



# Analysis of melting reconstruction treatment and cement solidification on ultra-risk municipal solid waste incinerator fly ash–blast furnace slag mixtures

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## Abstract

High temperature melting treatment and cement solidification are technologies currently used to reduce the leaching of heavy metals in municipal solid waste incinerator (MSWI) fly ash. In this paper, to ascertain the feasibility of melting MSWI fly ash with blast furnace (BF) slag, ultra-risk MSWI(U-MSWI) fly ash having high heavy metal (Zn, Pb, Cu, and Cr) contents were blended with BF slag, then melted and quenched into water to prepare reconstructed slag. The melting and solidification behaviors, phase composition and microstructure, and heavy metal leachability of reconstructed slag were studied. In addition, to study the further solidification and utilization of reconstructed slag in cement, the compressive strength and leaching concentration of cement composites with reconstructed slag were also investigated. The results indicate that the presence of heavy metals in the U-MSWI fly ash had a little influence on the microstructure and phase composition of reconstructed slag. The leaching concentration of heavy metals in the reconstructed slag increased with the increasing of U-MSWI fly ash content, and when the content of U-MSWI fly ash was less than 50 wt%, the reconstructed slag could meet the environmental requirements. The reconstructed slag further solidified by cement could be applied to landfill and construction materials. The technology of melting reconstruction treatment with cement solidification was a technical-economical choice for the industrial treatment of U-MSWI fly ash.

**Keywords** Heavy metals · Reconstructed slag · Utilization · Leaching concentration · Landfill and construction materials · Industrial treatment

## Introduction

Municipal solid waste incinerator (MSWI) fly ash belongs to a kind of secondary pollutant generated by incineration of municipal solid waste in large volume, and it is hazardous for both human health and natural environment because of its saturation with harmful heavy metals (such as Zn, Pb, Cu,

and Cr) (Mee 2016; Bie et al. 2016). Therefore, suitable treatment technologies of MSWI fly ash are required to ensure safe disposal to the environment.

Currently, there are many treatment methods have been done to reduce the MSWI fly ash harmful impact on human health and natural environment. Generally, the treatments are classified into three main categories (Phua et al. 2019): (i) separation processes, such as washing processes (Wang et al. 2009; Funari et al. 2017), leaching processes (Luo et al. 2019; Loginova et al. 2019), electrochemical processes (Pedersen et al. 2005), and thermal evaporation (Nowak et al. 2012); (ii) solidification/stabilization processes, such as chemical stabilization (Youcai et al. 2002; Sun et al. 2011), accelerated carbonation (Todorovic and Ecke 2006), and chemical fixation with cement binder (Aubert et al. 2007; Zhan et al. 2018); (iii) thermal processes, such as sintering (Wey et al. 2006; Huber et al. 2018a; Huber et al. 2018b), vitrification (Alhadj-Mallah et al. 2015), melting (Ni et al. 2012), and hydrothermal (Shi et al. 2017a; Shi et al. 2017b). Wherein,

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the separation processes are often operated as pretreatment methods to remove heavy metals from MSWI fly ash before the utilization or solidification/stabilization of MSWI fly ash. Solidification/stabilization is an effective technology to immobilize heavy metals via chemical and physical modification (Malviya and Chaudhary 2006), among them the method of solidification with cement binder is widely adopted due to the mature technology, simple process, low cost, and the main components (CaO, SiO<sub>2</sub>, and Al<sub>2</sub>O<sub>3</sub>) of MSWI fly ash which are similar to cement (Guo et al. 2014; Yang et al. 2018; Zhan et al. 2018). However, due to the low activity and high heavy metal content of MSWI fly ash, the effective solidification amount of MSWI fly ash in cementitious materials is greatly limited. Besides, the Cl in MSWI fly ash do great harm to the mechanical properties of cement products (Lederer et al. 2017). The thermal process of MSWI fly ash generally aims to produce a more stable product, which can be disposed or utilized. The leaching concentration of heavy metals in MSWI fly ash can be reduced during the melting process (Jiang et al. 2009). However, the main disadvantage of melting process is usually associated with high treatment cost and high energy consumption, and the obtained molten slag cannot be efficiently recycled due to its low activity. Therefore, both processes of solidification with cement binder and melting have their own limitations.

Currently, the research and development of MSWI fly ash treatment technology is mainly focused on combining the advantages of solidification with cement binder and melting processes and compensating for the shortcomings of both processes, which is significant for its further development and applicability. Therefore, reducing the energy consumption and cost of MSWI fly ash melting process, improving the activity of the molten slag, and increasing the effective solidification amount of MSWI fly ash in cement should be solved. Blast furnace (BF) slag, as the byproduct during the iron-making process, generated by melting of various fluxes along with the gangue minerals during smelting. Globally, over 2.10 million tons of BF slag is generated every year in China, and its discharged temperature is extremely high, about 1450–1550 °C, with the carried energy about 1700 MJ/t<sub>slag</sub> (Xie et al. 2019). Therefore, such an available high-temperature source may be as a free high-temperature carrier to melt U-MSWI fly ash, and can greatly reduce the energy consumption and cost of U-MSWI fly ash melting treatment. Moreover, this melting treatment process can mix BF slag with U-MSWI fly ash and may produce reconstructed slag with certain potential activity. In addition, the reconstructed slag generated can be used as an admixture for further solidification and application in cement, increasing the solidification effect of cement on heavy metals in U-MSWI fly ash and promoting the application of reconstructed slag in building materials. Therefore, the method of combining melting reconstruction treatment and cement solidification on U-MSWI fly ash is feasible.

In this study, the ultra-risk municipal solid waste incinerator (U-MSWI) fly ash formed by mixing heavy metal ions with MSWI fly ash was first prepared. Then, using blast furnace (BF) slag in high temperature molten state as a high-temperature carrier to melt U-MSWI fly ash, water quenched into reconstructed slag, and then mixed into cement at room temperature for re-solidification. The experimental method in this paper realized the double solidification of U-MSWI fly ash for combining melting reconstruction treatment and cement solidification. Through the analysis of heavy metal leachability and compressive strength, the final solidified product can meet the environmental requirements and can be used in landfills and construction materials. This study would lay a groundwork for the future safe and economical treatment and utilization of MSWI fly ash based on industrial processing.

## Materials and methods

### Materials

MSWI fly ash used in this study was collected from the cyclone of a mass-burning incinerator located in the western part of Zhengzhou, China. The MSWI fly ash samples were dried and then ground to a fine powder with surface area of 420 m<sup>2</sup> kg<sup>-1</sup>. Blast furnace (BF) slags with a specific surface area of 355 m<sup>2</sup> kg<sup>-1</sup> were provided by an iron-making procedure from Steel Holding Group Co., Ltd. in Shanxi, China. P. O 42.5 type of Portland cement with a specific surface area of 345 m<sup>2</sup> kg<sup>-1</sup> was used from Cement Co., Ltd. in Zhengzhou. All raw materials were sieved (passing through a #200 sieve) and stored in desiccators throughout the duration of the study. ZnO, ZnCl<sub>2</sub>, Zn(NO<sub>3</sub>)<sub>2</sub>·6H<sub>2</sub>O, CuO, CuCl<sub>2</sub>·2H<sub>2</sub>O, Cu(NO<sub>3</sub>)<sub>2</sub>·3H<sub>2</sub>O, PbO, PbCl<sub>2</sub>, Pb(NO<sub>3</sub>)<sub>2</sub>, CrO<sub>3</sub>, Cr<sub>2</sub>O<sub>3</sub>, CrCl<sub>3</sub>·6H<sub>2</sub>O, and Cr(NO<sub>3</sub>)<sub>3</sub>·9H<sub>2</sub>O were of analytical grade from Sinopharm Chemical Reagent Co., Ltd, China, which were used as the heavy metal reagents. Deionized water was used in all experiments.

### Sample preparation

#### Preparation of U-MSWI fly ash

To prepare U-MSWI fly ash, the sieved MSWI fly ash was mixed with heavy metal reagents at a mass ratio of 10:1. The heavy metal reagents were added according to a Zn:Cu:Pb:Cr mass ratio of 100:100:5:15. Among them, Zn was provided by ZnO, ZnCl<sub>2</sub>, and Zn(NO<sub>3</sub>)<sub>2</sub>·6H<sub>2</sub>O at a mass ratio of 1:1:1, Cu was provided by CuO, CuCl<sub>2</sub>·2H<sub>2</sub>O, and Cu(NO<sub>3</sub>)<sub>2</sub>·3H<sub>2</sub>O at a mass ratio of 1:1:1, Pb was provided by PbO, PbCl<sub>2</sub>, and Pb(NO<sub>3</sub>)<sub>2</sub> at a mass ratio of 1:1:1, and Cr was provided by CrO<sub>3</sub>, Cr<sub>2</sub>O<sub>3</sub>, CrCl<sub>3</sub>·6H<sub>2</sub>O, and Cr(NO<sub>3</sub>)<sub>3</sub>·9H<sub>2</sub>O at a mass

ratio of 1.5:1:1:1. The U-MSWI fly ash was ground into powders to pass through a #200 sieve with a specific surface of approximately 460 m<sup>2</sup> kg<sup>-1</sup>.

**Formulation of U-MSWI fly ash-BF slag mixtures**

The as-prepared U-MSWI fly ash were mixed with sieve BF slags with varied contents ranging from 0 to 100 wt% to form U-MSWI fly ash-BF slag mixtures, then the mixtures were mixed evenly in a ZKF-WF Temperature controlled flip oscillator and pre-homogenized at 20 °C for 24 h to use. The initial melting temperature and time of U-MSWI fly ash-BF slag mixtures (~ 2 g) were tested by RT-TS1500. High temperature microscope and the basicity (CaO/SiO<sub>2</sub> weight ratio) of mixtures are presented in Table 1.

**Preparation of reconstructed slag**

In the laboratory, in order to study the mechanism of high temperature melting and reconstruction features of U-MSWI fly ash-BF slag mixtures, the SLQ1600-30 Electrically Heated Furnace was used to carry out the melting reconstruction experiment of the U-MSWI fly ash-BF slag mixtures. The sample 0–8 (~ 150 g each) were melted in the electrically heated furnace according to the results from initial melting temperature and time determination of the U-MSWI fly ash-BF slag mixture were shown in Table 1. However, due to the increase in the amount of sample during the experiment, the inside of all samples could not reach to a molten state, and the samples had poor fluidity and were easy to stick to the crucible after being released from the furnace. Therefore, combining with the discharged temperature (1450–1550 °C) of BF slag in actual industry processing and further debugging of the complete melting temperature of each sample (0–8), the reconstructed slag was formed by melting at 1400 °C for 150 s and quenching into water. Then the reconstructed slag samples were all dried at 105 °C for 12 h and ground into powders

to pass through a #200 sieve with a specific surface of approximately 450.7 m<sup>2</sup> kg<sup>-1</sup>.

Moreover, in order to simulate the industrial preparation of reconstructed slag and prepare cement composites with reconstructed slag, KGPS-50KW-2500HZ Induction Heating Device was used. The specific simulation preparation steps of reconstructed slag are shown in Fig. 1. The core temperature of the completely molten U-MSWI fly ash-BF slag mixtures reached about 1400 °C through temperature measurement, and the reconstructed slag was formed by quenching into water. This is basically consistent with the results of experiments performed with the SLQ1600-30 Electrically Heated Furnace.

**Preparation of cement composites**

In this work, the cement composites with replaced materials of reconstructed slag with different content of U-MSWI fly ash were prepared. The designed proportion of these cement composites are listed in Table 2.

**Sample analysis**

**Composition and microscopic structure tests**

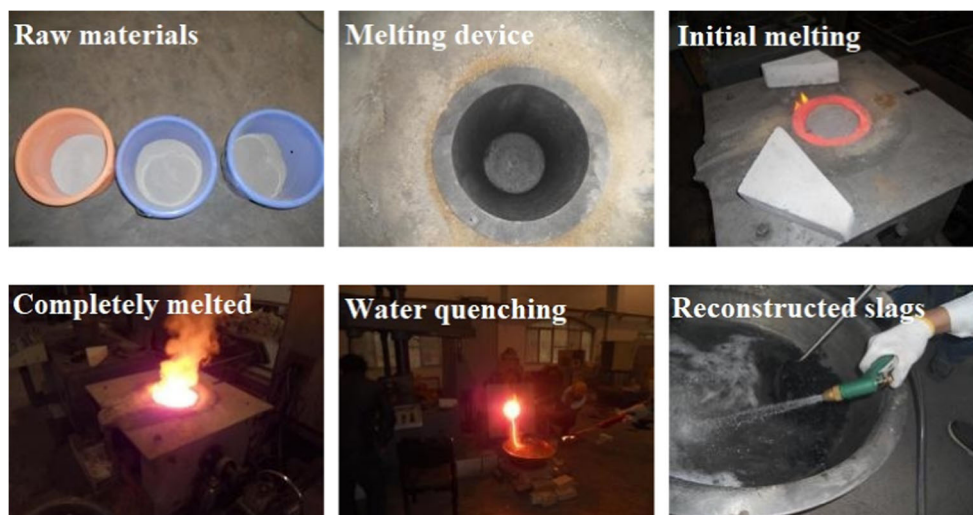
The chemical compositions of BF slag, MSWI fly ash and U-MSWI fly ash was determined by X-ray Fluorescence Spectrometry (XRF, Bruker, S4-Explore, Germany). The measured time was set for 100 s with the counting rate of 27,000 cps at 100 mA and 40 kV.

The phase compositions of BF slag, MSWI fly ash and reconstructed slag was identified by means of X-ray Diffraction (XRD, Panalytical, X’Pert PRO MPD, Holland), using CuK<sub>α</sub> radiation at 30 mA and 45 kV. Measurements were carried out in the 2θ angle range of 5–80°, with a step of 0.02° and counting time 2 s per step. Mineral identification was done in an X’Pert HighScore Plus application with PDF 2004.

**Table 1** Designed proportion of U-MSWI fly ash-BF slag mixtures and its melting characteristics

Sample	BF slag /wt%	U-MSWI fly ash/wt%	Initial melting temperature/°C	Initial melting time/s	Basicity (CaO/SiO <sub>2</sub> weight ratio)
0	100	0	1180	118	0.81
1	95	5	1173	117	0.84
2	90	10	1175	125	0.85
3	85	15	1180	118	0.88
4	80	20	1180	119	0.89
5	70	30	1204	145	0.99
6	50	50	1199	123	0.91
7	20	80	1171	117	0.80
8	0	100	1170	117	0.76

**Fig. 1** Specific simulation preparation steps of reconstructed slag



The microstructure of reconstructed slag was examined using the Scanning Electron Microscope (SEM, JEOL, JSM-6700F, Japan) with an acceleration voltage of 20 kV for gold-coated, loose powder samples. Images were taken under secondary electron mode.

Absorption ATR spectra of reconstructed slag was measured using a FT-IR Spectrometer (FT-IR, ALPHA, FT/IR-300E, Germany) with a KBr beamsplitter. The spectrometer was equipped with attenuated total reflection (ATR) accessory with heated diamond crystal plate and temperature controller. ATR instrument during measurements was purged with dry air to remove water vapor. The reconstructed slag was put to the ATR cell immediately after being mixed and then sealed with Teflon tape. Spectra were recorded in the region of  $4000\text{--}500\text{ cm}^{-1}$  at a resolution of  $2\text{ cm}^{-1}$ . The instrument was controlled by OPUS 5.5 software, and the baseline correction was carried out with its use.

The elements distribution on the surface of reconstructed slag was characterized by Electron Probe Microanalyzer (EPMA, JEOL, JXA-8230, Japan) equipped with Oxford Instruments Energy Dispersive Spectrometer (EDS, INCA, X-Act, England). The samples were coated with carbon before transferred for analysis.

#### Leaching behavior

The leaching behavior of MSWI fly ash, U-MSWI fly ash, reconstructed slag, and cement composites was evaluated by the Toxicity Characteristic Leaching Procedure (TCLP) test according to specification SW 846-1311 (US Environmental Protection Agency 1986). The leachates were analyzed by using inductively coupled plasma-atomic emission spectroscopy (ICP-AES, PLASMA300, America) to determine the concentrations of heavy metal Zn, Cu, Pb, and Cr according to the determination method of solid waste elements in the

**Table 2** Designed proportion of cement composites

Sample	Replaced material	Dosage of replaced materials/wt%	Dosage of PO42.5 OPC/wt%
F0	-	0	100
F1	#0 Reconstructed slag (0% U-MSWI fly ash)	50	50
F2	#2 Reconstructed slag (10% U-MSWI fly ash)	50	50
F3	#5 Reconstructed slag (30% U-MSWI fly ash)	50	50
F4	#6 Reconstructed slag (50% U-MSWI fly ash)	50	50
F5	#7 Reconstructed slag (80% U-MSWI fly ash)	50	50
F6	#8 Reconstructed slag (100%U-MSWI fly ash)	50	50

identification standards for hazardous wastes—identification for extraction toxicity (GB/T 5085.3-2007) (National Environmental Protection Agency 2007).

Leaching concentration ( $C_L$ ,  $\text{mg L}^{-1}$ ) was used to indicate the leaching toxicity of MSWI fly ash, U-MSWI fly ash, re-constructed slag, and cement composites. Fixation rate ( $R_F$ ) was used to indicate the content change of heavy metals, and leaching rate ( $R_L$ ) was used to indicate the leaching ratio of heavy metals:

$$R_F = [(c_1 \times m_1)/(c_0 \times m_0)] \times 100\% \tag{1}$$

$$R_L = [C_{m(\text{leaching})}/C_M] \times 100\% \tag{2}$$

where  $c_0$  and  $c_1$  ( $\text{mg kg}^{-1}$ ) are heavy metal content of U-MSWI fly ash-BF slag mixtures before and after melting, respectively,  $m_0$  and  $m_1$  (kg) are the mass of U-MSWI fly ash-BF slag mixtures before and after melting, respectively,  $C_{m(\text{leaching})}$  is the leaching content of heavy metals in unit mass samples,  $C_M$  is the total content of heavy metals in unit mass samples.

### Compressive strength test

Cement composites were prepared from water/binder with a mass ratio of 0.43~0.48 according to the cement standard consistency water consumption in the Chinese standard GB/T 1346-2001. The slurry was poured into a 40 mm × 40 mm × 40 mm cube model covered by the polythene sheet and left in the laboratory for 24 h. After then, the resulting material was demolded and then cured at  $20 \pm 2$  °C and  $95 \pm 2\%$  relative humidity prior for 3 and 28 days, respectively. The compressive strength of samples was measured in triplicate according to the Chinese standard GB/T 17671-1999.

## Results and discussion

### Characteristic of materials

The chemical compositions of BF slag, MSWI fly ash and U-MSWI fly ash are shown in Table 3. It can be obtained that  $\text{SiO}_2$ ,  $\text{Al}_2\text{O}_3$ , and  $\text{CaO}$  are the main components of BF slag, MSWI fly ash and U-MSWI fly ash. Wherein, MSWI fly ash

and U-MSWI fly ash also contain a certain amount of  $\text{ZnO}$ ,  $\text{CuO}$ ,  $\text{PbO}$ ,  $\text{Cr}_2\text{O}_3$ , and  $\text{Cl}$ .

Figure 2 shows the phase compositions of BF slag and MSWI fly ash. The X-ray diffraction pattern of BF slag shows an obviously broadened XRD pattern (a hump) between the  $2\theta$  of  $25^\circ$  and  $35^\circ$ , which is confirmed that the phase in the BF slag mainly exists in an amorphous form. From Fig. 2b, the phases in MSWI fly ash mainly contain quartz, anhydrite, calcite, and aluminosilicate, which reflect the resource utilization of MSWI fly ash, and there are also certain amount of  $\text{NaCl}$  and  $\text{KCl}$ . The X-ray diffraction pattern of U-MSWI fly ash is consistent with that of MSWI fly ash despite the addition of a certain proportion of heavy metals.

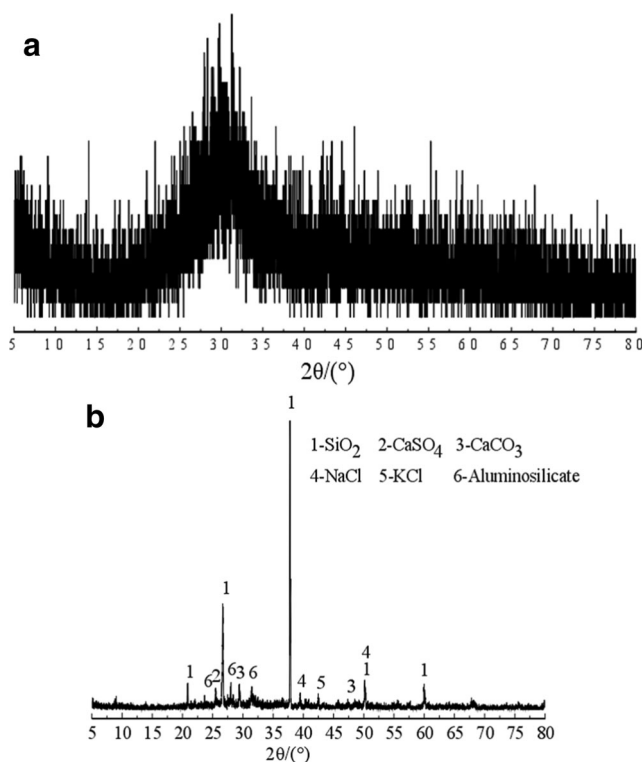
Heavy metal content and concentrations in MSWI fly ash and U-MSWI fly ash are given in Table 4; it is obtained that the leaching concentration of  $\text{Zn}$ ,  $\text{Cu}$ ,  $\text{Pb}$ , and  $\text{Cr}$  in MSWI fly ash did not exceeded the limit of Chinese standard GB/T 5085.3-2007, which may be due to the large-scale loss of heavy metal ions in MSWI fly ash caused by a long term stack. Correspondingly, the leaching concentrations of  $\text{Zn}$ ,  $\text{Cu}$ ,  $\text{Pb}$ , and  $\text{Cr}$  in U-MSWI fly ash all exceeded the limit of Chinese standard GB/T 5085.3-2007.

### The melting features of U-MSWI fly ash-BF slag mixtures

The basicity ( $\text{CaO}/\text{SiO}_2$  weight ratio), initial melting temperature ( $T$ ), and time ( $t$ ) of U-MSWI fly ash-BF slag mixtures are shown in Table 1. The initial melting temperature and time of the samples were generally in accordance with a linear relationship,  $T = 10t$ . This correlation between initial melting point and time was possibly due to a longer time was need when the mixture has a higher melting temperature. It can be obtained that the initial melting temperature (time) of mixtures increases from  $1180$  °C ( $118$  s) to  $1204$  °C ( $145$  s) by increasing the U-MSWI fly ash content from 0 to 30 wt%, and then decreases to  $1170$  °C ( $117$  s) with further increasing to 100 wt%. This may be due to the formation of trace transition minerals with higher melting point when the basicity ( $\text{CaO}/\text{SiO}_2$  weight ratio) of mixtures was close to 1.0 (Li et al. 2004). Li et al. (2004) also pointed out the optimal immobilization effect of heavy metals was achieved when the basicity of mixtures was about 1.0 because the mixtures tended to form amorphous slag.

**Table 3** Chemical compositions of BF slag, MSWI fly ash and U-MSWI fly ash (wt%)

Materials	$\text{SiO}_2$	$\text{Al}_2\text{O}_3$	$\text{Fe}_2\text{O}_3$	$\text{CaO}$	$\text{Na}_2\text{O}$	$\text{K}_2\text{O}$	$\text{MgO}$	$\text{P}_2\text{O}_5$	$\text{ZnO}$	$\text{CuO}$	$\text{PbO}$	$\text{Cr}_2\text{O}_3$	$\text{SO}_3$	$\text{Cl}$	$\text{IL}$
BF slag	32.20	17.45	0.31	34.76	0.89	0.64	10.34	--	--	--	--	--	2.00	--	1.41
MSWI fly ash	29.54	12.96	3.85	20.17	3.70	3.06	3.13	3.30	0.21	0.07	0.06	0.03	6.21	4.40	9.31
U-MSWI fly ash	22.39	10.68	3.64	20.18	2.19	2.47	2.59	2.1	5.56	5.43	0.22	0.92	5.21	6.81	9.61



**Fig. 2** XRD patterns of BF slag and MSWI fly ash. a BF slag and b MSWI fly ash

## The reconstruction features of reconstructed slag

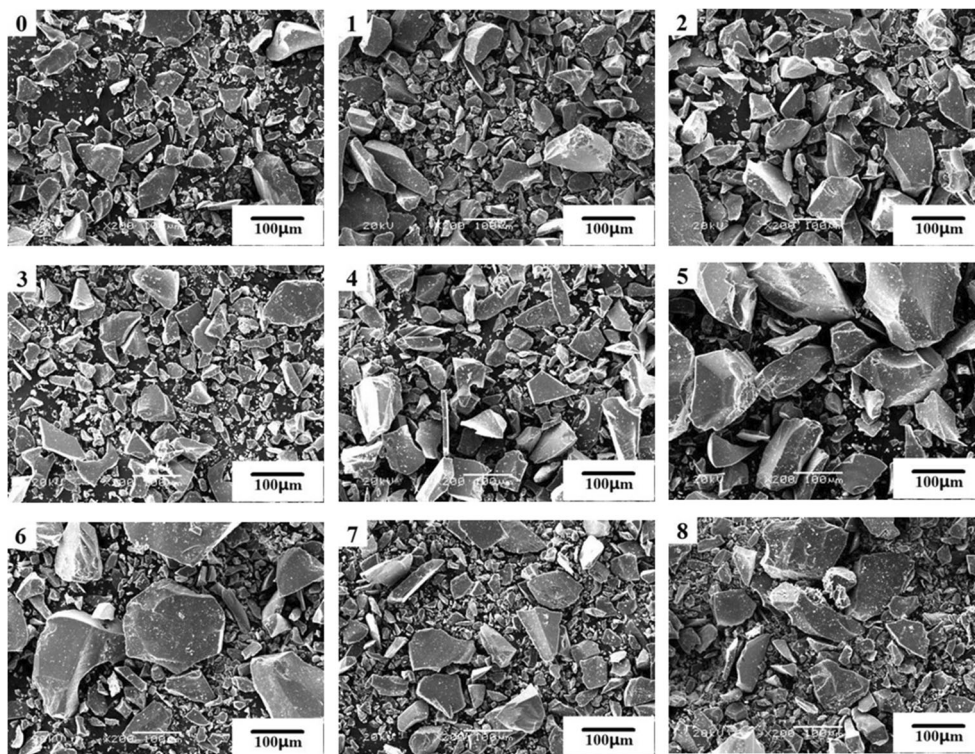
### Microstructure analysis

Microstructures of the reconstructed slag (sample 0–8 were melted at the 1400 °C for 150 s) are presented in Fig. 3. Irregular block polyhedrons and small sheet particles with a particle size of less than 100 μm can be seen in all quenched slag samples. In general, small particles with sharp features and irregular angularity present in amorphous dispersion state (Yuan 1996). Compared with the original BF slag (sample 0), the reconstructed slag exhibited loose vitreous features and the particle size became larger (especially samples 5 and 6). Further observation indicates that the reconstructed slag possessed a certain number of interconnected pores, which demonstrates that some internal and surface defects were generated during water quenching process at high temperature. These characteristics in reconstructed slag may bring out the occurrence of hydration reaction after being activated by alkali, which is similar to BF slag. However, compared with BF slag, the level of potential hydration activity of reconstructed slag is not clear, and further testing is should be required through mechanical properties. Simultaneously, it is unable to determine whether there are specific changes of the structure of reconstructed slag samples compared with that of the original BF slag on morphology alone, so further analyses are needed. The specific

**Table 4** Heavy metal content and concentrations in MSWI fly ash and U-MSWI fly ash

Heavy metals	MSWI fly ash		Heavy metal reagents		U-MSWI fly ash		The limit of concentration (mg l <sup>-1</sup> )
	Contents (mg kg <sup>-1</sup> )	TCLP Concentration (mg l <sup>-1</sup> )	Compounds	Mass (g kg <sup>-1</sup> MSWI fly ash)	Contents (mg kg <sup>-1</sup> )	TCLP concentration (mg l <sup>-1</sup> )	
Zn	1687.15	18.45	ZnO	18.86	37277.94	1686.00	100
			ZnCl <sub>2</sub>	31.59			
			Zn(NO <sub>3</sub> ) <sub>2</sub> ·6H <sub>2</sub> O	68.94			
Cu	559.22	16.95	CuO	18.97	36349.09	321.85	100
			CuCl <sub>2</sub> ·2H <sub>2</sub> O	40.64			
			Cu(NO <sub>3</sub> ) <sub>2</sub> ·3H <sub>2</sub> O	57.60			
Pb	464.16	2.53	PbO	0.82	2227.97	10.17	5
			PbCl <sub>2</sub>	1.02			
			Pb(NO <sub>3</sub> ) <sub>2</sub>	1.21			
Cr	273.68	1.14	CrO <sub>3</sub>	2.18	5610.52	29.96	15
			Cr <sub>2</sub> O <sub>3</sub>	1.66			
			CrCl <sub>3</sub> ·6H <sub>2</sub> O	11.65			
			Cr(NO <sub>3</sub> ) <sub>3</sub> ·9H <sub>2</sub> O	17.48			

Fig. 3 SEM micrographs of reconstructed slag sample 0–8



changes of the reconstructed slag samples can be determined by the following XRD and FT-IR analysis.

**Phase composition analysis**

XRD patterns of the reconstructed slag samples are shown in Fig. 4. A large and broad hump was observed in the original BF slag (sample 0) and reconstructed slag (sample 1–8), because after the U-MSWI fly ash with BF slag mixtures were melted at a high temperature and quenched into water, resulting in a change in the original crystal

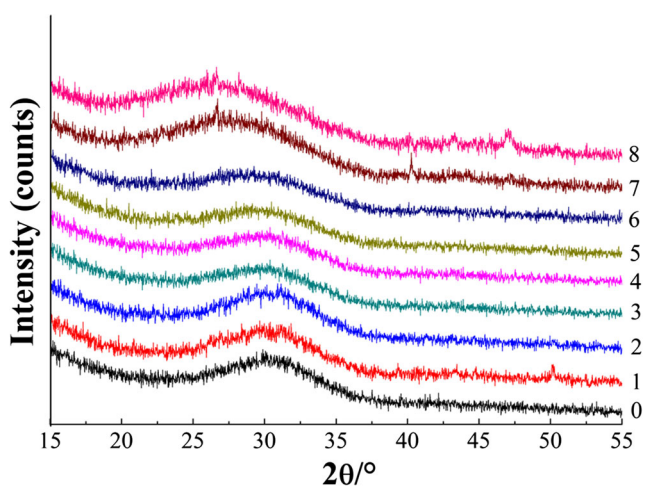


Fig. 4 XRD patterns of the reconstructed slag Sample 0-8

structure of U-MSWI fly ash to form an amorphous vitreous structure. No other obvious changes were found in these slag samples. The observation also suggests that the reconstructed slag possessed a certain hydration activity, which was similar to BF slag.

**Chemical bond analysis**

Figure 5 shows the FT-IR spectra of the original and reconstructed slag. The absorption valleys centered at 1628 (or 1630)  $\text{cm}^{-1}$ , and 3429 (or 3435)  $\text{cm}^{-1}$  were

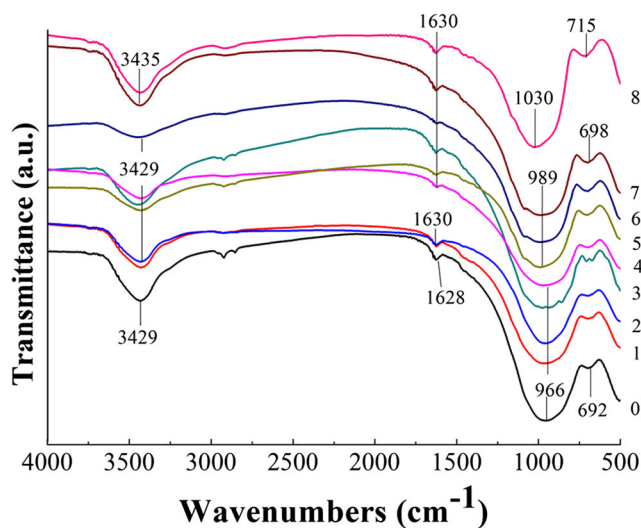


Fig. 5 FT-IR spectra of the reconstructed slag sample 0–8

respectively corresponding to bending and stretching vibration of water. Absorption valleys at  $860\text{--}1175\text{ cm}^{-1}$  and around  $700\text{ cm}^{-1}$  were ascribed to asymmetric silicon oxygen tetrahedron, aluminium oxygen tetrahedron, respectively. The strong and wide absorption peaks related to silicon oxygen tetrahedron of the slags indicates the different degree of polymerization (Wang et al. 2011). This difference in polymerization was caused by the fact that the slag comprises silicon oxygen tetrahedron monomer, dimer, and trimer. This observation suggests that the reconstructed slag could maintain a certain reactivity similar to the original BF slag, and can be used as construction materials. As shown in Fig. 5, compared with the original BF slag, the characteristic peaks at about  $700\text{ cm}^{-1}$  related to the asymmetric aluminium oxygen tetrahedron of the reconstructed slags shift to higher wave numbers. This observation indicates that the adhesion strength of the aluminum of the reconstructed slag sample was higher.

## The heavy metal leachability of reconstructed slag

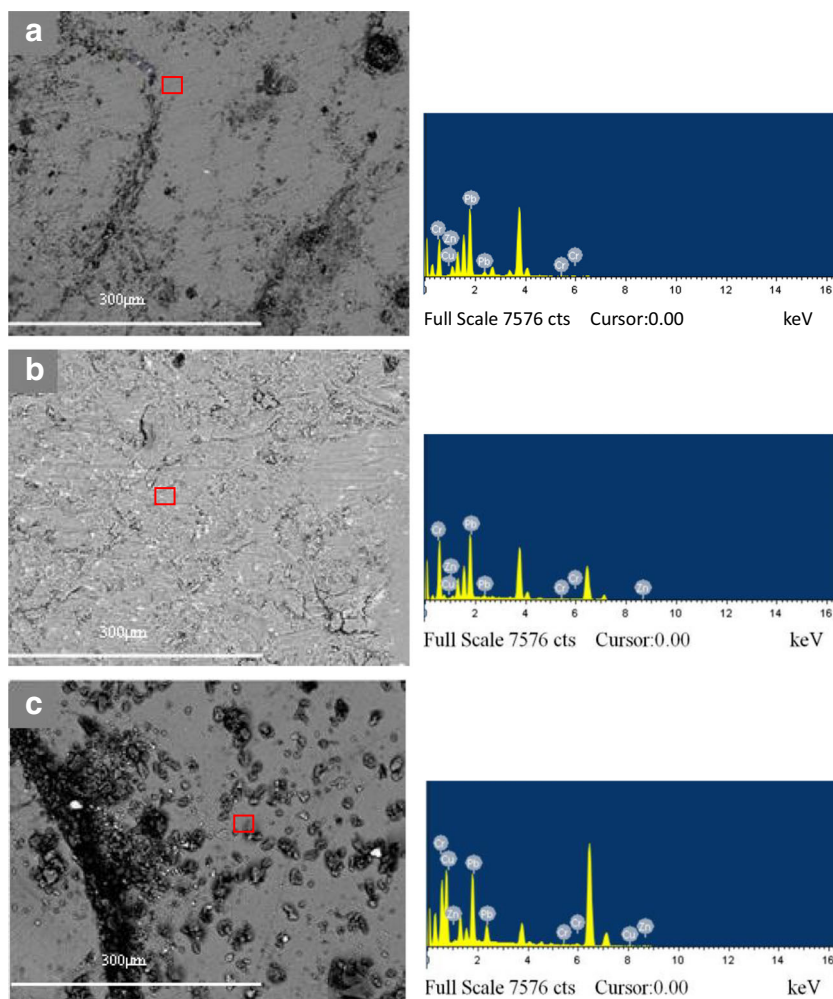
### The immobilization effect of heavy metal in reconstructed slag

Spot analyses with their related images for Zn, Cu, Pb, and Cr of reconstructed slag samples are shown in Fig. 6. The contents of Zn, Cu, Pb, and Cr in the slags were very low, probably due to the volatilization of these heavy metals during the melting process. EPMA of the samples indicates that the presence of relatively high amount of Cr than theirs Zn, Cu, and Pb contents, which also indicates that immobilization effect of Cr was probably much better than that of Zn, Cu, and Pb.

### The heavy metal fixation rate and leaching concentration of reconstructed slag

Table 5 shows the heavy metal fixation rate and leaching concentration of reconstructed slag. The fixation rates of

**Fig. 6** Electron probe microanalysis (EPMA) of the reconstructed slag samples. **a** Sample 2 (10 wt% U-MSWI fly ash + 90 wt% BF slag); **b** sample 5 (30 wt% U-MSWI fly ash + 70 wt% BF slag); **c** sample 6 (50 wt% U-MSWI fly ash + 50 wt% BF slag)





**Table 5** Heavy metal fixation rate (%) and leaching concentration ( $\text{mg L}^{-1}$ ) of reconstructed slag

NO.	Zn		Cu		Pb		Cr	
	R <sub>F</sub>	C <sub>L</sub>	R <sub>F</sub>	C <sub>L</sub>	R <sub>F</sub>	C <sub>L</sub>	R <sub>F</sub>	C <sub>L</sub>
1	9.02	0.03	13.14	0.70	3.91	--	94.08	0.08
2	13.80	0.77	15.83	6.34	3.60	--	96.40	0.12
3	16.80	0.83	24.04	15.50	3.48	--	92.87	0.23
4	25.82	1.07	23.63	24.84	3.48	--	95.43	0.36
5	46.60	1.12	26.71	68.52	4.02	--	96.69	0.41
6	50.37	1.61	24.68	98.54	4.22	--	93.12	0.46
7	34.92	2.19	18.12	184.27	4.29	0.08	94.22	5.06
8	41.99	3.33	18.48	246.59	3.77	0.11	93.89	11.82
Limit of GB/T 5085.3–2007	--	100	--	100	--	5	--	15

volatile heavy metals, i.e., Pb, Zn, and Cu were less than 51%, 27%, and 5%, respectively. The low fixation rates of Pb, Zn, and Cu were possibly due to the easy evaporation of these elements. Because of the high concentration of Cl in U-MSWI fly ash, the added heavy metal oxides or sulfates with high melting points formed low melting and boiling points chlorides (Xin et al. 2004). As a result, a large number of Zn, Cu, and Pb effectively partition into the vapor phase during melting at 1400 °C, leading to enhanced evaporation of Pb, Zn, and Cu. The influence of high Cl content on the volatilization of different heavy metals has been reported in previous work. It is found that high Cl content could effectively increase the volatilization ratio of Zn (Nowak et al. 2012). In comparison, the fixation rate of Cr is over > 93% after melting the U-MSWI fly ash-BF slag mixtures at 1400 °C. The results are in consistent with the conclusions from previous research (Chan et al. 2000; Chris and Donald 1999). Considering that Cr was a type of non-volatile heavy metal, and Cr-compounds in molten slag were  $\text{CrCl}_3$  (boiling point 1200–1500 °C) and  $\text{Cr}_2\text{O}_3$  (melting point 2266 °C), the high fixation rates of Cr suggest that Cr was effectively fixed in the reconstructed slag.

Since the volatilization of heavy metals might increase the complexity of the flue gas purification process and make it more difficult to manage, the fixation rates of heavy metals should be raised to maximize the safety utilization of the reconstructed slag and resources recovery.

Samples melted in air atmosphere when the temperature reached 1400 °C, ZnO reacted with  $\text{SiO}_2$ , and  $\text{Al}_2\text{O}_3$  formed non-volatile complexes easily, such as willemitite ( $\text{Zn}_2\text{SiO}_4$ ), and zinc aluminate gahnite ( $\text{ZnAl}_2\text{O}_4$ ), et al, and then its volatilization rate decreased. Another explanation is that the existence of aluminosilicate suppressed the precipitation of volatile substances from the molten slag, which exerted a negative influence on the volatility of heavy metals (Hong et al. 2000). The fixation rate of

Cr was very high. However, at melting temperature, Cr had low vapor pressure and very low volatilization rate. Thus, fixation rate was relatively high. Furthermore, the liquid phase in samples 1–7, which suppressed the volatilization of heavy metals from the molten slag, appeared earlier than that of sample 0 due to the fluxing action of Cu and Cr. Moreover, chemical duplex decomposition precipitation reactions occurred in microporosity of cured resin among heavy metal ions  $\text{Zn}^{2+}$ ,  $\text{Cu}^{2+}$ ,  $\text{Pb}^{2+}$ ,  $\text{Cr}^{3+}$ , and  $\text{OH}^-$  based on the alkaline condition offered by the reconstructed slag. At the same time, isomorphous replacement of heavy metal cations partial with  $\text{Ca}^{2+}$ ,  $\text{Al}^{3+}$ , and  $\text{Si}^{4+}$  in layered silicate lattice also played a certain role in solidifying heavy metals. Heavy metals solidification effect of sample 5, with a basicity of 0.99, was the best, and the result is consistent with the conclusion of Li et al. (2004).

Although the fixation rates of Zn, Cu, Pb, and Cr were varied (Table 5), the leaching concentrations of Zn, Cu, Pb, and Cr from samples 1–5 were very low. The lowered leaching concentration of Zn, Cu, Pb, and Cr might be due to the substitution of Zn, Cu, Pb, and Cr with  $\text{Ca}^{2+}$  and  $\text{Al}^{3+}$  of silicate and subsequent dissolution in the reticular basal body during the melting process. Additionally, Si-O reticular structure was formed during melting; the heavy metals in the molten slag were encapsulated and solidified in the dense grid of glassed material. As a result, the heavy metal leaching concentrations of the reconstructed slag, especially Samples 1–5, were reduced.

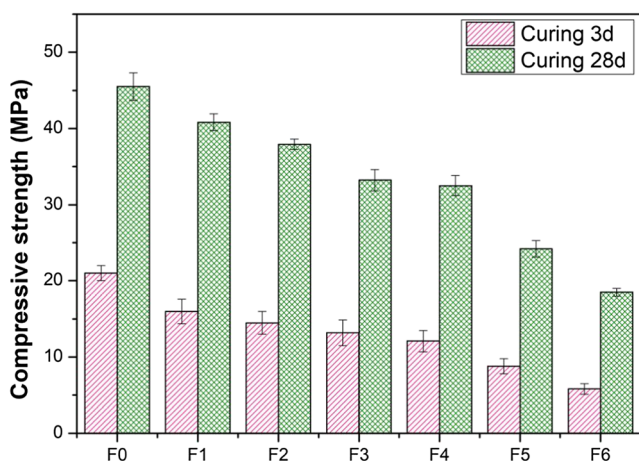
All the heavy metal leaching concentrations of reconstructed slag samples 1–5 were below the corresponding limit defined in the Chinese standard GB/T 5085.3-2007. But the leaching concentrations of Cu in the reconstructed slag samples 6–8 were above the standard limit; the requirements of environmental emission of these reconstructed slags (samples 6–8) were not met. Hence, in order to ensure reconstructed slag can be safely landfilled, the content of U-MSWI fly ash in reconstructed slag should be less than 50 wt%.

## Solidification and utilization of reconstructed slag in cement

As discussed in the above results, the reconstructed slag had certain resource utilization characteristics because its potential hydration activity was similar to BF slag. However, when the content of U-MSWI fly ash in reconstructed slag exceeded 50 wt%, the leaching concentration of heavy metals could not meet the requirement of GB/T 5085.3–2007, which limits the treatment and application of reconstructed slag. In order to comply with the safe disposal requirement and make use of its resource utilization characteristics, reconstructed slag needs to be further treated. Cement has several advantages in solidifying heavy metals and using pozzolanic materials (Bie et al. 2016; Lin and Lin 2006), so reconstructed slag can be further solidified and used in cement. Therefore, the compressive strength and leaching concentration of cement composites with reconstructed slag were investigated in this experiment, and the designed scheme is shown in Table 2.

### Compressive strength of cement composites with reconstructed slag

In order to investigate the effect of U-MSWI fly ash content on the utilization of reconstructed slag in cement composites, reconstructed slags with different content of U-MSWI fly ash were admixed into cement as partial replacement of OPC, respectively (whose scheme is shown in Table 2), and then the mechanical properties of cement composites generated were studied. Figure 7 show the compressive strength of cement composite samples after curing for 3 and 28 days. These samples exhibited a compressive strength between 5.8 and 45.5 MPa. Compared with pure OPC (sample F0), the strength properties of cement composites with reconstructed slag added became worse, mainly because in the liquid phase



**Fig. 7** Compressive strength of cement composites with reconstructed slag with different content of U-MSWI fly ash

environment of cement hydration; the potential pozzolanic properties of reconstructed slag could not be activated well.

For cement composites with reconstructed slag added, the increasing of U-MSWI fly ash content (from 0 to 100 wt%) generally reduced compressive strength: the 3-day and 28-day compressive strength decreased from 16.0 and 40.8 to 5.8 and 18.5 MPa, respectively. This possibly happened due to with the increase of U-MSWI fly ash content, the pozzolanic properties of reconstructed slag decreased, and the hydration rate of cement composites was reduced, less hydration products of  $\text{Ca}(\text{OH})_2$  were generated and less compounds in reconstructed slag could be activated by alkali to participate in the hydration reaction during the curing of cement composites, thus the compressive strength of cement composites with reconstructed slag added gradually decreased with the increasing of U-MSWI fly ash content in the reconstructed slag. In addition, the chlorides in U-MSWI fly ash can also cause volume expansion problems (Chen et al. 2019), but this phenomenon was not observed in this experiment, which is mainly related to the volatilization and solidification of chlorides in the melt reconstruction process of U-MSWI fly ash. Moreover, when the content of U-MSWI fly ash was increased to 80 wt% and 100 wt% (samples F5 and F6), the compressive strength drastically decreased to 8.8–24.2 MPa and 5.8–18.5 MPa, respectively. It is mainly also due to the retarding effect of heavy metals (He et al. 2006; Garcia-Lodeiro et al. 2016). With the large increase of U-MSWI fly ash content, the concentration of leached heavy metals in the reconstructed slag increased sharply (see Table 6), which is more detrimental to the hydration and solidification of cement composites, resulting in a worse mechanical properties. However, in general, admixing reconstructed slags into cement could make the samples obtain a certain compressive strength and more compact, and it is confirmed that reconstructed slag can be applied to landfill and construction materials by the further solidification of cement.

### Leaching concentrations of heavy metals in cement composites with reconstructed slag

Table 6 shows the heavy metal leaching concentration of cement composites with reconstructed slag curing 3 days and 28 days. With the prolonging of curing time, the leaching concentration of heavy metals in all cement composite samples with reconstructed slag decreased significantly. It indicates that during the hydration reaction of cement slurry, heavy metals could be stabilized in the hydration products or participate in the hydration reaction to form new hydration products (Mangialardi et al. 1999; Bie et al. 2016). Therefore, the leaching concentration of heavy metals tended to decrease significantly. However, with the increase of U-MSWI fly ash content in reconstructed slag, the leaching concentration of heavy metals in cement composites with reconstructed slag increased, especially when the content of U-MSWI fly ash

**Table 6** Heavy metal leaching concentration of cement composites with reconstructed slag (mg L<sup>-1</sup>)

NO.	Zn		Cu		Pb		Cr	
	Curing 3 days	Curing 28 days	Curing 3 days	Curing 28 days	Curing 3 days	Curing 28 days	Curing 3 days	Curing 28 days
F0	--	--	--	--	--	--	--	--
F1	--	--	--	--	--	--	--	--
F2	0.300	0.197	0.003	0.002	--	--	0.002	--
F3	0.370	0.264	0.006	0.004	--	--	0.036	0.026
F4	0.356	0.268	0.011	0.008	--	--	0.086	0.060
F5	0.872	0.624	0.652	0.310	0.032	0.009	0.394	0.210
F6	1.223	1.010	2.370	1.380	0.091	0.037	1.027	0.731
Limit of GB/T 5085.3-2007	100		100		5		15	

was increased to 80 wt% and 100 wt% (samples F5 and F6), the leaching concentration of Cu and Cr increased sharply. Which is due to the high leaching concentration of Cu and Cr in reconstructed slag with 80 wt% and 100 wt% U-MSWI fly ash added (samples 7 and 8) and the limited ability of cement to stabilize heavy metals. However, compared with the single melting reconstruction treatment process (see Table 5), the leaching concentrations of heavy metals after melting reconstruction treatment with cement solidification (Samples F2–F6) were all far below the corresponding limit defined in the Chinese standard GB/T 5085.3-2007 and meet the environmental requirement.

**Analysis of U-MSWI fly ash treatment cost**

In the experiment of this paper, the treatment process of U-MSWI fly ash mainly includes: raw materials preparation, high temperature melting treatment, mechanical grinding treatment, and cement solidification. The cost of U-MSWI fly ash in laboratory and industrial treatment processes are compared in Table 7. The treatment cost of U-MSWI fly ash was calculated by the formula:

The treatment cost of U-MSWI fly ash per unit (\$/t) = unit cost of raw materials (\$/t) × usage amount (t) + unit cost of energy consumption (\$/kWh) × electricity consumption (kWh).

Among the raw materials used in this experiment, the price of BF slag is \$ 43 per ton, and P.O42.5 Portland cement is \$ 65 per ton. Energy consumption price is \$ 0.17 per kWh.

In the laboratory treatment process of U-MSWI fly ash, the purchase of BF slag and the high temperature melting treatment process were charged to the cost. Wherein, when U-MSWI fly ash was treated by the single melting reconstruction treatment without cement solidification, according to the content of BF slag added (≥ 50 wt%) in reconstructed slag limited by the leaching concentration of heavy metals, melting temperature, and time, the lower treatment cost of U-MSWI fly ash per unit is \$ 623–815. When U-MSWI fly ash was treated by the melting reconstruction treatment with cement solidification, the energy consumption of the mechanical grinding of reconstructed slag and the cost of cement needed to be calculated, but the usage amount of BF slag could be greatly reduced; therefore, the lower treatment cost of U-MSWI fly ash per unit is \$ 485–526.

In the industrial treatment process of U-MSWI fly ash, the high temperature molten BF slag that has just been fired in the iron-making process could be directly used as the free melting reconstruction carrier of U-MSWI fly ash. This can not only make full use of the excess energy in industrial output, but also greatly reduce the cost of U-MSWI fly ash melt reconstruction. During the single melting reconstruction treatment

**Table 7** Analysis of U-MSWI fly ash treatment cost and comparison of reconstructed slag with other cement admixtures

Item	Laboratory treatment		Industrial treatment		Price of reconstructed slag (\$/t)	Price of slag (\$/t)	Price of fly ash (\$/t)
	##1	## 2	## 1	## 2			
Treatment methods							
Treatment cost (\$/t)	623-815	485-526	16-28	53-65	16-28	42-54	23-30
Comparison of products safety (leaching concentration)	Partly below the limit of GB/T 5085.3-2007	All below the limit of GB/T 5085.3-2007	Partly below the limit of GB/T 5085.3-2007	All below the limit of GB/T 5085.3-2007			

without cement solidification, the energy consumption of U-MSWI fly ash dumped into the molten BF slag was only needed to be calculated, and the lower treatment cost of U-MSWI fly ash per unit is \$ 16–28. When U-MSWI fly ash was treated by the melting reconstruction treatment with cement solidification process, the energy consumption of the mechanical grinding of reconstructed slag and the cost of cement are needed to be calculated, and the lower treatment cost of U-MSWI fly ash per unit is \$ 53–65.

The laboratory treatment process of U-MSWI fly ash is mainly used for theoretical research, while the actual safety treatment of a large amount of U-MSWI fly ash needs to rely on industrial treatment process, and the industrial treatment process according to this experimental method can provide significant economic benefits for the treatment and recycling of U-MSWI fly ash. In addition, comparing the two methods (melting reconstruction treatment with and without cement solidification) based on the industrial treatment process, the cost of melting reconstruction treatment without cement solidification method was lower, but this method has limited ability to solidify heavy metals in U-MSWI fly ash. However, the addition of cement can not only improve the ability to solidify heavy metals in U-MSWI fly ash, but also give full play to the resource characteristics of reconstructed slag as cement admixture. Therefore, comprehensively considering the above factors, the melting reconstruction treatment with cement solidification method provides a technical-economical possibility for industrial treatment of U-MSWI fly ash, and can improve the safety of products as a landfill or construction material. Moreover, as shown in Table 7, compared with other cement admixtures such as slag and fly ash, the price of reconstructed slag produced in this experiment is relatively economical, so it has certain production and utilization value for construction.

## 1: melting reconstruction treatment without cement solidification method

## 2: melting reconstruction treatment with cement solidification method

## Conclusions

Blast furnace (BF) slag in high temperature molten state from iron-making process can be used as a free high-temperature carrier to melt U-MSWI fly ash. The melting reconstruction features and heavy metals solidification behaviors of U-MSWI fly ash-BF slag mixtures were investigated according to phase composition, microstructure and heavy metals leachability. Moreover, the further solidification and utilization of reconstructed slag in cement were also studied. The conclusions can be drawn as following:

1. The main chemical compositions of U-MSWI fly ash are SiO<sub>2</sub> (22.39%), Al<sub>2</sub>O<sub>3</sub> (10.68%), CaO (20.18%), and Cl (6.81%), and the major phases are SiO<sub>2</sub>, CaSO<sub>4</sub>, CaCO<sub>3</sub>, NaCl, KCl, and aluminosilicate. The leaching concentration of Zn, Cu, Pb, and Cr is 1686 mg/L, 321.85 mg/L, 10.17 mg/L, and 29.96 mg/L, respectively.
2. The reconstructed slag formed by melting U-MSWI fly ash and BF slag, has a certain potential hydration activity similar to BF slag. However, when the content of U-MSWI fly ash exceeds 50 wt%, the leaching concentration of heavy metals in reconstructed slag cannot meet the environment requirement. Therefore, for the further solidification and utilization of reconstructed slag, the cement solidification process is strongly suggested.
3. Adding cement to reconstructed slag can generate hydration products, which can further solidify heavy metals and ensure the safe treatment and application of reconstructed slag. The method of melting reconstruction treatment with cement solidification can provide a technical-economical possibility for industrial treatment of U-MSWI fly ash, and can improve the safety of products as a landfill or construction material.

Further work for studying the durability of the products generated by melting reconstruction treatment with cement solidification treatment of U-MSWI fly ash should be done, and the use of reconstructed slag to prepare geopolymers, and other further solidification and utilization methods should be also investigated.

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