

Effects of taxonomy, sediment, and water column on C:N:P stoichiometry of submerged macrophytes in Yangtze floodplain shallow lakes, China

Haojie Su^{1,2} · Yao Wu^{1,2} · Ping Xie¹ · Jun Chen¹ · Te Cao¹ · Wulai Xia^{1,2}

Received: 4 April 2016 / Accepted: 8 August 2016 / Published online: 24 August 2016
© Springer-Verlag Berlin Heidelberg 2016

Abstract Carbon (C), nitrogen (N) and phosphorus (P) are the three most important essential elements limiting growth of primary producers. Submerged macrophytes generally absorb nutrients from sediments by root uptake. However, the C:N:P stoichiometric signatures of plant tissue are affected by many additional factors such as taxonomy, nutrient availability, and light availability. We first revealed the relative importance of taxonomy, sediment, and water column on plant C:N:P stoichiometry using variance partitioning based on partial redundancy analyses. Results showed that taxonomy was the most important factor in determining C:N:P stoichiometry, then the water column and finally the sediment. In this study, a significant positive relationship was found between community C concentration and macrophyte community biomass, indicating that the local low C availability in macrophytes probably was the main reason why submerged macrophytes declined in Yangtze floodplain shallow lakes. Based on our study, it is suggested that submerged macrophytes in Yangtze floodplain shallow lakes are primarily limited by low light levels rather than nutrient availability.

Keywords Submerged macrophyte · Sediment · Water column · C:N:P stoichiometry · Eutrophication

Responsible editor: Thomas Hein

✉ Ping Xie
xieping@ihb.ac.cn

¹ Donghu Experimental Station of Lake Ecosystems, State Key Laboratory of Freshwater Ecology and Biotechnology, Institute of Hydrobiology, The Chinese Academy of Sciences, 7# Donghu South Road, Wuhan 430072, China

² University of Chinese Academy of Sciences, Beijing 100049, China

Introduction

Carbon (C):nitrogen (N):phosphorous (P) stoichiometry provides a very powerful way to enhance our understanding of primary production, nutrient cycling, and population dynamics in freshwater and terrestrial ecosystems (Andersen et al. 2004; Elser and Urabe 1999; Evans-White and Lamberti 2006; Sterner and Hessen 2003). C, N and P are three essential elements of organisms and have strong interactions in biochemical functioning (Agren 2008). Many ecological processes such as photosynthesis (Elser et al. 2000; Reich et al. 1997), litter decomposition (Güsewell and Gessner 2009), predator-prey relationships (Ngai and Jefferies 2004; Tibbets and Molles 2005), community composition, and species diversity (Bedford et al. 2008; Gusewell et al. 2005) are related to C, N and P contents or C:N:P mass ratios. In freshwater ecosystems, for instance, N:P ratio can influence the productivity and species composition of plant communities (Hall et al. 2005; Harpole et al. 2011; Willby et al. 2001).

C:N:P stoichiometry of submerged vegetation could be influenced by sediment and water column nutrient availability because submerged macrophytes rely on the surrounding sediment and water to satisfy their N and P requirements (Cao et al. 2011; Madsen and Cedergreen 2002). Previous studies have addressed the relative importance of roots and leaves in nutrient uptake of rooted submerged macrophytes (Carignan and Kalff 1980; Rattray et al. 1991; Robach et al. 1995). Given that the concentrations of nutrients in sediments are usually higher than that of in water column, it is generally accepted that sediments are the major source of N and P for submerged macrophytes (Barko and Smart 1981; Best and Mantai 1978; Bristow and Whitcombe 1971; Carignan and Kalff 1980; Smith and Adams 1986).

The C:N:P stoichiometric signatures in plant tissue, however, depend not only on nutrient supply but also on

the light availability in water column. Light availability in water column can affect physiology, morphology, biomass allocation, community structure, or distribution of submersed macrophytes, causing great variation in the C, N, and P concentrations and C:N:P stoichiometry in plant (Cao et al. 2011; Chambers and Kalff 1987; Cronin and Lodge 2003; Xing et al. 2013). For instance, to alleviate low light availability in deep waters, *Potamogeton maackianus* and *Potamogeton malaianus* tend to enhance light harvesting by allocating more biomass to the stem, increasing shoot height and specific leaf area, whereas *Vallisneria natans* tend to allocate more biomass to the leaf than to the rosette (Fu et al. 2012). Furthermore, the C:N:P stoichiometry in plant tissue is not always directly correlated with sediment or water column N and P availability (Güsewell and Koerselman 2002). Indeed, plants will accumulate nutrients in excess of their cellular requirements when their growth is not limited by N and P availability (Cao et al. 2011; Demars and Edwards 2007).

In recent years, eutrophication induced by human activities has changed water and sediment physicochemical conditions extensively. Consequently, these changes have altered nutrient and light availability, leading to variations of nutrient uptake by plants. Thus, current anthropogenic changes in N:P

stoichiometry may have important implications for the ecological functioning of lake ecosystems. Additionally, since submersed macrophytes are different in taxonomy, growth forms, and C/N/P metabolic strategies, their stoichiometric signatures are species-specific (Yuan et al. 2016). Thus, which one of the three factors (taxonomy, sediment, and water column) is having the greatest influence on C:N:P stoichiometry of macrophytes is still unclear. In this paper, we aim to (i) explore how sediment and water column affect C, N, and P stoichiometry of submersed macrophytes and (ii) reveal the relative importance of taxonomy, sediment, and water column on the C:N:P stoichiometry of submersed macrophytes.

Materials and methods

Study sites and field sampling

Fourteen lakes along the mid-lower Yangtze River were investigated from June to August 2014 (Fig. 1). The longitudes of the studied lakes range from 112.2° E to 120.4° E and latitudes range from 29.7° N to 33.8° N. All the studied lakes are large shallow floodplain lakes (water depth ranged from 1.1 to 4.1 m) with different nutrient levels (Table 1).

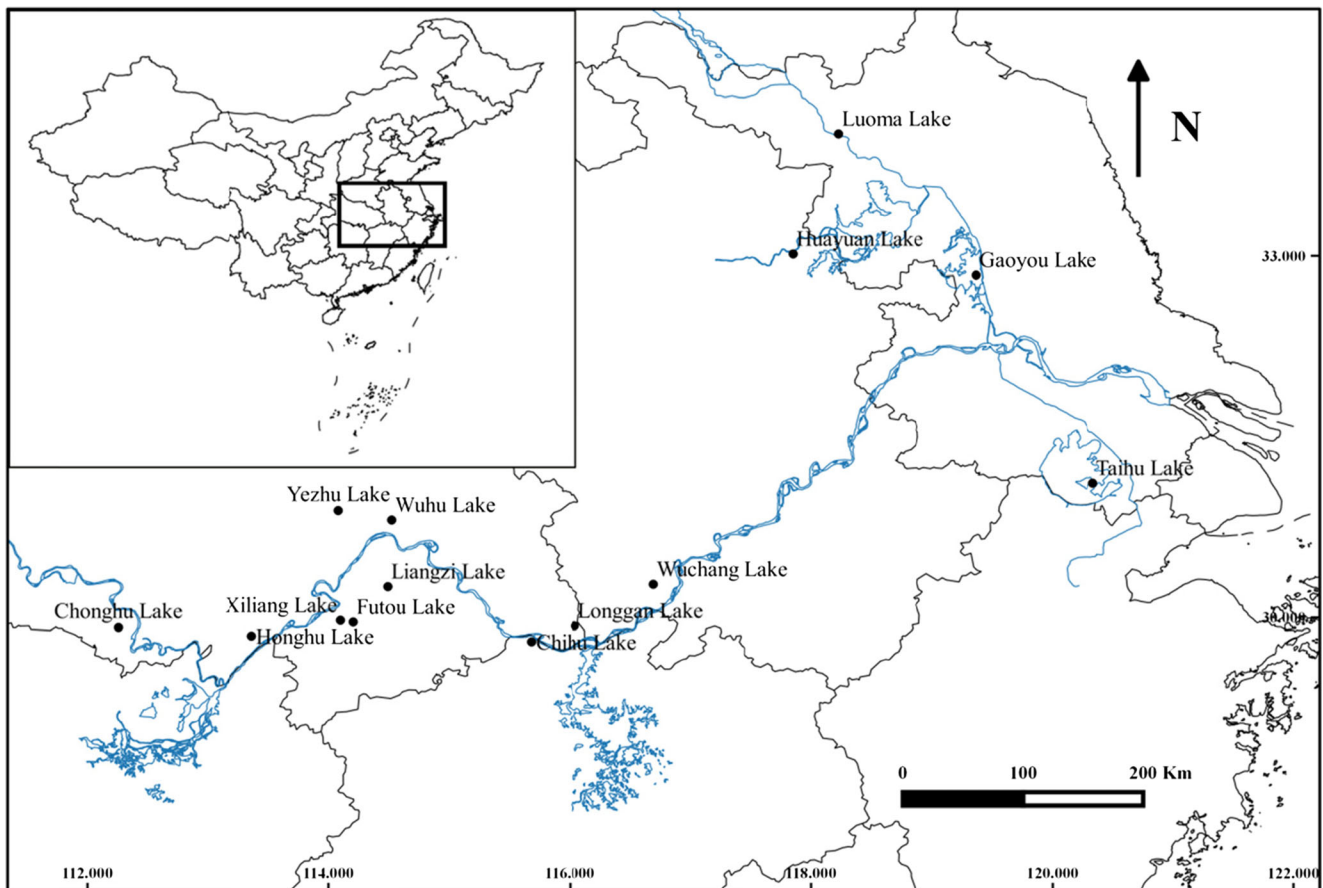


Fig. 1 Location of sampled lakes along the middle and lower Yangtze River, China

Table 1 Geographic and trophic parameters of the 14 lakes along the mid-lower Yangtze River

Lake	Latitude	Longitude	Area (km ²)	S-N (mg g ⁻¹)	S-C (mg g ⁻¹)	S-WC (%)	S-P (mg g ⁻¹)	WD (m)	SD (m)	TN (mg L ⁻¹)	TP (mg L ⁻¹)	Chl <i>a</i> (μg L ⁻¹)	<i>K</i> (m ⁻¹)
Taihu Lake	31.19	120.41	2537.2	1.12	6.1	38.6	0.41	2.1	1.2	0.69	0.02	13.7	1.87
Gaoyou Lake	32.84	119.33	639.2	1.09	11.1	48.9	0.62	1.9	0.7	0.71	0.04	23.6	3.71
Luoma Lake	33.75	118.22	290.9	1.35	18.4	37.2	0.61	2.1	0.9	0.73	0.04	39.9	2.20
Longgan Lake	29.96	116.03	280.5	3.46	28.8	62.1	0.75	2.8	0.6	1.08	0.04	50.6	2.63
Chihu Lake	29.79	115.69	62.8	3.48	45.8	69.0	0.58	4.1	1.6	0.49	0.01	10.9	1.18
Huayuan Lake	33.02	117.85	47.7	1.26	23.8	37.0	0.65	1.7	0.6	0.70	0.04	31.9	3.11
Wuchang Lake	30.24	116.67	100.5	1.32	7.1	40.3	0.55	3.1	1.7	0.58	0.01	6.0	1.74
Yezhu Lake	30.89	114.08	25.9	1.50	9.1	47.7	0.51	1.5	0.4	1.20	0.13	87.9	4.35
Wuhu Lake	30.81	114.50	27.5	2.87	20.6	50.0	0.63	1.7	1.0	1.05	0.04	10.2	1.61
Futou Lake	29.96	114.20	141.2	2.55	27.9	55.0	0.88	2.0	0.8	1.17	0.05	14.2	2.53
Liangzi Lake	30.25	114.55	351.8	2.25	17.6	48.9	0.37	2.0	0.7	0.66	0.04	21.9	2.04
Honghu Lake	29.83	113.36	340.1	2.79	29.9	56.8	0.69	1.5	0.9	0.74	0.06	27.4	2.63
Chonghu Lake	29.92	112.27	16.9	3.10	37.3	63.5	0.70	1.1	1.1	0.64	0.05	7.6	3.80
Xiliang Lake	29.98	114.10	98.1	2.83	33.8	60.7	0.56	1.9	1.9	0.34	0.02	3.9	1.26

S-N sediment nitrogen, S-C sediment carbon, S-WC sediment water content, S-P sediment phosphorus, WD water depth, SD Secchi depth, TN total nitrogen in water column, TP total phosphorus in water column, Chl *a* chlorophyll *a*, *K* light attenuation coefficient

The species we focused on are six common submerged macrophytes which are widely distributed in the Yangtze floodplain lakes: *Potamogeton maackianus*, *Potamogeton malaianus*, *Ceratophyllum demersum*, *Hydrilla verticillata*, *Myriophyllum spicatum*, and *V. natans*. Distribution of the collected submerged macrophytes at the 14 lakes was shown in Table 2. To reduce environmental stress, these macrophyte species take different growth forms and have various adaptive strategies, including morphological feature and metabolism (Ni 2001; Su et al. 2004; Titus and Adams 1979).

The abundance of submerged macrophytes was measured by random sampling (three replicates) at each site with a long-handled scythe-type sampler (0.2 m² in sampling area). All macrophytes were cleaned, sorted by species, and then weighed for wet biomass, respectively. Aboveground parts of the six plant species were collected and put into ziplock bags with waterproof labels. Corresponding water was taken from 0.5 m below the water surface. Surface sediments were collected from the top 0–10-cm layers of undisturbed sediments with a columnar sampling instrument (acrylic glass tube, 5 cm inner diameter). All the samples were put into icebox and taken back to laboratory for further analysis. Water depth (WD), transparency (SD), and light attenuation coefficient (*K*) were also measured at the same locations. WD was determined by a sounding lead. SD was measured by a Secchi disk. Photosynthetically active radiation (PAR) was measured by a LI-COR sensor (at water depth of 0, 0.5, 1.0, and 1.5 m, respectively) coupled with a data logger (Li-1400; LI-COR Company, Lincoln, NE, USA). *K* in the water column was calculated based on the equation: $K = (1 / Z) \ln(I_0 / I_z)$, where *I*_z is PAR at water depth *z* and *I*₀ is PAR at the water surface (Krause-Jensen and Sand-Jensen 1998).

Laboratory analysis

The collected submerged macrophytes were washed carefully to remove epiphytic algae, sediment, and invertebrates. The samples of macrophytes and sediments were oven-dried at 80 °C for 48 h to constant weights and then ground into fine powder using a planetary ball mill (Fritsch, Planetary Micro Mill PULVERISETTE 7 premium line, Germany) before the elemental analyses. The C and N concentrations of plant tissues and sediments were determined by an elemental analyzer (Flash EA 1112 series, CE Instruments, Italy). The total P was measured using a sulfuric acid/hydrogen peroxide digest and the ammonium molybdate ascorbic acid methods (Sparks et al. 1996). Sediment water content was determined gravimetrically by oven-drying at 105 °C to achieve constant weights. Total nitrogen (TN), total phosphorus (TP), and chlorophyll *a* (Chl *a*) of water samples were measured according to Chinese standard methods (Huang et al. 1999).

Table 2 Distribution of the collected submerged macrophytes at the 14 lakes along the mid-lower Yangtze River

Lake	<i>P. maackianus</i>	<i>C. demersum</i>	<i>M. spicatum</i>	<i>H. verticillata</i>	<i>P. malaianus</i>	<i>V. natans</i>
Taihu Lake	+	+	+		+	+
Gaoyou Lake		+	+	+	+	+
Luoma Lake	+		+			+
Longgan Lake		+				
Chihu Lake	+	+		+		
Huayuan Lake		+		+		+
Wuchang Lake						+
Yezhu Lake		+				
Wuhu Lake						+
Futou Lake			+			
Liangzi Lake	+	+	+	+		+
Honghu Lake	+	+	+	+	+	+
Chonghu Lake				+		
Xiliang Lake					+	

Data analysis

The tissue concentrations and the C:N, C:P, and N:P mass ratios at species level were compared with one-way ANOVA. Turkey's HSD was used to test the significance between means. Pearson's correlation analyses were performed using PASW (Predictive Analytics Software) Statistics 18 (SPSS, Inc., Chicago, IL, USA) between tissue elements and corresponding water and sediment parameters, respectively. Redundancy analyses (RDA) were performed with CANOCO for windows (version 5) to elucidate the relationship between plant C:N:P signatures and the environmental parameters. Variance partitioning based on partial redundancy analyses (pRDA) by vegan package in R (R Development Core Team 2014) was used to interpret the relative importance of taxonomy, sediment, and water column on the C:N:P stoichiometry of plant tissues. Monte Carlo permutation tests (Manly 2006) were used to test the significance of each environmental factor and all the constrained factors, respectively.

Results

Environmental conditions in sediment and water column

The average concentrations of sediment C, N, and P were 20.3 ± 14.78 , 2.03 ± 1.06 , and 0.56 ± 0.16 mg g⁻¹, respectively (Table 3). The average sediment C:N:P ratio was 38:4:1 by mass. Sediment water content ranged from 24.7 to 78.1 % and was significantly correlated with sediment C, N, and P concentrations, respectively (sediment C, $r = 0.71$, $p < 0.001$; sediment N, $r = 0.79$, $p < 0.001$; sediment P, $r = 0.35$, $p < 0.001$).

The mean concentrations of TN, TP, and Chl *a* in water column were 0.71 ± 0.19 mg L⁻¹, 0.04 ± 0.03 mg L⁻¹, and 24.0 ± 20.18 µg L⁻¹, respectively. The mean water depth where macrophyte sampling occurred and its associated SD was 2.0 ± 0.72 and 0.9 ± 0.45 m, respectively. The mean light extinction coefficient in water was 2.47 ± 1.09 m⁻¹ (Table 3).

Stoichiometric traits of submerged plant tissues

In our study, the average mass ratio of C:N:P in plant tissues was 191:12:1. The mean concentrations of C, N, and P

Table 3 Mean values of environmental parameters and the results of RDA analysis

	mean	sd	RDA1	RDA2	r^2	p
S-N (mg g ⁻¹)	2.03	1.06	-0.48	0.88	0.01	0.536
S-C (mg g ⁻¹)	20.3	14.78	-0.78	0.63	0.01	0.525
S-WC (%)	49.3	12.66	-0.97	0.23	0.02	0.153
S-P (mg g ⁻¹)	0.56	0.16	-0.94	-0.34	0.18	<0.001
S-CN	9.64	3.60	-0.79	-0.61	0.05	0.018
S-CP	37.5	28.99	0.74	0.68	0.04	0.027
S-NP	3.96	2.47	0.86	0.51	0.10	0.002
WD (m)	2.0	0.72	0.69	0.72	0.11	<0.001
SD (m)	0.9	0.45	0.55	-0.83	0.05	0.016
TN (mg L ⁻¹)	0.71	0.19	-0.85	-0.52	0.04	0.036
TP (mg L ⁻¹)	0.04	0.03	-0.98	-0.19	0.03	0.054
Chl <i>a</i> (µg L ⁻¹)	24.0	20.18	-0.92	-0.39	0.04	0.022
K (m ⁻¹)	2.47	1.09	-1.00	0.06	0.13	<0.001
SD:WD	0.50	0.27	-0.03	-1.00	0.10	<0.001
W-NP	25.73	21.67	0.38	0.93	0.05	0.013
Taxonomy	-	-	-	-	0.24	<0.001

The p values which <0.001 were emphasized in boldface

were 380.5 ± 26.60 , 25.1 ± 5.44 , and $2.50 \pm 1.25 \text{ mg g}^{-1}$, respectively (Table 4). The coefficient of variation (CV) of C, N, and P concentrations was 7.0, 21.7, and 50.0 %, respectively. The mean ratio of C:N was 16.0 ± 4.07 , the mean ratio of C:P was 191.3 ± 96.63 , and the mean ratio of N:P was 11.89 ± 5.01 . The lowest N and P concentrations were observed in *P. malaianus*, whereas the highest N and P concentrations were observed in *C. demersum* and *H. verticillata*, respectively. We also found that *P. malaianus* and *M. spicatum* had higher C:N and C:P than other submerged macrophyte species.

Effects of taxonomy, sediment, and water column on plant tissue stoichiometry

Sediment P concentrations had significant relationships with P, C:P, and N:P in plant tissue (all $p < 0.01$), whereas the concentrations of sediment C and sediment N had no significant relationship with all plant stoichiometric signatures (all $p > 0.05$) (Table 5). As for the factors in water column, both WD and K had significant relationships with P in plant tissues (all $p < 0.01$), whereas concentrations of TN and TP in water columns had no significant relationship with that of P in plant tissues (all $p > 0.05$).

The results of RDA were significant ($df = 20$, $F = 7.00$, $p < 0.001$) to explain C, N, and P stoichiometric signatures of plant tissues (Fig. 2). The whole model explained 46.4 % variation in total (adjusted explained variation was 39.7 %). Taxonomy, sediment P, WD, K, and SD:WD have high explanations of plant stoichiometry at significance of 0.001 level (Table 3).

The results of variance partitioning based on pRDA showed that taxonomy, sediment, and water column explained 18.6 %, 6.9 %, and 8.5 % of total variance of plant stoichiometry, respectively (Fig. 3 and Table 6). The interaction effects of sediment and water column explained 7.9 % of plant stoichiometric variance. Thus apart from taxonomy, factors in water column had a higher influence on plant C, N, and P stoichiometry.

Discussion

C:N:P stoichiometric characteristics of submerged macrophytes

In our study, the mean C:N:P mass ratio of macrophytes was approximately 191:12:1, lower than (except N:P ratio) the ratio of growth rate limitation (272:11:1) reported by Demars and Edwards (2007) for macrophytes in the River Spey in Great Britain. The mean concentration of tissue C in our study was much lower than 438 mg g^{-1} reported by Demars and Edwards (2007). This result was consistent with the previous study in eastern China and mid-lower Yangtze floodplain lakes (Xia et al. 2014; Xing et al. 2013). The low concentrations of C in macrophytes implied that metabolism and growth may be hampered by the local low C availability. In a community level, tissue C concentration could be used to examine patterns such as variation in productivity and community structure. In the present study, a significant positive relationship ($r = 0.49$, $p < 0.001$) was found between community C concentration and macrophyte community biomass (Fig. 4). Therefore, it was suggested that the decline of macrophyte community in Yangtze floodplain shallow lakes was associated with the local low C availability in macrophytes. We inferred this was triggered by low light levels because light could alter the capacity of C fixation by changing the intensity of photosynthesis. This result could potentially occur at the species level and the community level simultaneously. At the species level, biomass of submerged macrophyte species decreased as a result of low capacity of C fixation in low light availability. At the community level, macrophyte species declined generally from high C concentration species to low C concentration species. Previous study (Qiu and Wu 1997) had revealed that succession of submerged macrophyte communities in the process of eutrophication of the Yangtze floodplain shallow lakes is from *Potamogeton* to species dominated by *M. spicatum*, *V. natans*, and *C. demersum*. This

Table 4 Tissue C, N, and P concentrations and C:N, C:P, and N:P mass ratios of submerged macrophytes in Yangtze floodplain shallow lakes

	<i>n</i>	C	N	P	C:N	C:P	N:P
All species	183	380.5 ± 26.60	25.1 ± 5.44	2.50 ± 1.25	16.0 ± 4.07	191.3 ± 96.63	11.89 ± 5.01
<i>P. maackianus</i>	45	407.9 ± 15.50 a	25.4 ± 2.93 b	2.13 ± 0.80 bc	16.3 ± 2.02 b	219.5 ± 85.89 a	13.43 ± 4.90 a
<i>C. demersum</i>	29	366.7 ± 22.31 c	30.5 ± 4.35 a	2.91 ± 1.30 ab	12.3 ± 2.26 c	160.4 ± 92.42 ab	12.73 ± 6.42 a
<i>M. spicatum</i>	37	385.0 ± 22.76 b	20.7 ± 4.62 c	2.13 ± 0.97 bc	19.5 ± 4.47 a	226.2 ± 120.46 a	11.52 ± 5.09 ab
<i>H. verticillata</i>	26	362.7 ± 14.04 c	25.1 ± 4.25 b	3.70 ± 1.89 a	14.9 ± 2.60 b	130.5 ± 76.08 b	8.70 ± 4.79 b
<i>P. malaianus</i>	24	383.5 ± 18.25 b	20.7 ± 3.60 c	1.93 ± 0.73 c	19.0 ± 3.23 a	227.4 ± 81.99 a	11.79 ± 3.29 ab
<i>V. natans</i>	22	353.2 ± 19.75 c	29.4 ± 5.01 a	2.56 ± 0.75 bc	12.5 ± 2.99 c	148.5 ± 42.54 b	12.13 ± 3.14 ab

The number of species (*n*) and values (mg g^{-1}) (\pm sd) are shown. Differences between six submerged macrophytes were tested using one-way ANOVA with Turkey’s HSD post hoc test of significance. Different letters indicate significant differences at $p < 0.05$

Table 5 Pearson correlations between tissue C:N:P stoichiometric signatures and environmental parameters

	S-N	S-C	S-WC	S-P	S-CN	S-CP	S-NP	WD	SD	TN	TP	Chl <i>a</i>	K	SD:WD	W-NP
P-N	0.10	0.09	0.15*	0.10	0.02	-0.01	-0.05	-0.06	0.11	0.04	0.09	0.02	0.06	0.13	-0.17*
P-C	-0.08	-0.08	-0.09	-0.06	0.02	-0.03	-0.001	-0.07	0.22**	0.003	-0.05	-0.02	-0.17*	0.24**	-0.12
P-P	0.11	0.13	0.20**	0.45**	0.17*	-0.14	-0.27**	-0.27**	-0.003	0.08	0.13	0.07	0.36**	0.20**	-0.14
P-CN	-0.11	-0.11	-0.15*	-0.08	-0.03	-0.02	0.03	0.03	-0.05	-0.03	-0.10	-0.04	-0.09	-0.06	0.13
P-CP	-0.03	-0.06	-0.14	-0.42**	-0.19*	0.16*	0.29**	0.26**	0.10	-0.18*	-0.18*	-0.19*	-0.36**	-0.08	0.13
P-NP	-0.01	-0.04	-0.08	-0.44**	-0.20**	0.18*	0.30**	0.30**	0.16*	-0.18*	-0.15*	-0.19*	-0.37**	-0.06	0.08

Significance is indicated by * < 0.05, ** < 0.01

S-N sediment nitrogen, S-C sediment carbon, S-WC sediment water content, S-P sediment phosphorus, S-CN sediment C:N ratio, S-CP sediment C:P ratio, S-NP sediment N:P ratio, WD water depth, SD Secchi depth, TN total nitrogen in water column, TP total phosphorus in water column, Chl *a* chlorophyll *a*, K light attenuation coefficient, SD:WD Secchi depth to water depth ratio, W-NP N:P ratio in water column

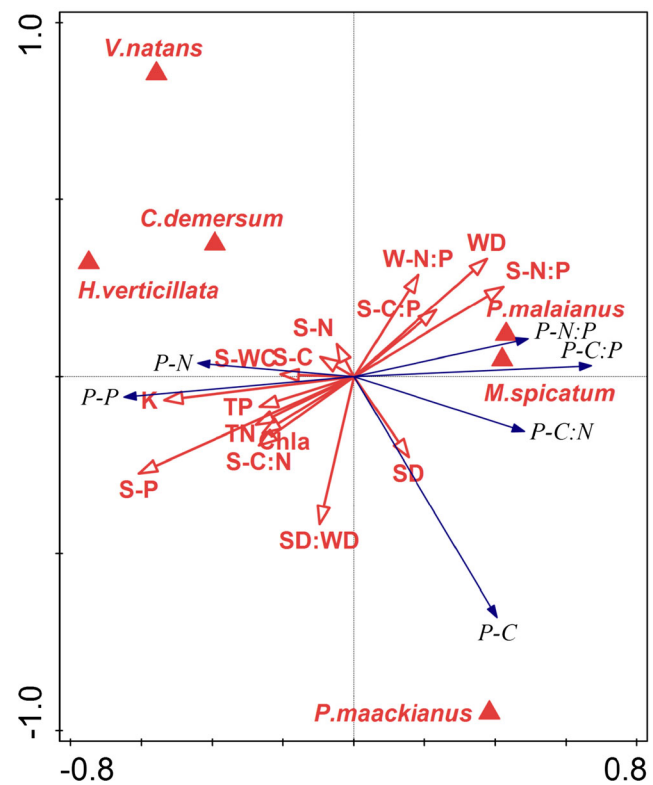


Fig. 2 RDA plot of tissue C:N:P stoichiometric signatures and explaining variables

succession pattern supported our result well because C concentration in *Potamogeton* was significantly higher than that of *V. natans* and *C. demersum* (Table 4).

Our results did not support the notion that plants were limited by nutrient availability because the obtained tissue N and P concentrations were much higher than the critical concentrations of 13 and 1.3 mg g⁻¹, respectively (Gerloff and Krombholz 1966). Furthermore, significant positive relationships between sediment P and plant tissue P concentration were observed. This suggested that plants would accumulate P even though the P availability in environment was sufficient (Demars and Edwards 2007; Garbey et al. 2004; Rattray et al. 1991). This “luxury” uptake could be an adaptive strategy which would benefit the plant once the limited factors were relieved. Therefore, P concentration of submerged vegetation was generally more variable than the N concentration, which is consistent with previous reports (Fernández-Alález et al. 1999; Güsewell and Koerselman 2002).

High interspecific variations of C:N:P stoichiometry were found between the six species. The species-specific variation of nutrient composition can be an important trait for identifying the ecological strategies and for predicting the result of competition in plant communities (Tilman 1982). *V. natans*, for instance, had lowest C concentrations which benefited its tolerance in low light stress and was consistent with its low light compensation point

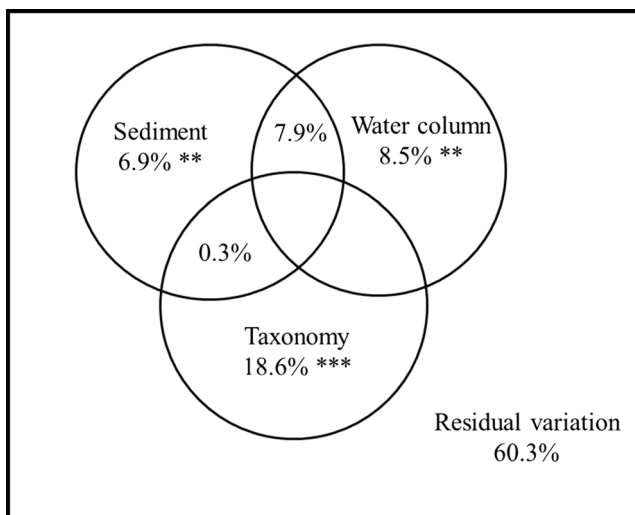


Fig. 3 Variance partitioning based on pRDA analysis for sediment, water column, and taxonomy on tissue C:N:P stoichiometric signatures

of photosynthesis (Su et al. 2004; Titus and Adams 1979). In addition, *V. natans* had lower C/N metabolic level and higher carbohydrate storage than other submerged macrophytes (Yuan et al. 2016). On the contrary, *P. malaianus* and *M. spicatum*, belonging to canopy formers, had higher C:N and C:P than other submerged macrophyte species. This might be another adaptive strategy which is the result of allocating more C on stem to elongate their shoots toward water surface (Ni 2001; Titus and Adams 1979). *C. demersum* is a submerged, rootless, free-floating aquatic plant. Unlike rooted vegetation, it requires nutrient uptake from the water (Denny 1987). However, a significant positive relationship ($r = 0.41$, $p = 0.027$) was still observed between sediment P and tissue P in *C. demersum*. This result implied that sediment might play a role in nutrient uptake in *C. demersum* because it may attach itself to the sediment sometimes. Besides, our results showed that tissue C concentration

of *C. demersum* generally had a higher coefficient of variation than that of other five submerged macrophytes. And previous study (Li et al. 2013) also revealed that the N and P concentrations in leaves and stems of *C. demersum* were relatively unstable. These results indicated that *C. demersum* had a low stoichiometric homeostasis.

Sources of variability in aquatic plant tissue stoichiometry

The results of pRDA revealed that both the sediment and water column had significant effects on C, N, and P concentrations and C:N:P stoichiometry. However, the explained variation was low compared to the variation explained by species composition, indicating that taxonomic effect was more important than that of sediment and water column at the community level. High interaction effects between sediment and water column in explaining stoichiometric variation were observed in this study. It was not surprising because the two components were usually not independent. For instance, re-suspension and nutrient release from sediment could alter light attenuation coefficient and stoichiometry of N and P in water column. Plant tissue P had significant positive relationship with sediment P, while it had no significant relationship with TP in water column, implying that submerged macrophyte uptake P was generally from sediment as mentioned in the “Introduction” (Barko and Smart 1981; Best and Mantai 1978; Carignan and Kalff 1980). However, this cannot be regarded as direct evidence that dominance of nutrient uptake cannot be mediated by leaves, especially in oligotrophic streams. For instance, there is no correlation between the total P concentration in sediment and tissues, but a positive relation to water P concentration was observed in Alsace Rhine streams, supporting the view that leaf uptake can be the main pathway for nutrient uptake in some streams (Robach et al. 1995). In water column, our study showed that water depth has a significant negative effect on plant tissue P which is

Table 6 Percentage (100 %) of explained variance based on variance partitioning for C:N:P stoichiometric signatures and for C, N, P, C:N, C:P, and N:P, respectively, in response to sediment, water column, and taxonomy

	<i>S</i>		<i>W</i>		<i>T</i>		<i>S</i> × <i>W</i>	<i>W</i> × <i>T</i>	<i>S</i> × <i>T</i>	<i>S</i> × <i>W</i> × <i>T</i>	Residuals
	(%)	<i>p</i>	(%)	<i>p</i>	(%)	<i>p</i>					
C:N:P stoichiometric signatures	6.9	0.003	8.5	0.002	18.6	<0.001	7.9	−1.8	0.3	−0.7	60.3
C	3.7	0.002	9.3	<0.001	45.9	<0.001	−3.9	4.0	−2.3	1.4	42.0
N	1.8	0.073	4.1	0.012	47.8	<0.001	1.3	−2.4	−1.2	−0.5	49.2
P	6.2	0.004	7.6	0.002	17.3	<0.001	11.6	0.7	−0.4	1.8	55.2
C:N	3.4	0.015	4.1	0.012	49.9	<0.001	−0.4	−3.6	−2.2	0.02	48.8
C:P	7.2	<0.001	8.5	<0.001	16.4	<0.001	8.8	−2.3	0.5	−0.8	61.7
N:P	9.3	0.002	5.0	0.029	3.3	0.03	8.8	1.7	0.4	0.8	70.8

The *p* values which <0.001 were emphasized in boldface

S sediment, *W* water column, *T* taxonomy

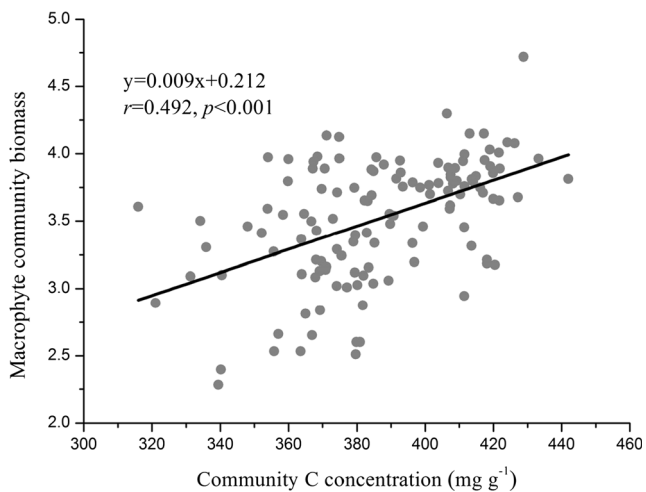


Fig. 4 Significant positive relationship between community C concentration and macrophyte community biomass. Macrophyte community biomass was \log_{10} transformed before analysis

consistent with previous study (Li et al. 2015). Indeed, in deep water, most submerged macrophytes tend to allocate more biomass to stem for shoot elongation to alleviate low light availability (Fu et al. 2012; Strand and Weisner 2001). Changes of this biomass allocation could lead to low tissue P and high C:P because the structural organs generally have higher C and lower P than leaves (Li et al. 2013).

In the present study, sediment had a lower effect on plant C:N:P stoichiometry than water column. All explanations of C:N:P stoichiometric signatures except N:P explained by water column were higher than that of sediment, even though sediment supplies major nutrient to the plants (Best and Mantai 1978; Carignan and Kalff 1980; Smith and Adams 1986). In addition, only factors associated with light availability (SD, *K*, and SD:WD) in water column were significantly correlated with plant C concentrations. These results suggested that growth of submerged macrophytes in Yangtze floodplain lakes was primarily limited by low light stress rather than nutrient availability. With the aggravation of eutrophication, we infer that this situation will be even worse than the present.

Conclusions

This study revealed the relative importance of taxonomy, sediment, and water column on the C:N:P stoichiometry of submerged macrophytes in Yangtze floodplain shallow lakes. Results showed that submerged macrophytes had a species-specific identity in determining C:N:P stoichiometry. Apart from taxonomy, water column had higher explanations on plant stoichiometry than that of sediment. In the present study, a significant positive relationship was found between community C concentration and macrophyte community biomass, indicating that the local low C availability was probably the

main reason why submerged macrophytes declined in Yangtze floodplain shallow lakes. Parameters of WD, *K*, and SD:WD associated with light availability had significant effects on C:N:P stoichiometry. It is suggested that submerged macrophytes in Yangtze floodplain shallow lakes are primarily limited by low light levels rather than nutrient availability.

Acknowledgments This work was supported by grant from the State Key Laboratory of Freshwater Ecology and Biotechnology (grant number 2014FBZ02).

Compliance with ethical standards

Conflict of interest statement The authors declare that there are no conflicts of interest.

References

- Agren GI (2008) Stoichiometry and nutrition of plant growth in natural communities. *Annu Rev Ecol Evol Syst*:153–170
- Andersen T, Elser JJ, Hessen DO (2004) Stoichiometry and population dynamics. *Ecol Lett* 7:884–900
- Barko JW, Smart RM (1981) Sediment-based nutrition of submersed macrophytes. *Aquat Bot* 10:339–352
- Bedford BL, Walbridge MR, Aldous A (2008) Patterns in nutrient availability and plant diversity of temperate North American. *Wetlands Ecology* 80:2151–2169
- Best MD, Mantai KE (1978) Growth of *Myriophyllum*: sediment or lake water as the source of nitrogen and phosphorus. *Ecology* 59:1075–1080
- Bristow JM, Whitcombe M (1971) The role of roots in the nutrition of aquatic vascular plants. *Am J Bot* 58:8–13
- Cao T, Ni L, Xie P, Xu J, Zhang M (2011) Effects of moderate ammonium enrichment on three submersed macrophytes under contrasting light availability. *Freshw Biol* 56:1620–1629
- Carignan R, Kalff J (1980) Phosphorus sources for aquatic weeds: water or sediments? *Science* 207:987–989
- Chambers PA, Kalff J (1987) Light and nutrients in the control of aquatic plant community structure. I In situ experiments *Journal of Ecology* 75:611–619
- Cronin G, Lodge DM (2003) Effects of light and nutrient availability on the growth, allocation, carbon/nitrogen balance, phenolic chemistry, and resistance to herbivory of two freshwater macrophytes. *Oecologia* 137:32–41
- Demars BOL, Edwards AC (2007) Tissue nutrient concentrations in freshwater aquatic macrophytes: high inter-taxon differences and low phenotypic response to nutrient supply. *Freshw Biol* 52:2073–2086
- Denny P (1987) Mineral cycling by wetland plants—a review. *Arch Hydrobiol Beih* 27:1–25
- Elser JJ et al. (2000) Biological stoichiometry from genes to ecosystems. *Ecol Lett* 3:540–550
- Elser JJ, Urabe J (1999) The stoichiometry of consumer-driven nutrient recycling: theory, observations, and consequences. *Ecology* 80:735–751
- Evans-White MA, Lamberti GA (2006) Stoichiometry of consumer-driven nutrient recycling across nutrient regimes in streams. *Ecol Lett* 9:1186–1197

- Fernández-Aláez M, Fernández-Aláez C, Bécares E (1999) Nutrient content in macrophytes in Spanish shallow lakes. *Hydrobiologia* 408–409:317–326
- Fu H, Yuan G, Cao T, Ni L, Zhang M, Wang S (2012) An alternative mechanism for shade adaptation: implication of allometric responses of three submersed macrophytes to water depth. *Ecol Res* 27:1087–1094
- Güsewell S, Gessner MO (2009) N:P ratios influence litter decomposition and colonization by fungi and bacteria in microcosms. *Funct Ecol* 23:211–219
- Güsewell S, Koerselman W (2002) Variation in nitrogen and phosphorus concentrations of wetland plants. *Perspectives in Plant Ecology, Evolution and Systematics* 5:37–61
- Garbey C, Murphy K, Thiébaud G, Muller S (2004) Variation in P-content in aquatic plant tissues offers an efficient tool for determining plant growth strategies along a resource gradient. *Freshw Biol* 49:346–356
- Gerloff G, Krombholz P (1966) Tissue analysis as a measure of nutrient availability for the growth of angiosperm aquatic plants. *Limnol Oceanogr* 11:529–537
- Güsewell S, Bailey K, Wj BB (2005) Nutrient limitation and botanical diversity in wetlands: can fertilisation raise species richness? *Oikos* 109: 71–80
- Hall SR, Smith VH, Lytle DA, Leibold MA (2005) Constraints on primary producer N:P stoichiometry along N:P supply ratio gradients. *Ecology* 86:1894–1904
- Harpole WS et al. (2011) Nutrient co-limitation of primary producer communities. *Ecol Lett* 14:852–862
- Huang X, Chen W, Cai Q (1999) Survey, observation and analysis of lake ecology standard methods for observation and analysis in Chinese Ecosystem Research Network, Series V. Standards Press of China, Beijing in Chinese
- Krause-Jensen D, Sand-Jensen K (1998) Light attenuation and photosynthesis of aquatic plant communities. *Limnol Oceanogr* 43:396–407
- Li W, Cao T, Ni L, et al. (2013) Effects of water depth on carbon, nitrogen and phosphorus stoichiometry of five submersed macrophytes in an in situ experiment. *Ecol Eng* 61:358–365
- Li W, Cao T, Ni L, et al. (2015) Size-dependent C, N and P stoichiometry of three submersed macrophytes along water depth gradients. *Environmental Earth Sciences* 74:3733–3738
- Madsen TV, Cedergreen N (2002) Sources of nutrients to rooted submersed macrophytes growing in a nutrient-rich stream. *Freshw Biol* 47:283–291
- Manly BF (2006) Randomization, bootstrap and Monte Carlo methods in biology, 3rd edn. Chapman & Hall/CRC, London
- Ngai JT, Jefferies RL (2004) Nutrient limitation of plant growth and forage quality in Arctic coastal marshes. *J Ecol* 92:1001–1010
- Ni L (2001) Growth of *Potamogeton maackianus* under low-light stress in eutrophic water. *J Freshw Ecol* 16:249–256
- Qiu D, Wu Z (1997) On the decline and restoration of submersed vegetation in eutrophic shallow lakes. *Journal of Lake Science* 9:82–88
- R Core Team (2014) R: a language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria
- Ratray M, Howard-Williams C, Brown J (1991) Sediment and water as sources of nitrogen and phosphorus for submersed rooted aquatic macrophytes. *Aquat Bot* 40:225–237
- Reich PB, Walters MB, Ellsworth DS (1997) From tropics to tundra: global convergence in plant functioning. *Proc Natl Acad Sci* 94: 13730–13734
- Robach F, Hajnsek I, Eglin I, Trémolières M (1995) Phosphorus sources for aquatic macrophytes in running waters: water or sediment? *Acta Botanica Gallica* 142:719–731
- Smith CS, Adams MS (1986) Phosphorus transfer from sediments by *Myriophyllum spicatum*. *Limnology & Oceanography* 31:1312–1321
- Sparks DL et al (1996) Methods of soil analysis. Part 3—Chemical methods. Soil Science Society of America, Madison, pp 869–919
- Sterner RW, Hessen DO (2003) Algal nutrient limitation and the nutrition of aquatic herbivores. *Annual Review of Ecology & Systematics* 25:1–29
- Strand JA, Weisner SE (2001) Morphological plastic responses to water depth and wave exposure in an aquatic plant (*Myriophyllum spicatum*). *J Ecol* 89:166–175
- Su W, Zhang G, Zhang Y, Xiao H, Xia F (2004) The photosynthetic characteristics of five submersed aquatic plants. *Acta hydrobiologica Sinica* 28:391–395
- Tibbets TM, Molles MC (2005) C:N:P stoichiometry of dominant riparian trees and arthropods along the Middle Rio Grande. *Freshw Biol* 50:1882–1894
- Tilman D (1982) Resource competition and community structure. Princeton University Press, Princeton
- Titus JE, Adams MS (1979) Coexistence and the comparative light relations of the submersed macrophytes *Myriophyllum spicatum* L. and *Vallisneria americana* Michx. *Oecologia* 40:273–286
- Willby NJ, Pulford ID, Flowers TH (2001) Tissue nutrient signatures predict herbaceous-wetland community responses to nutrient availability. *New Phytol* 152:463–481
- Xia CX, Yu D, Wang Z, Xie D (2014) Stoichiometry patterns of leaf carbon, nitrogen and phosphorus in aquatic macrophytes in eastern China. *Ecol Eng* 70:406–413
- Xing W, Wu H, Hao B, Liu G (2013) Stoichiometric characteristics and responses of submersed macrophytes to eutrophication in lakes along the middle and lower reaches of the Yangtze River. *Ecol Eng* 54:16–21
- Yuan G, Fu H, Zhong J, et al. (2016) Growth and C/N metabolism of three submersed macrophytes in response to water depths. *Environ Exp Bot* 122:94–99