

# Contrasting climate influences on *Nothofagus pumilio* establishment along elevational gradients

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**Abstract** Predicted warmer temperatures and more frequent extreme climatic events in the southern Andes may affect the dynamics of the Patagonian forests. These environmental changes may differentially alter the probability of *Nothofagus pumilio* establishment across its altitudinal range of distribution. We monitored fruit fall, seedling emergence and survival at altitudinal distribution range of *N. pumilio* forests in Santa Cruz (49°22' S—72°56' W), Argentina. Fruit fall, seedling emergence and survival were tested in relation to drought, based on the SPEI

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Instituto Ciencias del Ambiente, Sustentabilidad y Recursos Naturales, Universidad Nacional de la Patagonia Austral, Av. de los Mineros 1260, CP 9407 Río Turbio, Santa Cruz, Argentina (Standardized Precipitation Evapotranspiration Index), interacting with elevation. Fruit fall was consistently higher at low elevation and the abundance of fruits was not affected by drought intensity. Density of new-born seedlings was  $\sim 9$  to 24 times lower at low- than at high-elevation plots in our first observations (2014 and 2016), characterized by warm-dry climatic conditions in spring-early summers. In contrast, seedling abundance was  $\sim 1.5$  times larger at low elevations during relatively cold-wet growing seasons. Survival probability was explained by the interaction between SPEI and elevation. At low elevation and in dry periods, survival probability was lower (CI 54-72%) than in wet periods (CI 68-84%) but at high elevation similar survival was registered even with positive or negative SPEI values. Our results show interacting effects of elevation and drought on tree establishment at the elevation limits, with positive and negative drought effects at high and low elevations, respectively. Predicted increase of extreme drought events during the XXI century could be detrimental for N. pumilio establishment at dry, low-elevation forests.

**Keywords** Recruitment · Seedling emergence · Survival · Elevation range · Fruit fall

## Introduction

Understanding the consequences of climate change on mountain's forests requires a comprehensive view of tree life stages and species-specific processes along their ranges of distribution. At large-spatial scale, cold temperature is usually recognized as the limiting factor modulating plant performance at the upperelevation limits, whereas at the lower and warmer borders of distribution, water supply, indirectly regulated by temperature, is the main forcing of plant growth (Hampe and Petit 2005). Therefore, the consequences of climate warming on plant vital rates are expected to differ spatially for tree species with wide elevation ranges. Empirical knowledge on the performance of plant reproduction at elevation boundaries could help us to understand vegetation responses to climate changes before the potential expansion or contraction of their range occurs (Harsch et al. 2009; Harsch and Bader 2011; Cagnacci et al. 2020). Demographic information, in particular establishment of new individuals, can provide the basis to predict population performance, as well as changes in distribution and in turnover rates in tree populations. Seedling emergence and survival are the stages in the trees' life cycle most vulnerable to climate change (Grubb 1977; Peter and Ronald 1996; Hampe and Petit 2005; Engelbrecht et al. 2007; Tercero Bucardo et al. 2007).

Minor variations in environmental conditions at the elevational limits of the forest can induce changes in tree recruitment, but may affect to a lesser degree these demographic processes at intermediate sites with lower environmental stress (Kullman 2007; Kirdyanov et al., 2012). Whilst air temperature decreases along the elevation gradient with indirect effects on environmental humidity, atmospheric pressure gradually decreases and solar radiation increases with elevation (Körner and Paulsen 2004; Körner 2007a). Lower temperatures at higher elevations induce a shortening of the growing season, and in most cases, a reduction in the supply of nutrients from the soil, negatively impacting the growth rates with increasing elevations (Coomes and Allen 2007; Körner 2007b). In addition, trees at high elevation are more exposed to wind and experience longer persistence of snow cover than trees at lower elevation. In contrast, the increment in temperature at the lower forest boundary increases evapotranspiration, which in turn reduces soil water availability for tree recruitment and growth (Sánchez-Salguero et al. 2015). In this context, there is a greater possibility that environmental fluctuations on plant performance will manifest earlier at the extreme forest boundaries (Doak and Morris 2010), justifying the importance of conducting demographic studies in these focal locations.

*Nothofagus* (southern beech) forests in the Patagonian Andes from South America comprise one of the most singular and large mountain ecosystems in the planet (Donoso 1993). This genus dominates the upper treeline in the highest hostile Andean environment and their species are widely distributed throughout the region, from Central Chile (33.5°S) to Tierra del Fuego (56°S; Veblen et al., 1996). The *Nothofagus pumilio* ([Poepp. & Endl.] Krasser) forests are important as timber producers, regulators of water runoff and other ecosystem services to populations, like tourism and recreation, on both slopes of the Patagonian Andes (Ramírez et al. 1997; Martínez-Pastur et al. 2000; Bava and Rechene 2004).

Seed production and dispersal are prerequisites for successful establishment in the forest. A continuous decrease in seed production with increasing elevation has been reported for Nothofagus species, including N. solandri var. cliffortioides in New Zealand (Wardle 1980; Allen and Platt 1990) and N. pumilio in Southern Patagonia (Barrera et al. 2000; Cuevas 2000, 2002). Fruit fall, seed viability and seedling emergence in N. pumilio decrease with increasing elevation, suggesting that tree establishment at high elevation is regulated by two bottlenecks: seed production and seedling emergence (Cuevas 2000). In addition to seed production and germination, tree establishment could also be limited by seedling survival. The highest mortality rate of trees occurs during the first year of establishment since seedlings are extremely sensitive to variations in environmental factors (Maher and Germino et al. 2006; Soler et al. 2013; Cagnacci et al. 2020). Seedling mortality can be caused by desiccation, overheating, freezing, diseases or herbivory (Stevens and Fox 1991; Piper et al. 2006; Danby and Hik 2007). In addition, mortality can be also induced by competitive interactions between plants (Olofsson 2004). At the low-elevation Nothofagus forests in the ecotone with the Patagonian steppe, climate warming could impose an increased drought stress, limiting the survival of N. pumilio seedlings and trees. Conversely, seedling survival at higher elevations could benefit from the increase in temperature associated with the extension of the growing season and the stress reduction due to heavy snow accumulation (Daniels and Veblen 2004; Piper et al. 2013; Srur et al. 2016, 2018).

Based on the analysis of instrumental and proxy records, temperature in southern Patagonia has increased during the past century with a marked positive trend in the past few decades (Villalba et al. 2003; Falvey and Garreaud 2009; Masiokas et al. 2008, 2009; Lopez et al. 2010; Garreaud et al. 2013). According to Magrin et al. (2014), precipitation has decreased in southern Chile and Argentina since the 1960s concurrent with a warming trend close to 0.1°C per decade. First, we here describe and analyse the temperature and precipitation variations during 2006-2020 in order to show the recent recurrent droughts in the years of our study. We hypothesize that temperature and precipitation variations differentially modulate tree recruitment at high- versus low-elevation boundaries in the southern Andes. Therefore, in this study we evaluate how seedling emergence and survival at upper and lower elevations vary in relation to climate in southern Patagonia. We postulate that a warmer climate would ameliorate the harsh environmental conditions for tree recruitment imposed by high elevations, but intensified water stress (drought and overheating) for seedlings at lower elevation forests. We test this hypothesis by comparing fruit fall, seedling emergence and survival in N. pumilio forest at low and high elevations over 6 years of changing climate in the southern Andes.

## Materials and methods

#### Study area

The study area is located in the Río de las Vueltas basin (49° S–72° W), Santa Cruz, Argentina (Fig. 1). All the area is under the influence of strong winds from the west throughout the year. At a regional scale, temperature decreases with elevation, whilst precipitation increases in the east–west direction (Villalba et al. 2003). Meteorological data from Los Huemules meteorological station (49°13′09″ S, -72°57'38″ W), in the central sector of our study area, indicate a total annual precipitation over 1000 mm and a mean annual temperature of 6 °C.

N. pumilio is the dominant subalpine tree, endemic to the Patagonian Andes of Chile and Argentina (Veblen et al. 1996). The deciduous forests of N. pumilio belong to the SubAntarctic phytogeographical province of the Antarctic Region in southern South America (Cabrera 1976). In the study area, N. pumilio is the only tree species that constitute forest cover on the eastern slopes from  $\sim 500$  to 1100 m asl. These species poor forests mainly contain an herbaceous layer and few shrubs, like Berberis microphylla, Ribes magellanicum, Chilliotrichum difussum, Maytenus magellanica, Gaultheria mucronata and Myoschilos oblongum. Trees grow up to 20 m in height with erect trunks up to 1.5 m in diameter (Roig et al. 1985). Leaves are deciduous, elliptic, 2-4 cm long with regular lobed margins. The fruit is a smallnut (4-7 mm) dispersed by wind in early spring. Seeds do not form a persistent soil bank, they remain dormant during winter and start germination during spring and summer (after October), later they can persist as seedlings for many years until canopy opens (Donoso 1993). N. pumilio may form seedling banks as it was recorded at xeric sites of northern Patagonia (Heinemann et al. 2000; Heinemann 2007).

# Data collection

We collected data at 10 N. pumilio monospecific stands at low and high elevation at three sites along Rio Las Vueltas basin. Stands were located in private and national protected areas (Los Glaciares National Park, Estancia Los Huemules, and Estancia Río Toro) (Table 1; Fig. 1). Plots were located in old-growth, multi-aged erect stands without recent evidence of natural or anthropic disturbances (e.g. livestock grazing, fire or logging). None of our plots were located at N. pumilio "krummholz" stands, characterized by shrub-like stunted trees, usually present at stressed treelines or south-facing slopes with relatively short growing seasons due to a long snow cover period (Lara et al. 2005). Plots included 50 alive trees larger than 5 cm diameter at breast height (DBH) (Srur et al. 2008). To account for seedling emergence and seedling survival, inside each plot, 10 subplots  $(1 \times 1 \text{ m}, \text{ Fig. 2a})$  were established to mark and follow seedling's fate, biannually between 2012 and 2018. All seedlings were tagged with plastic numbers to follow individual fate. Due to accessibility limitations by a broken bridge given the river flood, seedling



◄ Fig. 1 a Map of the study area in Southern Patagonia, Argentina

survival could not be monitored in Ea. Río Toro for the year 2016. To estimate fruit fall we installed inside each plot 10 seed traps (diameter 25 cm wide and 35 cm in depth, Fig. 2b) and counted the number of fruits collected annually between February 2015 and 2018.

## Data analysis

#### Drought characterization

To identify the drought events in our study area, we used a multiscalar drought index based on instrumental climatic data. In the study area, unfortunately there are few meteorological stations and most are recent and often have missing data. To assess the reliability of the data from the meteorological station located in the centre of our study area at Ea. Los Huemules for the other sites we performed correlation analysis with the available data from the other stations in the area. One of the other stations is located in Los Glaciares National Park, which has temperature data from February 2015 to the present, but with missing data for the entire 2016. Unfortunately, this station has not recorded precipitation data. Another station is located at the outlet from Lago Milanesio, within the Ea. Río Toro and has temperature and precipitation records. Another station is located at the outlet from Lago Milanesio, within the Ea. Río Toro, and has temperature and precipitation records only from February 2016 to April 2020. Due to the lack of data for all our study period in those meteorological stations and considering the high correlation of data between all the stations located in the study area (r > 0.5, p < 0.001, see Mat. Supp. S1), we used only the data from the meteorological station at Ea. Los Huemules as it was the most complete and reliable, with records from 2006 to the present.

The Standardized Precipitation Evapotranspiration Index (SPEI) is calculated from the difference between the precipitation and the potential evapotranspiration. This last variable uses the Thornthwaite equation and the mean temperature data for the study area (McKee et al. 1993; Paulo et al. 2012). The SPEI was calculated using the SPEI package version 1.7 (Beguería and Vicente-Serrano 2017) of the free programming language R (R Core Team 2020). To describe trends in regional climate we analysed monthly precipitation and temperature deviations in relation to the period of 2006-2019 and show longterm climate in our study area. In our statistical model for fruit fall we calculated SPEI using climatic data from October to March (6 months), which was the current growing season when fruit were collected. To assess SPEI effect on seedling emergence and survival we used the SPEI calculated from October of two years before to March of the sampling year (18 months). Therefore, we included the growing

 Table 1
 Location and elevation of sites with monospecific Nothofagus pumilio where fruit fall, seedling emergence and survival were studied in Santa Cruz, Argentina

| Site             | Elevational boundary | ID site | Elevation (m asl) | Latitude S     | Longitude W   |
|------------------|----------------------|---------|-------------------|----------------|---------------|
| APN-Glaciares    | Low                  | PBB     | 774               | 49° 22′ 22.25″ | 72° 54′ 10.4″ |
| APN-Glaciares    | Low                  | PBB2    | 729               | 49° 21′ 38.5″  | 72° 53′ 47.4″ |
| APN-Glaciares    | High                 | PBA     | 896               | 49° 21′ 49.9″  | 72° 55′ 25.9″ |
| APN-Glaciares    | High                 | PBA2    | 908               | 49° 21′ 56.5″  | 72° 55′ 25.8″ |
| Ea. Los Huemules | Low                  | LHB2    | 508               | 49° 11′ 42.2″  | 72° 57′ 14.7″ |
| Ea. Los Huemules | Low                  | LHB3    | 466               | 49° 12′ 11.6″  | 72° 57′ 16.5″ |
| Ea. Los Huemules | High                 | ANI2    | 863               | 49° 11′ 52.55″ | 72° 59′ 6.0″  |
| Ea. Los Huemules | High                 | ANI5    | 869               | 49° 12′ 55.30″ | 73° 00′ 3.24″ |
| Ea. Rio Toro     | Low                  | TOB     | 664               | 49° 3′ 47.45″  | 72° 57′ 57.6″ |
| Ea. Rio Toro     | High                 | TOA     | 1008              | 49° 4′ 6.30″   | 72° 58′ 2.22″ |

(a)

**Fig. 2 a** Diagram of the field design to record *N. pumilio* demographic changes at the extremes of the elevation range. At high- and low-elevation forest boundaries, 10 subplots  $(1m^2)$  for monitoring the establishment were located inside a plot with 50

season from the previous year until the current growing season and so, for example, for the data taken in 2014 the period comprised was from October 2012 to March 2014.

# Statistical models

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We conducted all analyses with R version 4.0.2 (R Core Team, 2020). To evaluate fruit fall, seedling emergence and survival in relation to elevation and SPEI we applied Generalized Linear Models (GLMMs) using the *glmer.nb()* and *glmer()* function of the package lme4 (Bates et al. 2015). Fruit fall, seedling emergence and survival were tested in relation to elevation and SPEI as fixed factors. In all models site was included as random factor in order to control and consider variation of land management

adult trees. Inside each subplot all saplings and seedlings were tagged. Seedling emergence and survival were monitored every 2 years from 2012 to 2018. **b** Inside each area, 10 seed traps were used to quantify annual fruit fall between 2015 and 2018

and abiotic factors between locations. To test for variation in fruit fall and seedling emergence we assumed a Negative Binomial (log link) distribution of errors to remedy overdispersion (Crawley 2005). To test for variation in survival we assumed a Binomial (logit link) distribution of errors. Interactions between fixed factors were included and tested if required for empirical support in each model.

# Results

(b)

The study period was characterized by large climatic variability in terms of monthly temperature and precipitation (Fig. 3). In our study area, after 2012, recurrent positive deviations in temperature and negative precipitation deviations were registered



**Fig. 3 a** Monthly variations of precipitation and temperature deviations (in relation to the period of 2006–2020) from Los Huemules meteorological station  $(49^{\circ}13'09'' \text{ S}, -72^{\circ}57'38'' \text{ W})$ , southern Patagonia, Argentina. To emphasize the long-term climate fluctuations, variations in precipitation (blue) and temperature (red) are shown as 5-month moving averages.

along the growing seasons. Therefore, negative SPEI values were recurrent in our study period from 2012 to 2016 (Fig. 3).

Fruit fall was consistently more abundant at lowthan at high elevations between 2015 and 2018 (Negative binomial GLMM,  $\chi^2 = 10.41$ , *p* value < 0.05, Mat. Supp. S2) but the magnitude of this difference was variable between years (Fig. 4). There was no interaction between elevation and SPEI

**b** Long-term monthly variations of SPEI (black), dry periods (SPEI < 0) are highlighted in red. The initial experimental setting to monitor seedling emergence and survival was conducted in February 2012 and followed by periodic samplings in February 2014, 2016 and 2018

to explain fruit fall, neither SPEI affected fruit fall independently. Fruit production showed the lowest record at both elevations in 2016 and the maximum in 2017 (Fig. 4).

In average we estimated a density of  $8676 \pm 2188$ and  $6767 \pm 2231$  new seedlings/ha at low and high elevations, respectively, but the abundance of emerged seedlings showed temporal variation between high and low elevations. Seedling emergence



**Fig. 4** *Nothofagus pumilio* annual fruit fall mean ( $\pm$  SE) between 2015 and 2018, for low (grey) and high (black) elevation sites. Large interannual variability in the amount of fruit production was recorded over the interval 2015–2018, with a maximum in 2017

was explained by elevation interacting with SPEI  $\chi^2 = 51.88$ , (Negative binomial GLMM, p value < 0.001, Mat. Supp. S3), reflecting the importance of water deficit for regeneration. Seedling emergence was lower at low-elevation forests in 2014 (low:  $0.06 \pm 0.40$  vs. high:  $0.54 \pm 1.36$ ) and 2016 (low:  $0.01 \pm 0.11$  vs. high:  $0.24 \pm 0.65$ ), but higher in 2018 (low:  $1.46 \pm 3.30$  vs. high:  $0.95 \pm 1.66$ ). Therefore, in drought periods characterized by negative SPEI values (e.g. SPEI = -0.99), density of emerged seedlings was higher at high elevation but in wet periods (e.g. SPEI = 0.32) the abundance of emerged seedlings was similar at both elevation boundaries (Fig. 5). Seedling survival probability was explained by the interaction between elevation and SPEI index (Binomial GLMM,  $\chi^2 = 13.48$ , p value < 0.05, Mat. Supp. S4). At low elevation and in drought periods, survival probability was lower (CI = 0.54 - 0.72) than in wet periods (CI = 0.54 - 0.72)0.68–0.84). At high elevation similar survival probability was estimated with positive (CI = 0.67-0.82) or negative SPEI values (CI = 0.71-0.85) (Fig. 6).



Fig. 5 Density of seedling emergence, mean  $\pm$  CI 90%, explained by elevation and climate (SPEI index) interaction. In periods of drought (SPEI < 0) seedling emergence was higher at the upper-elevation boundary of *N. pumilio* forest, but differences in establishment were null in the absence of drought. The climatic anomalies used to show the response are consistent with SPEI -0.99 and 0.32 for our sampling period in 2014 and 2018, respectively, and were obtained predicting a Poisson GLMM



Fig. 6 Seedling survival probability, mean  $\pm$  CI 90%, in relation to elevation and climate (SPEI index) interaction. Drought conditions, with negative values in SPEI index, decreased survival for seedlings at low elevation

#### **Conclusion and discussion**

Our study reports different tree regeneration responses to climate variability at contrasting elevation boundaries of Patagonian mountain forests. Although fruit fall was consistently larger at lower than at higher elevations, seedling establishment along elevation varied over time, particularly in relation to drought conditions. Density of seedling emergence and seedling survival was lower at low elevation in periods with droughts, characterized by a negative SPEI index.

Across our study area, seedling survival probabilities were larger in the upper boundary of elevation during periods of drought. These findings contradict previous studies showing scarce tree recruitment at high-elevation erect N. pumilio forests (Barrera et al. 2000) in Tierra del Fuego (latitude ~ 54°S) or indistinctly along the elevation gradient at 40°S in northern Patagonia (Cagnacci et al. 2020). The difference in the results between these studies and ours could be associated with differences in the climatic conditions prevailing in each study during the monitoring period of establishment. In our region, the biannual periods 2012-2014 and 2014-2016 were characterized by a dry and warm spring-summer climate conditions, whilst abundant rainfall was recorded in the period 2016-2018. Seedling emergence was less abundant at the lower forest boundary during the periods with drought, suggesting that N. pumilio germination in low-elevation forests is limited by warm and dry climatic anomalies. Consistent with

our hypothesis, drought conditions favoured establishment at the high elevation but limited it at the lower forest edge. Our observations indicate that establishment patterns are strongly modulated by climatic variability interacting with elevation. Therefore, differences with Barrera et al. 2000 and Cagnacci et al. 2020 could be due to the specific climate conditions during the seedling monitoring, the shorter monitoring time, and the more humid and colder environments at the sampling sites in Tierra del Fuego or dryer climate in Northern Patagonia at 40°S.

The concept of increasing reproductive capacity in N. pumilio forests with increasing elevation was early suggested by Rusch (1993), who noted that although shorter growing seasons at high elevation limit tree productivity, recruitment is benefited by more abundant precipitations that mitigate soil water deficit. Consistent with Rush (1993) observations, the main cause of seedling mortality at low-elevation forest appears to be summer desiccation, as typified by the orange coloration of the seedling first leaves or cotyledons observed in the field. Most monitored seedlings in our study were from sexual reproduction since they were initially tagged with the presence of cotyledons. In seed-bearing plants, cotyledons are embryonic leaves, whereas plants from asexual sprouts do not shown them. Our results support the untested hypothesis from Rusch (1993), indicating that larger seedling survival probability could explain higher seedling density at high- than at low-elevation boundary.

Extreme climatic events, rather than the mean climate, seem to be more important for terrestrial ecosystem dynamics (Holmgren et al. 2006; Suarez and Kitzberger 2008; Amoroso et al. 2012). Extreme dry-warm summers between 2012 and 2016 at our study area exacerbated the recorded differences in establishment at both elevation extremes. Cagnacci et al. (2020) showed, with an experimental approach, that seedling mortality was regulated by air temperature during the first year of growth, but by relative air humidity during the second year of their experiments. Therefore, consistent with our results, Cagnacci et al. (2020) showed that establishment is largely dependent on climate conditions and emphasizes the need of long-term monitoring to properly characterize regeneration dynamics. In ecological studies, the use of a stationary temporal assumption could limit the applicability of results beyond the monitoring range. In turn, it has recently been emphasized that climate change is introducing unprecedented non-stationary changes that should be considered for forecasting the dynamics of biological systems (Wolkovich et al. 2014). Our findings highlight the benefits of long-term monitoring over time to understand the responses of forest species to changes in climate along elevational gradients.

Detrimental effects of climate warming for the conservation of plant populations at low elevation, called rear edges, have been reported in Europe (Peñuelas and Boada 2003). Our results support the rear edge hypothesis for forests at the lower elevation boundary in Southern Patagonia, particularly under recent prevalent warm-dry springs and summers. A challenge for future research is to monitor with a life-cycle approach how population dynamics of this Patagonian endemic tree would respond if climatic extremes increase their frequency as predicted for the near future.

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**Data availability** The data that support the findings of this study are available on request from the corresponding author.

**Code availability** The code that supports the findings of this study is available on request from the corresponding author.

## Declarations

Conflict of interest The authors declare no conflict of interest.

**Consent to participate** All the authors have approved the contents of this paper and agreed to submission policies.

**Consent for publication** All the authors agree to ask consideration for publication in Plant Ecology.

Ethical approval Not applicable.

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