

A world with CO₂ caps

Electricity production in consequential assessments

Göran Finnveden

Received: 11 March 2008 / Accepted: 15 March 2008 / Published online: 25 April 2008
© Springer-Verlag 2008

There is today a general agreement within the life cycle assessment (LCA) community that there are two types of LCA (e.g., Tillman 2000; Guinée 2002): These are often called attributional and consequential LCAs, although other terms sometimes are used as well (Ekvall and Weidema 2004). Consequential LCAs study the consequences of a choice. Ideally, the system boundaries and the data used should, then, reflect the actual changes taking place as a consequence of the choice. These may depend on the scale of the change and also on the time over which the changes occur. Attributional or accounting studies aim at describing a system as it actually was (or is, or would be) at a specific time. In such studies, the system boundaries and data should reflect what was actually happening in the system. A number of papers have been published discussing the methodology and case studies of consequential LCA (e.g., Andrae et al. 2007; Dalgaard et al. 2008; Ekvall and Andrae 2006; Frees 2008; Hofstetter and Norris 2003; Kløverpris et al. 2008; Lesage et al. 2007a, b; Sandén and Karlström 2007; Schmidt and Weidema 2008; Thiesen et al. 2008; Thrane 2006).

The same distinction is also relevant for other types of environmental systems analysis and sustainability tools (Finnveden and Moberg 2005; Ness et al. 2007). Some tools such as cost–benefit analysis and strategic environmental assessment are typically more oriented towards consequential types of analysis. Others, for example,

environmental audits and accounting systems, are generally of the attributional type (Finnveden and Moberg 2005). Other tools such as different types of energy analysis can be of both attributional and consequential types, although this is often not explicitly discussed.

One of the major differences between attributional and consequential LCAs concerns the choice between average and marginal data in the modeling of subsystems of the life cycle (Tillman 2000).

Average data can be national or regional average data for production, for example, of electricity. Marginal data represent the effects of a small change as a consequence of a decision. Average data are typically used within attributional LCAs and marginal data in consequential LCAs (Tillman 2000).

In many LCAs, energy-related impacts are important for the results; thus, the choice between average and marginal data and the type of marginal data can be important for the results. In many LCAs (and also other energy systems analysis), it has been assumed that the marginal source is a fossil fuel (e.g., coal or natural gas; Mathiesen et al. 2007). This may or may not have been correct up until now. However, this situation may now be radically changing with the introduction of the tradable carbon dioxide emission permits within the European Union (EU). If it works, it will provide a cap to the CO₂ emissions within the EU from the sectors that are included within the trading system. If the production increases within these sectors within the EU, the increase is not allowed to increase the CO₂ emissions.

If the cap would have been only over the electricity production system, this would have led to a CO₂-free marginal electricity production. This case was studied by Mattsson et al. (2003), who calculated a complex marginal electricity for the Nordic electricity system for a number

G. Finnveden (✉)
Division of Environmental Strategies Research—fms,
Department of Urban Planning and Environment,
School of Architecture and the Built Environment,
Royal Institute of Technology (KTH),
100 44 Stockholm, Sweden
e-mail: goran.finnveden@infra.kth.se

of different scenarios. One of these had a cap for the CO₂ emissions resulting in a very different marginal electricity production compared to other scenarios. Eriksson et al. (2007) used their study to define two alternative marginal electricity sources in a case study on district heating systems. The scenarios used were two extreme scenarios with a high amount of fossil fuels (here called “high gas price”) and a low amount of fossil fuels (“CO₂ cap”; Table 1).

In reality, the system is more complicated because the emission trading system also includes emissions from some industries. This means that emissions from the energy sector can increase if emissions from other sectors decrease, and vice versa. In a consequential LCA, these effects should also be included if they are consequences of the decision being made. But even so, it is clear that the marginal electricity source may be very different in the future compared to what has been normally assumed so far. It may very well be that the future marginal electricity source within the EU is virtually CO₂ free. This may have a very large impact on results from consequential LCAs.

The conclusion from this short paper is that the future marginal electricity sources need to be studied much more. It may very well be CO₂ neutral. In the meantime, we suggest that in cases where the electricity system is likely to have a significant influence on the results, two scenarios should be used: one with high CO₂ emissions and one with low or no CO₂ emissions, in consequential LCAs. A similar reasoning may be applied also for other types of products which are produced from industries included in the system of tradable carbon dioxide emission permits within the European Union, e.g., metal and paper products. The discussion here has focused on consequential LCAs. However, it is also valid for consequential studies in general.

This editorial can also be seen as a part of the discussion on consequential LCA vs. attributional LCA. A consequence of using consequential LCA may be that it seems like it does not really matter what we do. Although different

product systems may have different emissions, which can be seen in an attributional LCA, they may look similar in a consequential LCA. It may, of course, be argued that the approach taken here is too simplistic. Other types of consequences should also be included. For example, if we use products with lower emissions of CO₂, it may be easier for the politicians to lower the CO₂ cap in the future. This may or may not be true but is difficult to analyze and predict because political decisions are influenced by a number of different factors. In either way, it illustrates the complexities of consequential studies.

Acknowledgements This work is funded by the Foundation for Strategic Environmental Research (MISTRA). Thanks to Tomas Ekuell, IUL, for inspiring discussions.

References

- Andrae ASG, Itsubo N, Inaba A (2007) Global environmental impact assessment of the Pb-free shift. *Solder Surf Mt Technol* 19:18–28
- Dalgaard R, Schmidt J, Halberg N, Christensen P, Thrane M, Pengue WA (2008) LCA of soybean meal. *Int J LCA* 13(3):240–254
- Ekvall T, Weidema BP (2004) System boundaries and input data in consequential life cycle inventory analysis. *Int J LCA* 9(3):61–171
- Ekvall T, Andrae A (2006) Attributional and consequential environmental assessment of the shift to lead-free solders. *Int J LCA* 11:344–353
- Eriksson O, Finnveden G, Ekvall T, Björklund A (2007) Life cycle assessment of fuels for district heating: a comparison of waste incineration, biomass- and natural gas combustion. *Energy Policy* 35:1346–1362
- Finnveden G, Moberg Å (2005) Environmental systems analysis tools—an overview. *J Clean Prod* 13:1165–1173
- Frees N (2008) Crediting aluminium recycling in LCA by demand or by disposal. *Int J LCA* 13(3):212–218
- Guinée JB (ed), Gorée M, Heijungs R, Huppes G, Kleijn R, de Koning A, van Oers L, Wegener Sleswijk A, Suh S, Udo de Haes HA, de Bruijn JA, van Duin R, Huijbregts MAJ (2002): *Handbook on life cycle assessment: operational guide to the ISO standards*. Kluwer Academic, Dordrecht
- Hofstetter P, Norris GA (2003) Why and how should we assess occupational health impacts in integrated product policy? *Environ Sci Technol* 37:2025–2035
- Kløverpris J, Wenzel H, Nielsen PH (2008) Life cycle inventory modelling of land use induced by crop consumption. Part 1: conceptual analysis and methodological proposal. *Int J LCA* 13:13–21
- Lesage P, Ekvall T, Deschênes L, Samson R (2007a) Environmental assessment of Brownfield rehabilitation using two different life cycle inventory models. Part 1: methodological approach. *Int J LCA* 12:391–398
- Lesage P, Ekvall T, Deschênes L, Samson R (2007b) Environmental assessment of Brownfield rehabilitation using two different life cycle inventory models. Part 2: case study. *Int J LCA* 12:497–513
- Mathiesen BV, Münster M, Fruergaard T (2007) Energy system analyses of the marginal energy technology in life cycle assessments Proceedings of the SETAC Europe 14th LCA case study symposium, 3–4 December 2007, Göteborg, Sweden. SETAC Europe, Brussels, pp 15–19

Table 1 Complex marginal electricity production in two different scenarios for the future Nordic energy system (Mattsson et al. 2003)

Percentage	High gas price	CO ₂ cap
Wind	11.32	21.79
Nuclear	00.00	23.09
Biomass CHP	00.53	35.72
Coal CHP	59.99	00.77
Oil	03.03	-01.41
Natural gas CHP	25.33	19.95
Hydro power	-00.21	00.10

- Mattsson N, Unger T, Ekvall T (2003) Effects of perturbations in a dynamic system—the case of Nordic power production. In: Unger T (ed) Common energy and climate strategies for the Nordic countries—a model analysis, PhD thesis. Chalmers University of Technology, Göteborg
- Ness B, Urbel-Piirsalu E, Anderberg S, Olsson L (2007) Categorising tools for sustainability assessment. *Ecol Econ* 60:498–508
- Sandén B, Karlström M (2007) Positive and negative feedback in consequential life-cycle assessment. *J Clean Prod* 15:1469–1481
- Schmidt JH, Weidema BP (2008) Shift in the marginal supply of vegetable oil. *Int J LCA* 13(3):235–239
- Thiesen J, Christensen TS, Kristensen TG, Andersen RD, Brunoe B, Gregersen TK, Thrane M, Weidema BP (2008) Rebound effects of price differences. *Int J LCA* 13:104–114
- Thrane M (2006) LCA of Danish fish products: New methods and insights. *Int J LCA* 11:66–74
- Tillman A-M (2000) Significance of decision-making for LCA methodology. *Environ Impact Assess Rev* 20:113–123