

# Efficient photocatalytic degradation of tetracycline hydrochloride by $\text{Ag}_3\text{PO}_4$ under visible-light irradiation

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**Abstract** A facile, environmental-friendly  $\text{Ag}_3\text{PO}_4$ -PN photocatalyst was synthesized by a simple precipitation method at room temperature in the presence of ammonia and polyvinyl pyrrolidone (PVP). As-synthesized samples were characterized by scanning electron microscopy (SEM), X-ray diffraction (XRD), X-ray photoelectron spectroscopy (XPS), and UV–visible diffuse reflectance spectroscopy (UV–vis DRS). The enhancement of photocatalytic efficiency of  $\text{Ag}_3\text{PO}_4$ -PN is strongly dependent on the excellent photo-absorption capacity, sharp edges and corners, and synergistic effect of PVP and  $\text{NH}_3\cdot\text{H}_2\text{O}$ . The effects of catalyst dosage, TC concentration and solution pH were explored with tetracycline hydrochloride (TC) as target contamination. The mineralization was evaluated by total organic carbon (TOC) analysis and determination of the concentration of inorganic ions such as  $\text{NO}_3^-$  and  $\text{Cl}^-$ . Radical detection experiment indicated the  $\text{h}^+$  and  $\cdot\text{O}^{2-}$  are major active species in the degradation of TC by  $\text{Ag}_3\text{PO}_4$ -PN. Moreover, photocatalyst stability and regeneration experiments exhibited the favorable stability and rejuvenation ability, suggesting a promising prospect of practical application of  $\text{Ag}_3\text{PO}_4$  in the wastewater treatment.

**Keywords** Photocatalytic reaction ·  $\text{Ag}_3\text{PO}_4$  · Tetracycline hydrochloride · Radical detection · Catalyst regeneration

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## Introduction

As a “green” technology, semiconductor photocatalysis has been widely used for the treatment of water polluted by organic dye [Mahmoodi et al. 2011; Xiao et al. 2012].  $\text{TiO}_2$  was one of the most widely researched photocatalyst, with the advantages of non-toxicity, high stability, low-cost. Unfortunately,  $\text{TiO}_2$  can only utilize the ultraviolet ray consisting of about 5 % of the solar spectrum due to the relatively wide band gap and the high recombination rate, which seriously confine the photocatalysis performance and practical application of  $\text{TiO}_2$  [Katsumata et al. 2013]. From the view of using solar light for environmental remediation, exploring more efficient visible-light-induced photocatalysts has attracted considerable attention.

In recent years, silver orthophosphate ( $\text{Ag}_3\text{PO}_4$ ) is being thoroughly investigated by increasing researchers, because of excellent photocatalytic performance and efficient separation of photoexcited electrons and holes.  $\text{Ag}_3\text{PO}_4$  exhibited extremely excellent oxidative capacities for the evolution of  $\text{O}_2$  from water, as well as the degradation of organic dyes under the solar irradiation [Yi et al. 2010]. It was reported by Bi et al., that the shape and facet effects of  $\text{Ag}_3\text{PO}_4$  crystals on their photocatalytic properties, where the photocatalytic performance of single-crystalline  $\text{Ag}_3\text{PO}_4$  rhombic dodecahedrons with only {110} facets exposed was much higher than that of cubes bounded entirely by {100} facets, ascribing to the higher surface energy of {110} facets [Bi et al. 2011]. In addition, other researchers successfully synthesized different shapes of  $\text{Ag}_3\text{PO}_4$  single crystals, such as the cubes [Yang et al. 2014], tetrapods [Wang et al. 2013], myriapods [Teng et al. 2015; Wang et al. 2014c], branches [Li et al. 2014; Wang et al. 2014a], and spherical nanocrystals [Liu et al. 2014]. In spite of high efficiency of  $\text{Ag}_3\text{PO}_4$  previously reported, the self-corrosion and bad stability always limit the application

of  $\text{Ag}_3\text{PO}_4$  photocatalyst. Therefore, it becomes very important for  $\text{Ag}_3\text{PO}_4$  to successful regeneration.

As a type of wide pollutant, the pharmaceuticals and personal care products (PPCPs) are focused on increasing concern by scientists [Daughton and Ternes 1999; Ellis 2006; Ternes et al. 2004]. Among these, antibiotics are widely used for human and animal health, which are paid more comprehensive attention than other PPCPs [Golovko et al. 2014; Ortiz de Garcia et al. 2014].

In general, the  $\text{Ag}_3\text{PO}_4$  photocatalysis degradation under visible-light irradiation of organic pollutants has mainly been applied to colored dyes, including methylene blue, rhodamine B, and methyl orange [Hua et al. 2015], while few reports have focused on antibiotics compounds [Wang et al. 2015; Wu et al. 2013].

As the second most usage of antibiotics, tetracycline (TC) has been used as broad-spectrum antibacterial agent for human and animal health, leading to a large amount of residual in aquatic and soil environment, which in turn threatened human health for appearance of drug-resistant bacteria and clinical side effects on liver and kidney [Martins et al. 2014]. Therefore, it is essential to exploit cost-efficient and feasible treatment technologies to remove these contaminations in the aquatic environment.

In this paper, we successfully prepared high-efficiency, micron-scale polyhedral  $\text{Ag}_3\text{PO}_4$ -PN photocatalyst with sharp corners and edges with the synergetic assistances of PVP and  $\text{NH}_3\cdot\text{H}_2\text{O}$ . The catalysts were characterized by a series of methods. In this paper, tetracycline hydrochloride (TC), a kind of common broad-spectrum antibiotics, was selected as the target pollutant. Then, we investigated the photocatalysis performance and mineralization capacity of TC by as-synthesized  $\text{Ag}_3\text{PO}_4$ -PN. Moreover, the photocatalytic reaction mechanism, photocatalyst stability, and regeneration of obtain  $\text{Ag}_3\text{PO}_4$ -PN were also surveyed.

## Experimental

### Materials and methods

All the reagents used were of analytical purity and were used without further purification. Silver nitrate ( $\text{AgNO}_3$ ), disodium hydrogen phosphate dodecahydrate ( $\text{Na}_2\text{HPO}_4\cdot 12\text{H}_2\text{O}$ ) and ammonium oxalate (AO) were purchased from Tianjin Kermel Chem. Reagent Co., Ltd. Polyvinyl pyrrolidone (PVP, Mw 58000, K 29–32) was purchased from Shanghai Aladdin BioChem. Tech. Co., Ltd. Ammonia solution ( $\text{NH}_3\cdot\text{H}_2\text{O}$ ) was purchased from Luoyang Chem. Reagent Factory. Benzoquinone (BQ) and t-butyl alcohol (TBA) were Flukaimport reagents subpackaged by Shanghai Chemical Dispensing Factory and Shanghai Asus Fine Chemical Co.,

Ltd. Hydrogen peroxide ( $\text{H}_2\text{O}_2$  30 %) was purchased by Tianjin Fengchuan Chemical Reagent Technologies Co., Ltd.

In a typical synthesis, 0.2 g PVP was added in 300 ml deionized water with continuous stirring for 10 min. 0.6 g  $\text{AgNO}_3$  was dissolved in the above solution and stirred for 20 min in darkness. Ammonia solution (1:1, v:v) was added until solution became clear, then continuously stirred for 10 min. Then 50 ml, 0.15 M  $\text{Na}_2\text{HPO}_4$  solution was added drop-by-drop to above mixture solution, and stirred for 4 h. The yellow precipitate was centrifuged and washed with deionized water and absolute ethanol for several times and dried at 60 °C for 12 h and was designed as  $\text{Ag}_3\text{PO}_4$ -PN. And three similar procedures were used to prepare  $\text{Ag}_3\text{PO}_4$  without the presence of PVP,  $\text{NH}_3\cdot\text{H}_2\text{O}$ , PVP, and  $\text{NH}_3\cdot\text{H}_2\text{O}$ , and were designed as  $\text{Ag}_3\text{PO}_4$ -N,  $\text{Ag}_3\text{PO}_4$ -P, and  $\text{Ag}_3\text{PO}_4$ -A, respectively.

### Characterization

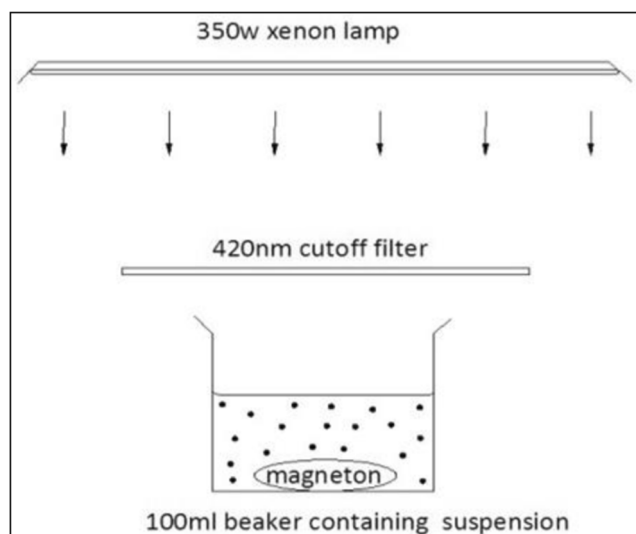
X-ray diffraction (XRD) analysis was performed on an X-ray diffractometer (X'Pert PRO MRD, PANalytical, the Netherlands) equipped with a  $\text{Cu K}\alpha$  X-ray source. The morphology of the samples was investigated by scanning electron microscopy (SEM, Quonxe-2000, Philips, the Netherlands). The surface elemental component and the chemical state of the sample were analyzed by X-ray photoelectron spectroscopy (XPS, PHI 5300, Perkin Elmer) with a monochromatized  $\text{Al K}\alpha$  X-ray source ( $h\nu = 1487$  eV). UV–vis diffuse reflection spectra (UV–vis DRS) of the samples were obtained using a UV–vis spectrophotometer (Caly 5000, Agilent, USA). The band gap energy ( $E_g$ ) of these samples was evaluated using the following equation [Guan et al. 2013; Liang et al. 2012]:

$$\alpha(h\nu) = A(h\nu - E_g)^2 \quad (1)$$

where  $\alpha$ ,  $\nu$ ,  $E_g$ , and A are the absorption coefficient, light frequency, band gap energy, and a constant, respectively [Yi et al. 2010].

### Measure of photocatalytic activity

The as-prepared photocatalyst activity was measured by the degradation of TC. A 100 mg photocatalyst was added to 50 mL of contamination solution as the initial concentration was 30 mg/L. Before illumination, the solution with a certain amount of photocatalyst was stirred for 30 min in darkness to reach the adsorption and desorption equilibrium. A 420-nm filter was placed between xenon lamp and reaction solution to filter ultraviolet. And the experimental device was illustrated in Fig. 1. At the process of the illumination of 350 W xenon lamp, a specified volume of solution was fetched at a given



**Fig. 1** The illustration of photocatalysis experiment device

time interval and filtered immediately with 0.45  $\mu\text{m}$  membrane filter. The concentration of TC solution were evaluated by UV–vis spectroscopy (UV-2450, Shimadzu, Japan), with 357.5 nm of the maximum absorption wavelength. The degradation efficiency was calculated by  $1 - c/c_0$ , while the  $c$  was instantaneous concentration and  $c_0$  was solution concentration after adsorption process. The data of degradation experiment was the average value of three times parallel experiments. And the standard deviations of the average of independent runs are all between 1.2 and 2.9 %.

The mineralization degree of TC and residual of long-lived organic intermediates in the solutions [Zhang et al. 2013] were measured by total organic carbon (TOC). The TOC concentration of filtrate was determined by using Total Organic Carbon Analyzer (TOC-V<sub>CPH/CPN</sub>, Shimadzu, Japan). The  $\text{NO}_3^-$  and  $\text{Cl}^-$  concentrations of TC solution were determined by ion chromatograph (IC), equipped with anion chromatography column (AS-14), anion guard column (AG-14), chemical suppressor (AMMS 30), and 8.0 mM  $\text{Na}_2\text{CO}_3/1.0$  mM  $\text{NaHCO}_3$  as a leachate with 0.8 mL  $\text{min}^{-1}$  of flow rate and dilute sulphuric acid as regeneration solution.

### Detection of reactive oxygen species

The scavenging experiments of reactive oxygen species were similar to the photodegradation experiments. Different quantity of scavengers was added into the TC solution prior to addition of catalysts. These experiments were carried out according to the previous study [Lin et al. 2012]. The concentrations of scavengers were  $1 \times 10^{-3}$  mol/L (7.2 mg),  $1.6 \times 10^{-2}$  mol/L (0.1 g), and  $1 \times 10^{-3}$  mol/L (5.4 mg) for TBA, AO, and BQ, respectively. The concentration of

solution was also measured by solution absorbance using a spectrophotometer (Shimadzu UV-2450).

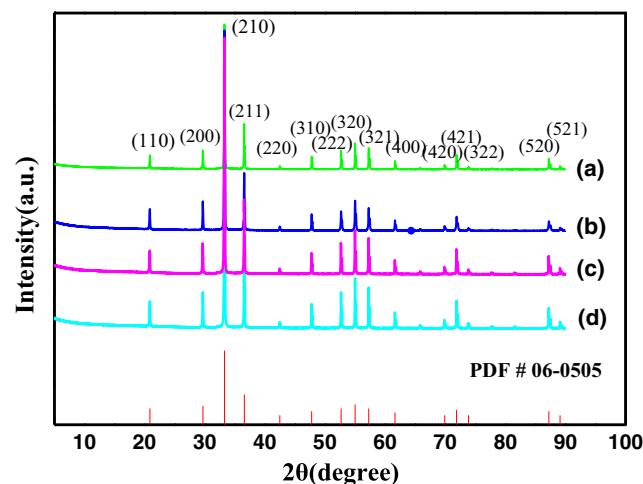
## Results and discussion

### Characterization of photocatalysts

#### Crystal structure

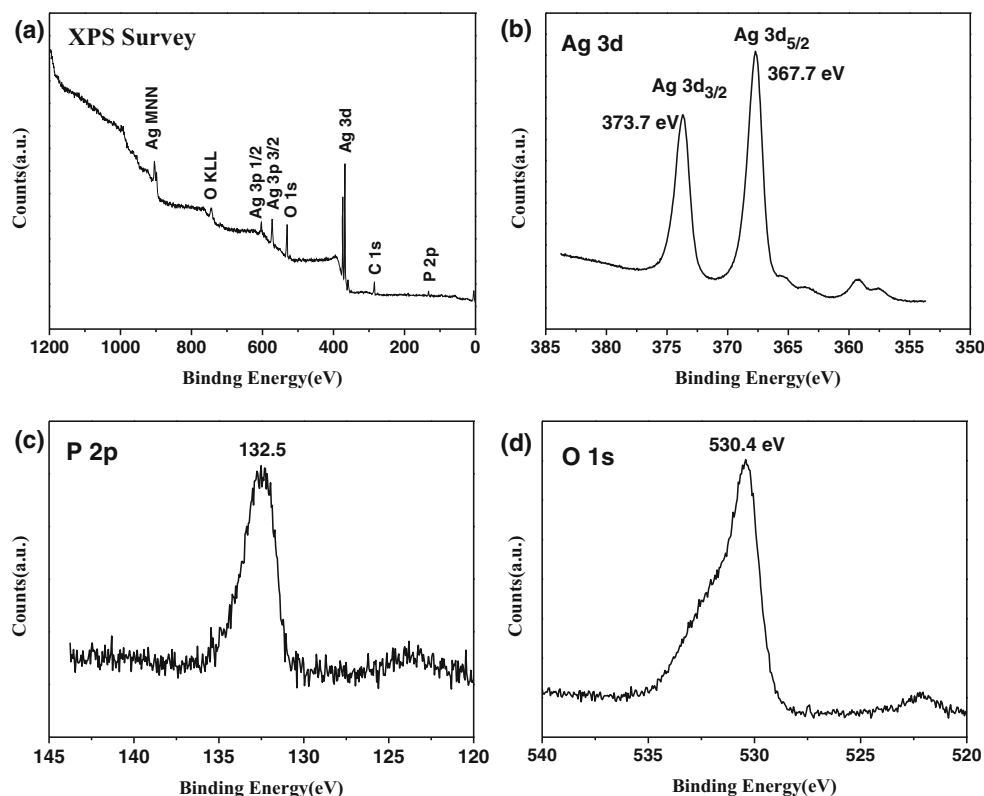
The X-ray diffraction (XRD) patterns of  $\text{Ag}_3\text{PO}_4$  samples (Fig. 2) show that all the diffraction peaks of the samples could be indexed to the body-centered cubic structure of  $\text{Ag}_3\text{PO}_4$  (JCPDS card No. 06-0505). No peaks assignable to other phases are observed, indicating a high purity of  $\text{Ag}_3\text{PO}_4$  samples. The sharp and strong diffraction peaks indicate that the  $\text{Ag}_3\text{PO}_4$  catalysts were well crystallized. In addition, the peak intensity of  $\text{Ag}_3\text{PO}_4\text{-N}$  and  $\text{Ag}_3\text{PO}_4\text{-PN}$  are stronger than those of  $\text{Ag}_3\text{PO}_4\text{-A}$  and  $\text{Ag}_3\text{PO}_4\text{-P}$ , indicating the addition of  $\text{NH}_3 \cdot \text{H}_2\text{O}$  is in favor of formation of crystals.

In order to investigate the valence states and surface chemical compositions of the  $\text{Ag}_3\text{PO}_4\text{-PN}$  microstructure, XPS analysis was employed and the results are shown in Fig. 3. Figure 3a presents an overview of the XPS spectrum of pure  $\text{Ag}_3\text{PO}_4\text{-PN}$  microstructures. Ag, P, O, and adventitious C coexist in the  $\text{Ag}_3\text{PO}_4\text{-PN}$  microstructures. Figure 3b–d display the high resolution XPS spectra of Ag, P, and O, respectively. In Fig. 3b, the peaks centered at 373.7 and 367.7 eV are ascribed to Ag  $3d_{3/2}$  and  $3d_{5/2}$ , respectively, revealing the Ag exists mainly in the +1 oxidation state in the sample. As shown in Fig. 3c, the peak of P 2p is located at 132.5 eV, indicating that the oxidation state of P is +3 in the sample [Bertrand 1981]. In addition, the spectrum of O 1s is located at 530.4 eV, which is assigned to the  $\text{O}_2^-$  in  $\text{Ag}_3\text{PO}_4$  [Katsumata et al. 2014].



**Fig. 2** XRD patterns of different  $\text{Ag}_3\text{PO}_4$  powders: **a**  $\text{Ag}_3\text{PO}_4\text{-A}$ ; **b**  $\text{Ag}_3\text{PO}_4\text{-P}$ ; **c**  $\text{Ag}_3\text{PO}_4\text{-N}$ ; **d**  $\text{Ag}_3\text{PO}_4\text{-PN}$

**Fig. 3** **a** Overview XPS spectrum of  $\text{Ag}_3\text{PO}_4$ -PN microstructure, **b–d** are the XPS spectra of  $\text{Ag}_3\text{PO}_4$ -PN microstructure for the Ag 3d spectrum, P 2p spectrum, and O 1s spectrum



*Crystal morphology*

The morphology and size of the  $\text{Ag}_3\text{PO}_4$  samples were surveyed by SEM. Figure 4a reveals the as-synthesized  $\text{Ag}_3\text{PO}_4$ -A is constituted of quasi-spherical particles with diameters around 600 nm. Figure 4b shows the quasi-tetrahedron morphology of  $\text{Ag}_3\text{PO}_4$ -P due to the addition of PVP as the structure inducer with the size in the range of 1–10  $\mu\text{m}$ .  $\text{Ag}_3\text{PO}_4$ -N has a cube structure with the size of 2–10  $\mu\text{m}$  (Fig. 4c), while  $\text{Ag}_3\text{PO}_4$ -PN exhibited a polyhedron structure with the size of 0.5–7  $\mu\text{m}$  (Fig. 4d). The formation of sharp edges and corners on the surfaces of cube and polyhedron may be benefit from the introduction of  $\text{NH}_3\cdot\text{H}_2\text{O}$  [Guo et al. 2015].

*Optical property*

The optical property of  $\text{Ag}_3\text{PO}_4$  samples was measured by UV–vis DRS, and the UV–vis spectra are exhibited in Fig. 5. The four samples are all  $\text{Ag}_3\text{PO}_4$  prepared only with different additives, so the optical absorption edges of the samples have rarely significant difference and it is largely related to electronic structure of materials. It can be clearly seen that all the  $\text{Ag}_3\text{PO}_4$  samples display photo-absorption in the UV and visible region. The absorption edges of  $\text{Ag}_3\text{PO}_4$ -A and  $\text{Ag}_3\text{PO}_4$ -PN are ca. 530 nm, while those of  $\text{Ag}_3\text{PO}_4$ -P and  $\text{Ag}_3\text{PO}_4$ -N are ca. 550 nm, which have slight shift to red. It

is noted that the absorbance of  $\text{Ag}_3\text{PO}_4$ -P sample is obviously lower than those of  $\text{Ag}_3\text{PO}_4$ -A,  $\text{Ag}_3\text{PO}_4$ -N, and  $\text{Ag}_3\text{PO}_4$ -PN samples in the range of 330–500 nm, which may adverse to degradation performance of  $\text{Ag}_3\text{PO}_4$ -P sample.

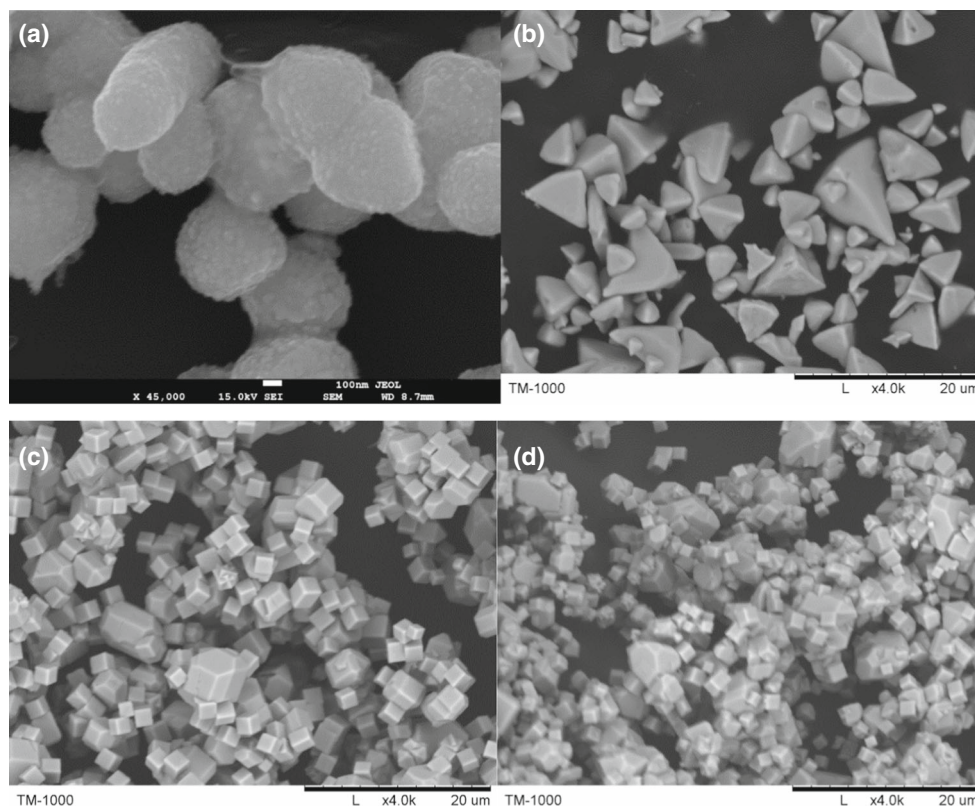
**Photocatalytic activity**

*Comparison of degradation activity of four  $\text{Ag}_3\text{PO}_4$  catalysts*

The photocatalytic activities of the four  $\text{Ag}_3\text{PO}_4$  samples were tested by photodegradation of TC under visible-light irradiation and the results are shown in Fig. 6a. The  $\text{Ag}_3\text{PO}_4$ -PN exhibited the highest degradation efficiency, arriving 91 % after 120 min of visible-light irradiation. The degradation efficiencies of TC by  $\text{Ag}_3\text{PO}_4$ -A,  $\text{Ag}_3\text{PO}_4$ -P, and  $\text{Ag}_3\text{PO}_4$ -N are 78, 72, and 85 %, respectively. The degradation of TC by  $\text{Ag}_3\text{PO}_4$  under visible-light irradiation followed pseudo-second-order kinetics and the reactive rate constants of four  $\text{Ag}_3\text{PO}_4$  samples are 0.0006, 0.0009, 0.0013, 0.0023  $\text{L mg}^{-1} \text{min}^{-1}$ , respectively, which are shown in Fig. 6b. The higher degradation efficiencies of  $\text{Ag}_3\text{PO}_4$ -N and  $\text{Ag}_3\text{PO}_4$ -PN are attributed to generation of smooth surface, sharp edges and corners, which might contribute to the photocatalysis process at the surface owing to [100] plane with high surface energies [Bi et al. 2011; Guo et al. 2015]. The enhanced photocatalytic activity of  $\text{Ag}_3\text{PO}_4$ -PN benefits from the synergistic effect of



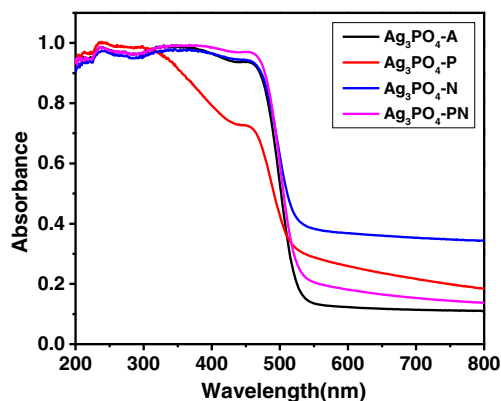
**Fig. 4** SEM of the obtained  $\text{Ag}_3\text{PO}_4$  samples: **a**  $\text{Ag}_3\text{PO}_4\text{-A}$ ; **b**  $\text{Ag}_3\text{PO}_4\text{-P}$ ; **c**  $\text{Ag}_3\text{PO}_4\text{-N}$ ; **d**  $\text{Ag}_3\text{PO}_4\text{-PN}$



PVP and  $\text{NH}_3\cdot\text{H}_2\text{O}$  which not only inhibits the crystal aggregation but also facilitates the order array of crystals. Besides, the strong visible-light absorption capacity and appropriate particle sizes of  $\text{Ag}_3\text{PO}_4\text{-PN}$  are also conducive to the improvement of photocatalytic activity. The lowest degradation efficiency of  $\text{Ag}_3\text{PO}_4\text{-P}$  is by reason of low absorbance in visible region, which is consistent with results in Fig. 5.

#### Effect of the reaction conditions

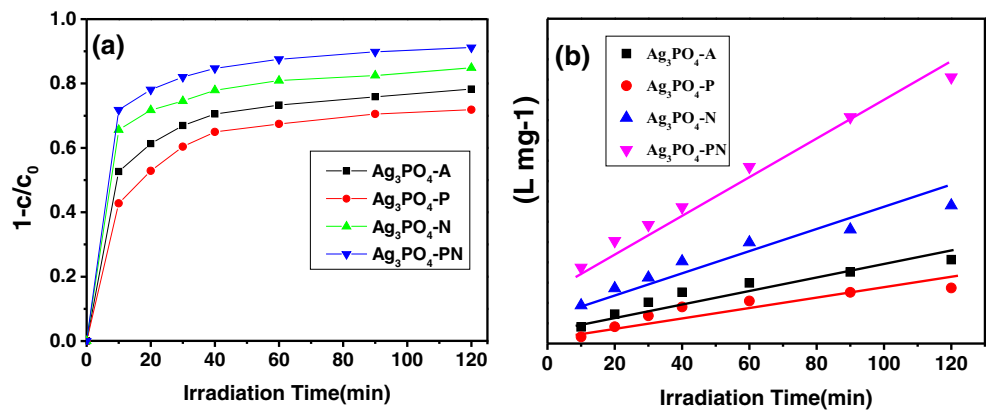
Since the enhanced photocatalytic activity,  $\text{Ag}_3\text{PO}_4\text{-PN}$  was selected to degrade contamination in the later experiments. To



**Fig. 5** UV-vis DRS of prepared  $\text{Ag}_3\text{PO}_4$  samples

investigate the effects of various reaction conditions on the photocatalytic efficiency of tetracycline hydrochloride (TC), three reaction conditions were researched, including catalyst dosage, TC concentration, and pH of TC solution. The catalyst dosage was ranged from 0.25 to 3.0 g/L. The result shows that the degradation efficiency of TC increased with the increase of catalyst dosage, while it barely continuously rose until catalyst dosage increasing to 2 g/L (Fig. 7a). It has been reported that in a certain dosage range, radicals increases with the increase of catalyst dosage owing to the increase of the generated carriers [Bertelli and Selli 2004]. But the scatter reflection will result in a low utilization of irradiation light because of an over dosage [Wang et al. 2014b]. To take both into account, the dosage of 2 g/L is used in the later experiments. The effect of TC concentration on the photocatalytic efficiency of TC was investigated (Fig. 7b) with the range of 20–60 mg/L. The results exhibited that the degradation efficiency of TC by  $\text{Ag}_3\text{PO}_4\text{-PN}$  increased with the solution concentration up to 30 mg/L, and then decreased with the increase of TC concentration. The effect of pH of TC solution on the photocatalytic efficiency of TC was also investigated (Fig. 7c). The result indicates that the photocatalytic efficiencies of TC under weak acid and neutral conditions were higher than those under strong acid and alkaline conditions, which may because of instability of  $\text{Ag}_3\text{PO}_4$  photocatalyst under strong acid and alkaline conditions. From above conclusion, the degradation efficiency of TC by  $\text{Ag}_3\text{PO}_4\text{-PN}$  was highest

**Fig. 6** Photocatalytic degradation efficiency (a) and degradation kinetic fitting (b) of different  $\text{Ag}_3\text{PO}_4$  samples under solar irradiation with the  $2.0 \text{ g L}^{-1}$  of catalyst concentration and  $30 \text{ mg L}^{-1}$  of initial concentration of TC



with  $2 \text{ g/L}$  of catalyst dosage,  $30 \text{ mg/L}$  of TC concentration, and  $5$  of pH of TC solution.

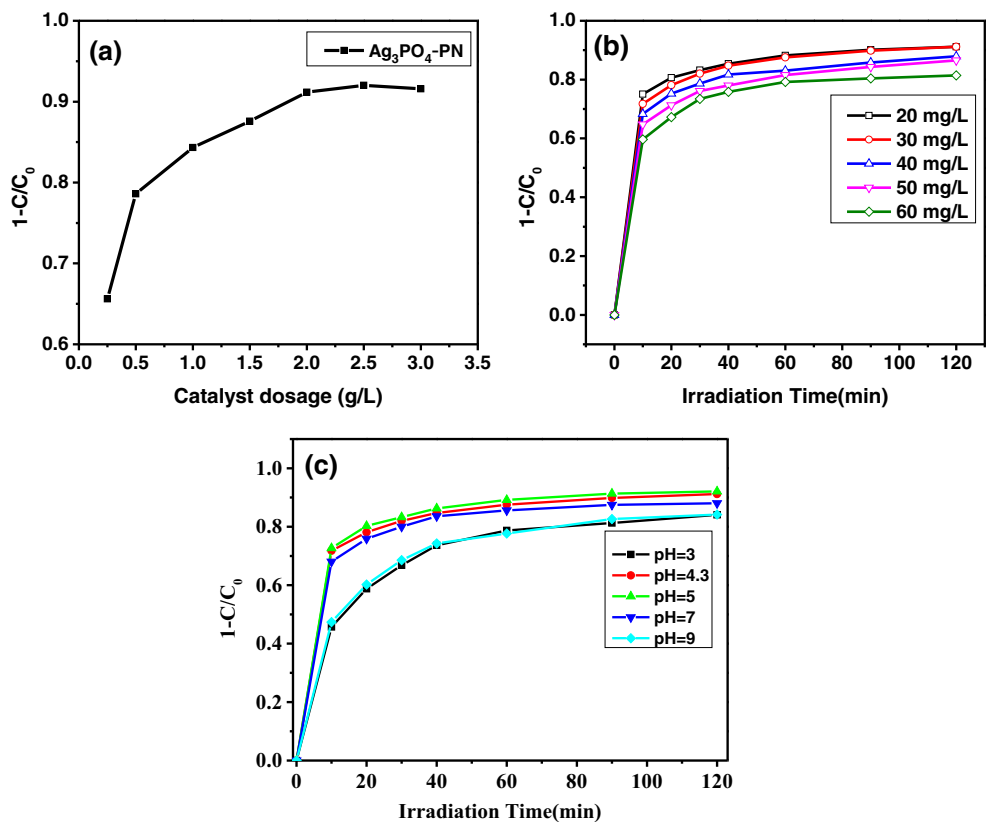
**Photocatalytic degradation mechanism**

*TOC and IC*

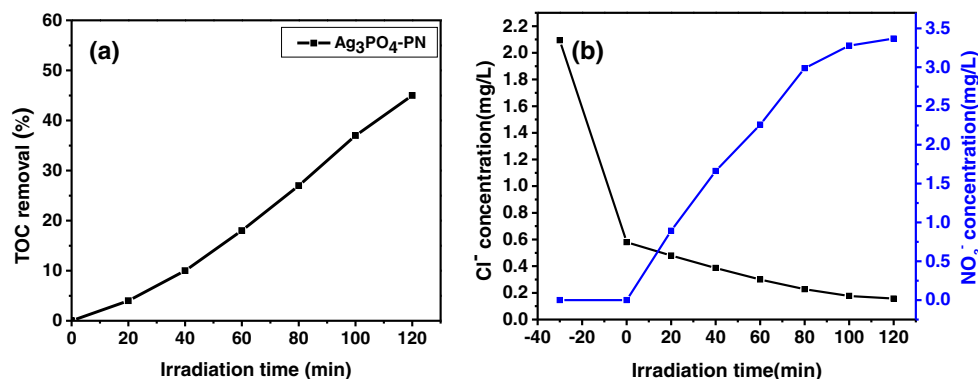
Total organic carbon (TOC) was frequently applied as an important index to represent the concentration of organic matter in contamination solution [Sakthivel et al. 2003]. Herein, the mineralization of TC is evaluated by TOC content and quantitative determination of inorganic ions by  $\text{Ag}_3\text{PO}_4\text{-PN}$  as the

photocatalyst. Figure 8a shows TOC removal increases with increase in irradiation time and arrived *ca.* 45 % through consecutive irradiation of visible light at 120 min. Moreover, due to the presence of nitrogen and chlorine elements, the detection of inorganic ions was also used to assess the mineralization degree of TC. The quantity measurements of  $\text{NO}_3^-$  and  $\text{Cl}^-$  were carried out by IC during the photodegradation process (Fig. 8b). The  $\text{Cl}^-$  concentration of TC solution rapidly decreases during dark adsorption process, which may be reason of the generation of  $\text{AgCl}$ . With the increase of irradiation time, the  $\text{Cl}^-$  concentration remains drop, owing to the catalytic oxidation. In addition, the  $\text{NO}_3^-$  concentration showed

**Fig. 7** Effects of the  $\text{Ag}_3\text{PO}_4\text{-PN}$  dosage (a), TC concentration (b) and pH of TC solution (c) on the degradation efficiencies of TC under visible-light irradiation, with irradiation for 120 min



**Fig. 8** TOC removal (a) and  $\text{Cl}^-$  and  $\text{NO}_3^-$  concentration (b) of TC solution vs. irradiation time with TC concentration of 30 mg/L and  $\text{Ag}_3\text{PO}_4\text{-PN}$  dosage of 2 g/L



a gradually increased tendency with the illumination irradiation, indicating the nitrogen of TC molecular was partly oxidized into  $\text{NO}_3^-$ . Previous studies [Bandara et al. 1997] have demonstrated the conversion of intermediate nitrite partly into ammonia and nitrate. The formation of  $\text{NH}_3$  which escapes may account for the loss of nitrogen in the mass balance.

#### Oxidative radical species

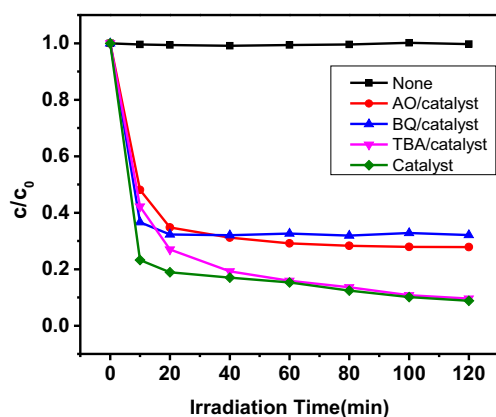
To investigate the photocatalytic mechanism of  $\text{Ag}_3\text{PO}_4\text{-PN}$  into more detail, the effects of active species scavengers on the degradation of TC were examined to shed light on the predominant reactive oxygen species in the photocatalytic process (Fig. 9). In this study, AO, BQ, and TBA were added into the reaction solution as  $\text{h}^+$ ,  $\cdot\text{O}_2^-$ , and  $\cdot\text{OH}$  scavenger, respectively [Lin et al. 2012]. Without the addition of the scavengers, the photocatalytic efficiency of TC is 91 % after 120 min of visible-light irradiation. As is clear from Fig. 9, the addition of TBA scarcely affected the degradation efficiency of TC by  $\text{Ag}_3\text{PO}_4\text{-PN}$ , which indicates that  $\cdot\text{OH}$  does not appear to be the main reactive oxygen species in the photocatalytic process of TC by  $\text{Ag}_3\text{PO}_4\text{-PN}$ .

The visible-light photocatalytic degradation of TC was significantly inhibited by BQ with 23 % of reduction of

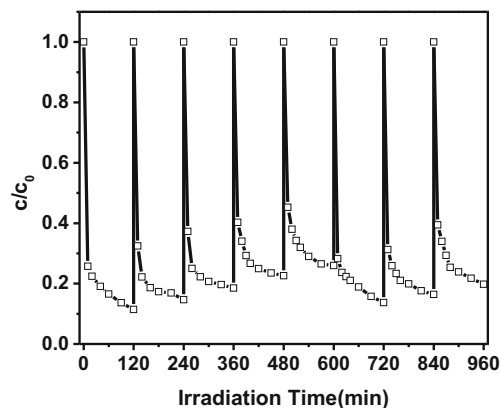
degradation efficiency, confirming that  $\cdot\text{O}_2^-$  indeed plays an important role in the degradation of TC by  $\text{Ag}_3\text{PO}_4\text{-PN}$ . Besides, the addition of AO likewise suppresses the photocatalytic degradation of TC with 19 % of decrease of degradation efficiency. It is, therefore, most likely that the photoinduced holes, with a high oxidation potential of +2.45 V vs. NHE, directly react with organic compounds [Yi et al. 2010]. Therefore, it may be concluded that  $\text{h}^+$  and  $\cdot\text{O}_2^-$  are major active species in the photodegradation of TC by  $\text{Ag}_3\text{PO}_4\text{-PN}$ .

#### Photocatalyst stability and regeneration

The photocatalyst stability was assessed by the cyclic photocatalytic reaction experiments of TC by  $\text{Ag}_3\text{PO}_4\text{-PN}$ . The cycle experiments were carried out at the same conditions as usual degradation tests and  $\text{Ag}_3\text{PO}_4\text{-PN}$  catalyst was washed twice by distilled water and dried before reused. As shown in Fig. 10, the photodegradation activity of  $\text{Ag}_3\text{PO}_4\text{-PN}$  was available held after five times loop experiments, indicating  $\text{Ag}_3\text{PO}_4\text{-PN}$  has relatively high stability under visible-light irradiation. Five degradation efficiencies of repeated experiments were 89, 85, 82, 77, and 74 %, respectively. Nevertheless, it was found that  $\text{Ag}_3\text{PO}_4$  materials always



**Fig. 9** Effect of various scavengers (AO, BQ and TBA) on the degradation of TC by  $\text{Ag}_3\text{PO}_4\text{-PN}$ , with TC concentration of 30 mg/L and catalyst dosage of 2 g/L



**Fig. 10** Five times recycling runs of  $\text{Ag}_3\text{PO}_4\text{-PN}$  and three times degradation experiment of regenerative  $\text{Ag}_3\text{PO}_4\text{-PN}$  for the degradation of TC under visible-light irradiation. Each run of photocatalytic reaction lasted for 120 min (catalyst dosage = 1.0 g/L, TC concentration = 20 mg/L, pH = 4.3)

possessed photo-corrosion characteristics, and reductive elemental silver wrapped around the surface of  $\text{Ag}_3\text{PO}_4$  catalyst may have a negative effect on the photo-absorption performance and photocatalytic activity of  $\text{Ag}_3\text{PO}_4$ , which prevented practical application of  $\text{Ag}_3\text{PO}_4$  as a recyclable high-efficient photocatalyst [Yi et al. 2010].

A facile wet chemical-oxidation method was utilized to rejuvenate  $\text{Ag}_3\text{PO}_4$ -PN from weakly active Ag, which was previously reported by Wang and his coauthors. During the rejuvenation process, weakly active Ag can be oxidized by dropwise adding  $\text{H}_2\text{O}_2$  into saturated  $\text{PO}_4^{3-}$  solution. While the acid environment of  $\text{NaH}_2\text{PO}_4$  and  $\text{H}_3\text{PO}_4$  aqueous solution can dissolve the  $\text{Ag}_3\text{PO}_4$ , and the fast decomposition of  $\text{H}_2\text{O}_2$  will occur under strong alkaline condition of  $\text{Na}_3\text{PO}_4$  solution, weaker alkaline  $\text{Na}_2\text{HPO}_4$  was chosen as the source for  $\text{PO}_4^{3-}$  ions [Wang et al. 2012]. The photodegradation efficiency of TC by rejuvenated  $\text{Ag}_3\text{PO}_4$ -PN during three times recycling runs were 86, 84, and 80 %, respectively, indicating a successful rejuvenation of  $\text{Ag}_3\text{PO}_4$ , which greatly promotes the development of  $\text{Ag}_3\text{PO}_4$  for practical applications.

### Conclusion

A facile, environmental-friendly  $\text{Ag}_3\text{PO}_4$ -PN photocatalyst was successfully synthesized by a simple precipitation method at room temperature with the presence of  $\text{NH}_3\cdot\text{H}_2\text{O}$  and PVP. The  $\text{Ag}_3\text{PO}_4$ -PN exhibits a superior photocatalytic activity compared to those of  $\text{Ag}_3\text{PO}_4$ -A,  $\text{Ag}_3\text{PO}_4$ -P, and  $\text{Ag}_3\text{PO}_4$ -N under visible-light irradiation. The enhancement of photocatalytic efficiency of  $\text{Ag}_3\text{PO}_4$ -PN is strongly dependent on the excellent photo-absorption capacity, sharp edges and corners, and synergistic effect of PVP and  $\text{NH}_3\cdot\text{H}_2\text{O}$ . The effect researches of reaction conditions indicated that the optimum condition were 2.0 g/L of catalyst dosage, 30 mg/L of TC solution concentration, and 5 of pH value of TC solution. Radical detection experiments indicate the  $\text{h}^+$  and  $\cdot\text{O}^{2-}$  are major active species in the degradation of TC by  $\text{Ag}_3\text{PO}_4$ -PN. TOC test and IC analysis indicate TC molecular was partially mineralized under the visible-light irradiation by  $\text{Ag}_3\text{PO}_4$ -PN. Moreover, photocatalyst stability and regeneration experiments exhibit a favorable stability and rejuvenation ability, suggesting a promising prospect of practical application of  $\text{Ag}_3\text{PO}_4$  in the wastewater treatment.

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