SUSTAINABLE DEVELOPMENT OF ENERGY, WATER AND ENVIRONMENT SYSTEMS



Effects of rainfall intensities on sediment loss and phosphorus enrichment ratio from typical land use type in Taihu Basin, China

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Abstract

Soil erosion induced by rainfall has become one of the major environmental disasters. It can lead to degradation of land productivity and water pollution, which limit human survival and development. Therefore, understanding the effects of rainfall intensities on soil nutrient and sediment transportation from farmland via surface runoff is important. This study simulated three types of rainfall intensities (low intensity 0.83 mm min⁻¹; medium intensity 1.17 mm min⁻¹ and 1.67 mm min⁻¹; and high intensity 2.50 mm min⁻¹). Field plots were established to evaluate the effects of rainfall intensities on sediment yield and phosphorus enrichment ratio (ER_P) in runoff sediment. Each plot was 2 m long and 1 m wide. The rainfall lasted for 30 min after effective runoff generation. Runoff and sediment, as well as the phosphorus (P) concentration of sediment, were measured. Results showed that the low rainfall intensity had a longer time to begin runoff than high rainfall intensity. The runoff volume increased with the rainfall intensity; however, no obvious differences were observed in the runoff rate under three rainfall intensities, except for 2.50 mm min⁻¹. The curve value of sediment concentration increased after 5-10 min of effective runoff decreased slightly, and then relatively became stable after 20 min of effective runoff. Although no obvious differences were observed in the dynamic changes of sediment concentration and rainfall intensity, a power function existed between rainfall intensity and sediment loss. The similar curves of ER_P value showed that it was higher in the initial stage, decreased gradually, and reached a stable level after 25 min of effective runoff, which varied around 1. During the rainfall-runoff event, no significant correlation was noted between ER_P and sediment yield. Whether any relationship existed between them needs to be further studied by setting up different field plots under large-scale conditions. The findings of this study provide valuable aid to farmers on how to manage farmland effectively to prevent soil nutrient loss and reduce water eutrophication.

Keywords Soil erosion · Rainfall intensity · Sediment · Runoff · Nutrient · Enrichment ratio · Phosphorus · Taihu Basin

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Introduction

Soil erosion is one of the main environmental issues in ecosystems (Eswaran et al. 2001; Zheng et al. 2005; Terranova et al. 2009; Zhang et al. 2011; Abrol et al. 2013). It can cause some serious consequences, including soil degradation, cropland productivity reduction, and water environment eutrophication (Kumar and Saha 2011; Yang et al. 2014; Liu et al. 2015a, b). The United Nations Environmental Program (1991) reported that soil degradation rate is estimated to be occurring on approximately 20×10^6 ha per year. The first census of water and soil conservation in the National Water Survey indicated that the soil erosion area in China is 295 million ha, which is up to 31.1% of the total census area (MEPPRC 2016). Soil degradation caused by water is the main type in China. Topsoil is washed away by rainfall–runoff, snowmelt, and irrigation. The area by water erosion is up to 56.4% of the total land area (Zou et al. 2001). Rainfall is the most important driving force of water erosion and has become a research focus in fields related to soil erosion (Oztas et al. 2003; Jin et al. 2004; Liu et al. 2015a, b; Napoli et al. 2017). Erosion caused by rainfall includes a two-phase process, that is, soil particle detachment and sediment movement by flowing water (Defersha et al. 2011), which are controlled by the kinetic energy of raindrop splash. Therefore, rainfall can change the condition of soil aggregates and lead to soil nutrient loss (Berger et al. 2010; Zhu 2012).

Among rainfall characteristics, rainfall intensity affects surface runoff generation and nutrient concentrations in runoff (Ran et al. 2012; Zhao et al. 2015; Zhou et al. 2016). Therefore, the effects of rainfall intensities on soil erosion have been discussed by many researchers (Bradford et al. 1987; Römkens et al. 2001; Parsons and Stone 2006). Mermut et al. (1997) observed that the load of rainfall splash under a rainfall intensity of 100 mm h⁻¹ was 2.5–4 times that of 40 mm h^{-1} . Margues et al. (2008) found that sediment yield in highly intensive events was obviously greater than that in medium intensive events. Ran et al. (2012) indicated that high rainfall intensity produces a high runoff peak. Raindrop splash of high rainfall intensity increases soil particle detachment and can transfer increased suspended sediment load because a high rainfall intensity results in a high runoff volume (Rose 1993). Mohamadi and Kavian (2015) found that a linear relationship exists between soil loss and rainfall intensity in low rainfall intensities. Many studies have explored the effects of rainfall intensities on sediments and nutrient discharging into water environment in farmlands, paddy fields, and pastures. However, the dynamic process of sediment and phosphorus (P) loss transported by runoff from vegetable lands in relation to rainfall intensity is not well known. Vegetable lands with high fertilizer application are distributed along tributaries, canals, and ditches of Taihu. Soil nutrients from vegetable lands easily enter the surrounding water bodies through surface runoff; this phenomenon is a considerable threat to water environment. As the third largest freshwater lake in China, Lake Taihu is a primary drinking water source. However, in recent years, frequent cyanobacteria outbreaks caused by agricultural non-point sources have resulted in serious water eutrophication in Lake Taihu. Therefore, this study aims to estimate the dynamic characteristics of sediment concentration and investigate the effects of rainfall intensities on sediment P concentration and enrichment ratio (ER) in vegetable lands. This study can provide valuable guide for farmers in farmland cultivation and management. The findings can also provide experimental parameters of non-point source pollution models for estimating sediment nutrient losses.

Materials and methods

Study area

The experiment was set up in a typical land use type of Hongsheng Town, Wuxi City, Jiangsu Province, in the upper river basin of Lake Taihu (Fig. 1). Given preponderant geographical location, many vegetables have been planted in the town to supply for Wuxi City. Vegetable land is considered a typical land use type in the Taihu Basin, which has high P content in soils because a large amount of fertilizers are amended into the soil with extensive plantation time. The study area has more than 10 years for experiment plots. The climate has an average annual temperature of 15.5 °C and an average annual rainfall of 1000 mm, with most of the rain falling in May to September. The agricultural land has a high multiple-crop index due to good natural conditions, which results in high fertilizer input. In view of the fertilization practice used by local farmers, rates of inorganic nitrogen fertilizer (carbamide), P fertilizer (superphosphate phosphate), and potassium fertilizer (potassium chloride) are 200, 75, and 100 kg ha⁻¹ by surface broadcast, respectively. The surface soil (0-15-cm depth) in vegetable field was collected. On the basis of the United States Department of Agriculture classification system, the soil texture in the plots is a silt loam type. The bulk gravity is 1.23 g cm⁻³. The soil has 29.75 g kg⁻¹ 1.23 g kg^{-1} , and $174.47 \text{ mg kg}^{-1}$ of organic matter, total P, and Olsen P. The pH and water content of the soil are 5.85 and 13.30%, respectively.

Simulated rainfall experiments

Rainfall was generated by a portable artificial rainfall simulator of BX-1 type (Institute of Soil and Water Conservation, Chinese Academy of Science). The equipment consisted of a nozzle, a driver, and dynamic water supply systems. By adjusting the nozzle sizes and changing water pressure, this equipment could splash different raindrops in the range of $10-332 \text{ mm h}^{-1}$. The height for this simulation equipment was 6 m. This instrument had a uniformity coefficient of >80% within 3×5 m². On the basis of the local historical rainfall records and average recurrence periods, the simulated rainfall intensities were set to low intensity (0.83 mm min^{-1}), medium intensity $(1.17 \text{ mm min}^{-1} \text{ and } 1.67 \text{ mm min}^{-1})$, and high intensity $(2.50 \text{ mm min}^{-1})$. Experimental plots were rectangular with a size of 1×2 m² with a random block design and uniform vegetable coverage density. The runoff frame borders for the sides of the test plots were constructed from PVC boards, whereas the bottoms were constructed from steel equipment. The top and side of each plot were surrounded by PVC borders. These PVC borders were inserted into the soil depth at approximately 15 cm, which



Fig. 1 Experimental site at Hongsheng Town, Wuxi City, Jiangsu Province

was 10 cm higher than the soil surface. Each experiment was conducted in three replicates, and a total of 10 simulated rainfalls were conducted. For different rainfall intensities, the rainfall lasted 30 min after effective runoff generation for each treatment. All the data used in the figure are the average values of the experimental data.

When the simulated rainfall began, the time for simulating rainfall and the initiation time for producing runoff were recorded. All runoff and sediment samples were collected in a marked plastic container on a 5-min interval for 30 min. Runoff volumes were measured. By stirring the samples, 500 ml of each sample was taken to measure the sediment concentration and sediment P loss. These samples were stored in a refrigerator at 4 °C for analyzing different items. During the simulated rainfall, rainwater was collected as a blank contrast.

The following properties of soil samples were measured: pH (water-to-soil ratio of 2.5:1, glass electrode), total P (HClO₄–H₂SO₄ digestion, molybdenum blue colorimetry), Olsen P (NaHCO₃, molybdenum blue colorimetry), organic matter (K_2 CrO₇–H₂SO₄ oxidation), bulk specific gravity (loop reamer), and soil containing water content (drying method). The methods are detailed in the "Analytical Handbook of Soil Agricultural Chemistry" (Lao 1988). Sediment concentration was determined by dry weight method. The total P in sediment was measured by HClO₄–H2SO₄ digestion and molybdenum blue colorimetry. The total P contents in parent soils and sediments were measured. The P enrichment ratio (ER_P) in sediments is calculated as follows (Li et al. 2016):

$$ER_p = P_{sediment} / P_{plot \cdot soil,}$$
(1)

where P_{sediment} is the P content in sediments, and $P_{\text{plot soil}}$ is the P content in plot soils.

Results

Effects of rainfall intensities on runoff rate and runoff yield

With the increase in rainfall intensity, the time to begin runoff had the following values: 31.20, 15.80, 10.40, and 4.27 min. The low rainfall intensity had a longer time to begin runoff than high rainfall intensity. The dynamic process of runoff rate showed no significant differences between the other three rainfall intensities, except for high intensity (2.50 mm min⁻¹; Fig. 2). In high intensity (2.50 mm min⁻¹), the runoff rate increased rapidly, the soil infiltration capacity decreased, the time when runoff began was relatively shortened, and the runoff increased considerably. During a 30-min rainfall after the runoff began, the high rainfall intensity (2.50 mm min⁻¹) produced 18.91 L m⁻² runoff yield, which was 3, 5, and 7

Fig. 2 Runoff rate with rainfall times since runoff began for different rainfall intensities



times that of the other three rainfall intensities. The effects of rainfall intensities on runoff yield were evident.

Effects of rainfall intensities on sediment loss

Fig. 3 Characteristics of sediment

concentration in runoff for different rainfall intensities

As shown in Fig. 3, no obvious differences were found in the dynamic change of sediment concentration. Except for 2.50 mm min⁻¹, sediment concentration was higher after 5– 10 min of effective runoff, and sediment concentration decreased slightly with the rainfall duration. The relatively stable stage of sediment concentration occurred after 20 min of effective runoff. During the rainfall duration, the average sediment concentration for 2.50 mm min⁻¹ was 2.85 g L⁻¹, which was highest in all rainfall intensities. Average sediment concentration for the other three rainfall intensities varied from

1.30 to 1.79 g L^{-1} . Therefore, sediment losses were more sensitive to the runoff caused by high intensity than for runoff caused by low and medium rainfall intensities.

When the rainfall intensity increased from 0.83 to 1.67 mm min⁻¹, the sediment loss increased by 2 times; however, sediment loss increased rapidly by 7.51 times when rainfall intensity increased from 1.67 to 2.50 mm min⁻¹. Hence, a critical point for sediment loss was observed between 1.67 and 2.50 mm min^{-1} . Therefore, rainfall intensity had an obvious effect on sediment loss.

Effects of rainfall intensities on ER_P in sediment

The curves of ER_P values in sediment under different rainfall intensities are shown in Fig. 4. The ER_P values from high to



Fig. 4 ER_P in sediment for different rainfall intensities



low rainfall intensity varied in the ranges of 1.20–1.70, 0.95– 1.38, 1.00–1.30, and 1.03–1.21; all of these values varied around 1. The ER_P curves had similar shapes. ER_P was higher in the initial stage, decreased gradually, and reached a stable level after 25 min of effective runoff because the fine clay content, which was detached and transported by rainfall–runoff, absorbed a high P nutrient in the initial stage. During the rainfall–runoff, the total P concentration in sediment for the four types of rainfall intensity decreased by 26%, 20%, 18%, and 12%. This result was mainly caused by the coarse soil particles, which were detached and transported by rainfall– runoff; these particles absorbed low P nutrient, and the runoff volume increased gradually, which had a dilution effect on the sediment P erosion. Although the total changing trend was similar for all rainfall intensities, the ER_P for 2.50 mm min⁻¹ was the highest in the 30 min of effective runoff, which showed a critical value between 1.67 and 2.50 mm min⁻¹. This critical value must be further studied. The average ER_P for different rainfall intensities increased with the rainfall intensity. In the entire rainfall event, no significant connection was found between ER_P and sediment yield (Fig. 5).

Discussion

The present study clearly demonstrates that rainfall intensity is an important factor that influences runoff, which affects not only the speed of runoff formation but also the runoff volume. Many researchers have analyzed the effect of rainfall on runoff. Some researchers have demonstrated that infiltration rate



Fig. 5 Changes in $\ln (ER_P)$ with $\ln (SED)$ under different rainfall intensities

increases with rainfall intensity, and runoff volume might decrease due to the spatial heterogeneity of soil infiltration (Römkens et al. 1990; Parsons and Stone 2006; Mu et al. 2015). Other researchers have indicated that runoff volume increases with rainfall intensity (Zhang and Liang 1995), which is similar to findings of the present study. At the beginning of rainfall, soil infiltration capacity is greater than that of rainfall volume. With the increase of rainfall, the soil surface water content increases gradually and the infiltration capability of soil gradually decreases due to soil crust formation. When the infiltration rate is equal to or less than the rainfall intensity during soil surface sealing, the soil surface begins to accumulate water and produces surface runoff (Römkens et al. 2001; Dijk et al. 2002). The surface runoff increases gradually with the rainfall duration and the decrease in infiltration rate. The results of this study show that rainfall intensity has an obvious effect on initial runoff time and runoff yield.

Sediment loss induced by rainfall-runoff events includes two physical processes, namely detachment and transportation of soil particles (Defersha et al. 2011; Sirjani and Mahmoodabadi 2012). Raindrop splash detaches soil particles from the soil surface and produces detached materials that can be transported by runoff (Sadeghi et al. 2017). Transportation rate resulting from transferring the splashed particles away from the soil surface depends on the force of water flowing by rainfall (Ali et al. 2012). The on-site consequences of detaching and transporting soil particles are mainly soil degradation, and the off-site consequences can result in sediment and nutrients being transferred by runoff away into natural water bodies (Blanco and Lal 2010; Noor et al. 2013). Therefore, nutrient attached in sediment is easily transported by runoff (Schoellhamer et al. 2007; Fazli and Noor 2013). Different rainfall intensities have varied effects on sediment loss due to the various dynamic energy of raindrop splash and detaching parent soil particles. The volume of loose soil particles detached by runoff greatly increases with rainfall intensity. Therefore, detached particles splashed by raindrop are rich in high rainfall intensity. The raindrop splash energy of high rainfall is greater than that of low rainfall intensity, which can lead to a relatively greater sediment transport capacity. In this study, a power function relationship is observed between rainfall intensity and sediment loss. Bouraima and He (2016) observed that the relationship between sediment and rainfall intensity is an exponential function. Wang et al. (2013) observed a positive correlation between rainfall intensity and runoff loss of soil organic carbon via an experiment. The growth of water algae is enhanced when P concentration exceeds 0.02 mg L^{-1} (Moore et al. 1992). However, in the entire rainfall-runoff process in this experiment, sediment-associated P concentration is higher than the limited concentration. At concentrations exceeding the threshold levels, nutrient transportation in surface runoff associated with sediment from vegetable fields can enhance the growth of aquatic plants and become a potential threat to the surrounding water environment.

The process of soil particle being transported by rainfall is essentially a selective process of particle size. Abundant nutrient content is stored in fine particles. These fine-textured particles are easily moved due to their features, which result in nutrient enrichment in sediments in comparison with that of the parent soil (Sharpley 1985; Quinton et al. 2001). ER has been defined as the ratio of nutrient content between the eroded sediment and the parent soil (Massey and Jackson 1952). On the basis of several experiments and models, the general conclusion is that ER_{nutrient} is greater than 1 (Panuska and Karthikeyan 2010). Young et al. (1986) found that the value of ER approaches 1 for different rainfall intensities, which correspond with the result of the experiment in the present study. Some researchers have shown specific linear equation relationships between ER and sediment loss (Menzel 1980; Young et al. 1989). However, no significant correlation between ER_P and sediment yield is observed in the present work. The effect of ER_P on sediment loss is unclear and should be further explored by setting up different experiments at a larger scale.

Conclusions

Low rainfall intensity had a longer time to begin runoff compared with high rainfall intensity. The runoff volume increased with rainfall intensity. No obvious differences in runoff rate were obtained among the other three rainfall intensities, except for high rainfall intensity (2.50 mm min⁻¹). No distinct difference was observed in the dynamic change of sediment concentration and rainfall intensity. The ER_P curves had similar shapes, and all ER_P values varied around 1. ER_P was higher in the initial stage, decreased gradually, and reached a stable level after 25 min of effective runoff. The average ER_P for different rainfall intensities increased with rainfall intensity. In the entire rainfall–runoff event, no distinct correlation was found between ER_P and sediment yield.

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References

Abrol V, Shainberg I, Lado M, Ben-Hur M (2013) Efficacy of dry granular anionic polyacrylamide (PAM) on infiltration, runoff and erosion. Eur J Soil Sci 64(5):699–705

- Ali M, Sterk G, Seeger M, Boersema M, Peters P (2012) Effect of hydraulic parameters on sediment transport capacity in overland flow over erodible beds. Hydrol Earth Syst Sci 16:591–601
- Berger C, Schulze M, Rieke-Zapp D, Schlunegger F (2010) Rill development and soil erosion: a laboratory study of slope and rainfall intensity. Earth Surf Proc Landf 35:1456–1467
- Blanco H, Lal R (2010) Principles of soil conservation and management. Springer science, Heidelberg, Springer Netherlands 9(1):199–200
- Bouraima AK, He B (2016) T Tian runoff, nitrogen (N) and phosphorus (P) losses from purple slope cropland soil under rating fertilization in three gorges region. Environ Sci Pollut Res 23(5):4541–4550
- Bradford JM, Ferris JE, Remley PA (1987) Interrill soil erosion processes: I Effect of surface sealing on infiltration, runoff, and soil splash detachment. Soil Sci Soc Am J 51(6):1566–1571
- Defersha MB, Quraishi S, Melesse A (2011) The effect of slope steepness and antecedent moisture content on interrill erosion, runoff and sediment size distribution in the highlands of Ethiopia. Hydrol Earth Syst Sci 15:2367–2375
- Dijk JM, Bruijnzeel LA, Rosewell CJ (2002) Rainfall intensity kinetic energy relationship: a critical literature appraisal. J Hydrol 261(1–4):1–23
- Eswaran H, Lal R, Reich PF (2001) Land degradation: an overview, responses to land degradation, in: Proceeding of 2nd international conference on land degradation and desertification. Oxford Press, New Delhi, India, pp,59–71
- Fazli S, Noor R (2013) Storm-wise sediment yield prediction using hillslope erosion model in semi-arid abundant lands. Soil Water Res 8: 42–48
- Jin K, Schiettecatte W, Verbist K, Gabriels D, Hartmann R (2004) Sediment load in runoff under laboratory and field simulated rainfall. J Integr Agric 3(1):31–36
- Kumar A, Saha A (2011) Effect of polyacrylamide and gypsum on surface runoff, sediment yield and nutrient losses from steep slopes. Agric Water Manag 98(6):999–1004
- Lao JZ (1988) Analytical handbook of soil agricultural chemistry. Agriculture Press, Beijing 240–213,229–275
- Li ZW, Nie XD, Chang XF, Liu L, Sun LY (2016) Characteristics of soil and organic carbon loss induced by water erosion on the loess plateau in China. PLOS One 11(4):1–13
- Liu DD, She DL, Yu SE, Shao GC, Chen D (2015a) Rainfall intensity and slope gradient effects on sediment losses and splash from a salinesodic soil under coastal reclamation. Catena 128:54–62
- Liu J, Zuo Q, Zhai LM, Luo CY, Liu HB, Wang HY, Liu S, Zou GY, Ren TZ (2015b) Phosphorus losses via surface runoff in rice-wheat cropping systems as impacted by rainfall regimes and fertilizer applications. J Integr Agric 15(3):667–677
- Marques MJ, Bienes R, Pérez-Rodríguez R, Jiménez L (2008) Soil degradation in Central Spain due to sheet water erosion by low- intensity rainfall events. Earth Surf Proc Land 33(3):414-423
- Massey HF, Jackson M (1952) Selective erosion of soil fertility constituents. Soil Sci Soc Am J 16:353–356
- Menzel RG (1980) Enrichment ratios for water quality modeling. In: Knisel, W.G. (Ed.), Creams. USDA Cons. Res. Rep, Washington DC 486–492
- MEPPRC (2016) Ministry of environmental protection of the People's Republic of China. China Environmental Status Bulletin. Environmental Protection Press, Beijing
- Mermut AR, Luk SH, Romkens MJM, Poesen JWA (1997) Soil loss by splash and wash during rainfall from two loess soils. Geoderma 75: 203–214
- Mohamadi MA, Kavian A (2015) Effects of rainfall patterns on runoff and soil erosion in field plots. Int Soil Water Conserv Res 3:273–281
- Moore PA, Reddy KR, Graetz DA (1992) Nutrient transformations in sediments as influenced by oxygen supply. J Environ Qual 21(3): 387–393
- Mu WB, Yu FL, Li C, Xie YB, Tian JY, Liu J, Zhao NN (2015) Effects of rainfall intensity and slope gradient on runoff and soil moisture

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content on different growing stages of spring maize. Water 7: 2990-3008

- Napoli M, Marta AD, Zanchi CA, Orlandini S (2017) Assessment of soil and nutrient losses by runoff under different soil management practices in an Italian hilly vineyard. Soil Tillage Res 168: 71–80
- Noor H, Fazli S, Alibakhshi M (2013) Evaluation of the relationships between runoff-rainfall-sediment related nutrient loss (a case study: Kojour watershed, Iran). Soil & Water Res 8(4): 172–177
- Oztas T, Koc A, Comakli B (2003) Change in vegetation and soil properties along a slope on overgrazed and eroded rangelands. J Arid Environ 55:93–100
- Panuska JC, Karthikeyan KG (2010) Phosphorus and organic matter enrichment in snowmelt and rainfall-runoff from three corn management systems. Geoderma 154:253–260
- Parsons AJ, Stone PM (2006) Effects of intra-storm variations in rainfall intensity on interrill runoff and erosion. Catena 67(1):68–78
- Quinton JN, Catt JA, Hess TM (2001) The selective removal of phosphorus from soil: is event size important? J Environ Qual 30:538–545
- Ran Q, Su D, Li P, He Z (2012) Experimental study of the impact of rainfall characteristics on runoff generation and soil erosion. J Hydrol 424:99–111
- Römkens MJM, Prasad SN, Parlange JY (1990) Surface seal development in relation to rainstorm intensity. Catena Suppl 17:1–11
- Römkens MJM, Helming K, Prasad SN (2001) Soil erosion under different rainfall intensities, surface roughness, and soil water regimes. Catena 46:103–123
- Rose CW (1993) Erosion, sedimentation. In: Bonell M, Hufschmidt M, Gladwell JS (eds) Hydrology and water management in the humid tropics. Cambridge University Press, Cambridge 115–129
- Sadeghi SH, Harchegani MK, Asadi H (2017) Variability of particle size distributions of upward/downward splashed materials in different rainfall intensities and slopes. Geoderma 290:100–106
- Schoellhamer D, Mumley TE, Leatherbarrow JE (2007) Suspended sediment and sediment-associated contaminants in San Francisco Bay. Environ Res 105:119–131
- Sharpley AN (1985) The selection erosion of plant nutrients in runoff. Soil Sci Soc Am J 49:1527–1534
- Sirjani E, Mahmoodabadi M (2012) Effects of sheet flow rate and slope gradient on sediment load. Arab J Geosci 7:203–210
- Terranova O, Antronico R, Coscarelli R, Iaquinta P (2009) Soil erosion risk scenarios in the Mediterranean environment using RUSLE and GIS: an application model for Calabria (southern Italy). Geomorphology 112:228–245
- United Nation Environmental Program (1991) Status of desertification and implementation of the UN plan of action to combat desertification. UNEP, Nairobi, Kenya 38–55
- Wang WX, Zhuang YL, Zhuang JY, Lv XN, Wu HQ (2013) Effects of downhill coverage on soil organic carbon loss under different rainfall intensities. J Soil Water Conserv 27(4):62–66
- Yang Y, Ye ZH, Liu BY, Zeng XQ, Fu H, Lu BJ (2014) Nitrogen enrichment in runoff sediments as affected by soil texture in Beijing mountain area. Environ Monit Assess 186:971–978
- Young RA, Olness AE, Mutchler CK, Moldenhauer WC (1986) Chemical and physical enrichments of sediment from cropland. Transactions ASAE 29:165–169
- Young RA, Onstad CA, Bosch DD, Anderson WP (1989) AGNPS: a nonpoint-source pollution model for evaluating agricultural watersheds. J Soil Water Conserv 44:168–173
- Zhang GH, Liang YM (1995) Study on runoff beginning time of artificial grassland in Loess hilly region. J Soil Water Conserv 9(3):78-83
- Zhang GH, Liu GB, Wang GL, Wang YX (2011) Effects of vegetation cover and rainfall intensity on sediment-bound nutrient loss, size

composition and volume fractal dimension of sediment particles. Pedosphere 21(5):676–684

- Zhao QH, Li DQ, Zhuo MN (2015) Effects of rainfall intensity and slope gradient on erosion characteristics of the red soil slope. Stoch Env Res Risk A 29:609–621
- Zheng FL, He XB, Gao XT, Zhang CE, Tang KL (2005) Effects of erosion patterns on nutrient loss following deforestation on the Loess Plateau of China. Agric Ecosyst Environ 108(1): 85–97
- Zhou J, Fu BJ, Gao GY, Lu YH, Liu Y, Lu N, Wang S (2016) Effects of precipitation and restoration vegetation on soil erosion in a semi-arid environment in the loess Plateau, China. Catena 137:1–11
- Zhu MY (2012) Soil erosion risk assessment with CORINE model: case study in the Danjiangkou Reservoir region, China. Stoch Env Res Risk A 26:813–822
- Zou YR, Zhang ZX, Yang CJ, Zhang ZK (2001) Analysis of soil erosion status of China land resource. J Soil Water Conserv 15(3):44–47