## ORIGINAL ARTICLE





# Diversity of freshwater hyphomycetes associated with leaf litter of Calophyllum brasiliense in streams of the semiarid region of Brazil

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Received: 30 October 2018 / Revised: 18 May 2019 /Accepted: 24 May 2019  $\copyright$  German Mycological Society and Springer-Verlag GmbH Germany, part of Springer Nature 2019

### Abstract

This study focused on freshwater hyphomycetes associated with submerged decaying leaves of Calophyllum brasiliense in three streams in the semiarid region of Brazil. Leaves were enclosed in litter bags and deployed into streams on two occasions (dry and wet seasons). Samples of leaf litter were collected every two months (November 2013 to January 2015). Unlike other studies, we specifically targeted ecologically distinct groups of freshwater hyphomycetes by using two methods to induce fungal sporulation: submerged incubation (SI) of leaf litter on an orbital shaker and incubation in moist chambers (MC). We aimed to analyze and compare the structure of freshwater hyphomycete communities in the streams using both methods, detect possible successional patterns, and evaluate if environmental variables have influenced fungal diversity and sporulation rates. Sixty-nine taxa of freshwater hyphomycetes were observed. Of these, 26 were found under SI and 56 in MC. We observed large differences in fungal communities recovered by SI vs. MC that demonstrates the importance of using several methodological approaches to maximize the number of taxa recovered in ecological studies of litter-associated fungi. The highest sporulation rates under SI were observed during the wet season. Results of community ordination suggested that environmental variables affected the structure of fungal communities; for example, water velocity showed a positive effect on fungal diversity, while higher oxygen availability was associated with lower diversity. This study advances our understanding of the freshwater hyphomycete communities in tropical streams where ecological studies of aquatic fungi are still uncommon.

Keywords Biodiversity · Ingoldian fungi · Sporulation · Streams · Tropical fungi

Section Editor: Dominik Begerow

Electronic supplementary material The online version of this article (<https://doi.org/10.1007/s11557-019-01501-6>) contains supplementary material, which is available to authorized users.

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## Introduction

Freshwater fungi are important decomposers of allochthonous plant litter entering rivers and streams from the riparian zone (Sridhar et al. [2010;](#page-13-0) Ferreira et al. [2014](#page-12-0)). Fungi mediate carbon, energy, and nutrient flow in freshwater ecosystems including the increase in the quality of decomposing leaf detritus, which makes it more palatable and nutritious food source to aquatic invertebrate consumers (Gulis et al. [2006\)](#page-12-0). In aquatic ecosystems, successional studies of fungal communities have been conducted on both submerged wood (Shearer and Webster [1991](#page-13-0); Ho et al. [2002](#page-12-0)) and leaves (Shearer and Webster [1985;](#page-13-0) Gessner et al. [1993;](#page-12-0) Schoenlein-Crusius and Milanez [1998;](#page-13-0) Newman et al. [2015\)](#page-13-0). On submerged leaves, both chytridiomycetes and zygomycetes can be detected, but the fungal communities are dominated by ascomycetes and basidiomycetes (Schoenlein-Crusius and Milanez [1989;](#page-13-0) Gessner et al. [1993](#page-12-0)).

Freshwater hyphomycetes include diverse aquatic fungi found in continental aquatic ecosystems that, as a group, have a worldwide distribution (Shearer et al. [2007\)](#page-13-0). Freshwater hyphomycetes encompass three ecological groups that have been studied for decades (Hyde et al. [2015](#page-12-0)). The most recognized group is called Ingoldian fungi in honor of Dr. Ingold for his pioneering contribution to their studies (Descals et al. [1977\)](#page-12-0). They represent the true aquatic hyphomycetes due to their unique adaptations to stream and river environments, such as tetraradial or sigmoid shape of conidia and fast germination that facilitate underwater dispersal and colonization of submerged plant litter. The aero-aquatic hyphomycetes display mycelial growth on submerged substrates; however, even though sporulation typically does not occur underwater, the conidia demonstrate adaptations to water dispersal (Fisher [1977\)](#page-12-0). Species of facultative aquatic hyphomycetes can sporulate below or above water but typically lack specific adaptation for water dispersal; they represent a continuous group with terrestrial hyphomycetes (with conidia often showing thick walls, i.e., being dematiaceous (Descals and Moralejo [2001\)](#page-12-0)).

The structure of freshwater hyphomycete communities can be assessed by a variety of techniques that are aimed to induce sporulation: incubation of submerged plant material on a shaker or in bubble chambers simulating stream conditions or, if fungi produce spores in air, moist chamber incubation (Graça et al. [2016\)](#page-12-0). Submerged incubation of plant material, mainly leaves, yields a large number of Ingoldian fungi or aquatic hyphomycetes (Chan et al. [2000](#page-12-0); Ghate and Sridhar [2016\)](#page-12-0), while incubation of leaf litter or wood in moist chambers reveals mainly facultative aquatic hyphomycetes, aeroaquatic hyphomycetes, and freshwater ascomycetes (Sridhar et al. [2010;](#page-13-0) Barbosa et al. [2013](#page-11-0)) that are quite diverse in the Neotropics (Castañeda-Ruiz et al. [2016](#page-12-0)).

Several studies have indicated that environmental variables, such as water temperature, oxygen concentration, turbulence, and pH, may influence the diversity, activity, and sporulation of freshwater hyphomycetes (Webster [1975](#page-13-0); Hissy et al. [1992;](#page-12-0) Chauvet and Suberkroop [1998](#page-12-0); Rajashekhar and Kaveriappa [2000](#page-13-0); Medeiros et al. [2009](#page-13-0)). The effects of dissolved inorganic nutrients (nitrogen and

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phosphorus), which are often elevated in rivers affected by human activity, have recently received considerable attention (Gulis and Suberkropp [2003](#page-12-0); Gulis et al. [2008](#page-12-0); Ferreira et al. [2015;](#page-12-0) Kominoski et al. [2015\)](#page-12-0). Another important factor affecting aquatic fungal communities is the diversity of plant litter and riparian vegetation (Gulis [2001;](#page-12-0) Chauvet et al. [2016\)](#page-12-0). For this study, Calophyllum brasiliense Cambess was chosen due to its high abundance and ecological importance in the riparian corridors in Chapada Diamantina including Rio de Contas.

In this study, we investigated communities of freshwater hyphomycetes associated with submerged decomposing leaves of C. brasiliense in three streams of the Rio de Contas basin. Unlike other studies, we specifically targeted ecologically distinct groups of freshwater hyphomycetes by using two methods to induce fungal sporulation: submerged incubation (SI) of leaf litter on an orbital shaker and incubation in moist chambers (MC). We aimed to (i) analyze and compare the structure of fungal communities in the experimental streams using both methods; (ii) detect possible successional patterns in the structure of freshwater hyphomycete communities; and (iii) evaluate if environmental variables (water temperature, pH, dissolved oxygen, water velocity, and total nitrogen and phosphorus) have influenced the diversity and sporulation rates of fungal communities.

# Materials and methods

## Study area

The present study was conducted in three streams of the Rio de Contas basin: Rio de Contas (RC), Patricinho (P1), and Patricio (P2), located in Piatã, Serra da Tromba, Chapada Diamantina, a semiarid region in the northeastern Brazil (Table 1; Fig. S1). The RC has predominantly grassland vegetation, some trees, shrubs, and lianas in the riparian corridor, but it does not have a closed canopy; P1 has less grassland vegetation in the riparian corridor, while P2 has even more trees and nearly closed canopy (Table S2).





#### <span id="page-2-0"></span>Sampling and laboratory analyses

In September 2013 and May 2014, we visited the study area to collect leaves of Calophyllum brasiliense, an evergreen tree. Green leaves were collected directly from trees in the riparian forest for two of our experiments. Experiment 1 (E1) was carried out from November 2013 to May 2014 and experiment 2 (E2) from July 2014 to January 2015, approximately 8 month duration each to allow sufficient leaf litter decomposition and fungal succession. The rainfall in the study area was approximately two times higher than historical means during December to April period (2013–2014), and the collections made from January 2014 to May 2014 and in January 2015 were categorized as "wet season" while the rest of the samples were considered as "dry season" (Fig. 1). The rainfall and temperature data were obtained from the Instituto Nacional de Metereologia (INMET).

For both experiments, five leaves were enclosed in each fine-mesh litter bag  $(30 \times 30$ -cm, 0.5-mm mesh size) which were then deployed in streams by tying with nylon line to trees or other riparian vegetation. Every two months, we collected six litter bags from each stream (6 litter bags  $\times$  3 streams  $\times$  4 collection dates  $\times$  2 experiments for the total of 144 bags and 720 leaves). Litter bags were placed in plastic bags and transported to the Laboratory of Mycology (LAMIC), Universidade Estadual de Feira de Santana in a cooler. At each sampling date, we also collected water samples (1 L) that were later analyzed for total nitrogen and phosphorus according to Koroleff ([1976\)](#page-13-0) in the Laboratory of Environmental Microbiology, Universidade Federal da Bahia (UFBA). The environmental variables: water temperature, pH, and dissolved oxygen, were measured in the field on each sampling date with portable equipment (Horiba U-50) and the water

velocity with the flow meter (Global Water FP111 flow probe).

In the laboratory, leaf litter samples were processed using two methodological approaches: to induce sporulation of freshwater fungi: (1) submerged incubation and (2) incubation in moist chambers.

Submerged incubation following Bärlocher ([2005](#page-11-0)): on each sampling date, leaves from three litter bags per stream were rinsed and twenty 12-mm-diameter discs were cut with a cork borer from each litter bag. These discs were incubated for 48 h at 18–20 °C on a shaker (100 rpm) in 100-mL Erlenmeyer flasks containing 30 mL of sterile distilled water to induce fungal sporulation. Spore suspensions were filtered through membrane filters (5-μm pore size, Millipore), and the filters were then mounted on slides with cotton blue in lactic acid. In addition, the leaf discs were also mounted on slides and examined under compound microscope. All spores on filters and slides with leaf discs were identified and counted.

Incubation in moist chambers following Castañeda-Ruiz et al. [\(2016](#page-12-0)): leaf litter samples (from three litter bags per stream) were rinsed, placed in Petri dishes (moist chambers) and incubated in a polystyrene box with sterile water plus glycerol for 30 days. Fungal structures were mounted on slides with lactic acid and sealed with nail polish. Fungi from each slide were identified and counted. Some fungi were also isolated into pure culture to confirm identification.

### Data analysis

Litter-associated fungal communities from experimental streams were evaluated using Simpson diversity index and evenness (Magurran [1988](#page-13-0)). Chao 1 was used to estimate the number of taxa expected in the communities (Chao [1984\)](#page-12-0).

Fig. 1 Rainfall and average temperature during the study period. E1 and E2 indicate the experiments 1 and 2, respectively (\*start of leaf litter incubation for E1, \*\*collection of leaves, \*\*\*start of incubation for E2, coincided with final litter collection for E1)



While analyzing the relative importance of individual taxa in the communities, the relative abundance was expressed as a percentage of total conidia produced by all taxa (SI), and a percentage of total fungi slides by taxa (MC). Analysis of variance (ANOVA) was performed to test for the effects of stream, experiment, and collection day on fungal sporulation rate (SI). Similarity percentage analysis (SIMPER) was performed to identify the most influential taxa that contributed to the dissimilarity of fungal communities among streams. ANOSIM (Analysis of Similarity) permutation test was also performed to assess the dissimilarities of communities among streams (Clarke [1993\)](#page-12-0). Non-metric Multidimensional Scaling (NMDS) was used to analyze the effects of environmental variables (nutrient availability, oxygen, pH, temperature, and water velocity) on communities of fungi, using Bray-Curtis dissimilarity matrix (Clarke [1993\)](#page-12-0). Successional changes in fungal communities were evaluated through the constancy of species, which was calculated as:  $C = p \times 100/P$ , (where  $p =$ number of collections in which the taxon was present, and  $P =$ total number of collections.). We used the following constancy classes: accidental taxon ( $C \leq 25\%$ ), accessory taxon (25% <  $C \leq 50\%$ ), and constant taxon  $(C > 50\%)$  (Dajoz [1983](#page-12-0)). To compare the methodologies (SI vs. MC), we performed Unweighted Pair Group Method with Arithmetic Mean (UPGMA) cluster analysis based on the Dice similarity matrix.

Analyses were conducted in Past 3.14 (Hammer et al. [2013\)](#page-12-0) and Primer 5.2.6 software (Clarke and Gorley [2006](#page-12-0)).

# **Results**

## Environmental variables

RC, P1, and P2 are oligotrophic streams, with moderate-tohigh concentrations of dissolved oxygen and slightly acidic pH. P1 had the highest temperatures while P2 had slightly cooler waters (Table [1\)](#page-1-0).

## Sporulation rates in SI

Sporulation rates of freshwater hyphomycetes under submerged incubation conditions (Fig. 2) were significantly higher in E1 than in E2 (ANOVA,  $F_{1,60} = 14.1$ ,  $P = 0.0004$ ). In E1, high sporulation rates were found in RC and P2 on day 120, and P1 on day 180, while in E2, the highest peaks in all streams were achieved on days 180 or 240. These patterns generally corresponded to increased precipitation during the wet seasons (Fig. [1](#page-2-0)). No differences in sporulation rates among streams or collection days were found (ANOVA,  $P > 0.05$ ).



Fig. 2 Sporulation rates of freshwater hyphomycetes under submerged condition from leaf material incubated in three streams (RC, Rio de Contas; P1, Patricinho; P2, Patricio) during the experiment 1 and experiment 2 (E1 and E2)

#### Fungal community structure

SI

Taxa richness and diversity Twenty-six taxa of freshwater hyphomycetes were recorded from submerged leaves of C. brasiliense: one was an aero-aquatic hyphomycete, 13 were Ingoldian fungi, and 12 were facultative aquatic hyphomycetes (Table [2](#page-4-0), S3–S5). Taxa richness was comparable among streams with 16, 19, and 22 species, while 11 species were shared by all three streams (Table [2\)](#page-4-0). Dendrosporium lobatum was recorded only in RC, while Scutisporus brunneus was found only in P1, and four species were exclusive to P2: Dactylellina ellipsospora, Dictyotrichocladium aquaticum, Spirosphaera caricigraminis, and Subulispora procurvata. The highest taxa richness was observed during E1 that mostly coincided with the wet season, while fewer species were recorded during E2 (mostly dry season) (Fig. [3](#page-5-0), Table [2\)](#page-4-0). Estimates of Chao 1 were close to the actual number of taxa recorded in our samples suggesting an adequate sampling effort. Diversity estimates (Simpson index) and evenness of spore distribution among taxa were generally higher during the E[2](#page-4-0) than E1 (Table 2).

Succession The temporal changes in relative abundances of dominant taxa for all streams and both experiments (E1 and E2) are illustrated in Fig. [4.](#page-5-0) Triscelophorus acuminatus was a dominant or co-dominant species in

<span id="page-4-0"></span>Table 2 Relative abundances (%) of conidia of freshwater hyphomycetes associated with submerged leaves of C. brasiliense in streams RC, P1, and P2 during the two experiments (E1 and E2) assessed using submerged incubation technique



a Ingoldian fungi

<sup>b</sup> Facultative aquatic hyphomycetes

<sup>c</sup> Aero-aquatic hyphomycete

most cases, with its relative abundance increasing at later stages of decomposition by the end of E2. Morphotype 1 displayed high relative abundance during the middle stages of decomposition (day 60 and day 120), especially during the E2 and in P2. Monotosporella sp. was the dominant species in P1 during E1, while during the E2, it was mostly recorded during the middle stages of decomposition. Overall, Ingoldian fungi were mostly associated with submerged leaves of Calophyllum brasiliense during early-tomiddle stages of succession in all streams during E1 and in RC in E2 (Tables S3–S5), while they were more abundant during the later successional stages in P1 and P2 during E2. The constancy of fungal taxa in each stream and experiments (Table S6) shows that during E1, all streams had more constant taxa  $(2, 5,$  and 5 in RC, P1, and P2, respectively) than during E2 (3, 4, and 3 taxa), while the number of accessory and accidental species did not show a clear pattern.

Dissimilarity of communities and effects of environmental variables Similarity percentage analysis (SIMPER) showed that Triscelophorus acuminatus, Monotosporella sp., and Morphotype 1 contributed the most to the differentiation of fungal communities among streams (Table S7). No significant differences in freshwater hyphomycete communities among streams were found by ANOSIM.

<span id="page-5-0"></span>Fig. 3 Number of taxa of freshwater hyphomycetes under submerged condition (SI) and moist chamber incubations (MC) from leaf material collected in three streams (RC, Rio de Contas; P1, Patricinho; P2, Patricio) during the experiments 1 and 2 (E1 and E2)



UPGMA cluster and NMDS analyses generally showed separation of experiments based on their fungal communities (Figs. [5](#page-6-0) and [6](#page-6-0)). NMDS analysis also allowed to relate fungal communities to environmental variables. Oxygen availability was important in separating fungal communities along coordinate 1 axis, and it was mainly related to the communities found in P1 and P2 during the dry season. Fungal communities were separated along coordinate 2 axis, mainly due to pH being slightly higher in P2 vs. RC on most days. Higher water velocity was associated with more diverse fungal communities (mainly RC\_E1\_2 and P1 E1 3).



Fig. 4 Relative abundance of taxa common under submerged incubation conditions from leaf material collected in three streams (RC, Rio de Contas; P1, Patricinho; P2, Patricio) during the experiments 1 and 2 (E1 and E2)

<span id="page-6-0"></span>

**Coordinate 1**

Fig. 5 Results of NMDS analysis showing ordination of leaf-associated freshwater hyphomycete communities from submerged incubations. The effects of environmental factors are also shown: total nitrogen (TN), total phosphorus (TP), water temperature (Temp), pH, dissolved oxygen (O2),

The UPGMA cluster analysis with samples incubated under submerged conditions and in moist chambers (see below) showed clear separation by the type of and water velocity (WV). For sample labels, RC, P1, and P2 refer to streams (Rio de Contas, Patricinho, and Patricio, respectively); E1 and E2 to experiment 1 and 2; and numbers 1 through 4 to sampling days 60, 120, 180, and 240, respectively

incubation (Fig. 6). Moreover, the similarity among samples was higher in case of submerged incubation  $(> 65\%).$ 



Fig. 6 Results of UPGMA cluster analysis of leaf-associated freshwater hyphomycete communities from submerged incubations and moist chambers. Treatments are defined by stream (RC, Rio de Contas; P1,

Patricinho; P2, Patricio), experiments (E1 or E2), and methodology (SI for submerged incubations and MC for moist chambers)

#### MC

Taxa richness and diversity Fifty-six taxa of freshwater hyphomycetes were recorded in moist chambers; of these, six were aero-aquatic hyphomycetes, six were Ingoldian fungi, and 44 were facultative aquatic hyphomycetes (Table [3,](#page-8-0) S8–S10). RC had the highest taxa richness with 41 species, while samples from P1 and P2 yielded 32 and 19 species, respectively. All streams shared 10 taxa, while RC, P1, and P2 had 18, 9, and 3 exclusive taxa, respectively. RC and P2 showed the highest taxa richness during E2 (Table [3\)](#page-8-0), while in P1, it was the same in both experiments. In contrast to submerged incubations, estimates of Chao 1 for all streams and both experiments were considerably higher than the number of taxa actually recorded. Simpson diversity indices were comparable across streams and experiments, while the evenness among taxa was considerably lower in E2 than E1 in RC and P2 (Table [3\)](#page-8-0).

Succession Overall, the highest number of taxa was generally found at later stages of decomposition (day 240) in all streams and in both experiments (Tables  $S8-S10$ ). In RC, species of Thozetella were co-dominant in E1, being especially important during earlier stages of decomposition (Fig. [7](#page-9-0)). In P2, Kionochaeta spissa and Chalara alabamensis were highly important on day 180 of E1, while the latter was the dominant species on the very last sampling date in E2. Beltrania rhombica was an important early colonizer in P2 in both experiments, while Dictyochaeta simplex had high relative abundances in P1 throughout both experiments. Overall, the number of constant taxa varied from 2 to 5 among streams and experiments, while the number of accidental taxa was generally higher during E2 than E1 (e.g., up to 24 accidental taxa in RC) (Table S11).

Dissimilarity of communities and effects of environmental variables SIMPER analysis showed that Dictyochaeta simplex, Triscelophorus acuminatus, and Chalara alabamensis contributed the most to the differentiation of fungal communities among streams (Table S12). ANOSIM showed that streams significantly affected the structure of fungal communities ( $R = 0.47$ ,  $P =$ 0.0001). In contrast to data from submerged incubations, NMDS analysis based on fungal data from moist chambers was not significant and did not provide meaningful ordination (not shown).

# **Discussion**

A total of 69 species of freshwater hyphomycetes have been recorded in this study using two methodologies. However, since SI and MC targeted (and yielded) largely different ecological communities of fungi, the observed patterns will be discussed separately.

# SI

#### Taxa richness and diversity

Most species of freshwater hyphomycetes recovered from C. brasiliense under submerged conditions are Ingoldian fungi (13 species). This was expected since the methodology of Bärlocher [\(2005\)](#page-11-0) promotes sporulation of this group. The taxa richness recorded is slightly higher than what was previously reported from submerged leaves in Brazil (Schoenlein-Crusius and Milanez [1989;](#page-13-0) Schoenlein-Crusius and Milanez [1998](#page-13-0); Sales et al. [2014](#page-13-0)). However, some studies in the Neotropics have found a higher number of Ingoldian fungi, e.g., Bärlocher et al. [\(2010\)](#page-11-0) reported 32 taxa from streams in Panama.

Jabiol et al. ([2013](#page-12-0)) studied aquatic hyphomycete communities along a broad latitudinal gradient and found relatively low diversity in the tropics, which corroborates the earlier observation of Shearer et al. [\(2007\)](#page-13-0) that aquatic hyphomycetes display the highest diversity in temperate regions. Duarte et al. [\(2016](#page-12-0)) reviewing the distribution of aquatic hyphomycetes worldwide came to the same conclusion. Graça et al.  $(2016)$  $(2016)$ suggested some explanations of the low diversity of aquatic hyphomycetes in tropical streams. One of those is the relatively low dissolved nutrient availability, which was also the case in our study. Another factor that may explain the low diversity in tropical streams is the low quality of substrate, such as leaf litter. Many species of plants in tropical regions produce compounds (tannins and phenols) that inhibit the colonization by fungi and decreasing litter decomposition rates (Canhoto and Graça [1996;](#page-11-0) Hoorens et al. [2003](#page-12-0); Moretti et al. [2007](#page-13-0)). Calophyllum brasiliense used in this study produces phenolic compounds and calcium oxalate crystals (Junior et al. [2005](#page-12-0)).

## Succession

Most Ingoldian fungi together with some facultative hyphomycetes were found in pioneer stage of succession on decaying leaves of C. brasiliense. Earlier studies of succession on submerged leaves of Ficus microcarpa, Quercus robur, Alchornea triplinervia, Caesalpinia echinata, and Campomanesia phaea conducted in the Atlantic forest of Brazil indicated that Ingoldian fungi typically occur in later stages of succession while facultative aquatic or terrestrial fungi are found during the early phases (Schoenlein-Crusius and Milanez [1989](#page-13-0); Schoenlein-Crusius et al. [1990;](#page-13-0) Schoenlein-Crusius and Milanez [1998](#page-13-0); Moreira [2011](#page-13-0)). Gessner et al. [\(1993](#page-12-0)) working with alder leaves in temperate region also found higher richness of Ingoldian fungi at later successional stages. In our study, we did not find clear replacement of species through stages of succession; some taxa were constant from pioneer stages until the end of the study, while a few taxa occurred just in early stages. Gessner et al. [\(1993](#page-12-0))

<span id="page-8-0"></span>Table 3 Relative abundances (%) of freshwater hyphomycetes associated with submerged leaves of C. brasiliense in streams RC, P1, and P2 during two experiments (E1 and E2) assessed using moist chamber technique

| Taxa/samples                              | $RC_E1$ | $P1_E1$                  | $P2_E1$                  | $RC_E2$                  | $P1$ <sub><math>-E2</math></sub> | $P2$ <sub>_E2</sub>      |
|---|---------|--------------------------|--------------------------|--------------------------|----------------------------------|--------------------------|
| Anungitopsis triseptata <sup>b</sup>      | 3.13    | 1.33                     |                          | 2.83                     |                                  |                          |
| A. speciosa <sup>b</sup>                  | 1.56    |                          |                          |                          |                                  |                          |
| Ardhachandra aequilatera <sup>b</sup>     | 4.69    |                          |                          |                          |                                  |                          |
| Beltrania africana <sup>b</sup>           |         | 4.00                     |                          |                          |                                  |                          |
| B. rhombica <sup>b</sup>                  |         | $\equiv$                 | 2.86                     | 21.70                    | 10.91                            | 11.58                    |
| Beltraniella botryospora <sup>b</sup>     |         |                          | 2.86                     |                          |                                  |                          |
| B. fertilis <sup>b</sup>                  | 3.13    |                          | $\overline{\phantom{0}}$ | 0.94                     | 3.64                             |                          |
| B. portoricensis <sup>b</sup>             |         | 2.67                     | 8.57                     | $\overline{\phantom{0}}$ |                                  |                          |
| Beverwykella clathrata <sup>c</sup>       |         |                          |                          |                          |                                  | 3.16                     |
| Candelabrum microsporum <sup>c</sup>      | 1.56    |                          |                          | 1.89                     |                                  |                          |
| Chalara alabamensis <sup>b</sup>          | 1.56    | 2.67                     | 14.29                    | 0.94                     |                                  | 31.58                    |
| Cladosporium sp. <sup>b</sup>             |         |                          |                          | 1.89                     |                                  |                          |
| Corynesporopsis antillana <sup>b</sup>    |         |                          |                          | $\equiv$                 | 7.27                             |                          |
| Cylindrocladiella infestans <sup>b</sup>  |         |                          |                          | 0.94                     | 1.82                             | 7.37                     |
| Dactylaria fusifera <sup>b</sup>          |         |                          |                          | 0.94                     | 3.64                             |                          |
| $D.$ naviculiformis $^{b}$                |         | 8.00                     |                          | $\equiv$                 | $\overline{\phantom{0}}$         |                          |
| Dactylellina ellipsospora <sup>b</sup>    |         | 2.67                     |                          | $\overline{\phantom{0}}$ |                                  |                          |
| Dendrosporium lobatum <sup>a</sup>        |         | $\overline{\phantom{0}}$ |                          | 1.89                     |                                  |                          |
| Dendrosporomyces prolifer <sup>a</sup>    | 3.13    |                          |                          |                          | 3.64                             |                          |
| Dictyochaeta fertilis <sup>b</sup>        |         |                          |                          | 2.83                     | 1.82                             | 1.05                     |
| D. heteroderae <sup>b</sup>               |         |                          |                          | 0.94                     |                                  |                          |
| $D.$ simplex $^{b}$                       |         | 20.00                    |                          | 0.94                     | 12.73                            |                          |
| Filosporella sp. <sup>a</sup>             |         |                          |                          | 0.94                     |                                  |                          |
| Fusticeps laevisporus <sup>c</sup>        |         | 1.33                     | 2.86                     | $\equiv$                 |                                  | 1.05                     |
| Geniculospora inflata <sup>a</sup>        |         |                          |                          | 0.94                     |                                  |                          |
| Helicoon myosuroides <sup>c</sup>         | 1.56    |                          |                          | 0.94                     |                                  |                          |
| Hemibeltrania decorosa <sup>b</sup>       |         | 2.67                     |                          | $\overline{\phantom{0}}$ | 1.82                             |                          |
| Henicospora coronata <sup>b</sup>         | 1.56    | 6.67                     | 22.86                    | 0.94                     | 5.45                             | 4.21                     |
| Ingoldiella hamata <sup>a</sup>           | 1.56    | $\overline{\phantom{0}}$ |                          | $\overline{\phantom{0}}$ |                                  |                          |
| Isthmolongispora sp.1b                    | 1.56    | $\overline{\phantom{0}}$ |                          | 5.66                     | $\overline{\phantom{0}}$         | 1.05                     |
| $I.$ sp. $2^b$                            | 1.56    | 2.67                     |                          | 1.89                     | 5.45                             | $\overline{\phantom{0}}$ |
| $I.$ sp.3 <sup>b</sup>                    | 1.56    | $\equiv$                 | 5.71                     | 7.55                     | 7.27                             | $20.0\,$                 |
| Kionochaeta spissa <sup>b</sup>           |         | 1.33                     | 22.86                    |                          |                                  | 7.37                     |
| Mariannaea elegans <sup>b</sup>           | 1.56    |                          |                          | 0.94                     | 5.45                             |                          |
| Mirandina uncinata <sup>b</sup>           |         |                          |                          | 0.94                     |                                  |                          |
| Morphotype 1 <sup>b</sup>                 |         | 1.33                     |                          | $\overline{\phantom{0}}$ |                                  |                          |
| Paliphora porosa <sup>b</sup>             | 3.13    |                          |                          |                          |                                  |                          |
| Pseudaegerita websteri <sup>c</sup>       |         |                          |                          |                          | 1.82                             |                          |
| Roselymyces brasiliensis <sup>b</sup>     | 4.69    |                          |                          |                          |                                  |                          |
| Selenosporella curvispora <sup>b</sup>    |         |                          |                          | 0.94                     |                                  |                          |
| S. minima <sup>b</sup>                    |         | 4.00                     |                          |                          | 3.64                             |                          |
| Spirosphaera carici-graminis <sup>c</sup> |         | 10.67                    |                          | 15.09                    | 3.64                             | 1.05                     |
| Sporendocladia bactrospora <sup>b</sup>   |         |                          |                          | 0.94                     | 1.82                             |                          |
| Subulispora procurvata <sup>b</sup>       |         |                          | 5.71                     |                          | $\overline{\phantom{0}}$         | 4.21                     |
| Synchaetomella aquatica <sup>b</sup>      | 7.81    | 1.33                     |                          | 2.83                     | 1.82                             |                          |
| Thozetella acerosa <sup>b</sup>           | 1.56    |                          |                          |                          |                                  |                          |
| T. canadensis <sup>b</sup>                |         |                          |                          | 0.94                     |                                  | 1.05                     |
| T. cristata <sup>b</sup>                  | 10.94   | 1.33                     |                          | 8.49                     | 3.64                             | 1.05                     |
| T. pinicola <sup>b</sup>                  | 26.56   | 1.33                     | 8.57                     | 0.94                     | 10.91                            | 3.16                     |
| T. queenslandica <sup>b</sup>             |         |                          |                          | 0.94                     |                                  | -                        |
| $T.$ radicata $^{b}$                      |         |                          |                          | 1.89                     |                                  | 1.05                     |

## <span id="page-9-0"></span>Table 3 (continued)



a Ingoldian fungi

<sup>b</sup> Facultative aquatic hyphomycetes

c Aero-aquatic hyphomycetes

also did not observe clear species replacement as expected from classical succession studies.

All streams in E1 showed pioneer communities with high species richness. The impoverished communities at the later stages of decomposition were in agreement with a pattern suggested by Yanna et al. ([2002](#page-13-0)). In E2, we did not generally observe this pattern probably because of delay in colonization due to the presence of green spots in collected leaves.

According to Bärlocher [\(1991\)](#page-11-0), when the leaching does not occur, the colonization by aquatic hyphomycetes could be delayed. Beltrania rhombica and Thozetella havanensis were ob-

served only in early stages (until 120 days) of succession. According to Hyde et al. [\(2015\)](#page-12-0), some species found in terrestrial situations cannot survive for a long time under submerged conditions.



Fig. 7 Relative abundance of taxa common in moist chambers from leaf material collected in three streams (RC, Rio de Contas; P1, Patricinho; P2, Patricio) during the experiments 1 and 2 (E1 and E2)

#### MC

#### Taxa richness and diversity

Moist chamber incubations favor sporulation by taxa of facultative aquatic hyphomycetes and aero-aquatic hyphomycetes due to simulation of a terrestrial humid environment. Some Ingoldian fungi may also sporulate as previously observed from Brazilian Caatinga and Amazon (Fiuza et al. [2017\)](#page-12-0).

Most studies of freshwater hyphomycetes in moist chambers that dealt with submerged wood showed higher diversity of facultative aquatic hyphomycetes (up to 115 taxa) than what was found in the present study, while observations from leaf litter yielded comparable results (Tsui et al. [2001](#page-13-0); Castañeda-Ruiz et al. [2010](#page-11-0), [2016](#page-12-0); Fiuza et al. [2015](#page-12-0); Hyde et al. [2015](#page-12-0); Kodsueb et al. [2016\)](#page-13-0).

## Succession

In most cases, we found the highest richness of freshwater hyphomycetes at later stages of the experiments, which may correspond to a mature stage of succession as predicted by Dix and Webster [\(1995\)](#page-12-0). Communities of saprotrophic fungi often demonstrate species replacement as succession progresses (Gessner et al. [1993](#page-12-0); Rayner and Todd [1979](#page-13-0)), and we have found a similar pattern in moist chambers as suggested by high number of accessory or rare species in all streams.

Cylindrocladiella infestans that occurred during the first sampling date (day 60) in all streams in E2 is often isolated as endophyte or pathogen from leaves (Brown and Ferreira [2000;](#page-11-0) Evans et al. [2003\)](#page-12-0). Since our leaves had large green spots when collected, it is likely that it continued growing and then sporulated in moist chambers, a saprotrophic pattern described for some endophytes earlier (Promputtha et al. [2007,](#page-13-0) [2010](#page-13-0)).

#### Comparison between methodologies

We did not see a great overlap between submerged and moist chamber incubations with respect to the recovered fungal taxa (only 13 of 69 taxa were recorded by both techniques). This suggests that using several methodological approaches may help maximize the number of taxa recovered in ecological studies of litter-associated fungi in streams. Sridhar et al. [\(2010\)](#page-13-0) studying aquatic fungi on wood using damp incubation and bubble chamber methodologies have found even smaller overlap of just one species.

In most cases, moist chamber incubations yielded higher fungal diversity (Simpson index) than submerged incubations because of the greater degree of dominance displayed by samples in the latter method. Moreover, evenness was also higher in moist chambers. UPGMA results clearly showed separation of samples based on methodology.

# Sporulation rates, multivariate analyses, and effects of environmental variables

We observed the highest sporulation rates and species richness during E1 that was started at the end of the dry season and spanned all of the wet season. In Brazil, the litterfall occurs at the end of the dry season due to hydrological stress, and with the beginning of the wet season, increases in leaf litter standing stock in streams coincide with inputs of nutrients from the riparian zone stimulating microbial activity (Gonçalves et al. [2014](#page-12-0); Rezende et al. [2016](#page-13-0)). Sales et al. [\(2014\)](#page-13-0), who also worked in Bahia (Chapada Diamantina), found peaks of fungal sporulation during the wet season. Ghate and Sridhar [\(2016\)](#page-12-0) studying aquatic hyphomycetes on leaves of palm trees also detected the highest richness during the wet season. Thus, it appears that both in tropical and temperate regions, the seasonal patterns of activity of leaf-associated fungi in streams are driven by the availability of resources (litterfall). However, in this and some other studies in the tropics, the timing of the litterfall was determined by seasonal precipitation patterns, while in the temperate regions, it is driven by temperature declines in the fall (Shearer et al. [2007](#page-13-0); Sridhar and Sudheep [2010;](#page-13-0) Graça et al. [2016](#page-12-0)). Graça et al. [\(2016\)](#page-12-0) also suggested that increased water turbulence during the wet season in tropical regions stimulates activity; sporulation; and, hence, concentration of conidia of aquatic hyphomycetes in water, which facilitates colonization of new substrates.

In moist chamber incubations, the highest richness of freshwater hyphomycetes was observed in streams RC and P2 during the second experiment (E2), which was conducted mostly during the dry season, when the stream flow is low and some plant litter may become partially emergent. According to SIMPER results, fungal communities in these incubations were dominated by fungi common in terrestrial ecosystems, with the exception of *Triscelophorus acuminatus*. For example, Dictyochaeta simplex is usually found on submerged or terrestrial leaves, twigs, and grasses (Wong and Hyde [2001;](#page-13-0) Paulus et al. [2006\)](#page-13-0), while Chalara alabamensis has been recorded several times on leaves or wood in terrestrial situations (Morgan-Jones and Ingram [1976](#page-13-0); Rambelli et al. [2004\)](#page-13-0).

According to SIMPER analysis, Triscelophorus acuminatus was the dominant or co-dominant species in both submerged and moist chamber incubations in all streams. It is a species with worldwide distribution (Fiuza and Gusmão [2013\)](#page-12-0); however, it is often reported from tropical areas or from temperate streams during the warmer seasons (Jabiol et al. [2013](#page-12-0)). In tropical areas of India, Sridhar and Sudheep ([2010\)](#page-13-0) observed the high production of spores by T. acuminatus, as well as Amniculicola longissima,

<span id="page-11-0"></span>Flagellospora curvula, Lunulospora curvula, and T. monosporus, which are also reported in the current study.

Temperature is known to affect sporulation rate of aquatic hyphomycetes (Chauvet and Suberkropp [1998](#page-12-0)), plant litter decomposition (Boyero et al. 2016), and fungal communities in streams due to differences in temperature optima displayed by fungal species (i.e., cold-water vs. warm-water species) (Duarte et al. [2013\)](#page-12-0). In our study, most species were typical of warm waters, water temperature did not vary greatly throughout the experiments, and thus, NMDS analysis did not show a great effect of temperature on fungal communities. On the other hand, NMDS revealed an apparent effect of oxygen concentration in structuring fungal communities driven mostly by samples from the second experiment with higher  $O<sub>2</sub>$ concentrations. These fungal communities (P1 and P2 in dry season) had lower diversity and abundance of freshwater hyphomycetes conidia. Experimental studies that examined the wide range of oxygen concentrations found positive effects on diversity and sporulation of aquatic hyphomycetes (Medeiros et al. [2009](#page-13-0)). However, in our experiments, stream oxygen availability did not vary to the same extent, and it is also possible that our observed pattern could be due to the confounding effects of other variables. The positive effect of water velocity on NMDS ordination was driven mostly by fungal communities during the first experiment (mainly RC\_E1\_2 and P1\_E1\_3). Sanders and Webster [\(1980](#page-13-0)) found that higher flow can stimulate fungal sporulation due to enhanced water turbulence. Higher pH and total nitrogen in water were related to fungal communities in P2 during the first experiment. The effect of total phosphorus was rather minor. Several studies on the effects of nutrient enrichment (Gulis and Suberkropp [2003,](#page-12-0) [2004;](#page-12-0) Ferreira et al. [2015](#page-12-0)) found that dissolved inorganic nutrients can greatly stimulate the activity of aquatic litterassociated fungi, including their sporulation rate with concomitant changes in community structure, and ultimately affect fungi-driven decomposition of leaf litter in streams and rivers.

# Conclusions

Our data demonstrate the importance of using several methodological approaches to maximize the number of taxa recovered in ecological studies of litter-associated fungal communities. Submerged and aerial incubations provided complementary insights into fungal community structure since they favored fungal groups adapted to different sporulation environments. Successional patterns also differed between the two methodologies: the submerged incubations showed the highest richness of freshwater hyphomycetes during pioneer and mature stages of succession, while in moist chambers, many taxa were observed at the final stages of the experiment. Overall, the richness (69 taxa) of freshwater hyphomycetes

associated with submerged leaves of C. brasiliense was lower than often reported from temperate streams. At the same time, we found that samples in submerged incubations were more sensitive to differences in environmental variables driven by seasonal precipitation. Fungal communities differed between the experiments, and the highest fungal sporulation rates were observed during the wet season. The input of leaves, changes in oxygen concentration, and water velocity likely affected the communities of freshwater hyphomycetes associated with submerged leaf litter.

This study improves our understanding of the freshwater hyphomycete communities in streams of the semiarid regions of Brazil and the environmental variables that control the structure of these communities, which is especially important since ecological studies of aquatic fungi in Brazil and the tropical regions in general are still uncommon.

Funding information The authors are grateful to the "Programa de Pesquisa em Biodiversidade"—(PPBio Semi-arid/MCTI/CNPq) for financial support. POF thanks Coordenação de Aperfeiçoamento de Pessoal de Nível Superior—Programa de Doutorado Sanduíche no Exterior (CAPES-PDSE) for scholarship "Sciences Without Borders" (Proc. 99999.000984/2015-09) that enabled her stay at the Coastal Carolina University. POF also thanks "Programa de Pós-graduação em Botânica PPGBot/UEFS." LFPG is grateful to CNPq for financial support (Proc. 303062/2014-2).

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