ADSORPTION PROCESSES FOR ENVIRONMENTAL ISSUES



Isotherms, kinetics and thermodynamics of industrial dye acid red 27 adsorption on Sugarcane Bagasse Ash

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Abstract

In this study, sugarcane bagasse ash (SCBA), obtained as residue from the sugar mill, was used as an adsorbent for Acid Red 27 (AR27) removal from aqueous solutions. The ash characterization data showed 23.63% of organic compounds and silica (α -SiO₂) as the most expressive inorganic compound (confirmed by X-ray diffractogram), the BET surface area had a value of 62.79 m².g⁻¹ and the pH_{pzc} was 8.45. Regarding the adsorptive tests, the optimal initial pH to the dye removal was 2.0. The adsorption equilibrium reached in about 4 h contact time and optimum SCBA dosage was found to be 4 g.L⁻¹. The pseudo-second order model best represented the adsorption kinetics. The Freundlich equation presented the best fit to the equilibrium data for the removal of AR27 by ash, with maximum adsorption capacity of 15 mg.g⁻¹ at pH 2.0. Thermodynamic study indicate that AR27 adsorption on SCBA occurs through a physisorption mechanism, with $\Delta H^o_{ads} < 15 \text{ kJ.mol}^{-1}$. The ΔH^o_{ads} evaluated by Vant' Hoff equation was explained as a combination of water desorption enthalpy, ΔH^o_W and isosteric like enthalpy, ΔH^o_D for the dye adsorption in liquid environment. The $\Delta H^o_D = 9.2 \text{ kJ.mol}^{-1}$ was calculated from Clausius-Clapeyron approach. The effects of coexisting anions on the adsorption and regeneration and reuse of the adsorbent were also investigated. This study suggests that SCBA, which was used without any pretreatment, has the potential to be applied as a low-cost adsorbent to mitigate effluents contamination with AR27 dye at low concentrations.

Keywords Dye adsorption \cdot Decontamination \cdot Sustainability \cdot Adsorption kinetics \cdot Isotherms \cdot Sugar mill waste \cdot Isosteric enthalpy

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Highlights

- Acid Red 27 dye adsorption at sugarcane bagasse ash (SCBA) has been investigated.
- Isosteric enthalpy analogy for adsorption in liquids was applied to better understanding the thermodynamic parameters.
- Regeneration of SCBA with NaOH (0.1 and 0.3 M) allowed recuperation of at least 75 % of the dye adsorbed.
- Adsorption with coexisting anions evidenced high influence of bicarbonate for decrease the SCBA adsorptive capacity.

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Introduction

Brazil is the world's largest producer of sugarcane (FAOSTAT 2023), an extremely versatile product, and major player in the sugar-energy sector. It is estimated that sugarcane production in the harvest of 2022/2023 has reached 610.1 million tons (CONAB 2023a). In that same harvest, alcohol production totaled 27.3 billion liters, with 15.7 billion of the hydrated type, and 11.6 billion of the anhydrous type. The sugar production was 37 million tons; where, of this total, 27.7 million were exported (CONAB 2023b). This performance confirms the importance of sugarcane as a national commodity, placing Brazil as the second largest producer of ethanol and the largest producer of sugar.

In the sugar and alcohol industry, bagasse is generated in large quantities, approximately 25 a 30% of the weight of the total sugarcane crushed (Gar et al. 2017; Souza et al. 2011). This biomass is burned in high-pressure boilers to generate steam and electricity, making some industries self-sufficient

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in energy and offering the possibility of selling the surplus. The bagasse burning process generates sugarcane bagasse ash (SCBA) as a residue, which, if not disposed correctly, would be a serious environmental problem. By weight, it is estimated that SCBA corresponds to 0.6% of the weight of processed sugarcane (Andrade Neto et al. 2021; Matos et al. 2021). Leading the Brazilian industry to allocate 3.7 million tons of ash in the 2022/2023 harvest.

Presenting a high silica content, SiO₂, bagasse ash has been widely studied for construction purposes, such as, a mineral additive in the manufacture of cements and ceramics, in the manufacture of reinforced concrete (in order to reduce corrosion by chlorides), as well as in the preparation of pastes and mortars (Patel 2020). Other studies have shown that ash can also be used in the production of adsorbents, mesoporous silica, zeolites, immobilizers in reactors and secondary abrasives (Patel 2020; Sriatun et al. 2018; Rahman et al. 2015).

Furthermore, the use of ash as an adsorbent can be a very promising alternative, especially since it is cheap, readily available and with potential application by the industry as an input or product to mitigate impacts generated by contaminated effluents (Mane et al. 2007; Mall et al. 2005, 2006; Gupta et al 2000).

Dyes have been widely used in various industrial fields, such as pharmaceutical, textile, cosmetics, food, pulp and paper, photographic etc., and their disposal must be constantly monitored. The presence of dyes in natural waters inhibits the penetration of sunlight, reducing the photosynthetic reaction and interfering with the biological degradation of water impurities, resulting in ecological imbalance. Furthermore, some may be toxic and even carcinogenic or mutagenic (Ebrahimpoor et al. 2019; Pinedo-Hernández et al. 2012).

Several technologies have already been proposed for the removal of dyes from water, such as electrochemical methods, chemical precipitation, advanced oxidation processes, biological treatment and membrane filtration (Arabkhani et al. 2021). As synthetic dyes have a complex aromatic structure, they are rarely removed by biological methods due to their low biodegradation (Salman-Naeem et al. 2018). On the other hand, combined techniques sometimes show good performance in color removal, but most have high costs, high technical requirements and complicated practice. In addition, some of them cause the treated effluent to remain with high values of organic carbon dissolved due to resulting by-products, which can be more toxic than the previous molecule (Metivier-Pignon et al. 2007; Nassar 2010). In fact, adsorption is seen as a simple technique, with excellent cost-effectiveness, ease of operation when using efficient and widely available adsorbents, which is why it has been widely studied and industrially applied (Arabkhani et al. 2021; Ebrahimpoor et al. 2019).

The dye Acid Red 27 (AR27) has wide industrial applications, being used in household products and cosmetics, to dye synthetic and natural fibers, leather, wood, paper, and in photography (Chhabra et al 2015). However, the specialized literature shows few studies on its decontamination in water (Nassar 2010; Al-Aoh et al. 2013; Salman-Naeem et al. 2018; Ebrahimpoor et al. 2019; Yusof et al. 2020), and no study was found in our research on its adsorption in sugarcane bagasse ash.

The objectives of this work are to characterize the ashes obtained from the combustion of sugarcane bagasse in highpressure boilers and to study the adsorption of the dye AR27 by the SCBA. The applicability of kinetic and isothermal models was evaluated, determining the influence of adsorbent dosage, initial pH, and the presence of anions on its adsorption capacity. The temperature influence on adsorption isotherms was studied an applied for understanding the thermodynamic parameters and the isosteric enthalpy behavior of the dye adsorption in SCBA. The study of its regeneration and reuse was also carried out.

Experimental

The chemicals used in this study were hydrochloric acid P.A. (37%, Quimica Moderna, Brazil), sodium hydroxide P.A. (>98%, Dinâmica, Brazil), sodium chloride P.A. (>99.5%, Fmaia, Brazil), anhydrous sodium sulfate P.A. (>99.0%, Vetec, Brazil), sodium bicarbonate P.A.-ACS (>99.7% Química Moderna, Brazil), and the dye acid red 27 (Polycrom, Brazil). All chemicals were used without further purification.

Sugarcane bagasse ash

The ashes used to this work was kindly provided by Petribu S.A. sugarcane mill (Lagoa de Itaenga, PE, Brazil). Through the industrial process, the sugarcane bagasse was burned in high-pressure boilers at 380 °C, resulting in ashes that falls in a water channel, which was dragged to a settling pool. After its draining, the SCBA were collected, dried at 105 °C for about 24 h. Then, it was sieved, and the part with particle size lower than 0.25 mm was used as adsorbent.

Adsorbate

The synthetic dye Acid Red 27 [CAS 915–67-3, C.I. 16185, chemical formula = $C_{20}H_{11}N_2Na_3O_{10}S_3$, molecular weight = 604.5 g.mol⁻¹, pK_a = 6.5, λ_{max} = 520 nm (Fig. S1)]. A quantity of dye was weighed in a Quimis Q500B210C analytical balance and dissolved in distilled water to prepare a stock solution of 1000 mg.L⁻¹. Experimental solutions at the desired concentration were obtained by successive

dilutions with distilled water. The structure of the AR27 is illustrated in Fig. 1.

Characterization of the sugarcane bagasse ash

The chemical composition of SCBA was determined by X-ray fluorescence (XRF), using a Rigaku ZSX Primus II equipped with Rh tube and 7 analyzer crystals. Scanning electron microscope (SEM) images were obtained with a VEGA3 (Tescan, CZ) Microscope, operating at acceleration voltage from 5 to 20 kV. Each sample was placed on a small metallic support (stub) with double-sided carbon tape. The samples were then metallized with a nanometric layer of gold using a Desk V (DentonVacuum, US). The thermogravimetric curve was obtained in an instrument TGA O50 V6.7 Build 203 and performed in the range of 27 to 1000 °C with a flow of air and a heating rate of 10°C/min. The N₂ adsorption-desorption isotherms were obtained at 77 K and relative pressures up to 0.01, using a Quantachrome Instruments NOVA STATION A. The specific surface area (SBET) of the adsorbent was calculated from the adsorption data using the Brünauer-Emmett-Teller (BET) method. The micropore area and volume (S_{mic} and V_{mic}) were obtained by the t-plot method. The mesopore volume (Vext) was calculated as the difference between the total pore volume (V_T) and the micropore volume (V_{mic}), obtained at P/P_o = 0.99. X-ray diffraction (XRD) patterns were obtained using a Shimadzu XRD-6000 instrument operated with Cu-K_a radiation ($\lambda = 0.1546$ nm) in the 20 range from 5 to 80°, with voltage and current of 30 kV and 30 mA, respectively. FTIR spectra were recorded from 4000 to 400 cm⁻¹ using



Fig. 1 Molecular structure of Acid Red 27

a Shimadzu IR PRESTIGE-21 spectrometer. The point of zero charge (pH_{pzc}) of the adsorbent was obtained according to an adapted methodology (Mahmood et al. 2011), using NaCl (0.1 M) saline solution and adjusting the initial pH (pH_0) with solutions of HCl (0.1 M) or NaOH (0.1 M). The pH was monitored using a pHmeter Instrutherm PH-1500.

Adsorption experiments

To study the adsorption of the AR27 dye using SCBA as adsorbent, experiments in batches were conducted in a shaker (Marconi MA-420) at 30 ± 1 °C. For each experimental run, 25 ml of dye solution of known concentration and a measured amount of adsorbent were placed in 125 ml Erlenmeyer flasks, under stirring speed of 300 rpm. At the end of each experiment, the samples were centrifuged to separate the adsorbent and the final concentrations determined using a UV–visible spectrophotometer (Thermo Fisher Scientific Genesys 10-S) at a wavelength of 520 nm.

For kinetic experiments, the contact time ranged of 5—360 min with dye solutions concentration of 30 mg.L⁻¹. The models evaluated in this work were Pseudo-first-order (PFO), Pseudo-second-order (PSO), Elovich and Intraparticle Diffusion (IPD, proposed by Weber and Morris). The equations are given in Table S1.

The effect of important parameters that influence adsorption was studied. The effect of initial pH (pH₀) on dye removal was studied over a pH range of 2 to 12. The pH₀ was adjusted by adding dilute aqueous solutions of HCl (0.1 M) or NaOH (0.1 M). The effect of adsorbent dosage on dye removal percentage and adsorption capacity was analyzed by changing the adsorbent dosage from 1.2 to 12 g.L⁻¹, keeping other factors unchanged. To define the adsorbent dosage, a dye solution with a defined concentration (30 mg.L⁻¹) was placed in contact with different amounts of adsorbents (0.05 g to 0.8 g) until equilibrium was reached.

For equilibrium experiments, different dye concentrations ranging from 2.5 to 240 mg.L⁻¹ (2.5; 5; 20; 40; 50; 60; 80; 100; 120; 140;160; 180; 200; 220; 240 mg.L⁻¹) were used. The vials were placed on a shaker until equilibrium was reached. The experimental data were fitted to the Langmuir, Freundlich and Temkin models, considering linear and nonlinear fits. The equations are given in Table S2.

The non-linear fitting to the data was performed in Origin Pro©. Chi-square tolerance value of 1.0×10^{-9} (difference between reduced chi-square values of two successive iterations) was used as the convergence criterion.

Thermodynamic evaluations were carried out at temperatures of 303, 321 and 338 K using the incubator shaker, for dye concentrations ranging from 20 to 200 mg.L⁻¹ (20; 40; 60; 80; 100; 120; 140;160; 180; 200 mg.L⁻¹).

In addition, the effects of coexisting anions (Cl⁻, SO₄⁻, HCO₃⁻) at concentrations of 0.1 M on AR27 adsorption

by SCBA were investigated. The salts (NaCl, Na₂SO₄, NaHCO₃) were added separately to the dye solution at a defined concentration (30 mg.L⁻¹) until equilibrium was reached.

For the regeneration and reuse of the adsorbent, 0.1 g of the SCBA was added into 25 mL of 30 mg.L⁻¹ dye solution $(pH_0=2)$ and placed under constant stirring (300 rpm) at 30 ± 1 °C for 4 h. The final concentration was measured and the adsorbent was separated by centrifugation. Alkaline solutions are used to elute negatively charged dyes; alkaline eluent sodium hydroxide solution are most used for desorption study to elute anionic dye (Patel 2021). So, the regeneration possibility of exhausted adsorbent was investigated using distilled water and NaOH solution at different concentrations (0.01 M, 0.1 M and 0.3 M). The adsorbent used was added into 10 mL of eluent in a shaker (300 rpm stirring) for 2 h at 30 ± 1 °C; after that, the SCBA was washed with distilled water to neutral pH to remove the residual NaOH. The regenerated adsorbent was tested for AR27 adsorption (initial dye concentration: 30 mg.L⁻¹; $pH_0 = 2$; contact time: 4 h; adsorbent dosage: 4 g.L⁻¹; temperature: 30 ± 1 °C). The SCBA adsorbent was regenerated by three cycles. The amount of AR27 adsorbed after the cycles was analysed and recorded.

All experiments were performed in duplicate.

The removal efficiency (%) of dye was calculated using the Eq. 1. The adsorption capacity, q_e (mg.g⁻¹), was calculated from the mass balance (Eq. 2) after the attainment of equilibrium.

Removal efficiency =
$$\frac{(C_0 - C_e)100}{C_0}$$
 (1)

$$q_e = \frac{\left(C_0 - C_e\right)V}{W} \tag{2}$$

where C_0 (mg.L⁻¹) is the initial concentration, C_e (mg.L⁻¹) is the equilibrium sorbate concentration, V is the volume of the solution (L) and w is the mass (g) of the adsorbent.

Results and discussion

Characterization of the sugarcane bagasse ash

Ash, in general, has a very high concentration of silica and contains aluminum, iron, alkali and alkaline earth metal oxides as minor components (Teixeira et al. 2008). Table 1 shows the results of XRF analysis of SCBA sample along with its chemical composition. The results indicate that SCBA mainly contains SiO₂ (52.40%), CaO (5.26%), K₂O (5.00%), Fe₂O₃ (4.09%), among other minor constituents. SiO₂ is the major component of SCBA, being consistent

$\begin{array}{c c} SiO_2 & 52.404 \\ SiO_2 & 52.404 \\ CaO & 5.262 \\ K_2O & 5.002 \\ Fe_2O_3 & 4.086 \\ P_2O_5 & 3.979 \\ MgO & 2.100 \\ SO_3 & 1.298 \\ Na_2O & 0.993 \\ TiO_2 & 0.519 \\ Cl & 0.275 \\ MnO & 0.206 \\ ZnO & 0.053 \\ SrO & 0.046 \\ BaO & 0.046 \\ CuO & 0.031 \\ ZrO_2 & 0.031 \\ ZrO_2 & 0.031 \\ Cr_2O_3 & 0.023 \\ RbO & 0.015 \\ As_2O_3 & 0.008 \\ LOI & 23.632 \end{array}$	Table 1Chemical constituents(%) of SCBA determined by	Chemical constituent	(%)
$\begin{array}{cccc} CaO & 5.262 \\ K_2O & 5.002 \\ Fe_2O_3 & 4.086 \\ P_2O_5 & 3.979 \\ MgO & 2.100 \\ SO_3 & 1.298 \\ Na_2O & 0.993 \\ TiO_2 & 0.519 \\ Cl & 0.275 \\ MnO & 0.206 \\ ZnO & 0.053 \\ SrO & 0.046 \\ BaO & 0.046 \\ BaO & 0.046 \\ CuO & 0.031 \\ ZrO_2 & 0.031 \\ Cr_2O_3 & 0.023 \\ RbO & 0.015 \\ As_2O_3 & 0.008 \\ LOI & 23.632 \end{array}$	XRF	SiO ₂	52.404
$\begin{array}{cccc} K_2 O & 5.002 \\ Fe_2 O_3 & 4.086 \\ P_2 O_5 & 3.979 \\ Mg O & 2.100 \\ SO_3 & 1.298 \\ Na_2 O & 0.993 \\ TiO_2 & 0.519 \\ Cl & 0.275 \\ MnO & 0.206 \\ ZnO & 0.053 \\ SrO & 0.046 \\ BaO & 0.046 \\ BaO & 0.046 \\ CuO & 0.031 \\ ZrO_2 & 0.031 \\ ZrO_2 & 0.031 \\ Cr_2 O_3 & 0.023 \\ RbO & 0.015 \\ As_2 O_3 & 0.008 \\ LOI & 23.632 \end{array}$		CaO	5.262
$\begin{array}{cccc} Fe_2O_3 & 4.086 \\ P_2O_5 & 3.979 \\ MgO & 2.100 \\ SO_3 & 1.298 \\ Na_2O & 0.993 \\ TiO_2 & 0.519 \\ Cl & 0.275 \\ MnO & 0.206 \\ ZnO & 0.053 \\ SrO & 0.046 \\ BaO & 0.046 \\ BaO & 0.046 \\ CuO & 0.031 \\ ZrO_2 & 0.031 \\ ZrO_2 & 0.031 \\ Cr_2O_3 & 0.023 \\ RbO & 0.015 \\ As_2O_3 & 0.008 \\ LOI & 23.632 \end{array}$		K ₂ O	5.002
$\begin{array}{cccc} P_2 O_5 & 3.979 \\ Mg O & 2.100 \\ SO_3 & 1.298 \\ Na_2 O & 0.993 \\ TiO_2 & 0.519 \\ Cl & 0.275 \\ MnO & 0.206 \\ ZnO & 0.053 \\ SrO & 0.046 \\ BaO & 0.046 \\ BaO & 0.046 \\ CuO & 0.031 \\ ZrO_2 & 0.031 \\ ZrO_2 & 0.031 \\ Cr_2 O_3 & 0.023 \\ RbO & 0.015 \\ As_2 O_3 & 0.008 \\ LOI & 23.632 \end{array}$		Fe ₂ O ₃	4.086
$\begin{array}{cccc} MgO & 2.100 \\ SO_3 & 1.298 \\ Na_2O & 0.993 \\ TiO_2 & 0.519 \\ Cl & 0.275 \\ MnO & 0.206 \\ ZnO & 0.053 \\ SrO & 0.046 \\ BaO & 0.046 \\ BaO & 0.046 \\ CuO & 0.031 \\ ZrO_2 & 0.031 \\ ZrO_2 & 0.031 \\ Cr_2O_3 & 0.023 \\ RbO & 0.015 \\ As_2O_3 & 0.008 \\ LOI & 23.632 \end{array}$		P_2O_5	3.979
$\begin{array}{cccc} SO_3 & 1.298 \\ Na_2O & 0.993 \\ TiO_2 & 0.519 \\ Cl & 0.275 \\ MnO & 0.206 \\ ZnO & 0.053 \\ SrO & 0.046 \\ BaO & 0.046 \\ CuO & 0.031 \\ ZrO_2 & 0.031 \\ ZrO_2 & 0.031 \\ Cr_2O_3 & 0.023 \\ RbO & 0.015 \\ As_2O_3 & 0.008 \\ LOI & 23.632 \end{array}$		MgO	2.100
$\begin{array}{cccc} Na_2O & 0.993 \\ TiO_2 & 0.519 \\ Cl & 0.275 \\ MnO & 0.206 \\ ZnO & 0.053 \\ SrO & 0.046 \\ BaO & 0.046 \\ CuO & 0.031 \\ ZrO_2 & 0.031 \\ ZrO_2 & 0.031 \\ Cr_2O_3 & 0.023 \\ RbO & 0.015 \\ As_2O_3 & 0.008 \\ LOI & 23.632 \end{array}$		SO ₃	1.298
$\begin{array}{cccc} {\rm TiO}_2 & 0.519 \\ {\rm Cl} & 0.275 \\ {\rm MnO} & 0.206 \\ {\rm ZnO} & 0.053 \\ {\rm SrO} & 0.046 \\ {\rm BaO} & 0.046 \\ {\rm CuO} & 0.031 \\ {\rm ZrO}_2 & 0.031 \\ {\rm ZrO}_2 & 0.031 \\ {\rm Cr}_2{\rm O}_3 & 0.023 \\ {\rm RbO} & 0.015 \\ {\rm As}_2{\rm O}_3 & 0.008 \\ {\rm LOI} & 23.632 \end{array}$		Na ₂ O	0.993
$\begin{array}{ccc} Cl & 0.275 \\ MnO & 0.206 \\ ZnO & 0.053 \\ SrO & 0.046 \\ BaO & 0.046 \\ CuO & 0.031 \\ ZrO_2 & 0.031 \\ Cr_2O_3 & 0.023 \\ RbO & 0.015 \\ As_2O_3 & 0.008 \\ LOI & 23.632 \end{array}$		TiO ₂	0.519
$\begin{array}{ccc} MnO & 0.206 \\ ZnO & 0.053 \\ SrO & 0.046 \\ BaO & 0.046 \\ CuO & 0.031 \\ ZrO_2 & 0.031 \\ Cr_2O_3 & 0.023 \\ RbO & 0.015 \\ As_2O_3 & 0.008 \\ LOI & 23.632 \end{array}$		Cl	0.275
$\begin{array}{cccc} ZnO & 0.053 \\ SrO & 0.046 \\ BaO & 0.046 \\ CuO & 0.031 \\ ZrO_2 & 0.031 \\ Cr_2O_3 & 0.023 \\ RbO & 0.015 \\ As_2O_3 & 0.008 \\ LOI & 23.632 \end{array}$		MnO	0.206
$\begin{array}{ccc} SrO & 0.046 \\ BaO & 0.046 \\ CuO & 0.031 \\ ZrO_2 & 0.031 \\ Cr_2O_3 & 0.023 \\ RbO & 0.015 \\ As_2O_3 & 0.008 \\ LOI & 23.632 \end{array}$		ZnO	0.053
$\begin{array}{ccc} BaO & 0.046 \\ CuO & 0.031 \\ ZrO_2 & 0.031 \\ Cr_2O_3 & 0.023 \\ RbO & 0.015 \\ As_2O_3 & 0.008 \\ LOI & 23.632 \end{array}$		SrO	0.046
$\begin{array}{ccc} CuO & 0.031 \\ ZrO_2 & 0.031 \\ Cr_2O_3 & 0.023 \\ RbO & 0.015 \\ As_2O_3 & 0.008 \\ LOI & 23.632 \end{array}$		BaO	0.046
$\begin{array}{ccc} ZrO_2 & 0.031 \\ Cr_2O_3 & 0.023 \\ RbO & 0.015 \\ As_2O_3 & 0.008 \\ LOI & 23.632 \end{array}$		CuO	0.031
$\begin{array}{c} Cr_2O_3 & 0.023 \\ RbO & 0.015 \\ As_2O_3 & 0.008 \\ LOI & 23.632 \end{array}$		ZrO ₂	0.031
RbO 0.015 As ₂ O ₃ 0.008 LOI 23.632		Cr ₂ O ₃	0.023
As ₂ O ₃ 0.008 LOI 23.632		RbO	0.015
LOI 23.632		As_2O_3	0.008
		LOI	23.632

with results obtained conventionally for sugarcane ashes (Patel 2020; Zhang et al. 2020; Patil et al. 2021). Variations in ash composition are related to differences in the chemical composition of the soil and the variety of cultivated sugarcane.

The SEM images of the SCBA are shown in Fig. 2. The porous morphology of SCBA may be due to inherent structure in the original raw material, the sugarcane bagasse, which mainly contains cavities on its fiber composition. A large number of fiber-like macrostructures are also seen in the micrograph (Fig. 2a). At a higher magnification (Fig. 2b), it is possible to observe irregular microsized particles with round open pores, similar to that observed for activated carbon (Deng et al. 2021; Gan 2021). SEM analysis at higher magnification was carried out to show in more details the nano-micro structure of the SCBA (Fig. S2).

In Fig. 3, in which the XRD pattern is displayed, it can be observed an overall amorphous behavior of the ashes, characterized by a broad peak at a maximum of 23°. There is also a crystalline character, indicated by the peak at 26.7°, and others small peaks. These reveal the presence of quartz (α -SiO₂) as the predominant crystalline material. This information is consistent by the chemical analysis of XRF, which indicates silicon as an element present in high quantity, and with other works (Patel 2020; Ribeiro and Morelli 2014).

The porosimetry characterization of the SCBA (Table 2) shows a BET specific surface area of $62.79 \text{ m}^2.\text{g}^{-1}$, a small value, but it must be taken into account that the ashes did not undergo activation processes. The average pore diameter



Fig. 2 Scanning electron microscopy of SCBA with magnification: a) 700 x; b) 5000 x



Fig. 3 X-ray diffraction pattern of SCBA

 Table 2
 Data for the pore structure properties of the sugarcane bagasse ash
 Description
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 Description

BET specific surface area $(m^2.g^{-1})$	62.79
Total pore volume for pores with Diameter less than 261.13 nm at P/Po = 0.992609 (cm ³ .g ⁻¹)	0.05283
Average pore diameter (nm)	3.366

calculated was 3.37 nm, classified, according to IUPAC, as mesoporous (Rouquerol et al. 1994).

In Fig. 4, the FTIR spectrum shows a band between 3800 and 3000 cm^{-1} , assigned to vibration bands due to the group



Fig. 4 The FTIR spectra of the SCBA

hydroxyl (OH⁻) (Frías et al. 2011; Mor et al. 2019). The presence of three bands in the range of OH-group vibrations was noted in the FT-IR spectrum. A narrows peaks at 3696 and 3640 cm⁻¹ can be attributed to OH stretching vibration by hydroxyl groups (Salman et al. 2011a) derived from the regularly distributed group OH⁻ in the structure. The wide band at 3180 cm⁻¹ is related to a randomly placed hydroxyl group in the structure (Kalembkiewicz et al. 2018). Between 2830–2965 cm⁻¹ peak showed stretching correspond to asymmetric and symmetric vibrations of -CH₂ group (Ahmad et al. 2018). The peak at 2340 cm⁻¹ denotes C=C stretching (Salman et al. 2011a). The band corresponding to the carboxylic group C=O is observed as small shoulder at 1720 cm⁻¹, whereas the band due to O–H deformation coupled with C–O stretching vibrations can be observed at 1237 cm⁻¹. The peak at 1060 cm⁻¹ indicates the existence of C–O stretching vibrations in alcohols, phenols, acids, ethers, or esters (Trivedi et al. 2016). Silverstein et al. (2005) reported that strong Si–O bands appear at 830–1110 cm⁻¹, which is consistent with 1040 cm⁻¹ band found on the FTIR spectra of SCBA. The band at approximately 798 cm⁻¹ is attributed to the Si–O–Si symmetric stretching vibration of quartz (Yan et al. 2012). The band at 467 cm⁻¹ in the IR spectrum is related to the bending vibrations of SiO₄ tetrahedra, related to the presence of silica glass and quartz (Jovanovski and Makreski 2016). All those intense transmittance bands evidence the large amount of silica on composition of the SCBA.

The TGA/DTG curves of the SCBA are shown in Fig. 5. As noted, there are four main thermal events: at 42.14 °C, 312.38 °C, 534.75 °C and 916.11 °C. Under air atmosphere, the first and second zones correspond to removal of water moisture and light volatiles up to 423 °C, with a total loss of about 7.6%. Then, there is the active pyrolysis and oxidation zone from 423 °C to 615 °C with total degradation of 16.3%. The sample contains carbonaceous matter produced as a consequence of the incomplete combustion of bagasse, in such a way that part of the fibres was decomposed and carbonized in the boilers (Cordeiro et al. 2009). Subsequently, the sample weight remains almost constant with total degradation of 0.4% up to 1000 °C. The residue left at 1000 °C is about 75.7% of the original sample weight. Therefore, the total mass loss was 24.3%, which is consistent with the loss on ignition analysis (23.6%).

The point of zero charge has traditionally been defined as the pH value at which the surface density of positive charges equals that of negative charges (ie, zero net electric charge); is a characteristic parameter for a given surface of a solid in a given aqueous solution (electrolyte) (Rey et al. 2017; Sposito 1998). Their knowledge is important to predict the ionization of surface functional groups and their interaction with species in solution; for pH values lower than pH_{pre}, the surface charge is positive and the adsorption of anions is favored (Freitas et al. 2015). The graph of initial pH versus ΔpH was obtained (Fig. 6) and the null intersection point corresponded to the zero charge potential. The pH_{pzc} for the adsorbent in this study was 8.45. Anionic dyes have components that cause their molecule dissociate into negatively charged ions in an aqueous solution (Salleh et al. 2011). The Acid Red 27 contain sulfonic groups, present as sodium sulfonates, which give the dye molecules a negative charge, and making the AR27 an anionic dye. At pH values lower than pH_{pzc}, the adsorbent surface charge is positive, favoring the adsorption of anions (Freitas et al. 2015). Since the AR27 dye is anionic, the range to be used in further studies is pH < 8.45.

Adsorption results

Adsorption kinetics and effect of contact time

The adsorption kinetics is an important factor to evaluate the performance of the adsorbent against a certain adsorbate. It describes the rate of adsorbate removal in the fluid medium and the time required for the adsorption to complete. The pseudo-first-order (PFO), pseudo-second-order (PSO), Intraparticle Diffusion and Elocivh models were used to fit the time-concentration profile of adsorption of AR27 onto SCBA adsorbent. The models were evaluated using residual sum of squares (RSS), coefficient of determination (R²) and chi-square factor (χ^2). The kinetic parameters obtained from the models are presented in Table 3.

The kinetic experiment showed that the steady state was reached in 240 min (Fig. 7). The initial slope observed in the



Fig. 5 TGA/DTG curve of SCBA



Fig. 6 Determination of pH_{pzc} for SCBA

 Table 3
 Kinetic parameters for AR27 dye adsorption by SCBA adsorbent and residual sum of squares (RSS) from nonlinear regression analysis

Adsorption kinetics	Parameter	Value
Pseudo-first-order	Predicted $q_e (mg.g^{-1})$	1.8214 ± 0.0553
	$k_1 (min^{-1})$	0.0203 ± 0.0020
	RSS	0.1159
	\mathbb{R}^2	0.9662
	χ^2	0.0116
Pseudo-second-order	Predicted $q_e (mg.g^{-1})$	2.1240 ± 0.0511
	$k_2 (g.mg^{-1}.min^{-1})$	0.0113 ± 0.0012
	RSS	0.0380
	\mathbb{R}^2	0.9889
	χ^2	0.0038
Intraparticle diffusion	$k_{id} (mg.g^{-1}.min^{-1/2})$	0.0982 ± 0.0097
	$C (mg.g^{-1})$	0.2970 ± 0.1071
	RSS	0.3366
	\mathbb{R}^2	0.9018
	χ^2	0.0336
Elovich	α	0.0883 ± 0.0126
	β	2.1005 ± 0.1301
	RSS	0.0582
	\mathbb{R}^2	0.9830
	χ^2	0.0058



Fig. 7 Kinetic models plots for the removal of AR27 by SCBA

first 200 min of contact reflects rapid initial surface adsorption due to availability of SCBA sorption sites for the AR27 dye, before a period of slow adsorption as available sites decrease and equilibrium is reached after site saturation. The pseudo-second order model best fitted the experimental data, with a coefficient of determination (R^2) equal to 0.989, greater than the others, and a chi-square factor (χ^2) equal to 0.0038 and RSS equal to 0.03798, smaller than the others. In this model, it is assumed that the rate-limiting step may be surface adsorption, where the removal of the compound from a solution is due to physicochemical interactions between adsorbent and adsorbate (Robati 2013; Ho and McKay 1999).

Effect of solution pH on dye adsorption

The adsorption behavior of the AR27 dye by the SCBA at pH 2 to 12 are shown in Fig. 8. The SCBA had a maximum adsorption of the AR27 dye equal to 77% at $pH_0=2.0$; At $pH_0=3.0$ it dropped to 35.3%. Removal dropped to an average of approximately 29.0%, remaining nearly constant over the pH_0 range of 4 to 9. From this pH value onwards, removal dropped to 2.3% at $pH_0=12$.

Under highly basic conditions, the dye removal efficiency and the adsorption capacity decreased due to the accumulation of negative charge on the adsorbent surface, generating an electrostatic repulsion with the deprotonated dye molecules (sulfonate groups). This repulsive force between the anionic dye molecules and the SCBA surface was mainly responsible for the very low dye removal at high pH.

In a solution with a very acidic pH, it was expected that the removal capacity would decrease if the electrostatic interaction were the only mechanism for dye adsorption (Al-Degs et al. 2008), since the adsorbent would be positively charged and with its carboxylic groups protonated and the dye molecules would be neutral or partially positively charged. At this pH, the sulfonate groups of the dyes would be practically all protonated (—SO₃H, that is, neutral) and protonation of nitrogen atoms is also likely. Therefore, it is possible that the adsorption of AR27 on SCBA also occurs through hydrogen bonds and ion- π like interaction. The neutral characteristic of the dye salt AR27 focuses on practically



Fig. 8 Effect of solution pH on extent of dye adsorption (Experimental conditions: initial dye concentration 30 mg.L⁻¹; adsorbent dosage 4 g.L⁻¹; stirring speed 300 rpm; temperature: 30 ± 1 °C)

do not change the pH_{pzc} and the ΔpH curve shape of the SCBA.

Effect of Adsorbent Dosage

From Fig. 9, it is clear that the percentage of dye removal increased as the amount of SCBA was increased, up to a dosage of 6 g.L⁻¹; thereafter, the removal percentage underwent little change, as the incremental dye removal becomes very low as its concentration on the surface and in the solution reach equilibrium each other. The initial increase in removal efficiency can be attributed to the fact that more surface area is available, with consequent availability of a greater number of active sites, with increasing adsorbent dose. However, continuously increasing the dosage maintains or even may cause a drop in the removal value due to adsorbent particle interactions such as agglomeration, leading to a decrease in the total available surface area of the adsorbent (Mosoarca et al. 2020). Indeed, the adsorption capacity of AR27 decreased with increasing adsorbent dosage, a reduction in the adsorption efficiency of the adsorbent per mass unit. Therefore, to weigh the material saving and the adsorption capacity and removal percentage, the optimal adsorbent dosage was selected as 4 $g.L^{-1}$ for subsequent experiments.

Adsorption isotherm

Linear and nonlinear isotherm models such as Freundlich, Langmuir, Temkin and were used to fit adsorption experimental data, Fig. 10. The residual sum of squares (RSS), coefficient of determination (\mathbb{R}^2) and chi-square factor (χ^2) were used to test all models used in this study. The parameters of investigated isotherms obtained from the models are presented in Table 4. The adsorption equilibrium experiments showed that the experimental adsorptive

Fig. 9 Effect of adsorbent dosage on the adsorption of dye AR27 by SCBA (Experimental conditions: initial dye concentration 30 mg.L⁻¹; pH₀=2,0; stirring speed 300 rpm; temperature = 30 ± 1 °C, t=4 h)

capacity reaches 10.20 mg.g⁻¹. It is observed that the Freundlich model was the one that best adjusted to the equilibrium data. The Freundlich isotherm is convenient when intending to describe a non-ideal and reversible adsorptive system, on heterogeneous surfaces (adsorbent surface sites have a spectrum of different binding energies), not restricted to monolayer formation, and can be applied to multilayer adsorption (Ayawei et al. 2017; Pezoti et al. 2016). This result corroborates the fact that ash comes from sugarcane bagasse, which has a heterogeneous microstructure and an extremely turbulent burning process.

Table 5 lists a comparison of maximum monolayer adsorption capacity of dye AR27 with various adsorbents and some dyes with the bagasse ash as adsorbent. From this table, it is important to highlight that depending on the adsorbent-adsorbate interactions and the surface area available for adsorption, the q_{max} for a certain adsorbent can vary considerably. As expected, activated carbon presents the best adsorption results for the different adsorbates presented, except for the adsorption of Basic green 4 on commercial activated carbon. For sugarcane ash, the value found here is consistent with other works, presenting a maximum value of 15.01 mg.g⁻¹, one order of magnitude. Above that we have the adsorption of basic green 4 and methylene blue, with two orders of magnitude for ash. Despite its low adsorptive capacity for Acid Red 27 is important to consider that bagasse ash is an industrial tailing, would be acquired for free and used without pretreatment. In the search for an environmentally friend economy this lower environmental impact issue is of most importance. Nevertheless, a more detailed study is needed to quantify the real impact for the environment and economy with the application of SCBA for water and effluents treatments.







Fig. 10 Equilibrium isotherms for the adsorption of AR27 by SCBA: (a) Lagmuir, Freundlich and Temkin non-linearized; (b) Lagmuir linearized; (c) Freundlich linearized; (d) Temkin linearized (Experimen-

tal conditions: adsorbent dosage=4 g.L⁻¹; mixing rate=300 rpm; $T=30\pm1$ °C; initial conc.: 2.5—240 mg.L.⁻¹; contact time=4 h)

Table 4 Coefficients obtained for sorption isotherm models	Isotherm model	Parameter	Linear	Nonlinear
for AR27 dye adsorption by SCBA and RSS, R^2 and χ^2 from	Langmuir	$q_{max} (mg.g^{-1})$	14.3781 ± 0.7318	15.0083 ± 0.8738
		$K_L (L.mg^{-1})$	0.0545 ± 0.0145	0.03675 ± 0.0080
analysis (Experimental		RSS	9.0355	13.4012
conditions: adsorbent		\mathbb{R}^2	0.9649	0.9432
dosage = 4 g.L ⁻¹ ; mixing rate = 300 rpm; T = 30 ± 1 °C; initial conc.: 2.5—240 mg.L ⁻¹ ; contact time = 4 h)		χ^2		1.0309
	Freundlich	$K_{F}([(mg.g^{-1}).(mg.L^{-1})^{-1/n}])$	2.6029 ± 0.0899	2.4595 ± 0.2216
		n	3.1268 ± 0.0810	3.0029 ± 0.1779
		RSS	0.0021	3.5165
		\mathbb{R}^2	0.9920	0.9851
		χ^2		0.2705
	Temkin	K_{T} (L.mg ⁻¹)	0.7098 ± 0.5926	0.7098 ± 0.2178
		$B = \frac{RT}{L}$	2.7014 ± 0.2003	2.7014 ± 0.2156
		RSS	6.5676	8.9913
		\mathbb{R}^2	0.9378	0.9619
		χ^2		0.6916

Adsorbent	Preparation	Dye	$q_{max} (mg.g^{-1})$	Ref
Sugarcane bagasse ash	Residue of burning in high pressure boiler	Acid Red 27	15.01	This work
Activated Carbon Web Prepared from Acrylic Fibrous Waste	Physical activation under the layer of charcoal in high temperature furnace (800–1200 °C)	Acid Red 27	21.68	Salman-Naeem et al. 2018
Chitosan-polyethyleneimine beads	Graft copolymerization under micro- wave irradiation	Acid Red 27	48.30	Yusof et al. 2020
Activated Carbon Prepared from Coco- nut Husk Fiber	Activation with ZnCl ₂ at high tempera- ture	Acid Red 27	333.33	Al-Aoh et al. 2013
Commercial Activated Carbon	Commercial product	Acid Red 27	166.67	Al-Aoh et al. 2013
γ -Fe ₂ O ₃ nanoadsorbents	Commercial product	Acid Red 27	16.32	Nassar 2010
Polypyrrole/SrFe ₁₂ O ₁₉ /graphene oxide nanocomposite	The SrFe ₁₂ O ₁₉ (SrM) nanoparticles were synthesized by citrate sol–gel process. The GO particles were synthesized by Hummers' method. The PPY/SrM/GO nanocomposite powder was fabricated through in situ polymerization	Acid Red 27	294.11	Ebrahimpoor et al. 2019
Bagasse fly ash	Residue of burning in high pressure boiler	Basic green 4	116.28	Mane et al. 2007
Bagasse fly ash	Residue of burning in high pressure boiler	Basic Green 4	170.33	Mall et al. 2005
Activated carbons commercial grade	Commercial product		8.27	
Activated carbons laboratory grade	Commercial product		42.18	
Bagasse fly ash	Residue of burning in high pressure boiler	Acid Orange 10	18.80	Mall et al. 2006
		Basic violet 1	26.25	
Bagasse fly ash	Residue of burning in high pressure boiler	Methylene blue	64.61	Gupta et al. 2000

 Table 5
 Comparison of the maximum monolayer adsorption of dyes onto adsorbents

Adsorption thermodynamics and effect of temperature

Adsorption thermodynamic parameters were derived from the linear form of the van't Hoff equation and the thermodynamic definition of chemical equilibrium (Lima et al. 2020), depicted in the Eqs. 3 and 4, respectively.

$$\ln b = \frac{\Delta S^{\circ}}{R} - \frac{\Delta H^{\circ}}{RT}$$
(3)

$$\Delta G^{\circ} = -RTlnb \tag{4}$$

where ΔG° is the change in the Gibbs free energy (kJ. mol⁻¹); ΔH° is the change in enthalpy (kJ.mol⁻¹), and ΔS° is the change in entropy (J.mol⁻¹.K⁻¹), R is the gas constant (8.314 J.mol⁻¹.K⁻¹), T is the thermodynamic temperature (K). The adsorption equilibrium constant (b) was derived from the isotherm data at different temperatures using the Langmuir and Temkin nonlinear models (K_L and K_T, expressed in L. mg⁻¹). The equilibrium constants K_L and K_T must become dimensionless for use in the Vant' Hoff equation; for this purpose, Eq. 5 was used (Lima et al. 2019).

$$b = K_{ads}^{\circ} = \frac{(1000.K.molecular weight of adsorbate).[Adsorbate]^{\circ}}{\gamma}$$
(5)

where γ is the coefficient of activity (dimensionless, it is considered that the adsorbate solution is very diluted to consider that its value is unitary), [Adsorbate]° is the standard concentration of the adsorbate (1 mol.L⁻¹), K is the and K°_{ads} is the thermodynamic equilibrium constant that is dimensionless.

The parameters of the investigated isotherms obtained by the linear and nonlinear Langmuir, Freundlich, Tenkim models are presented in Table S3 for the three temperatures. It can be seen that for these temperatures the model that best fits the experimental data were also the Freundlich models. Furthermore, we have added the BET type II isotherm model to proceed the thermodynamic calculation with better fitted data, as Freundlich parameters does not have thermodynamic correlation to Vant' Hoff equation. The parameters of BET isotherm model are listed in Table S4. The BET model is an extension of Langmuir's theory and considers that adsorption can occur



Fig. 11 Effect of temperature on AR27 adsorption performance by SCBA (Experimental conditions: adsorbent dosage=4 g.L⁻¹; mixing rate=300 rpm; $T=30\pm1$ °C; initial conc.: 2.5—240 mg.L⁻¹; contact time=4 h)

 $\label{eq:table_formula} \begin{array}{l} \mbox{Table 6} & \mbox{Thermodynamic parameters for adsorption of AR27 onto SCBA} \end{array}$

Adsorption modeling		Temperature (K)		
		303	321	338
Langmuir	$\Delta G^{o}_{ads} (kJ.mol^{-1})$	-33.47	-35.46	-37.34
	$\Delta H^{o}_{ads} (kJ.mol^{-1})$		7.33	
	$\Delta S^{o}_{ads} (J.mol^{-1}.K^{-1})$		110.49	
Temkin	$\Delta G^{o}_{ads} (kJ.mol^{-1})$	-48.20	-51.07	-53.77
	$\Delta H^{o}_{ads} (kJ.mol^{-1})$		14.48	
	$\Delta S^{o}_{ads} (J.mol^{-1}.K^{-1})$		159.14	
BET	$\Delta G^{o}_{ads} (kJ.mol^{-1})$	-10.25	-11.49	-12.71
	$\Delta H^{o}_{ads} (kJ.mol^{-1})$		11.47	
	$\Delta S^{o}_{ads} (J.mol^{-1}.K^{-1})$		71.53	

through multiple layers of adsorbate on the surface of the adsorbent. A non-linear version of this model, the BET type II isotherm is used to describe adsorptive processes in the liquid phase, and its mathematical model is given by Equation S11 in Table S2. For the BET model, the equilibrium constant is at dimensionless form, and can be applied directly to the Vant' Hoff equation.

Figure 11 shows the adsorption isotherms for AR27 dye on SCBA at three different temperatures. Table 6 shows the values of the thermodynamic parameters ΔG^{o}_{ads} , ΔH^{o}_{ads} , ΔS^{o}_{ads} for the adsorption of AR27, using the K_L (Langmuir) K_T (Temkin) and K_B (BET)constants obtained from the isotherms at temperatures of 303, 321 and 338 K. In Fig. S3 we can observe the graph of ln b versus T⁻¹ (K⁻¹) used in the determination of thermodynamic parameters.

The magnitude of the enthalpy (less than 40 kJ.mol⁻¹) suggests that the adsorption process occurs through a physisorption (physical process), since it does not assume sufficient values to indicate ruptures and formation of chemical bonds. The low values of the enthalpy evidence that the interaction between AR27 dye and the SCBA may be weak, as there is a tendency that as greater the energy involved as greater the propensity for formation of chemical bonds between adsorbent-adsorbate (Sahmoune 2018).

The negative ΔG°_{ads} and positive ΔS°_{ads} confirm a spontaneous adsorption process and suggests affinity of the SCBA for AR27 dye, with an increase in spontaneity as temperature increases from 303 to 338 K. The endothermic value obtained for ΔH^{o}_{ads} appears to be inconsistent with the decrease of adsorption as the increase of the temperature. Some apparently inconsistent data have been founded in literature, suggesting an endothermic process, however the maximum capacity of adsorption decrease with the temperature increasing (Al-Aoh et al. 2013; Salman et al. 2011b, 2011a; Alver and Metin 2012; Mittal 2006). This behavior have been liked to the isosteric enthalpy, in analogy to the physisorption of gases but not described in depth (Mane et al. 2007). As expected for a physisorption process, the increase of the temperature enhance the distribution of dye molecule with energy higher than the electrostatic potential. which reflect at the decrease of the dye adsorption (Nuhnen and Janiak 2020).

In this way, its suggest need to understand the physisorption in aqueous environment as a combination of two main processes, the desorption of water ΔH^o_W and the adsorption of the dye ΔH^o_D at the ash surface. Thus, the constant of the equilibrium determined by the isothermal modeling need to involve the contribution of both processes, as follow Eqs. 6–8, where: W is the water; S is the substrate; and D is the dye.

$$WS \overleftrightarrow W + S \tag{6}$$

$$D + S \rightarrow DS$$
 (7)

$$\mathbf{b} = \mathbf{b}_{\mathrm{w}} \cdot \mathbf{b}_{\mathrm{dye}} = \frac{[\mathbf{W}][\mathbf{S}]}{[\mathbf{WS}]} \cdot \frac{[\mathbf{DS}]}{[\mathbf{D}][\mathbf{S}]} \tag{8}$$

Above consideration implies the ΔH°_{ads} of Van't Hoff be a contribution of $\Delta H^{\circ}_{Wdes} + \Delta H^{\circ}_{Dads}$. Although, this enthalpy consider only the energy involved in the equilibrium of the reaction. However in analogy of gas physisorption its possible to involve the isosteric enthalpy of the adsorption (as a condensation like process) of the dye, ΔH_D (q_e), onto the SCBA surface. Therefore, for a dye physisorption the ΔH_D (q_e) assumed can be calculated using the adapted Clausius-Clapeyron approach, Eq. 9, for P=RTC_e. The slope m of the straight line with the three data points at ln(T₁C_{e1})|1/T₁, $\ln(T_2C_{e2})|1/T_2$ and $\ln(T_3C_{e3})1/T_3$ at equal loading q_e will give $\Delta H_D(q_e)$ according to Eq. 10 (Nuhnen and Janiak 2020).

$$\frac{d(RTC_e)}{dT} = \frac{C_e \cdot \Delta H_D(q_e)}{-T}$$
(9)

$$\Delta H_{\rm D}(q_{\rm e}) = \mathbf{m} \cdot \mathbf{R} \tag{10}$$

A last consideration can be made for the "first few molecules" adsorbed $(q_e \rightarrow 0)$ (Nuhnen and Janiak 2020). In this stage, the ΔV of the "dye condensation" onto surface of the adsorbent tend to zero $(\Delta V \rightarrow 0)$ and is reasonable to consider that $\Delta H^o_{Dads} \approx \Delta H_D (q_e \rightarrow 0) = \Delta H^o_D$. In conclusion, when $\Delta H^o_{Dads} < 0$, the seeming discussed inconsistence happens for $|\Delta H^o_{Dads}| < |\Delta H^o_{Wdes}|$, resulting in a positive value of the ΔH^o_{ads} in the Vant' Hoff equation; also for $\Delta H^o_{Dads} > 0$.

Additionally, the ΔH^{o}_{D} value can differ distinctly from ΔH_{D} for the bulk adsorption. For the AR27 adsorption onto SCBA, the approach of the $\Delta H_{D}(q_{e})$ values based on Freundlich model gives the values of $\Delta H^{o}_{D} = 9.2$ kJ.mol⁻¹, described in Supplementary Information (S4). Considering the value of the $\Delta H^{o}_{ads} = 11.5$ kJ.mol⁻¹ obtained from BET approaches, the value for the ΔH^{o}_{Wdes} was 2.3 kJ.mol⁻¹. $\Delta H_{ads}(q_{e}) = \Delta H^{o}_{Wdes} + \Delta H_{D}(q_{e})$ was plotted in Fig. S11, considering the ΔH^{o}_{Wdes} as a constant.

Applying this same procedure for the others apparently inconsistent values of the ΔH^{o}_{ads} enthalpy available in literature, Table S5, the values of ΔH^{o}_{Wdes} calculated were between 3.5 and 31.0 kJ.mol⁻¹. Despite some adsorbents initially presented an endothermic behavior, analyzing the Fig. S5 to S10 was clear that increasing the dye adsorbed amount the isosteric enthalpy changes reaching to be an exothermic behavior which explain the unfavorable adsorption with the increase of the temperature to high adsorbed amount.

Influence of coexisting anions

Coexisting anions such as chloride, sulfate and bicarbonate are typically present in effluents. These anions can compete with the adsorbate and interfere in the removal efficiency. As a result, the effects of these coexisting anions on the AR27 adsorption are shown in Fig. 12.

The presence of chloride anions proved to have little effect on the removal efficiency of AR27, as the removal decreased from 66.8% to 63.6%. On the other hand, the effects of sulfate and bicarbonate anions on the removal of AR27 were significant. In the presence of sulfate at 0.1 M, the removal efficiency of AR27 decreased to 43.5%, while at a concentration of 0.1 M of bicarbonate, the removal of AR27 decreased to 28.2%. In other words, sulfate and bicarbonate compete more with AR27 for active sites in the SCBA than does the chloride anion. In the concomitant



Fig. 12 Effect of coexisting anions on AR27 removal at fixed initial concentration of AR27=30 mg.L⁻¹, coexisting anion concentration=0.1 M, adsorbent dose=4 g.L⁻¹, shaking rate=300 rpm, contact time=4 h, pH=2, T=30 \pm 1 °C



Fig. 13 Batch adsorption/desorption cycles after regeneration with NaOH (Experimental conditions: initial concentration of AR27=30 mg.L⁻¹, adsorbent dose=4 g.L⁻¹, shaking rate=300 rpm, contact time=4 h, pH=2, T=30 \pm 1 °C)

presence of the mentioned anions (0.1 M of each), the removal fell to 30.9%.

Regeneration and reuse of SCBA

Three SCBA adsorption-regeneration cycles were performed as shown in Fig. 13. For regeneration with distilled water, the adsorbent lost around 70% of its removal efficiency after three consecutive cycles of adsorption–desorption experiments. For regeneration with 0.01 M NaOH, the efficiency of the first cycle remained at the same level as the original cycle, but decreased by 35% after the next two cycles. For regeneration with 0.1 and 0.3 M NaOH, the removal efficiency increased slightly after the first cycle, decreasing in the next two, but managing to maintain > 50% removal after the third cycle.

Therefore, SCBA has great potential for use in the treatment of effluents containing AR27 dye, because it can be reused through simple regeneration, and combined with its ability to maintain removal efficiency after two cycles, it becomes even more economical and sustainable.

Conclusions

The present study shows that sugarcane bagasse ash, a residue obtained from the sugar mill, can be used as an adsorbent for the removal of Acid Red 27 dye in aqueous solution. The kinetics study showed that equilibrium was reached in 4 h, with the adsorption kinetics best represented by the pseudo-second order model. The pH influence to the adsorptive equilibrium showed that the maximum adsorption of the dye by the ash occurred at pH_0 equal to 2.0. For the equilibrium study, the results showed that the Freundlich isotherm described well the adsorption behavior of the dye in the studied concentration ranges, with adsorption capacity of 14 mg.g⁻¹ to the most concentrated solution studied at optimal pH₀; the maximum adsorption capacity of Langmuir was 15 mg.g $^{-1}$. The thermodynamic adsorption parameters evaluated by Vant' Hoff equation with fitted parameter for Langmuir, Temkin and BET demonstrate an endothermic, spontaneous and indicate a physisorption mechanism, with ΔH°_{ads} < 15 kJ.mol⁻¹. Clausius-Clapeyron approach adapted for adsorption in liquid environment was successful applied to determine the isosteric enthalpy behavior of Acid Red 27 at sugarcane bagasse ash, which the $\Delta H^{\circ}_{D} = 9.2 \text{ kJ.mol}^{-1}$ was founded. However, ΔH_D for the bulk adsorption is exothermic reaching absolute values higher than 25 kJ.mol⁻¹. The effect of coexisting anions were studied for the presence of chloride, sulfate and bicarbonate. The results showed small interference of chloride but high influence of bicarbonate, with adsorption capacity reaching reduction of almost 60%. Regeneration of SCBA with sodium hydroxide was able to preserve its initial removal capacity for two cycles. Sugarcane bagasse ash seems to have potential to be used as a cheap adsorbent in the removal of Acid Red 27 dye in low concentrated solutions.

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Author contributions All authors contributed to the study conception and design. Material preparation, data collection and analysis were performed by Joana Eliza de Santana, Fábio Gabriel Silva de Andrade and Aldebarã Fausto Ferreira. The review, supervision and project administration were performed by Marcos Gomes Ghislandi and Maurício Alves da Motta Sobrinho. The first draft of the manuscript was written by Joana Eliza de Santana and all authors commented on previous versions of the manuscript. All authors read and approved the final manuscript.

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Data Availability The authors declare that the data supporting the findings of this study are available within the paper and its Supplementary Information files. Should any raw data files be needed in another format they are available from the corresponding author upon reasonable request.

Declarations

Ethical approval Not applicable.

Consent to participate Not applicable.

Consent to publish Not applicable.

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