org/10.1007/s13280-020-01330-y) [org/10.1007/s13280-020-01330-y](https://doi.org/10.1007/s13280-020-01330-y)
- <span id="page-7-20"></span>Bartoń K (2019) MuMIn: multi-model inference. R package version 1.43.6. <https://CRAN.R-project.org/package=MuMIn>
- <span id="page-7-23"></span>Bowler DE, Bjorkman AD, Dornelas M et al (2020) Mapping human pressures on biodiversity across the planet uncovers anthropogenic threat complexes. People Nature. [https://doi.org/10.1002/](https://doi.org/10.1002/pan3.10071) [pan3.10071](https://doi.org/10.1002/pan3.10071)
- <span id="page-7-10"></span>Bruner AG, Gullison RE, Rice RE, Da Fonseca GAB (2001) Efectiveness of parks in protecting tropical biodiversity. Science 291:125– 128.<https://doi.org/10.1126/science.291.5501.125>
- <span id="page-7-12"></span>Brussaard L, Caron P, Campbell B et al (2010) Reconciling biodiversity conservation and food security: scientifc challenges for a new agriculture. Curr Opin Environ Sustain 2:34–42. [https://doi.](https://doi.org/10.1016/j.cosust.2010.03.007) [org/10.1016/j.cosust.2010.03.007](https://doi.org/10.1016/j.cosust.2010.03.007)
- <span id="page-7-19"></span>Burnham KP, Anderson DR (2002) Model selection and multimodel inference: a practical information-theoretic approach. Springer-Verlag, New York
- <span id="page-7-2"></span>Butchart SHM, Walpole M, Collen B et al (2010) Global biodiversity: indicators of recent declines. Science 328:1164–1168. [https://doi.](https://doi.org/10.1126/science.1187512) [org/10.1126/science.1187512](https://doi.org/10.1126/science.1187512)
- <span id="page-7-22"></span>Carson RT (2010) The environmental Kuznets curve: seeking empirical regularity and theoretical structure. Rev Environ Econ Policy 4:3–23.<https://doi.org/10.1093/reep/rep021>
- <span id="page-7-15"></span>Chape S, Harrison J, Spalding M, Lysenko I (2005) Measuring the extent and efectiveness of protected areas as an indicator for meeting global biodiversity targets. Philos Trans R Soc B Biol Sci 360:443–455. <https://doi.org/10.1098/rstb.2004.1592>
- <span id="page-7-11"></span>Cincotta RP, Wisnewski J, Engelman R (2000) Human population in the biodiversity hotspots. Nature 404:990–992. [https://doi.](https://doi.org/10.1038/35010105) [org/10.1038/35010105](https://doi.org/10.1038/35010105)
- <span id="page-7-4"></span>Coetzee BWT, Gaston KJ, Chown SL (2014) Local Scale comparisons of biodiversity as a test for global protected area ecological performance: a meta-analysis. PLoS ONE 9:e105824. [https://doi.](https://doi.org/10.1371/journal.pone.0105824) [org/10.1371/journal.pone.0105824](https://doi.org/10.1371/journal.pone.0105824)
- <span id="page-7-0"></span>Corlett RT (2015) The anthropocene concept in ecology and conservation. Trends Ecol Evol 30:36–41. [https://doi.org/10.1016/j.](https://doi.org/10.1016/j.tree.2014.10.007) [tree.2014.10.007](https://doi.org/10.1016/j.tree.2014.10.007)
- <span id="page-7-21"></span>Defries RS, Rudel T, Uriarte M, Hansen M (2010) Deforestation driven by urban population growth and agricultural trade in the twentyfrst century. Nat Geosci 3:178–181. [https://doi.org/10.1038/ngeo7](https://doi.org/10.1038/ngeo756) [56](https://doi.org/10.1038/ngeo756)
- <span id="page-7-7"></span>Di Marco M, Ferrier S, Harwood TD et al (2019) Wilderness areas halve the extinction risk of terrestrial biodiversity. Nature. [https](https://doi.org/10.1038/s41586-019-1567-7) [://doi.org/10.1038/s41586-019-1567-7](https://doi.org/10.1038/s41586-019-1567-7)
- <span id="page-7-6"></span>Di Marco M, Venter O, Possingham HP, Watson JEM (2018) Changes in human footprint drive changes in species extinction risk. Nat Commun 9:4621.<https://doi.org/10.1038/s41467-018-07049-5>
- <span id="page-7-13"></span>Dudley N (2008) Guidelines for applying protected area management categories. IUCN, Gland
- <span id="page-7-8"></span>Dudley N, Ali N, Kettunen M, MacKinnon K (2017) Protected areas and the sustainable development goals. Parks 23:9–12. [https://doi.](https://doi.org/10.2305/IUCN.CH.2017.PARKS-23-2ND.en) [org/10.2305/IUCN.CH.2017.PARKS-23-2ND.en](https://doi.org/10.2305/IUCN.CH.2017.PARKS-23-2ND.en)
- <span id="page-7-14"></span>Dudley N, Jonas H, Nelson F et al (2018) The essential role of other efective area-based conservation measures in achieving big bold conservation targets. Global Ecol Conserv 15:e00424. [https://doi.](https://doi.org/10.1016/j.gecco.2018.e00424) [org/10.1016/j.gecco.2018.e00424](https://doi.org/10.1016/j.gecco.2018.e00424)
- <span id="page-7-16"></span>Eklund J, Cabeza M (2017) Quality of governance and efectiveness of protected areas: crucial concepts for conservation planning. Ann N Y Acad Sci 1399:27–41.<https://doi.org/10.1111/nyas.13284>
- <span id="page-7-1"></span>Ellis EC, Klein Goldewijk K, Siebert S et al (2010) Anthropogenic transformation of the biomes, 1700 to 2000. Glob Ecol Biogeogr 19:589–606.<https://doi.org/10.1111/j.1466-8238.2010.00540.x>
- <span id="page-7-18"></span>Fox J, Weisberg S (2011) An R companion to applied regression, 2nd edn. Sage, Thousand Oaks
- <span id="page-7-3"></span>Fuller RA, McDonald-Madden E, Wilson KA et al (2010) Replacing underperforming protected areas achieves better conservation outcomes. Nature 466:365–367.<https://doi.org/10.1038/nature09180>
- <span id="page-8-0"></span>Gaston KJ, Jackson SF, Cantú-Salazar L, Cruz-Piñón G (2008) The ecological performance of protected areas. Annu Rev Ecol Evol Syst 39:93–113. [https://doi.org/10.1146/annurev.ecolsys.39.11070](https://doi.org/10.1146/annurev.ecolsys.39.110707.173529) [7.173529](https://doi.org/10.1146/annurev.ecolsys.39.110707.173529)
- <span id="page-8-5"></span>Geldmann J, Coad L, Barnes MD et al (2018) A global analysis of management capacity and ecological outcomes in terrestrial protected areas. Conserv Lett 11:1–10.<https://doi.org/10.1111/conl.12434>
- <span id="page-8-9"></span>Geldmann J, Manica A, Burgess ND et al (2019) A global-level assessment of the efectiveness of protected areas at resisting anthropogenic pressures: supplementary material. Proc Natl Acad Sci USA.<https://doi.org/10.1073/pnas.1908221116>
- <span id="page-8-3"></span>Gray CL, Hill SLL, Newbold T et al (2016) Local biodiversity is higher inside than outside terrestrial protected areas worldwide. Nat Commun 7:12306.<https://doi.org/10.1038/ncomms12306>
- <span id="page-8-33"></span>Grueber CE, Nakagawa S, Laws RJ, Jamieson IG (2011) Multimodel inference in ecology and evolution: challenges and solutions. J Evol Biol 24:699–711. [https://doi.org/10.111](https://doi.org/10.1111/j.1420-9101.2010.02210.x) [1/j.1420-9101.2010.02210.x](https://doi.org/10.1111/j.1420-9101.2010.02210.x)
- <span id="page-8-31"></span>Hegyi G, Garamszegi LZ (2011) Using information theory as a substitute for stepwise regression in ecology and behavior. Behav Ecol Sociobiol 65:69–76. <https://doi.org/10.1007/s00265-010-1036-7>
- <span id="page-8-4"></span>Jones KR, Venter O, Fuller RA et al (2018) One-third of global protected land is under intense human pressure. Science 360:788– 791.<https://doi.org/10.1126/science.aap9565>
- <span id="page-8-7"></span>Joppa LN, Pfaf A (2011) Global protected area impacts. Proc R Soc B Biol Sci 278:1633–1638.<https://doi.org/10.1098/rspb.2010.1713>
- <span id="page-8-21"></span>Karp DS, Rominger AJ, Zook J et al (2012) Intensive agriculture erodes β-diversity at large scales. Ecol Lett 15:963–970. [https://](https://doi.org/10.1111/j.1461-0248.2012.01815.x) [doi.org/10.1111/j.1461-0248.2012.01815.x](https://doi.org/10.1111/j.1461-0248.2012.01815.x)
- <span id="page-8-15"></span>Leader-Williams N, Albon SD (1988) Allocation of resources for conservation. Nature 336:533–535.<https://doi.org/10.1038/336533a0>
- <span id="page-8-35"></span>Leberger R, Rosa IMD, Guerra CA et al (2019) Global patterns of forest loss across IUCN categories of protected areas. Biol Cons. <https://doi.org/10.1016/j.biocon.2019.108299>
- <span id="page-8-27"></span>Leroux SJ, Krawchuk MA, Schmiegelow F et al (2010) Global protected areas and IUCN designations: do the categories match the conditions? Biol Cons 143:609–616. [https://doi.org/10.1016/j.](https://doi.org/10.1016/j.biocon.2009.11.018) [biocon.2009.11.018](https://doi.org/10.1016/j.biocon.2009.11.018)
- <span id="page-8-39"></span>Leverington F, Costa KL, Pavese H et al (2010) A global analysis of protected area management efectiveness. Environ Manage 46:685–698.<https://doi.org/10.1007/s00267-010-9564-5>
- <span id="page-8-30"></span>Lira-Noriega A, Soberón J (2015) The relationship among biodiversity, governance, wealth, and scientifc capacity at a country level: disaggregation and prioritization. Ambio 44:391–400. [https://doi.](https://doi.org/10.1007/s13280-014-0581-0) [org/10.1007/s13280-014-0581-0](https://doi.org/10.1007/s13280-014-0581-0)
- <span id="page-8-19"></span>Luck GW (2007) A review of the relationships between human population density and biodiversity. Biol Rev 82:607–645. [https://doi.](https://doi.org/10.1111/j.1469-185X.2007.00028.x) [org/10.1111/j.1469-185X.2007.00028.x](https://doi.org/10.1111/j.1469-185X.2007.00028.x)
- <span id="page-8-11"></span>Mammides C (2018) Do satellite-derived data on forest loss correlate with indices of small-scale logging measured in the feld? Afr J Ecol 56:390–394.<https://doi.org/10.1111/aje.12434>
- <span id="page-8-26"></span>Mammides C (2020) Evidence from eleven countries in four continents suggests that protected areas are not associated with higher poverty rates. Biol Cons 241:108353. [https://doi.org/10.1016/j.](https://doi.org/10.1016/j.biocon.2019.108353) [biocon.2019.108353](https://doi.org/10.1016/j.biocon.2019.108353)
- <span id="page-8-16"></span>McCarthy DP, Donald PF, Scharlemann JPW et al (2012) Financial costs of meeting global biodiversity conservation targets: current spending and unmet needs. Science 338:946–949. [https://doi.](https://doi.org/10.1126/science.1229803) [org/10.1126/science.1229803](https://doi.org/10.1126/science.1229803)
- <span id="page-8-18"></span>McKee JK, Sciulli PW, Fooce CD, Waite TA (2004) Forecasting global biodiversity threats associated with human population growth. Biol Cons 115:161–164. [https://doi.org/10.1016/S0006](https://doi.org/10.1016/S0006-3207(03)00099-5) [-3207\(03\)00099-5](https://doi.org/10.1016/S0006-3207(03)00099-5)
- <span id="page-8-17"></span>Miller DC, Agrawal A, Roberts JT (2013) Biodiversity, governance, and the allocation of international aid for conservation. Conserv Lett 6:12–20.<https://doi.org/10.1111/j.1755-263X.2012.00270.x>
- <span id="page-8-34"></span>Mills JH, Waite TA (2009) Economic prosperity, biodiversity conservation, and the environmental Kuznets curve. Ecol Econ 68:2087– 2095. <https://doi.org/10.1016/j.ecolecon.2009.01.017>
- <span id="page-8-23"></span>Molotoks A, Stehfest E, Doelman J et al (2018) Global projections of future cropland expansion to 2050 and direct impacts on biodiversity and carbon storage. Glob Change Biol 24:5895–5908. [https://](https://doi.org/10.1111/gcb.14459) [doi.org/10.1111/gcb.14459](https://doi.org/10.1111/gcb.14459)
- <span id="page-8-6"></span>Nagendra H (2008) Do parks work? Impact of protected areas on land cover clearing. Ambio 37:330–337. [https://doi.](https://doi.org/10.1579/06-R-184.1) [org/10.1579/06-R-184.1](https://doi.org/10.1579/06-R-184.1)
- <span id="page-8-25"></span>Nelson A, Chomitz KM (2011) Efectiveness of strict vs. multiple use protected areas in reducing tropical forest fres: a global analysis using matching methods. PLoS ONE 6:e22722. [https://doi.](https://doi.org/10.1371/journal.pone.0022722) [org/10.1371/journal.pone.0022722](https://doi.org/10.1371/journal.pone.0022722)
- <span id="page-8-38"></span>Newbold T, Hudson LN, Arnell AP et al (2016) Has land use pushed terrestrial biodiversity beyond the planetary boundary? A global assessment. Science 353:288–291. [https://doi.org/10.1126/scien](https://doi.org/10.1126/science.aaf2201) [ce.aaf2201](https://doi.org/10.1126/science.aaf2201)
- <span id="page-8-2"></span>Newbold T, Hudson LN, Hill SLL et al (2015) Global effects of land use on local terrestrial biodiversity. Nature 520:45–50. [https://doi.](https://doi.org/10.1038/nature14324) [org/10.1038/nature14324](https://doi.org/10.1038/nature14324)
- <span id="page-8-10"></span>Peres CA, Barlow J, Laurance WF (2006) Detecting anthropogenic disturbance in tropical forests. Trends Ecol Evol 21:227–229. [https](https://doi.org/10.1016/j.tree.2006.03.007) [://doi.org/10.1016/j.tree.2006.03.007](https://doi.org/10.1016/j.tree.2006.03.007)
- <span id="page-8-22"></span>Ramankutty N, Mehrabi Z, Waha K et al (2018) Trends in global agricultural land use: implications for environmental health and food security. Annu Rev Plant Biol 69:789–815. [https://doi.](https://doi.org/10.1146/annurev-arplant-042817-040256) [org/10.1146/annurev-arplant-042817-040256](https://doi.org/10.1146/annurev-arplant-042817-040256)
- <span id="page-8-20"></span>Rockström J, Williams J, Daily G et al (2017) Sustainable intensifcation of agriculture for human prosperity and global sustainability. Ambio 46:4–17.<https://doi.org/10.1007/s13280-016-0793-6>
- <span id="page-8-28"></span>Sanderson EW, Jaiteh M, Levy MA et al (2002) The human footprint and the last of the wild. Bioscience 52:891. [https://doi.](https://doi.org/10.1641/0006-3568(2002)052[0891:THFATL]2.0.CO;2) [org/10.1641/0006-3568\(2002\)052\[0891:THFATL\]2.0.CO;2](https://doi.org/10.1641/0006-3568(2002)052[0891:THFATL]2.0.CO;2)
- <span id="page-8-29"></span>Schmiedeknecht MH (2013) Environmental Sustainability index: encyclopedia of corporate social responsibility. In: Idowu SO, Capaldi N, Zu L, Gupta A Das (Eds). Springer, Berlin, pp 1017–1024
- <span id="page-8-24"></span>Smith P, Gregory PJ, van Vuuren D et al (2010) Competition for land. Philos Trans R Soc B Biol Sci 365:2941–2957. [https://doi.](https://doi.org/10.1098/rstb.2010.0127) [org/10.1098/rstb.2010.0127](https://doi.org/10.1098/rstb.2010.0127)
- <span id="page-8-13"></span>Smith RJ, Muir RDJ, Walpole MJ et al (2003) Governance and the loss of biodiversity. Nature 426:67–70. [https://doi.org/10.1038/](https://doi.org/10.1038/nature02025) [nature02025](https://doi.org/10.1038/nature02025)
- <span id="page-8-14"></span>Smith RJ, Walpole MJ (2005) Should conservationists pay more attention to corruption? Oryx 39:251–256. [https://doi.org/10.1017/](https://doi.org/10.1017/S0030605305000608) [S0030605305000608](https://doi.org/10.1017/S0030605305000608)
- <span id="page-8-8"></span>Spracklen BD, Kalamandeen M, Galbraith D et al (2015) A global analysis of deforestation in moist tropical forest protected areas. PLoS ONE 10:e0143886. <https://doi.org/10.1371/journal.pone.0143886>
- <span id="page-8-32"></span>Symonds MRE, Moussalli A (2011) A brief guide to model selection, multimodel inference and model averaging in behavioural ecology using Akaike's information criterion. Behav Ecol Sociobiol 65:13–21. <https://doi.org/10.1007/s00265-010-1037-6>
- <span id="page-8-37"></span>Tscharntke T, Clough Y, Wanger TC et al (2012) Global food security, biodiversity conservation and the future of agricultural intensifcation. Biol Cons 151:53–59. [https://doi.org/10.1016/j.bioco](https://doi.org/10.1016/j.biocon.2012.01.068) [n.2012.01.068](https://doi.org/10.1016/j.biocon.2012.01.068)
- <span id="page-8-12"></span>Tucker MA, Böhning-Gaese K, Fagan WF et al (2018) Moving in the anthropocene: global reductions in terrestrial mammalian movements. Science 359:466–469. [https://doi.org/10.1126/scien](https://doi.org/10.1126/science.aam9712) [ce.aam9712](https://doi.org/10.1126/science.aam9712)
- <span id="page-8-36"></span>United Nations (2017) World population prospects: The 2017 revision, key fndings and advance tables. Department of Economic and Social Afair, Population Division, ESA/P/WP/248
- <span id="page-8-1"></span>Venter O, Sanderson EW, Magrach A et al (2016a) Sixteen years of change in the global terrestrial human footprint and implications

for biodiversity conservation. Nat Commun 7:1-11. [https://doi.](https://doi.org/10.1038/ncomms12558) [org/10.1038/ncomms12558](https://doi.org/10.1038/ncomms12558)

- <span id="page-9-3"></span>Venter O, Sanderson EW, Magrach A et al (2016b) Global terrestrial Human Footprint maps for 1993 and 2009. Sci Data 3:160067. <https://doi.org/10.1038/sdata.2016.67>
- <span id="page-9-1"></span>Visconti P, Butchart SHM, Brooks TM, et al (2019) Protected area targets post-2020. Science 364:eaav6886. doi: 10.1126/science. aav6886
- <span id="page-9-6"></span>Waldron A, Miller DC, Redding D et al (2017) Reductions in global biodiversity loss predicted from conservation spending. Nature 551:364–367. <https://doi.org/10.1038/nature24295>
- <span id="page-9-5"></span>Waldron A, Mooers AO, Miller DC et al (2013) Targeting global conservation funding to limit immediate biodiversity declines. Proc Natl Acad Sci 110:12144–12148. [https://doi.org/10.1073/](https://doi.org/10.1073/pnas.1221370110) [pnas.1221370110](https://doi.org/10.1073/pnas.1221370110)
- <span id="page-9-2"></span>Watson JEM, Dudley N, Segan DB, Hockings M (2014) The performance and potential of protected areas. Nature 515:67–73. [https](https://doi.org/10.1038/nature13947) [://doi.org/10.1038/nature13947](https://doi.org/10.1038/nature13947)
- <span id="page-9-0"></span>Watson JEM, Venter O (2019) Mapping the continuum of humanity's footprint on land. One Earth 1:175–180. [https://doi.org/10.1016/j.](https://doi.org/10.1016/j.oneear.2019.09.004) [oneear.2019.09.004](https://doi.org/10.1016/j.oneear.2019.09.004)
- <span id="page-9-7"></span>West P, Brockington D (2006) An anthropological perspective on some unexpected consequences of protected areas. Conserv Biol 20:609–616.<https://doi.org/10.1111/j.1523-1739.2006.00432.x>
- <span id="page-9-8"></span>West P, Igoe J, Brockington D (2006) Parks and peoples: the social impact of protected areas. Annu Rev Anthropol 35:251–277. [https](https://doi.org/10.1146/annurev.anthro.35.081705.123308) [://doi.org/10.1146/annurev.anthro.35.081705.123308](https://doi.org/10.1146/annurev.anthro.35.081705.123308)
- <span id="page-9-4"></span>Wright SJ, Sanchez-Azofeifa GA, Portillo-Quintero C, Davies D (2007) Poverty and corruption compromise tropical forest reserves. Ecol Appl 17:1259–1266.<https://doi.org/10.1890/06-1330.1>
- <span id="page-9-10"></span>Zhang D (2018) rsq: R-squared and related measures. R package version 1.1. <https://CRAN.R-project.org/package=rsq>
- <span id="page-9-9"></span>Zuur AF, Ieno EN, Elphick CS (2010) A protocol for data exploration to avoid common statistical problems. Methods Ecol Evol 1:3–14. <https://doi.org/10.1111/j.2041-210X.2009.00001.x>

**Publisher's Note** Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional afliations.