RESEARCH ARTICLE

Co-remediation of cadmium-polluted soil using stainless steel slag and ammonium humate

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Abstract

Introduction and methods This study investigated the remediation of cadmium-polluted soil using a combination of stainless steel slag and ammonium humate. These remedial agents were added to an artificially polluted garden soil to inhabit cadmium toxicity in soil by changing the physical and chemical properties of soil in a pot experiment.

Results and conclusions The results showed that the coapplication of ammonium humate and stainless steel slag significantly decreased the total and available soil cadmium concentrations, with maximum decreases of 16.30% and 58.04%, respectively. The co-application of an adequate dose of these remedial agents can significantly increase soil pH. The soil organic matter and cation exchange capacity, as well as the amount of soil aggregates, were also significantly increased by the addition ammonium humate, but not stainless steel slag.

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1 Introduction

Cadmium is a high-toxicity soil pollutant that originates mainly from mining and smelting activities (Tang et al. 2006). The chemical properties of Cd are similar to zinc, and both are stable, not readily eliminated from the environment and accumulate in soil (Gupta et al. 2004). Cadmium is absorbed from soil by crop roots, transferred to seeds, and may enter the human body through the food chain, potentially threatening human health. Cadmium tends to accumulate in the liver, kidney, pancreas, thyroid, and bones, causing kidney lesions, nerve pain, endocrine disorders, and other illnesses (Hrudey et al. 1995). The earliest example of Cd poisoning is the "pain disease event" in Toyama Prefecture, Japan, in the 1960s, where Cd containing wastewater was discharged from a lead and zinc smelter into rice paddies (Gupta and Sharma 2003; Gupta and Ali 2004). The local residents routinely consumed the contaminated rice and drinking water, and exhibited severe symptoms of Cd poisoning. Cadmium damages victims' skeletal systems, and bones become brittle and easily broken, causing dozens of people to die from the pain (Moreno et al. 2002). Soil microorganisms and enzyme activities are also affected by toxicity caused by high levels of soil cadmium (Karaca et al. 2002; Moreno et al. 2006). A number of physicochemical and biological remediation methods have been developed for reclaiming metal-contaminated soils, including soil washing, soil flushing, phytoremediation, and electrokinetics (Giannis et al. 2007; Gupta et al. 2007a,b).

Ammonium humate is used as a humic acid fertilizer. This complex organic substance includes various functional groups such as carboxyl, carbonyl, amino, aldehyde, phenolic hydroxyl, and sulfur-containing groups (Pehlivan and Arslan 2006). These groups can undergo complex reactions with heavy metals, altering the metal ions' mobility, migration, adsorption, and bioavailability (Gupta and Rastogi 2008a, b, c). The ammonium nitrogen present in ammonium humate also improves soil fertility by replenishing nitrogen (Gupta et al. 1999, 2000, 2010).

Stainless steel slag is generated in large quantities during the steel refining process and is commonly used in road paving, as a building material and for cement production. Stainless steel slag is rich in Ca, Mg, Si, Fe, and other elements and has been used for soil improvement and as a fertilizer in agriculture (Wang and Cai 2006). Stainless steel slag is also potentially useful for remediation of heavy metal pollution due to its porosity, large surface area, and good adsorptivity. Some studies have shown that stainless steel slag is a good adsorbent of heavy metals, mostly via precipitation and adsorption of the metal oxide on the slag surface (Gupta et al. 2006; Kim et al. 2008). Granulated steel slag has also been tested and found to be effective at adsorbing dyes and metal ions (Liu et al. 2010). Heavy metal leaching experiments have confirmed the safety of stainless steel slag in the environment, so environmental pollution control is an effective use for stainless steel slag.

Various amendment materials have been used to inhibit the absorption of cadmium by plants, including zeolite (Li et al. 2009), calcite (Thakur et al. 2006), humic acid (Ghabbour et al. 2006; Gupta et al. 2001), and sewage sludge (Moreno et al. 2003). However, the combined effect of stainless steel slag and ammonium humate to remediate heavy-metal-contaminated soil has not been investigated. The objective of this study was to evaluate the effectiveness of co-remediation of Cdpolluted soil with stainless steel slag and ammonium humate.

2 Materials and methods

2.1 Stainless steel slag and ammonium humate properties

The stainless steel slag was obtained from Taiyuan Iron and Steel (Group) Co., Ltd., Taiyuan, Shanxi. It was crushed to <0.178 mm for chemical analysis and used in the remediation trials. The stainless steel slag was alkaline (pH=13.96), with high concentrations of CaO (51.84%), FeO (6.17%), Al₂O₃ (4.39%), MgO (9.74%), MnO (0.29%), S (0.18%), SiO₂ (28.2%), and TiO₂ (0.35%); the apparent density, porosity, surface area, and most probable pore size of stainless steel slag were 2.88 g cm⁻¹, 16.20%, 2.61 m² g⁻¹, and 1.72 μ m⁻¹, which were analyzed by Mercury Porosimeter (PoreMaster-60).

The ammonium humate is a traditional chemical organic fertilizer that was obtained from Yongfeng chemical fertilizer

plant, Taiyuan, Shanxi. That was sieved to <0.02 mm and alkaline (9.21). The effective components are humic acid (52.22%) and total N (5.60%).

2.2 Soil properties

The soil pH was determined potentiometrically in water. The organic matter was determined using Tiurin's method with dichromate oxidation. The soil cation exchange capacity (CEC) was determined using the NH₄Cl–NH₄COOH method; and the amount of soil aggregates was measured by the dry sieving method.

2.3 Pot experiments

An uncontaminated loess-derived, calcareous cinnamon soil was obtained, finely ground, and then sieved to <2 mm. The soil properties are shown in Table 1.

The soil (250 g) was placed in polyethylene pots and mixed with $3CdSO_4 \cdot 8H_2O$ to give a soil Cd concentration of 150 mg kg⁻¹. Various amounts (0, 100, 200, and 300 mg g⁻¹) of both the stainless steel slag and ammonium humate were blended into the treated soil, as shown in Table 2. Each treatment was performed in triplicate and incubated for 60 days. The moisture content of each pot was monitored by weighing the pot every 3 days. The experiment was laid out in a random block design.

2.4 Analysis of total and available Cd in soils

For the available Cd analysis, an extractant, consisting of 5 mmol L^{-1} DTPA, 10 mmol L^{-1} CaCl₂, and 0.1 mol L^{-1} TEA (triethanolamine), was prepared in deionized water, and the pH was adjusted to 7.3 with 1:1 HCl. Triplicate extractions were carried out using 5 g soil and 25 mL of extractant in 60 mL HDPE sample bottles. The samples were shaken for 2 h (120 cycles/min) in a water bath at 20°C after which the extract was filtered (Whatman no.42 filer paper) and the supernatant was acidified for analysis prior to analysis by atomic adsorption spectrophotometer (Shimadzu-AA6800).

Tabl	e	1	Soil	properties
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Properties	
pH(H ₂ O)	8.70
Organic matter (%)	0.615
$CEC \pmod{kg^{-1}}$	7.16
Total N (%)	0.03
Readily available P (mg kg ⁻¹)	6.25
Readily available K (mg kg ⁻¹)	113.8
Total Cd (mg kg ⁻¹)	0.50

0

100

200

300

100

200

300

0

Table 2 Experimental treatments (uniform design U16 (4^2)) Ammonium humate dose (g kg^{-1}) Stainless steel slag dose ($g kg^{-1}$) 0 0 100 0 0 200 0 300 100 0 100 100 200 100 300 100

200

200

200

200

300

300

300

300

For the total Cd analysis, soil samples were digested in an HNO_3 -HF-HClO₄-HCl mixture. The total Cd concentration in the extract was determined by atomic absorption spectrophotometer (Shimadzu-AA6800).

2.5 Statistical analysis

A one-way variance analysis (ANOVA) was performed on all data, with different concentrations of Cd in soils for different amounts of stainless steel slag and ammonium humate determined using Duncan's test. Statistical significance was defined as p < 0.05.

3 Results

3.1 Total cadmium in soils

The results shown in Table 3 indicate that increasing the amounts of stainless steel slag and ammonium humate led to

significant decreases in the total soil cadmium concentration (p < 0.05). For each dose of ammonium humate, the increasing amounts of stainless steel slag continually led to significantly reduced total soil cadmium concentrations (p < 0.05). When the amount of ammonium humate was 0 and 100 g kg⁻¹, increasing the amount of stainless steel slag from 100 to 200 and 300 g kg⁻¹, significantly decreased the total soil cadmium concentration (p < 0.05). Similarly, when the amount of ammonium humate was 200 and 300 g kg⁻¹, increasing the amount of stainless steel slag from 200 to 300 g kg⁻¹, also led to significant decreases in the total soil cadmium concentration (p < 0.05).

With different amounts of stainless steel slag, as the amount of ammonium humate was increased from 100 to 200 g kg⁻¹, the total soil cadmium concentration decreased significantly (p < 0.05).

Overall, the addition of stainless steel slag and ammonium humate resulted in removal rates for total cadmium from 0.93% to 16.30%. The maximum total cadmium removal rate was achieved when the amounts of both stainless steel slag and ammonium humate were 300 g kg⁻¹.

3.2 Available cadmium in soils

Table 4 shows that, when the amounts of both stainless steel slag and ammonium humate increased, the available soil cadmium concentrations decreased significantly (p<0.05). For each ammonium humate dose, increasing the amount of stainless steel slag resulted in significant decreases in the available cadmium concentrations (p<0.05). When the amount of ammonium humate was 0 and 100 g kg⁻¹, increasing the stainless steel slag dose significantly reduced the available soil cadmium concentration (p<0.05). When the amount of ammonium humate was 200 and 300 g kg⁻¹, increasing the stainless steel slag dose from 0 to 100 g kg⁻¹ significantly decreased the available soil cadmium (p< 0.05); but increasing the stainless steel slag dose from 100 to 200 g kg⁻¹ did not lead to a significant decrease in available soil cadmium concentration (p>0.05).

For each amount of added stainless steel slag, as the ammonium humate dose increased, the decreases in the available soil

Table 3 Effect of stainless steel slag and ammonium humate doses on total Cd in soil

Stainless steel slag dose (g kg ⁻¹) ^a	Ammonium humate dose (g kg ⁻¹)					
	0	100	200	300		
0	133.03±1.01aA	131.79±1.49aA	124.42±1.59aB	116.20±1.04aC		
100	131.04±1.02aA	127.62±1.04bB	117.19±1.23bC	114.94±1.33aC		
200	124.49±0.95bA	121.83±1.20cA	117.20±1.23bB	114.48±1.22aB		
300	117.90±0.91cA	116.11±0.80dA	112.15±1.58cB	111.34±1.71bB		

^a Differences between treatments was tested by one-way ANOVA; data on the same row with different capital letters and in the same column with different lowercase letters are significantly different at p < 0.05 (Duncan's method)

Stainless steel slag dose (g kg ⁻¹) ^a	Ammonium humate dose (g kg ⁻¹)					
	0	100	200	300		
0	32.46±0.11aA	32.38±0.73aA	32.26±0.54aA	31.77±0.07aA		
100	22.43±0.93bA	21.82±0.56bA	23.17±0.64bA	22.95±0.72bA		
200	16.39±0.11cA	17.30±0.61cA	$23.21{\pm}0.93bB$	22.82±0.64bB		
300	13.89±0.30dA	13.62±0.44dA	16.11±0.33cB	20.96±1.15bC		

 Table 4
 Effect of stainless steel slag and ammonium humate doses on available Cd in soil

^a Differences between treatments were tested by one-way ANOVA; data on the same row with different capital letters and in the same column with different lowercase letters were significantly different at p<0.05 (Duncan's method)

cadmium concentrations were insignificant (p>0.05). When the amount of stainless steel slag was 0 and 100 g kg⁻¹, increasing the ammonium humate dose did not significantly reduce the available soil cadmium concentration (p>0.05). Furthermore, when the amount of stainless steel slag was 200 and 300 g kg⁻¹, increasing the ammonium humate dose from 100 to 200 g kg⁻¹ significantly increased the available soil cadmium concentration (p<0.05).

Overall, the treatments using different doses of stainless steel slag and ammonium humate removed 0.25% to 58.04% of the available soil cadmium. The maximum removal of available cadmium occurred with a stainless steel slag dose of 300 g kg⁻¹ and ammonium humate dose of 100 g kg⁻¹.

3.3 Soil physical and chemical properties

Figure 1 shows that, as both the stainless steel slag and ammonium humate doses increased, the soil pH also increased significantly (p<0.05). For all ammonium humate doses, increasing the amount of stainless steel slag significantly increased the soil pH (p<0.05). For all stainless steel slag doses, increasing the amount of ammonium humate from 0 to 200 g kg⁻¹ increased the soil pH significantly (p<0.05), but increasing the ammonium humate dose from 200 to 300 g kg⁻¹ led to a slight decrease in



Fig. 1 Effect of different stainless steel slag and ammonium humate doses on soil pH

soil pH. Overall, with different stainless steel slag and ammonium humate doses, the maximum increase in the soil pH was 34.94%, which occurred when the stainless steel slag dose was 300 g kg^{-1} , and the ammonium humate dose was 200 g kg^{-1} .

Figure 2 shows that the addition of stainless steel slag along did not significantly increase the soil organic matter content (p>0.05). However, the combined application of stainless steel slag and ammonium humate significantly increased the soil organic matter content (p<0.05). When the ammonium humate dose increased from 0 to 100 g kg⁻¹, the increase in the soil organic matter was 129.12%.

Figure 3 shows that the addition of stainless steel slag alone did not increase the CEC in soils (p>0.05), but the addition of ammonium humate at increasing doses significantly enhanced the CEC (p<0.05). When the ammonium humate dose increased from 0 to 100 g kg⁻¹, the soil CEC increased by 58.84%.

Figure 4 shows that the addition of 100 g kg⁻¹ of stainless steel slag significantly decreased the amount of soil aggregates (p < 0.05) compared with the control. Increasing the stainless steel slag dose from 100 to 300 g kg⁻¹ did not to reduce the amount of soil aggregates further (p > 0.05). In contrast, the addition of ammonium humate at increasing doses increased the amount of soil aggregates (p < 0.05). The maximum increase in soil aggregates was 42.42%, when the stainless steel slag dose was 0 g kg⁻¹ and the ammonium humate dose was 300 g kg⁻¹.



Fig. 2 Effect of different stainless steel slag and ammonium humate doses on soil organic matter



Fig. 3 Effect of different stainless steel slag and ammonium humate doses on CEC

4 Discussion

4.1 Effect on total Cd and available Cd in soils

Tables 3 and 4 indicate that treatment with stainless steel slag and ammonium humate led to maximum decreases in the total and available soil cadmium concentrations of 16.30% and 58.04%, respectively. These results are lower than the decreases reported by Deng et al. (2011), which may be related to the shorter incubation time used in their study, as a longer time may be necessary to remove cadmium to low concentrations.

Some researchers indicated soil pH and CEC were the dominant factors for decreasing availability of heavy metal (Shi et al. 2009a, b). Some studies showed the available soil cadmium concentration was significantly affected by the soil pH (Li et al. 2008) and found some chemical amendments could decrease soil Cd availability through rising soil pH (Liang et al. 2005). Others suggested remediation of soils was also affected by the CEC (Castaldi et al. 2005).

The removal of both total and available cadmium by stainless steel slag may be attributed to the increased high pH (13.96), porosity (14.50%), and large surface area (2.31 m² g⁻¹) that are necessary for satisfactory adsorption (Gupta 1998; Ali and Gupta 2006; Gupta and Rastogi 2009). In this study, Figs. 1 and 3 indicated stainless steel slag addition significantly increased soil pH, then decrease total and available cadmium.

Some scientists found organic amendment could decrease the bioavailability of heavy metals in soil (Tordoff and Baker 2000; Shi et al. 2009a; Gupta et al. 2009), thus



Fig. 4 Effect of different stainless steel slag and ammonium humate doses on the amount of soil aggregates

permitting the re-establishment of vegetation at contaminated sites. This relies on the ability of the humic substance to re-distribute heavy metals from available form to nonavailable ones (McGrath and Cegarra 1992; Gupta et al. 1998, 2004, 2007c). In this study, ammonium humate effectively decreased total cadmium in soil in this study. Addition of ammonium humate significantly increase soil CEC due to its effective components humic acid (52.22%) and total N (5.60%), which contains numerous active functional groups, including carbonyl, carboxyl, hydroxyl, and phenolic hydroxyl groups, many of which can form complex reactions with cadmium (Wang et al. 2005; Chen and Wang 2006). The reduction in total soil Cd most resulted from the formation of Cd complexes with ammonium humate. Ammonium humate is alkaline; the addition can increase soil alkaline (Fig. 1) and so decrease total cadmium in soil.

But ammonium humate could not effect on the soil available cadmium almost when the stainless steel slag dose was low $(<100 \text{ g kg}^{-1})$. And when the amount of stainless steel slag was 200 or 300 g kg⁻¹, increasing dose of the ammonium humate led to significant increase of available cadmium in soil. That because: (1) when the stainless steel dose was low, the preferential adsorption of Cd on the stainless steel slag, as the stainless steel slag has a greater capacity for Cd adsorption than ammonium humate (Gupta et al. 1997); (2) when the stainless steel dose was high, ammonium humate as a competitive adsorbent lead to the desorption of some Cd ions from the surface of the stainless steel slag, with some of the Cd ions remaining in the soil, elevating the available soil cadmium (Srivastava et al. 1997); (3) The addition of ammonium humate could increase water-soluble fraction of heavy metal in soil due to bioavailability of organic matter (Shi et al. 2009a).

4.2 Effect on soil physical and chemical properties

Figure 1 shows the increasing soil pH with the addition of stainless steel slag, which is similar to the results of study in which silicate induced pH rise in the soils (Liang et al. 2005). The soil pH increased at lower ammonium humate doses, but decreased at higher doses (200 and 300 g kg⁻¹). The maximum soil pH was found when the ammonium humate dose was 200 g kg⁻¹. The ammonium humate used in this study had a pH of 9.72 and increased soil pH on its own. Humic substances are a heterogeneous mixture of polyacidic compounds with a considerable buffering capacity over a wide pH range, due to the dissociation of the acidic functional groups (Ceppi et al. 1999). The buffering capacity of soil is of considerable practical importance in most processes that control the optimum supply of nutrients to crops, most of which act within a narrowly defined pH range (García-Gil et al. 2004). In this study, increasing the ammonium humate dose enhanced the soil buffering capacity, which explains slight decrease in soil pH at high ammonium humate doses.

Figures 2, 3, and 4 indicate that the addition of stainless steel slag had no effect on soil organic matter and CEC, and decreased the amount of soil aggregates. The stainless steel slag could not organically fuse with soil and influence on all aspects of soil physical and chemical properties like organic or mineral addition, but that can play an independent assistant, which is not fused with soil in essence, for remediating Cd-polluted soil utilizing high pH, porosity, and large surface area (Wang and Cai 2006; Shi et al. 2009a; Gupta et al. 2009).

In contrast, the addition of ammonium humate increased the soil organic matter, CEC, and amount of soil aggregates. It is well known that humic substances (HA) contribute up to 85-90% of soil organic matter and can improve soil properties such as aggregation, water holding capacity, microbial growth, organic matter mineralization, and the solubility and availability of micronutrients (e.g., Fe, Zn, and Mn) (Chen and Aviad 1990; Ayuso et al. 1996; Sharif et al. 2002). Consequently, HA is utilized in agriculture as a fertilizer, plant growth promoter, nutrient carrier, and soil conditioner (Nisar and Mir 1989). Humic substances can integrate nitrogen into their structure either directly, through chemical reactions, or indirectly, through microbial activities and the subsequent decomposition of microbial biomass (Clinton et al. 1995). Several reports indicate that NH₄⁺–N could be fixed to soil organic matter and that increasing the amount of nitrogen in soil facilitates the conversion of organic matter.

5 Conclusions

Both ammonium humate and stainless steel slag could be applied individually for the remediation of heavy-metalcontaminated soils; due to their physical and chemical properties, each material only contributes to some aspects of the soil remediation process. Individually, neither material meets all the requirements for soil remediation.

In this study, we found that the co-application of ammonium humate and stainless steel slag not only significantly decreased the total and available soil cadmium concentrations but also improved the soil physical and chemical properties. In particular, the co-application of these materials increased soil pH, organic matter, and CEC and promoted the formation of soil aggregates. This study has proven the feasibility of these materials for the remediation of Cd-polluted soil.

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