RESEARCH ARTICLE



Study on harmless treatment of electrolytic manganese residue by low temperature thermochemical method

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Abstract

The Electrolytic Manganese Residue (EMR) is a by-product of the electrolytic manganese metal (EMM) industry, containing high concentrations of potential pollutants such as NH_4^+ -N and soluble Mn^{2+} . These components pose a serious threat to the ecological environment. To explore accurate, efficient, and harmless treatment methods for EMR, this study proposes a low-temperature thermochemical approach. The orthogonal experiment design investigates the effects of reaction temperature, reaction time, quicklime (CaO), sodium carbonate (Na₂CO₃), sodium phosphate (Na₃PO₄) (Reviewer #3), and water consumption on manganese solidified and ammonia removal from EMR. The results indicate that optimal conditions are a reaction temperature of 60 °C (Reviewer #3) and a reaction time of 10 min. CaO precipitates Mn^{2+} as manganese hydroxide (Mn(OH)₂) (Reviewer #3), achieving effective manganese solidified and ammonia removal. The addition of Na₂CO₃ causes Mn^{2+} to form manganesecarbonate (MnCO₃) (Reviewer #3) precipitate, while Na₃PO₄ makes Mn^{2+} form Manganese phosphate trihydrate ($Mn_3(PO_4)_2$ ·3H₂O) (Reviewer #3). Increased water consumption enhances the interaction adequacy between ions. Under optimal conditions (CaO 10%, Na₂CO₃ 1%, Na₃PO₄ 0.5%, and 80% water consumption), the removal rate of ammonium ions reaches 98.5%, and the solidification rate of soluble Mn^{2+} is 99.9%. The order of influence on ammonium ion removal is CaO > water consumption > Na₃PO₄ > Na₂CO₃. Therefore, this study provides a new method for low-cost process disposal and efficient harmless treatment of EMR (Reviewer #3).

Graphical Abstract



Keywords Low temperature thermochemistry \cdot Electrolytic manganese residue \cdot Harmless treatment \cdot Solidification of soluble manganese \cdot Ammonium ion removal (Reviewer #3)

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Introduction

Manganese stands as an exceptionally vital industrial raw material, with manganese-containing alloy products finding extensive applications in non-ferrous metallurgy, electronics, chemicals, and various other industries (Ghosh et al. 2016). Presently, China holds the distinction of being the world's foremost producer, consumer, and exporter of electrolytic manganese metal (EMM), contributing over 97% to the global production (Du et al. 2014). Electrolytic manganese residue (EMR), a byproduct of manganese carbonate sulfuric acid leaching, ammonia water neutralization, and pressure filtration in EMM plants, poses a significant environmental challenge (He et al. 2021a) (Reviewer #3). Each ton of EMM production generates approximately 10-12 tons of EMR (Wang et al. 2016), resulting in an annual emission of around 10 million tons of EMR from Chinese EMM plants. This issue is exacerbated by the declining manganese ore grades, leading to an escalating yearly increase in EMR production, with the current volume exceeding 120 million tons (Shu et al. 2018).

To date, a substantial portion of EMR has been directly deposited in slag yards without proper pretreatment, causing severe harm to local ecological environments and impeding the global EMM industry's progress. The pollutants NH_4^+ -N and Mn^{2+} in EMR pose a threat by infiltrating nearby soil and migrating to groundwater (Shu et al. 2016a).

Once the high concentration of ammonia nitrogen flows into the surrounding water environment, it will cause eutrophication of the water body (Zhang et al. 2020; Zulkifli et al. 2022), resulting in the excessive growth of microorganisms and algae and the depletion of dissolved oxygen in the water, so that a large number of aquatic organisms are anoxic and die (Yan et al. 2020), resulting in the destruction of ecological balance (Zhang et al. 2020). In addition, excessive ammonia nitrogen in the human body can also lead to blue baby syndrome, stomach and liver damage (Shu et al. 2019a). Drinking water contaminated with manganese may cause manganism, hallucinations, amnesia and nerve damage (Yang et al. 2022). At the same time, manganese can also lead to bronchitis, muscle weakness, headache, insomnia, even lead to schizophrenia (Hui et al. 2023), and potential risks to human health (Liu et al. 2017; Dey et al. 2022) (Reviewer #3).

Recognizing these risks, government bodies and relevant enterprises have closely monitored the issue since 2021. HJ 1241–2022 mandates that EMR must be 100% harmless before comprehensive utilization (Agency 2022), with the Ministry of Ecology and Environment setting a deadline for all disposal to be completed by 2025. Therefore, harmless disposal is imminent. Currently, research on the harmless treatment of EMR is ongoing. Fang (2014), Zhao et al (2017) have explored the washing method, which includes traditional slurry washing, slurry filter cake washing, and filtration combined with washing, effectively removing NH_4^+ -N and Mn^{2+} from EMR. However, this method is not suitable for insoluble substances and increases the cost of treating waste liquid. Tian et al (2019) (Reviewer #3), Shu et al (2016b), investigated electric field-enhanced leaching and electrokinetic remediation, which effectively removes soluble manganese and ammonia nitrogen but has drawbacks such as long reaction times and high energy consumption (Liu et al 2020) (Reviewer #3).

Additionally, the chemical solidification method, a primary approach for treating toxic and hazardous solid waste (He et al. 2021b) (Reviewer #3), involves encapsulating pollutants in solid waste with inert or cementitious materials to create stable compounds (Li et al. 2023) (Such as cement (Liu et al. 2023), oligomers (Guo et al. 2017), etc.) (Reviewer #3). Luo et al (2017) studied EMR treatment by adding quicklime, but a lengthy standing time was required. Shu et al (2018; 2020) used alkaline materials and phosphate to immobilize soluble manganese and ammonia nitrogen simultaneously, yet faced challenges with a 48-h reaction time. Du et al (2015a; b) employed CaO and NaHCO₃ for synergistic solidification of Mn²⁺ in EMR, with a prolonged reaction and drying cycle. In light of the current research, there remains a lack of an efficient (short reaction time) and widely implemented method for achieving the harmless pretreatment standard of EMR before comprehensive utilization.

In summary, in order to explore the precise, efficient and harmless treatment methods of EMR, this study proposes a low-temperature thermochemical method to pretreat EMR, and uses three different alkaline reagents (quicklime, sodium carbonate, sodium phosphate) to ensure the accuracy and efficiency of harmless treatment. The main research contents are as follows: The samples are heated at low temperature (40 \sim 100 °C), and the EMR is harmlessly treated by chemical reaction at this temperature. The impact of factors such as reaction temperature, time, lime dosage, sodium carbonate dosage, sodium phosphate dosage, and water consumption on EMR's harmless treatment effectiveness. The primary indicators for harmlessness are to reduce the ammonium ion and soluble manganese content in the solid. (Reviewer #4) This study provides a new method for the low-cost process disposal and efficient harmless treatment of EMR, and can provide effective support for the EMR harmless treatment production line, and promote the next resource utilization after EMR pretreatment (Reviewer #3).

Materials and methods

Materials

The EMR utilized in the experiment originated from an electrolytic manganese plant in Quanzhou, Guilin, Guangxi, China (111° 05' 38.21" E, 25° 57' 47.85" N) (Reviewer #3), displaying a black mud paste appearance (Zhou et al. 2023) (Reviewer #3). Table 1 presents the results obtained through X-ray fluorescence spectrometry (XRF, ZSX Primus II) (Reviewer #3) analysis of the electrolytic manganese residue. Prior to conducting the harmless experiment, the EMR samples underwent drying to a constant weight at 60 °C (Reviewer #3). Subsequently, the dried EMR underwent grinding using a ball mill and sieving through a 100-mesh sieve, with a sieve margin of less than 2%. The analytical chemical reagents (purity \geq 99.7%) (Reviewer #3) employed in the experiment included lime (CaO), sodium carbonate (Na₂CO₃), and sodium phosphate dodecahydrate (Na₃PO₄·12H₂O). These reagents were purchased from Xilong Scientific Co., Ltd (Reviewer #3). Deionized water was utilized in the experiments as well.

The results of X-ray diffraction (XRD, X'Pert PRO) experiments which are shown in Fig. 1. on EMR (Reviewer #3). The mineral components present in electrolytic manganese residue, as illustrated in Fig. 1, primarily consist of CaSO₄·2H₂O (dihydrate gypsum), SiO₂ (quartz), (NH₄)[Fe₃(OH)₆(SO₄)₂] (ammonium jarosite), $Fe(NH_4)_2(SO_4)_2 \cdot 6H_2O$ (Ammonium iron sulfate), MnFeO₂ (Manganese ferrite), $(NH_4)_2Mn(SO_4)_2$ (ammonium manganous sulfate), FeO(OH) (Iron hydroxide oxide) (Reviewer #3) and others. Notably, the characteristic spectral lines of CaSO₄·2H₂O and SiO₂ exhibit sharpness, indicating good crystallinity (Wu et al. 2024) (Reviewer #3). The characteristic peak of (NH₄)[Fe₃(OH)₆(SO₄)₂] is distinct and exhibits a certain degree of crystallinity. CaSO₄·2H₂O originates from the reaction between CaCO₃ and sulfuric acid (H_2SO_4) during the acid leaching process of rhodochrosite (Su et al. 2023) (Reviewer #3). The chemical formula is as follows (1) (Reviewer #3):

$$CaCO_3 + H_2SO_4 + H_2O \rightarrow CaSO_4 \cdot 2H_2O + CO_2$$
(1)

Compounds such as $(NH_4)_2Mn(SO_4)_2$, Fe $(NH_4)_2(SO_4)_2$ · $6H_2O$, MnFeO₂, among others, suggest that after the rhodochrosite leaching, the Mn²⁺ not completely filtered out and the ammonia added to adjust the pH are transferred to the electrolytic manganese residue during the subsequent pressure filtration process (Duan et al. 2010).



Fig. 1 XRD pattern of EMR

In this study, EMR undergoes a harmless treatment using low-temperature heating and thermochemical methods, as illustrated in Fig. 2. (Reviewer #3) After drying and grinding, the EMR is mixed with the corresponding reagent in the mixing tank, and then the appropriate amount of water is added to the reaction vessel to react to the corresponding time to obtain the harmless sample (Reviewer #4).

Experimental

The experimental parameters are as follows: the reaction temperature is set at 40, 60, 80, and 100 °C, and the reaction time is varied between 5, 10, 15, 20 min. (Note: Due to the sample's thinness, sampling at 40 °C is performed from 25 min onwards.) The reagent dosage is adjusted to 0%, 3%, 6%, 9%, and 12%, respectively, as outlined in Table 2. The investigation into the harmless treatment efficiency of EMR covers four key aspects: CaO, Na₂CO₃, Na₃PO₄, and water consumption (Shu et al. 2020) (Reviewer #3). The corresponding ratios are detailed in Tables 3, 4, 5, and 6.

Analysis and detection methods

Determination of EMR Moisture Content: The assessment was conducted in accordance with GB/T 14949.8 (Agency 2018). The recorded average moisture content stands at 17%.

Table	1 EMR	composition of
raw m	aterials,	wt.%

Raw materials	SiO ₂	SO ₃	Fe ₂ O ₃	CaO	Al ₂ O ₃	MnO	MgO	P ₂ O ₅	Others
EMR	35.53	16.60	22.53	8.04	5.29	7.35	1.90	0.83	1.93

Fig. 2 Schematic diagram of harmless treatment of EMR (Reviewer #3)



Table 2 Reaction Temperature, Time Experimental Ratio, wt.%

Number	CaO (%)	Na ₂ CO ₃ (%)	Na ₃ PO ₄ (%)	W (%)
1	10	1	0.5	80

 Table 3
 CaO in electrolytic manganese residue Addition ratio, wt.%

Number	CaO(%)	Na ₂ CO ₃ (%)	Na ₃ PO ₄ (%)	W (%)
Ca-0	0	2	2	40
Ca-3	3			
Ca-6	6			
Ca-9	9			
Ca-12	12			

Table 4 Na₂CO₃ in EMR Addition ratio, wt.%

Number	CaO (%)	Na ₂ CO ₃ (%)	Na ₃ PO ₄ (%)	W (%)
N2-0	3	0	2	40
N2-3		3		
N2-6		6		
N2-9		9		
N2-12		12		

Table 5	Na_3PO_4	in EMR	Addition	ratio,	wt.%
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Number	CaO(%)	Na ₂ CO ₃ (%)	Na ₃ PO ₄ (%)	W(%)
N3-0	3	2	0	40
N3-3			3	
N3-6			6	
N3-9			9	
N3-12			12	

Table 6 Water in EMR Addition ratio, wt.%

Number	CaO(%)	Na ₂ CO ₃ (%)	Na ₃ PO ₄ (%)	W(%)
W-3	6	2	2	30
W-5				50
W-7				70
W-9				90
W-12				120

- (2) Referring to GB / T 39701 (Agency 2020), the standard ammonium ion limit was 210 mg/kg, and the content of EMR ammonium ion was 6101.88 mg/kg (Reviewer #3).
- (3) The soluble manganese content was 1784 mg / L measured by EMR according to HJ 766 (Agency 2015). The samples to be tested were leached of soluble manganese ions (Mn²⁺) according to HJ 557 (Agency 2010). A fitting curve correlating absorbance and soluble manganese content was generated according to GB 11906 (Agency 1989) (refer to Fig. 3). According to the curve, the content of soluble manganese was obtained. It is noteworthy that the soluble manganese content of harmless samples should not exceed the 2 mg/L limit specified in GB 8978 (Agency 1996) (Reviewer #3).
- (4) XRD Analysis: The samples underwent analysis using an X'Pert PRO X-ray diffractometer, employing Cu target Kα-ray (λ=1.54060 Å). The scanning angle range covered 5° to 80°, with a scanning duration of 4 min (Li et al. 2022). (Reviewer #3).
- (5) SEM-EDS Characterization: Microstructure characterization and energy spectrum analysis of the samples were conducted using the FESM S-4800 (Li et al. 2022). (Reviewer #3).



Fig. 3 Mn²⁺ concentration versus absorbance linear fitting curve

Result and Discussion

The effects of reaction temperature and reaction time on the removal of ammonia from (Reviewer #4) EMR were studied

Figure 4 (a) and (b) illustrate the impact of reaction temperature and reaction time on ammonia removal by EMR. Maintaining a specify values (Reviewer #3) an observable decrease in the levels of ammonium ions and soluble manganese in the sample is evident with the progression of both reaction temperature and reaction time. Regarding ammonia removal, higher temperatures correspond to a more significant reduction in ammonium ions in the initial stages. This phenomenon can be attributed to the accelerated volatilization of NH_4^+ into NH_3 under alkaline conditions as the temperature rises (Deng et al. 2018). (Reviewer #3).

Concerning manganese fixation, the soluble manganese content remains extremely low at this specify values (Reviewer #3). This observation underscores the notion that, within the context of the harmless EMR process, the challenge of ammonia removal surpasses that of manganese fixation. Notably, at a reaction temperature of 60 °C and a reaction time of 10 min, the ammonium ion content is measured at 93.29 mg/kg, yielding an impressive removal efficiency of 98.5%, well below the national standard limit of 210 mg/ kg (Agency 2020) (Reviewer #3). Simultaneously, the soluble manganese content is a mere 0.07 mg/L, attaining an exceptional curing rate of 99.9%. Considering these factors comprehensively, this specific combination of reaction temperature and time is deemed optimal and is consequently selected for subsequent experiments.

Effect of CaO addition on manganese fixation and ammonia removal of EMR

Figure 5 (a) and (b) show the effect of lime addition on EMR manganese fixation and ammonia removal. Under the premise of constant Na_2CO_3 , Na_3PO_4 and water consumption, with the increase of CaO content (0%, 3%, 6%, 9%, 12%) (Reviewer #3), the removal rate of ammonia nitrogen and the solidification rate of soluble manganese are significantly improved.

In Fig. 5 (a), the increase in CaO content correlates with a linear decrease in ammonium ions, declining from the initial level of 5536.48 mg/kg to 205.39 mg/kg. At this point, the ammonium ion content aligns with the stipulated limit of 210



Fig. 4 Effect of reaction temperature and time on the content of ammonium ions (a) and soluble manganese (b) in EMR



Fig. 5 Effect of CaO addition on the content of ammonium ion (a) and soluble manganese (b) in EMR (reaction temperature: 60 °C and reaction time: 10 min) (Reviewer #4)

mg/kg outlined in GB/T 39701 (Agency 2020) (Reviewer #3). The removal of ammonium ions is attributed to the escape of NH_4^+ in EMR in the form of NH_3 through the generation of OH^- resulting from the reaction between CaO and water (Zhou et al. 2013). The reaction equation is presented as follows (2~3) (Reviewer #3). Additionally, due to the exothermic nature of the CaO-water reaction and the low-temperature heating environment, this reaction proceeds swiftly.

$$CaO + H_2O \rightarrow Ca(OH)_2$$
 (2)

$$\mathrm{NH_4}^+ + \mathrm{OH}^- \xrightarrow{\Delta} \mathrm{NH_3} \uparrow \tag{3}$$

In Fig. 5 (b), as the CaO content increases, the concentration of Mn^{2+} undergoes a substantial reduction from 1784 mg/L to 0.1 mg/L. Specifically, when the CaO content reaches 6%, the Mn^{2+} concentration diminishes to 0.32 mg/L, complying with the specified limit of 2.0 mg/L in GB 8978 (Agency. 1996) (Reviewer #3). The primary mechanism behind this phenomenon is the formation of $Mn(OH)_2$ precipitation by Mn^{2+} under the alkaline conditions facilitated by CaO. Subsequently, $Mn(OH)_2$ undergoes oxidation by O_2 in the air, leading to the formation of stable compounds such as MnOOH, Mn_2O_3 , and MnO_2 (Luo et al. 2017). The reaction equation, presented below (4~7) (Reviewer #3), illustrates how this process achieves the goal of solidifying soluble Mn^{2+} .

$$Mn^{2+} + 2OH^{-} \to Mn(OH)_{2} \downarrow$$
(4)

$$4Mn(OH)_2 + O_2 \rightarrow 4MnOOH + 2H_2O$$
(5)

$$6Mn(OH)_2 + O_2 \rightarrow 2Mn_3O_4 + 6H_2O \tag{6}$$

$$2\mathrm{Mn}(\mathrm{OH})_2 + \mathrm{O}_2 \to 2\mathrm{MnO}_2 + 2\mathrm{H}_2\mathrm{O} \tag{7}$$

In examining the XRD results presented in Fig. 6, it is evident that during the low-temperature thermochemical treatment of EMR, a portion of dihydrate gypsum (CaSO₄·2H₂O) undergoes conversion into hemihydrate gypsum (CaSO₄·0.5H₂O), as revealed through phase analysis. This transformation is likely attributed to the increased temperature inducing a change in the crystal form of gypsum (Duan et al. 2023) (Reviewer #3). Furthermore, the augmentation of CaO content correlates with a decrease in the crystallinity of (NH₄)₂Mn(SO₄)₂ and Fe(NH₄)₂(SO₄)₂·6H₂O



Fig. 6 XRD plots of samples harmless to EMR with different CaO additions

in EMR. This change is possibly linked to the escape of ammonium ions in the form of NH_3 under alkaline conditions (Zhang et al. 2023) (Reviewer #3), leading to alterations in these two phases (He et al. 2022) (Reviewer #3). Additionally, there is a discernible shift in the form of soluble manganese, characterized not only by the formation of $Mn(OH)_2$ precipitation but also by manganese binding with Ca^{2+} and SiO_3^{2-} to create stabilizers like $CaMn(SiO_3)_2$, $Ca_3Mn_2O_7$, and others (He et al. 2022). The corresponding reaction equations are provided below (8~10) (Reviewer #3).

$$\operatorname{SiO}_2 + 2\operatorname{OH}^- \to \operatorname{SiO}_3^{2-} + \operatorname{H}_2\operatorname{O}$$
(8)

$$\operatorname{Ca}^{2+} + 2\operatorname{SiO}_{3}^{2-} + \operatorname{Mn}^{2+} \to \operatorname{Ca}\operatorname{Mn}(\operatorname{SiO}_{3})_{2} \downarrow$$
(9)

$$6\mathrm{Ca}^{2+} + 4\mathrm{Mn}^{2+} + 7\mathrm{O}_2 \rightarrow 2\mathrm{Ca}_3\mathrm{Mn}_2\mathrm{O}_7 \downarrow \tag{10}$$

Effect of Na₂CO₃ content on manganese fixation and ammonia removal of EMR

Figure 7 (a) and (b) depict the impact of sodium carbonate addition on EMR manganese fixation and ammonia removal. Maintaining constant CaO, Na_3PO_4 , and water consumption, the rise in Na_2CO_3 content notably enhances the solidification rate of soluble manganese. However, the effect on ammonia removal is not as pronounced as observed with CaO.

In Fig. 7 (a), the increase in Na_2CO_3 content correlates with a substantial decrease in ammonium ion content, dropping from 3502.41 mg/kg to 1630.89 mg/kg. At a dosage of 12%, the removal rate reaches 73.27%, although the ammonium ion content still falls short of the stipulated limit of 210 mg/kg in GB/T 39701 (Agency. 2020) (Reviewer #3). This phenomenon can be attributed to the alkalinity provided by Na₂CO₃ still influencing EMR ammonia removal, albeit to a lesser extent compared to CaO, resulting in a limited effect (Li et al. 2018) (Reviewer #3).

In Fig. 7 (b), the elevation of Na₂CO₃ content leads to a significant reduction in the concentration of Mn^{2+} , decreasing from 18.4 mg/L to 0.04 mg/L. At a Na₂CO₃ content of 3%, the Mn²⁺ concentration diminishes to 0.7 mg/L, meeting the specified limit of 2.0 mg/L in GB 8978 (Agency. 1996) (Reviewer #3). The primary mechanism behind this observation is the formation of insoluble MnCO₃ (Du et al. 2015a) when Mn²⁺ combines with CO₃²⁻ under alkaline conditions. The corresponding reaction equation is provided below (11) (Reviewer #3), illustrating how this process achieves the desired effect of mitigating soluble Mn²⁺.

$$\mathrm{Mn}^{2+} + \mathrm{CO}_3^{2-} \to \mathrm{Mn}\mathrm{CO}_3 \downarrow \tag{11}$$

Examining Fig. 8, it is evident through phase analysis that, consistent with prior findings, some dihydrate gypsum (CaSO₄·2H₂O) undergoes conversion into hemihydrate gypsum (CaSO₄·0.5H₂O). Furthermore, with the inclusion of Na₂CO₃, the appearance of the characteristic peak of MnCO₃ aligns with previous analyses (Du et al. 2015a) (Reviewer #3), indicating that soluble manganese is fixed in the form of MnCO₃. Additionally, as Na₂CO₃ content increases, the characteristic peak of CaCO₃ emerges. This occurrence may be attributed to the fact that the solubility product constant (K_{sp}) of MnCO₃ (K_{sp} (MnCO₃)=1.8×10⁻¹¹) is significantly lower than that of CaCO₃ (K_{sp} (CaCO₃)=2.8×10⁻⁹). Consequently, in the harmless treatment process, MnCO₃ precipitate



Fig. 7 Effect of Na₂CO₃ addition on the content of ammonium ion (**a**) and soluble manganese (**b**) in EMR (reaction temperature: 60 °C and reaction time: 10 min) (Reviewer #4)



Fig. 8 XRD plots of samples harmless to EMR with different $\mathrm{Na_2CO_3}$ additions

forms first (Li et al. 2021). Subsequently, the excess CO_3^{2-} in Na_2CO_3 reacts with Ca^{2+} in EMR to form $CaCO_3$ precipitate. The reaction equation representing this process is provided below (12) (Reviewer #3).

$$Ca^{2+} + CO_3^{2-} \to CaCO_3 \downarrow \tag{12}$$

Effect of Na₃PO₄ content on manganese fixation and ammonia removal of EMR

Figure 9 (a) and (b) illustrate the impact of sodium phosphate addition on EMR manganese fixation and ammonia removal. Under constant CaO, Na_2CO_3 , and water consumption, the increase in Na_3PO_4 content exhibits a discernible, albeit less pronounced, effect on the removal of EMR ammonium ions and the solidification of soluble manganese compared to CaO and Na_2CO_3 .

In Fig. 9 (a), the rise in Na_3PO_4 content leads to a noticeable decrease in ammonium ion content, dropping from 3571.92 mg/kg to 2039.57 mg/kg. At a dosage of 12%, the removal rate reaches 66.57%, yet the ammonium ion content still falls short of the stipulated limit of 210 mg/kg in GB/T 39701 (Agency. 2020) (Reviewer #3). The primary reason for this occurrence is that the weak alkalinity provided by Na_3PO_4 influences EMR ammonia removal, albeit to a lesser extent than CaO, resulting in a limited effect (Shu et al. 2019b) (Reviewer #3).

In Fig. 9 (b), the augmentation of Na_3PO_4 content significantly reduces the concentration of Mn^{2+} , decreasing from 9.8 mg/L to 0.15 mg/L. At a Na3PO4 content of 3%, the Mn^{2+} concentration diminishes to 0.7 mg/L, meeting the specified limit of 2.0 mg/L in GB 8978 (Agency. 1996) (Reviewer #3). The primary mechanism behind this observation is the formation of insoluble $Mn_3(PO_4)_2$ when Mn^{2+} combines with PO_4^{2-} under alkaline conditions (Chen et al. 2019). The corresponding reaction equation is provided below (13) (Reviewer #3), illustrating how this process achieves the desired effect of mitigating soluble Mn^{2+} . Concurrently, a comparison between N3-0 and N2-0, as well as Ca-0 groups, indicates that the curing effect of Na₃PO₄ on



Fig. 9 Effect of Na₃PO₄ addition on the content of ammonium ion (**a**) and soluble manganese (**b**) of EMR (reaction temperature: 60 °C and reaction time: 10 min) (Reviewer #4)

soluble Mn^{2+} is not as effective as Na_2CO_3 , suggesting a potential reduction in its dosage in subsequent experiments.

$$3\mathrm{Mn}^{2+} + 2\mathrm{PO}_4^{2-} \to \mathrm{Mn}(\mathrm{PO}_4)_2 \downarrow \tag{13}$$

The XRD of EMR harmless samples with raw residue and Na₃PO₄ content is shown in Fig. 10. Through phase analysis, in addition to the transformation of dihydrate gypsum (CaSO₄·2H₂O) and hemihydrate gypsum (CaSO₄·0.5H₂O), the characteristic peak of Mn₃(PO₄)₂·3H₂O appeared, which is consistent with the previous analysis (Chen et al. 2019) (Reviewer #3). Soluble manganese and PO₄^{2–} combined to



Fig.10 XRD plots of samples harmless to EMR with different $\mathrm{Na_3PO_4}$ additions

form insoluble $Mn_3(PO_4)_2$ precipitation to achieve the purpose of curing soluble manganese.

Effect of water addition on manganese fixation and ammonia removal of EMR

Figure 11(a) and (b) illustrate the impact of water consumption on EMR manganese fixation and ammonia removal. Maintaining constant levels of CaO, Na_2CO_3 , and Na_3PO_4 , the variations in water consumption exert a discernible effect on the removal of EMR ammonium ions and the solidification of soluble manganese.

In Fig. 11(a), as water consumption increases, there is a substantial initial decrease in ammonium ions in EMR. However, beyond a water consumption level of 70%, the reduction in ammonium ions becomes marginal. This phenomenon arises because the heightened water content affects the degree of ion interaction, but there is a discernible upper limit. Further increasing water volume beyond this point does not significantly enhance the removal of ammonium ions in EMR.

In Fig. 11(b), with a constant reagent dosage, the concentration of Mn^{2+} remains consistently low, meeting the specified limit of 2.0 mg/L in GB 8978 (Agency. 1996) (Reviewer #3). As water consumption increases, there is a modest improvement in the manganese fixation efficiency of the sample. This is attributed to the fact that water content also impacts the degree of ion reaction.

Figure 12 displays the XRD results of EMR harmless samples influenced by raw slag and water consumption. Through phase analysis, it becomes evident that as water consumption increases, the characteristic peak intensity of dihydrate gypsum (CaSO₄·2H₂O) diminishes, concomitant



Fig. 11 Effect of water addition on the content of ammonium ion (a) and soluble manganese (b) in EMR (reaction temperature: 60 $^{\circ}$ C and reaction time: 10 min) (Reviewer #4)



 $\ensuremath{\mbox{Fig.12}}$ XRD patterns of samples harmless to EMR with different water additions

with the appearance of the characteristic peak of hemihydrate gypsum (CaSO₄·0.5H₂O). This observation may be attributed to the specific harmless treatment method employed in this study. Furthermore, the decrease in crystallinity of $(NH_4)_2Mn(SO_4)_2$ and $Fe(NH_4)_2(SO_4)_2 \cdot 6H_2O$ in EMR aligns with the aforementioned findings (Zhou et al. 2013) (Reviewer #3). Notably, the characteristic peak intensity of $Mn(OH)_2$, $MnCO_3$, and $Mn_3(PO_4)_2 \cdot 3H_2O$ increases with rising water consumption, underscoring that an elevated water content impacts the adequacy of ion reactions.

Orthogonal experiment

Through the conducted research, it becomes apparent that achieving the prescribed standard for the removal rate of ammonium ions is more readily accessible. However, for achieving the curing rate of soluble manganese at equivalent reagent dosages is more challenging. The impact of four variables impact on the removal degree of EMR ammonium ions was further investigated and the optimal mix ratio for harmless EMR was determined through an orthogonal experiment using L₉ (3⁴) (Reviewer #3). The selected factor levels are as follows: CaO (6%, 8%, 10%), Na₂CO₃ (0.5%, 1%, 1.5%), Na₃PO₄ (0.5%, 1%, 1.5%), and water consumption (40%, 60%, 80%). The specific ratios are outlined in Table 7.

The evaluation indices for the orthogonal experiment were the ammonium ion content and soluble manganese content of the treated EMR, aiming to analyze the impact of each level factor on these two measures. The results of the orthogonal test and subsequent range analysis are presented in Tables 7 and 8. The greater the range (R value, $R=K_{avg}max-K_{avg}min$) (Reviewer #3), the more pronounced the influence of factors on the test results. Analyzing the K value corresponding to different levels of each factor helps identify the level closest to satisfactory test results for the index, facilitating the determination of the optimal level combination for each factor. (K value is the sum of experimental data at the same level of a certain factor. K_{avg} is the corresponding average, and the following values are the K_{avg} .) (Reviewer #3).

Table 7 R value (Reviewer #3) reveals that the lowest ammonium ion content in treated EMR was observed in Z-8 at 93.29 mg/kg. The ammonium ion content exhibited varying degrees of change with alterations in CaO, Na₃PO₄, Na₂CO₃, and water consumption. Notably, the curing rate of manganese ions in the orthogonal experimental group remained consistently high, underscoring the greater difficulty in achieving high removal of ammonium ions compared to achieving high curing of soluble manganese. Subsequently, the ammonium ion content was analyzed based on the results of the orthogonal test.

The interaction curve of the four factors Kavg in Table 8 is shown in Fig. 13. It can be seen from the figure that CaO and water consumption have a great influence on the removal

Table 7 Orthogonal experimental mix ratios and results wt.%

Number	A(CaO)	B(Na ₂ CO ₃)	C(Na ₃ PO ₄)	D(W)	NH ₄ ⁺ concentra- tion (mg/kg)	NH ₄ ⁺ removal rate (%)	Mn ²⁺ concentra- tion (mg/L)	Mn ²⁺ solidifi- cation rate (%)
Z-1	6	0.5	0.5	40	2197.14	63.99	0.18	99.84
Z-2	6	1	1	60	1516.01	75.16	0.1	99.91
Z-3	6	1.5	1.5	80	1498.42	75.44	0.07	99.93
Z-4	8	0.5	1	80	391.12	93.6	0.04	99.96
Z-5	8	1	1.5	40	1444.5	76.33	0.15	99.87
Z-6	8	1.5	0.5	60	666.83	89.07	0.12	99.89
Z-7	10	0.5	1.5	60	541.15	91.13	0.09	99.92
Z-8	10	1	0.5	80	93.29	98.47	0.07	99.94
Z-9	10	1.5	1	40	1284.10	78.96	0.09	99.92

Table 8Effect of factors on ammonium ion content of EMR wt.%

Samples	NH_4^{+} concentration (mg/kg)							
	A CaO (%)	B Na ₂ CO ₃ (%)	C Na ₃ PO ₄ (%)	D W(%)				
K1	1737.19	1043.14	985.75	1641.91				
K2	834.15	1017.93	1063.74	908.00				
K3	639.51	1149.78	1161.36	660.94				
R	1097.68	131.85	175.61	998.97				

of ammonium ions in the sample. In order to ensure that the ammonium ion content of the sample is low, a high content of CaO and water consumption (10% CaO and 80% W) is required. On the contrary, Na₂CO₃ and Na₃PO₄ have little effect on the removal of ammonium ions in the sample, and the minimum value of K_{avg} can be taken (1% Na₂CO₃ and 0.5%Na₃PO₄). Therefore, Z-8 group is the best harmless group.

From the Table 8 R value, it can also be seen that the CaO content has the most significant effect on the removal effect of EMR ammonium ion (Reviewer #3), while Na₂CO₃ has a relatively minor impact on the removal effectiveness. The hierarchy of factors influencing the removal of ammonium ions by EMR is as follows: CaO > water consumption > Na₃PO₄ > Na₂CO₃. The dominant impact of CaO can be attributed to its high alkalinity, allowing for the rapid conversion of ammonium ions into ammonia volatilization under low-temperature heating conditions. The quantity of water accelerates ion reactions, with a more pronounced effect in dry reactions.

Conversely, the low alkalinity of Na₂CO₃ and Na₃PO₄ imparts a comparatively minor influence on the removal

of ammonium ions by EMR. Notably, the higher removal of ammonium ions by Na_3PO_4 compared to Na_2CO_3 may stem from the complexation of $(PO_4)^{3-}$ with a small amount of NH_4^+ (Chen et al. 2019).

Based on the phase analysis of Z-8 and EMR in the qualified group, as depicted in Fig. 14, it becomes evident that the characteristic peak of $(NH_4)_2Mn(SO_4)_2$ disappears with the removal of ammonium ions. Mn^{2+} undergoes solidification, taking the forms of $Mn(OH)_2$, MnO_2 , $MnCO_3$, and $Mn_3(PO_4)_2 \cdot 3H_2O$. Additionally, the characteristic peak of $CaSO_4 \cdot 0.5H_2O$ emerges due to the low-temperature thermochemical harmless method. The augmentation of the characteristic peak of $CaCO_3$ may be attributed to the reaction of OH^- with CO_2 in the air, resulting in the formation of CO_3^{2-} , which then reacts with Ca^{2+} to produce a $CaCO_3$ precipitate.

The SEM and EDS surface scan analysis in Fig. 15, comparing EMR samples (a, b, c) with the qualified group Z-8 (d, e, f), reveals distinctive features. In the EMR samples, irregular massive, columnar, and flaky substances identified as CaSO₄·2H₂O are observed, arranged alternately and disorderly stacked. Some irregular fine particles, likely composed of SiO₂ or other minerals rich in Al and Fe, may be attached to the CaSO₄·2H₂O structure (Xue et al. 2020). Post-harmless treatment, the EMR retains $CaSO_4 \cdot 2H_2O$, with an increase in small particles covering the surface of $CaSO_4 \cdot 2H_2O$. These small particles may be carbonates and hydroxide precipitates formed through the reaction of heavy metal ions with OH^{-} and CO_{3}^{2-} (Shu et al. 2020). This observation further underscores the efficacy of solidifying heavy metal ions in EMR during the harmless treatment process.



Fig. 13 The interaction curve diagram of various factors(Reviewer #3)



Fig. 14 XRD patterns of samples harmless to EMR



Fig. 15 SEM-EDS images of EMR and harmless to EMR samples

Conclusion

In this study, a low-temperature thermochemical method was employed to investigate the impact of reaction temperature, reaction time, CaO, Na₂CO₃, Na₃PO₄, and water consumption on the efficacy of EMR manganese fixation and ammonia removal. Orthogonal experiments were designed to assess the influence of four variables on the removal of EMR ammonium ions, ultimately leading to the identification of the optimal mix ratio. The key findings of the experiments are summarized as follows:

- The increase in reaction temperature and reaction time facilitates ion reactions, resulting in a varying degree of reduction in ammonium ion and soluble manganese content in the sample. Considering all factors, a reaction temperature of 60 °C and a reaction time of 10 min were selected as the optimal conditions for achieving harmlessness.
- The high alkalinity and substantial heat release provided by CaO in water not only efficiently remove ammonium ions in EMR but also solidify soluble Mn²⁺ in the form

of precipitates, such as $Mn(OH)_2$, $CaMn(SiO_3)_2$, and $Ca_3Mn_2O_7$.

- 3) Na₂CO₃'s alkalinity has a relatively low removal efficiency for EMR ammonium ions, but it exhibits good manganese fixation efficiency. This is attributed to the smaller $K_{sp}(MnCO_3)$ compared to $K_{sp}(CaCO_3)$, causing CO_3^{2-} to induce the formation of MnCO₃ precipitate. The addition of Na₃PO₄ induces Mn²⁺ to form Mn₃(PO₄)₂·3H₂O precipitation. The increase of water consumption promotes the reaction between ions(Reviewer #3).
- 4) In the orthogonal experiment, the order of influence of various factors on the removal of ammonium ions by EMR is CaO > water consumption > $Na_3PO_4 > Na_2CO_3$. Under The optimal mixing ratio(Reviewer #3), the ammonium ion content of the sample is 93.29 mg/kg, with a removal rate of 98.5%. The soluble manganese content is 0.07 mg/L, achieving a curing rate of 99.9%, well below the national standard limit.

This study addresses a new harmless process-low temperature thermochemical method was used to treat EMR efficiently and harmlessly, and solves the issue of EMR pollution stemming from the EMM industry, providing novel research methodologies for the harmless treatment of EMR (Reviewer #3). It realizes an economical and efficient harmless treatment of EMR. The harmless products will be used for the resource utilization of supplementary cementitious materials and cement retarders (Reviewer #4). However, continued research on short-term and long-term curing properties, as well as cost control, is crucial, and ongoing attention to these areas is warranted.

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Declarations

Ethical approval and consent to participate Written informed consent for publication of this paper was obtained from all authors.

Consent to publish The work described has not been published before; that it is not under consideration for publication anywhere else; that its publication has been approved by all co-authors.

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