RESEARCH ARTICLE

Fine and ultrafine particles emitted from laser printers as indoor air contaminants in German offices

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Abstract

Purpose Various publications indicate that the operation of laser printers and photocopiers may be associated with health effects due to the release of gaseous components and fine and ultrafine particles (UFP). However, only sparse studies are available that evaluate the possible exposure of office workers to printer emissions under real conditions. Therefore, the aim of our study was to assess the exposure of office workers to particulate matter released from laser printers and photocopiers.

Methods Concentrations of fine particles and UFP were measured before, during, and after the operation of laser printing devices in 63 office rooms throughout Germany. Additionally, the particles were characterized by electron microscopy and energy-dispersive X-ray spectroscopy.

Results A significant increase of fine particles and UFP was identified in ambient workplace air during and after the printing processes. Particle fractions between 0.23 and 20 μ m emitted by the office machines significantly affect particle mass concentrations while printing 500 pages, i.e., during the printing process, PM_{0.23-20}, PM_{2.5}, and PM₁₀ concentrations increased in 43 out of the evaluated 62 office rooms investigated. Additionally, a significant increase was observed in submicrometer particles, with median particle

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79106 Freiburg, Germany e-mail: julia.hurrass@uniklinik-freiburg.de number concentrations of 6,503 particles/cm³ before and 18,060 particles/cm³ during the printing process.

Conclusions Our data indicate that laser printers and photocopiers could be a relevant source of fine particles and particularly UFP in office rooms.

Keywords Laser printers · Photocopiers · Emissions · Office room measurements · Ultrafine particles · Health effects

1 Introduction

In recent years, exposure to toner dust and volatile organic compounds (VOC) emitted into the air of office rooms during operation and maintenance of laser printers or photocopiers has been discussed as a possible cause of health complaints, not only those affecting the respiratory tract, but also the immunological and nervous systems (Ewers and Nowak 2006; Gminski and Mersch-Sundermann 2006; Wolkoff et al. 1992).

While so far there is no evidence showing a relationship between emissions from printers and photocopiers and health effects, test chamber investigations and indoor air measurements have been conducted to define typical emissions from these office machines. The studies have shown that not only significant amounts of fine particles and ultrafine particles (UFP) are released into ambient air, but also gases such as ozone and various VOC (Caesar and Schmitt 2009; He et al. 2007; Kagi et al. 2007; Lee et al. 2001; Morawska et al. 2009; Schripp et al. 2008; Wensing et al. 2008; Wolkoff 1999). Additionally, a study published by He et al. (2007) found that almost one third of 62 printers investigated for particle emissions released high levels of UFP with diameters <100 nm. Numerous other studies in different parts of the world have found similar results (Kagi et al. 2007; Lee and Hsu 2007; Schripp et al. 2009; Wensing et al. 2008).

Particles with an aerodynamic diameter of $<2.5 \ \mu m$ (and VOC) reach the lower respiratory tract, and consequently, the small airways of the lungs after inhalation of polluted indoor air. Under consideration of studies dealing with the biological effects of fine particles and UFP (Nel et al. 2006; Schulz et al. 2005), inhaled particles of laser printers may pose health risks in humans. Symptoms such as rhinitis, sore throat, initiation of asthma attacks, and pseudoallergic inflammations of the respiratory tract as well as irritations of the skin and eyes, headache, and sick building syndrome are suspected to be associated with exposure to laser printer emissions (Gminski and Mersch-Sundermann 2006).

The results of the chamber emission studies mentioned above demonstrate the need to investigate the influence of laser printer and photocopier operation on the quality of indoor air under real office conditions. However, resilient studies on printer and photocopier emissions in office rooms are still relatively scarce (Kagi et al. 2007; Schripp et al. 2009; Stefaniak et al. 2000; Wensing et al. 2008). Currently, only two field studies (Fiedler et al. 2009; McGarry et al. 2011) are available, of which the former was only published in the German language.

Therefore, the objective of the present study, which was commissioned by the German Federal Institute for Risk Assessment (BfR), was to investigate the quality of indoor air at the workplace in a representative number of office rooms to obtain information on the concentrations of airborne particulates emitted by laser printers and photocopying machines (referred to as printers in the following) during the printing process.

2 Methods

Particle release into the ambient air of 63 German office rooms before, during, and after printer operation was investigated between January and October 2006.

2.1 Sampling sites and buildings

Sixty-three offices in 9 multifloor, frequently air-conditioned office buildings were investigated at four locations in Germany, i.e., Gießen (Central), Freiburg (South), Trier (West), and Berlin (North). All the offices were government-run. Information on general conditions in the offices were collected by a standardized questionnaire, which included important facts such as the size of the rooms, the number of occupants, type and material of furniture, types of printers and photocopiers used, etc.

2.2 Laser printer and photocopier models used

A total of 59 laser printers and 4 photocopiers of different makes and models were tested. Most were tabletop units; some, however, were high-throughput, floor-mounted devices. All the devices were from current series production.

2.3 Measurement of particle concentrations and their characteristics

2.3.1 Particle monitoring

Continuous measurements were taken of number concentrations of particles with an optical diameter between 0.23 and 20 μ m (PM_{0.23-20}) using an optical laser aerosol spectrometer (LAS; Dust Monitor, Model 1.108, Grimm Technologies Inc., Ainring, Germany). The number concentration of the particles was determined for 15 different channels (sampling interval, 6 s). Data conversion from particle number (PN) concentrations to mass concentrations was achieved under the premise of spherical particles using Eq. 1:

$$M_i = \frac{N_i \pi \rho D_i^3}{6} \tag{1}$$

where N_i is the number concentration of particles, M_i is the mass concentration, and D_i is the geometric mean diameter of the *i*-th channel. ρ is the mean density estimated by gravimetry performed before printer measurements in selected rooms of the four locations using the gravimetric unit integrated in the Grimm aerosol spectrometer.

 $PM_{2.5}$ and PM_{10} were calculated using Eq. 2 as described by Grimm Technologies Inc.:

$$PM_{2.5/10} = \sum_{i=1}^{15} M_i F_i \tag{2}$$

where F_i is the weighting factor of the *i*-th channel. To calculate PM₁₀, the factor is 1 for channels 1 to 7 and 0.942, 0.922, 0.893, 0.8345, 0.724, 0.4486, 0.041, and 0, respectively, for channels 8 to 15. To calculate PM_{2.5}, the factor is 1 for channels 1 to 6 and 0.995, 0.855, and 0.48, respectively, for channels 7 to 9; the factor is 0 for channels 10 to 15.

In recent studies, the Grimm model 1.108 was compared to other optical particle counters, and PM_{10} and $PM_{2.5}$ measurements were determined to be accurate (Cheng 2008; Peters et al. 2006).

For continuous measurement of the total number concentrations of submicrometer particles (0.01 to 1 μ m), a condensation particle counter (CPC; Model TSI 3007, TSI Incorporated, St. Paul, USA) was used (measurement interval, 1 s).

2.3.2 Particle morphology and element composition analysis

Particle morphology, geometric size, and element composition were analyzed by scanning electron microscopy/ energy-dispersive X-ray spectroscopy (SEM-EDX; Hitachi S2300, Hitachi, Japan) and transmission electron microscopy (TEM; Hitachi H600, Hitachi, Japan). SEM imaging was performed with 1,000-fold and 10,000-fold magnification; standardized TEM imaging was performed with 40,000-fold magnification as described by Rödelsperger et al. (2003). Both toner powder from the cartridges and air samples from the office rooms were analyzed. Particles in air samples were collected with nucleopore filters (flow rate, 1.2 mL/min; sampling time, between 30 and 50 min). For direct analyses of the filter deposits by TEM, 50 vision fields were evaluated for each filter. The concentrations of primary particles (PT), aggregates and agglomerates (A+A) as well as the mass-weighted PT diameters were determined.

2.4 Study design

To monitor the particle concentrations in the office rooms, the Grimm particle counter 1.108 and the TSI CPC 3007 were placed in the middle of a desk to include the office worker's normal breathing zone. The two devices were placed about 1 m apart. Before each measurement, the windows and doors of the respective office were closed for the whole night. Staff were not permitted to enter the room during the night or before the measurement. This standard procedure for indoor air measurements is useful to minimize the influence of outdoor particles on the indoor environment. During standby and printing measurements, the experimenter was advised to minimize any movement to avoid particle release from the floor and other surfaces. Additionally, analytical staff wore particle-free protective clothing.

According to the definition given by the Standard ECMA-328 (ECMA International 2007), the printers under test were switched on during the preoperation period. Continuous particle measurement by LAS and TSI CPC 3007 was initiated about 30 min before starting the printing process, i.e., during the standby period. Thus, measurements were done before (standby phase), during (printing phase), and after (working phase) the printing process. In order to simulate worst case but realistic conditions commonly encountered in normal office rooms, 500 pages were printed with each printer during the printing phase.

For the printing procedure, a standard page with 5% black covering was used according to DIN 33870. The printers operated at normal speed; thus, it took 10 to 100 min to complete the printing process (see Supplemental

Material Table 1). If available, the room ventilation system was switched on during the measurements.

SEM and TEM analyses were performed exemplarily for three offices in which particle collection was carried out during the printing process.

2.5 Statistical methods

All the statistical analyses were conducted with SPSS software, version 15.0^{TM} (SPSS Inc., Chicago, IL, USA). Paired *t* tests on particle mass (PM) and PN concentrations were performed at a 95% confidence level to examine the differences in results obtained for the standby, printing, and working phases. The Kolmogorov–Smirnov test, which considers the normal distribution of data, was conducted beforehand to ensure that data on the differences between the three phases could be processed by paired *t* test. Significance was accepted for *P*<0.05.

3 Results

3.1 Conditions in the office rooms

The conditions in the office rooms varied considerably in terms of size, furnishing, and equipment. The office volumes ranged from 20 to 170 m^3 . One of the investigated offices was an open-plan room with a volume of about 1,000 m³. The offices were occupied by 1 or 2 persons, and in the case of the open-plan office, by 17 persons. The furnishing consisted of file cabinets, chairs, and desks. As electronic devices, desktop computers, laser printers, photocopiers, and telephone/fax machines were present in most of the office rooms. Some rooms had wood or carpet floorings, while others had PVC or linoleum. Smoking was not permitted in any of the office rooms.

Neither the background particle concentrations nor their development after starting the printing process showed a relationship to the recorded room conditions (Supplemental Material Table 1).

3.2 Particle mass concentrations

Altogether, PM concentrations in the office rooms ranged between 19.1 and 231 μ g/m³. In comparison to the standby phase, 43 of the 62 investigated office rooms from which a full data set was obtained, i.e., about 70%, showed an increase in PM_{0.23–20} particles during the printing process. In 16 of 62 (26%) office rooms, the PM_{0.23–20} concentration decreased slightly during the printing process, while 3 rooms (5%) did not show any significant difference between the standby and printing phases (Fig. 1). Thus, significantly higher (*P*<0.001) mean concentrations were

Fig. 1 $PM_{0.23-20}$ concentrations of operation-specific phases (standby phase, printing phase, and working phase) expressed as the mean±SD in the indoor air of 62 office rooms (A1–A12 rooms in the office building in Gießen, B13–B26 rooms in the six office buildings in Berlin, C27–C37 rooms in the office building in Freiburg, D38–D63 rooms in the two office buildings in Trier). Due to a technical failure, data from C-30 are missing



observed during the printing and working phases than during the standby phase related to all PM parameters (Table 1).

3.3 Particle number concentrations

Especially the PN concentrations of the small particles increased in the course of the printing phase (Supplemental Material Tables 1 and 2, showing the mean PN concentrations of the smallest fractions measured by LAS). When the office staff entered the room and started working immediately after the printing phase—often opening the windows because of poor air quality—the average values of these small particle fractions further considerably increased.

Figure 2 shows, on the example of room D-48, the timedependent changes in the PN concentrations regarding the different particulate fractions. Soon after printing started, the $PM_{0.23-0.3}$ concentration increased from a background concentration of about 150 particles/cm³ to 270 particles/cm³ at the end of the printing process. After printing stopped, the airborne particle count for this size fraction increased to a maximum of 320 particles/cm³ and then slowly decayed. This example demonstrated that small particles may persist in the office air after printing and may, therefore, be responsible for at least part of the particle burden.

Determination of submicrometer particles was accomplished in 31 out of the 63 offices investigated. Altogether, the PN concentrations of UFP ranged between 1,000 and 80,000 particles/cm³. In comparison to the standby phase, the PN concentration of UFP was found to be significantly higher (P<0.001) during the printing process in 30 of the 31 rooms (Supplemental Material Fig. 1). During the working phase, the mean PN concentration of UFP unexpectedly but significantly decreased in comparison to the printing phases (Table 1).

In Fig. 3, the time-dependent changes in UFP concentrations are exemplarily shown for office room B-14. In this case, the PN concentrations increased as an "initial burst" to a nearly 20-fold increase in UFP PN concentration immediately after print start. Maximum UFP concentrations were reached after approximately 2 min of printing, with 190,000 particles/cm³ being detected in the ambient air. In contrast, only 10,000 particles/cm³ were found before and after printing.

Table 1 Operation-specific mean PM and PN concentrations for the denoted size fractions

	Standby phase			Printing phase			Working phase		
	Av. ^a	SD^b	Med. ^c	Av.	SD	Med.	Av.	SD	Med.
PM concentrations of $PM_{0.23-20}$ (µg/m ³)	61	26	57	71	28	68	81	39	74
PM concentrations of PM_{10} (µg/m ³)	37	17	33	44	18	40	53	31	45
PM concentrations of $PM_{2.5}$ (µg/m ³)	22	13	18	27	13	24	36	27	30
PN concentrations of PM _{0.23-0.3} (particles/cm ³)	199	171	139	220	161	165	470	528	336
PN concentrations of UFP (particles/cm ³)	11,054	14,050	6,503	23,647	18,444	18,060	18,923	10,565	15,539

^a Arithmetic average of the averages calculated for the time-resolved measurement data of the three operation phases of the different office rooms ^b Standard deviation

^c Median of the averages calculated for the time-resolved measurement data of the three operation phases of the different office rooms

Fig. 2 Temporal changes in PN concentrations and particle sizes in the office room D-48 before, during, and after the printing process



3.4 SEM-EDX and TEM results

The toner particle diameters were in the range of 10 μ m, while the measurements of particles from indoor air merely show UFP (Fig. 4). However, submicrometer particles can also be observed in the toner powder, as shown in Fig. 4, especially on the surfaces of the coarser particles. This observation is in agreement with SEM analyses of toner powder conducted by Caesar and Schmitt (2009) and Gminski et al. (2011), which show that the micrometer-sized toner particles are covered by UFP.

The element composition measured by EDX shows that iron is the most abundant element in the investigated toner material (Supplemental Material Fig. 2). Silicon was also detected in the original toner powders. While the iron originates from the black colorant, silica particles are used as charge control additives (Pettersson and Fogden 2005). In contrast to toner material, iron did not play a role in the aggregates and agglomerations (A+A) found in the air samples; those particles consisted mainly of silicon and aluminum (Supplemental Material Fig. 4). In order to explain this discrepancy, both toner materials and air samples were reanalyzed with TEM. EDX element spectra showed element contents in the order Fe>>Al>Si for the large black toner particles, whereas Fe, Si>Al was identified for the optically smaller toner components. If it represents the dominant fraction in the large black particles, iron may possibly interfere with the analysis of other elements.

Both the EDX results for the sampled aerosols and the toner material agree with the XRF analysis results of Barthel et al. (2011). However, due to the better sensibility





Fig. 4 SEM image of the original toner powder used in the printer of office room B-20 with a magnification of ×10,000. Additionally, an image of a particle which was sampled from the office room air during the printing phase is included in the picture

of their method, they detected additional elements, e.g., Ca, Cl, and Br, which were attributed to the paper and to the flame retardants used in the printing devices.

As shown in Table 2, the CPC measurements produced significantly higher PN concentrations, i.e., around 100 times higher for A+A and about 10 times higher for PT concentrations in comparison to the quantitative UFP analyses by TEM.

4 Discussion

The present study revealed elevated concentrations of fine particles and UFP at office workplaces due to the use of laser printers. However, a large variation in PM concentrations was observed between the different office rooms investigated. Various factors, e.g., printer, toner, paper type, and maintenance cycles or air exchange rate, may affect particle concentrations. Similar conclusions have also been drawn by authors of recent publications (Fiedler et al. 2009; Schripp et al. 2009; Stefaniak et al. 2000; Wensing et al. 2008).

The results of particle measurements showed that the mean concentrations of particulate pollution in the investigated office rooms were in the same range as reported for other rooms, e.g., classrooms (Fromme et al. 2007), private residential buildings (Morawska and Salthammer 2003), or office rooms in Australia (He et al. 2007). A study of other typical indoor activities in a 60-m³ laboratory room under controlled conditions yielded comparable or higher PM and PN concentrations for processes such as candle burning or onion frying (Glytsos et al. 2010). However, it should be noted that, depending on the distance of the workplace to the printer exhaust vent, for study design reasons, the measurements were either taken "near to source" or "far from source." Therefore, from a statistical point of view, the results may by all means not reflect comparable mean concentrations in the offices under investigation. However, that was not the goal of the study.

The study allows an assessment of the exposure levels of the office workers under investigation; also, the results can be transferred to other office rooms with similar architecture. Destaillats et al. (2008) pointed out that aerosol particle levels were close to background room levels at a distance of 1.5 m from the printer. McGarry et al. (2011) also found a clear dependence of PN concentrations from the distance of the measurement points to the printers they investigated in office rooms. Therefore, possible health effects caused by emissions can be prevented by positioning the office devices at a safe distance.

Some studies have suggested that aerosolized toner powder, normally of a diameter >10 µm, can be emitted from printers (Caesar and Schmitt 2009; Lee et al. 2001; Wolkoff 1999). In contrast, Lee and Hsu (2007) found that the mean aerodynamic particle diameter shifted from 10.4 µm during the standby phase to 7.1 µm during the printing phase. In our study, the partial increase in the PM concentration of coarse particles, i.e., particles with a diameter of $>10 \mu m$, seems to be consistent with results of the studies suggesting that particulate matter emitted by laser printers consists of aerosolized toner powder that does not adhere to the photoconductive drum. However, if the PN is assessed instead of the PM, the fine particle fraction, i.e., particles with a diameter of <0.5 µm, shows a significantly stronger increase than the coarse fraction, pointing at further, albeit printing-related, sources of particle emission.

Generally, it is more probable that fragments or aggregates melted in the fuser unit are released into the

Table 2 CPC and TEM results of the quantitative particle	Office room	B-20	B-16	B-21
office rooms	PT/cm ³ analyzed by TEM	2,099	11,045	2,337
	A+A/cm ³ analyzed by TEM	300	600	275
	Mean PT diameter (nm) analyzed by TEM	71.1	30.0	38.2
	PT/A+A analyzed by TEM	7	18.4	8.5
	Average PN concentration (particles/cm ³) measured by CPC	17,073	59,564	35,976

indoor air rather than whole, undamaged toner particles. This assumption is in accordance with results obtained by Morawska et al. (2009), Caesar and Schmitt (2009), and Barthel et al. (2011). Comprehensive examination of the chemical composition of the toner material and the particles emitted from the printer in the latter two studies revealed distinct signs for a toner-based origin of a part of the emitted particles.

In addition to the fine particles, we measured distinct release of submicrometer UFP in the office rooms during the printing process. This is also in good accordance with other printer emission studies for office rooms (Fiedler et al. 2009; He et al. 2007; McGarry et al. 2011; Schripp et al. 2009). The authors of these studies noted that UFP emissions varied substantially even among the same printer models. Fiedler et al. (2009) and Schripp et al. (2009) concluded that there are many factors influencing the temporal development and distribution of particulate printer emissions, such as the printer use history, type of paper used, air flow and exchange rate in the room, and background concentrations, and that it is not possible to transfer test chamber results to real room conditions. On the basis of our results, we agree with this statement.

Furthermore, the number of pages printed may have an impact on the resulting maximum particle concentrations. But, according to the study by Wensing et al. (2008) and the temporal change of particle concentrations observed in the studied office rooms (Figs. 2 and 3), we assume that comparable curve progressions would result and that only the maximum concentrations reached after the initial phase of concentration increase would differ, if a different number of pages had been printed.

In many offices, the submicrometer PN concentrations increased significantly by about 1.1 to 42 times after printing process start and, in most cases, decreased to near background level by the end of the printing phase. This phenomenon of high peak concentrations soon after printing operation start was first identified as an "initial burst" within the scope of our investigations and has meanwhile been confirmed by controlled chamber measurements (He et al. 2007; Koivisto et al. 2010; Morawska et al. 2009; Schripp et al. 2008, 2009; Wensing et al. 2008). The origin and significance of the "initial burst" have not yet been explained, but may be caused by condensation of VOC. The suggestion was supported by the results obtained by Morawska et al. (2009) and He et al. (2010), which indicate a correlation between temperature fluctuations in the fuser roller and peaks in particle release. Presumably, the measured aerosols are at least partially made up of condensates and microaggregates of various substances, e.g., silicone oils, VOC, or semivolatile organic compounds (SVOC) or their oxidation products.

Such UFP are assumed to be formed during toner fixation. This process involves heating the fuser roller and the toner powder adhering to it as well as the paper and may result in a release of VOC, which are then possibly nucleated to form particles. The formation mechanisms of the printer-based UFP and their chemistry were intensely discussed by Lee and Hsu (2007), Morawska et al. (2009), and Wensing et al. (2008). In order to cast more light on this phenomenon, we performed electron microscope analyses of selected toners and office air samples.

In accordance with other studies (Barthel et al. 2011; Caesar and Schmitt 2009; Gminski et al. 2011), our SEM and TEM results show that the toner powders also contain submicrometer particles. Under consideration of the EDX results, the size distribution of solid particles sampled during the printing process in the office room air—with a large part in the UFP range—suggests that at least part of the printer emissions originates from toner powder.

However, the distinctly higher UFP concentrations measured by CPC than by electron microscopy indicate that liquid or volatile aerosols presumably account for the greatest portion of laser printer emissions. The aerosol formation from condensation and coagulation processes of VOC and SVOC originating from the chassis, fuser chemicals, paper, or toner that remained on the fuser roller is supported by reports on the missing correlation between particle emissions and toner consumption. Thus, the concentrations of UFP emitted showed no correlations with the toner coverage on the printed paper or the number of pages printed (Morawska et al. 2009; Schripp et al. 2008). As shown by Wensing et al. (2008), even the use of a modified printer without paper and toner resulted in no significant differences in particle emission behavior. Additionally, investigations of the chemical composition of the particles emitted from printers (Morawska et al. 2009; Wensing et al. 2008) and of VOC and particles released from individual heated components of printers (Barthel et al. 2011; Morawska et al. 2009) indicate that the greatest proportion of printer emissions can be explained in this way.

If our results for the office rooms are compared to other measurements, it becomes evident that comparable UFP levels have been found in other office rooms as well as inside residential and school buildings. For instance, Fromme et al. (2007) found a median particle concentration of 5,660 particles/cm³ (particle size between 10 and 500 nm) in Bavarian classrooms. Morawska and Salthammer (2003) measured particles ranging from 70 to 800 nm in Australian apartments at concentrations of between 12,400 and 18,200 particles/cm³. For Australian office workplaces, 8-h time-weighted background particle concentrations between 1,700 and 12,000 particles/cm³ with concentration peaks up to 99,000 particles/cm³

associated with discrete printing events were determined by McGarry et al. (2011). The office room UFP concentrations modeled by Koivisto et al. (2010) with maxima up to 26,000 particles/cm³ are also consistent with this order of magnitude.

However, the similar order of magnitude of PN concentrations in these rooms does not provide any information about the comparability of the qualitative aspects of the particles, i.e., physical and chemical properties. For this reason, the results of almost all studies published so far offer sparse information on the possible significance of these particles for room air hygiene or any hazards they may pose to human health.

Health effects from inhaling UFP depend on PN, size, surface area, shape, solubility, and chemical reactivity (Marconi 2006; Nel et al. 2009) and may range from sensory irritation of mucous membranes to more severe illness such as cardiovascular diseases or cancer. Three recent reviews (BfR 2008: Ewers and Nowak 2006: Gminski and Mersch-Sundermann 2006) summarized and evaluated studies that illuminate the significance of health effects caused both by direct exposure to toner particles and exposure to emissions from laser printers. The reviews critically analyzed data on exposure values by chamber and indoor measurements generated with in vitro studies using bacteria and cell models, in vivo studies with animals, human exposure studies, investigations dealing with human effect markers (biomonitoring), as well as epidemiological and occupational studies. From these studies, it can be surmised that emission rates and exposure to chemical compounds or substance classes released during operation of laser printers were related to biological threshold values.

However, the results of studies on the biological effects caused by exposure to toners outlined in these reviews are contradictory. Whereas the in vivo studies dealing with direct exposure to toner dusts led to the assumption that neither acute nor chronic oral, dermal, or inhalation toxicity can be expected even at high concentrations (Lin and Mermelstein 1994), human biomonitoring studies provided convincing indications for irritative and genotoxic effects related to exposure to the emissions from laser printers and photocopiers (Gadhia et al. 2005; Goud et al. 2001, 2004; Wolkoff et al. 1992).

While numerous studies dealing with the direct toxicity and inhalation toxicity of toner dust are available (Gminski et al. 2011; Lin and Mermelstein 1994; Nakadate et al. 2006), investigations evaluating the effects of exposure to emissions from office machines are scarce. The first report on human health effects due to printer emissions was published in 1992 (Wolkoff et al. 1992). In this study, 30 human subjects were exposed to increased levels of ozone, formaldehyde, TVOC, and particulate matter in a chamber experiment with operating office equipment. The exposed participants reported significantly increased perception of headache, mucous membrane irritation, dryness in the eyes, nose, and throat, and dry and tight facial skin. A population-based cross-sectional study (Jaakkola et al. 2007) showed that exposure to paper dust and to exhaust air from printers was associated with upper respiratory and skin symptoms, breathlessness, tonsillitis, and middle ear infections. Muñoz et al. (2007) found evidence that toner can cause asthma and even vocal cord dysfunction. However, Hänninen et al (2010) performed a risk estimate based on published printer emission rates and general toxicological and epidemiological data of nanoparticles and concluded that health effects due to printer particles are negligible.

5 Conclusions

Laser printers and photocopiers are a relevant source of fine particles and particularly UFP found as air contaminants in office rooms. However, with respect to UFP, this study indicates that liquid or volatile aerosols presumably account for the greatest portion of laser printer emissions.

There are only slight indications of health effects, and due to the lack of data, many questions regarding the association between emissions from laser printers and health effects cannot as yet be answered. While a comprehensive study is still required to provide a better database on laser printer emissions, exposure, and chemical characterization, this and other studies imply that particle concentrations in office rooms can be reduced by proper choice of printers (He et al. 2007) and probably by use of appropriate filter techniques (Caesar and Schmitt 2009; Wensing et al. 2008). However, to date, it remains unknown which component of printer emissions is responsible for the assumed adverse health effects. Since laser printers are being used increasingly, not only in offices but also in private homes, additional studies, such as in vitro studies with human cell lines, controlled human exposure studies using reliable biological markers, and epidemiological surveys are indispensable for human risk assessment.

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