

Single-step solvothermal synthesis of mesoporous anatase $TiO₂$ -reduced graphene oxide nanocomposites for the abatement of organic pollutants

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Abstract In the present work, we have fabricated a novel mesoporous $TiO₂-rGO$ nanocomposite by a facile one-step solvothermal method using titanic sulfate as the $TiO₂$ source. The as-prepared composites were characterized by transmission electron microscopy, X-ray diffraction; UV–Vis diffuse reflectance spectra, X-ray photoelectron spectroscopy and photoluminence spectra. In situ nucleation and anchoring of TiO₂ nanoparticles onto a graphene sheet is favorable fpr forming an intimate interfacial contact, and the chemically bonded $TiO₂$ -rGO nanocomposites commendably enhanced their photocatalytic activity in the photodegradation of rhodamine B and phenol. The high photocatalytic activity of the as-synthesized nanocomposites are primarily ascribed to the mesoporous structure, efficient charge transportation and separation with enhanced visible light absorption, which come from the appealing nanoarchitecture, for instance, ultra-dispersed and ultra-small $TiO₂$ nanocrystals along with intimate and absolute interfacial contact between the $TiO₂$ nanocrystals and the graphene sheet.

Keywords Mesoporous $TiO₂ \cdot In situ growth \cdot Solvothermal method \cdot Graphene \cdot$ Photocatalytic activity

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In the recent years, semiconductor nanocrystals have drawn considerable interest on account of their exceptional physicochemical features and strong prospects for applications in solar energy conversion, supercapacitors, and the photodegradation of hazardous pollutants $\left[1\text{-}3\right]$. TiO₂ photocatalysts have been regarded as one of the most efficient and reliable materials due to its superior photocatalytic performance, non-toxicity, low cost, biological and chemical inertness and long-term physical and chemical stability [\[4–6](#page-13-0)]. Undesirably, the fast recombination rate of photogenerated charge carriers within $TiO₂$ results in its low quantum productivity, consequently deterring its real-world applications. In previous years, numerous approaches have been devised in an endeavor to improve the photocatalytic process and to enhance the photocatalytic performance of $TiO₂$ photocatalysts [[7\]](#page-13-0). In particular, carbon– titania nanocomposites have been attaining considerable responsiveness as a stateof-the-art class of photocatalysts [[8,](#page-13-0) [9\]](#page-13-0), which can potentially offer an appropriate ability for sorting out electron–hole pairs.

Recently, graphene, a single layer of carbon atoms, has enthralled researchers because of its extraordinarily specific surface area, thermal and mechanical properties, and better electron mobility, and many efforts have been made to use graphene as a substrate for producing very large numbers of functional nanohybrids with improved catalytic performance [\[10–](#page-13-0)[15\]](#page-14-0). Although many exhilarating advancements have been accomplished in the synthesis of $TiO₂$ reduced graphene nanocomposites, there still exist numerous complications hampering the further improvement of their photoactivity. The aggregated or large $TiO₂$ nanoparticles are a big concern, which meaningfully reduced the surface– volume ratio of $TiO₂$ and lengthens the diffusion distance of radiative electrons and holes, consequently resulting in a decline in photocatalytic activity [\[16](#page-14-0)]. Additionally, the random distribution of $TiO₂$ nanoparticles, with a notable degree of exposed and undecorated graphene sheet or irreversible aggregation of graphene sheets encouraged by means of robust $p-p$ interactions, lead to a substantial shrinkage of the available surface area and substrate approachability, hence losing material efficiency over a brief time frame [[17–19\]](#page-14-0). Furthermore, most of the existing approaches for the preparation of $TiO₂$ nanocrystals on graphene sheets involve environmentally detrimental surfactants (e.g., P123 [\[20](#page-14-0)], SDS [\[21\]](#page-14-0), SDBS [[22](#page-14-0)], CTAB [[23\]](#page-14-0), and PVP [[24\]](#page-14-0)), and complicated synthetic steps such as a two-step solution process, gas phase synthesis pathway, direct deposition and sol–gel methods [[25–32](#page-14-0)] also profoundly restrict the industrialized expansion of the nanohybrids. The ability to control the size, morphology, dispersion density and strong interaction of inorganic nanocrystals on graphene sheets, permitting one to tailor nanocomposites toward a specific application, is still an imperative challenge, and demands more grit and determination, because it is extremely challenging to match their compatibilities and interactions as well as regulating the reduction of graphene oxide (GO) simultaneously in the synthetic process, which could thwart the transfer of electrons and thus impede applications in energy storage and catalysis [[33,](#page-14-0) [34\]](#page-14-0). Compared to physically blending two

compounds together, in which the pathetic interaction between $TiO₂$ and graphene sheets is a foremost shortcoming, in situ anchoring of $TiO₂$ nanocrystals onto graphene sheets through chemical bonds improves the stability of the nanocomposite and encourages charge separation and mutual electron transport between $TiO₂$ and graphene [[35](#page-14-0)]. It has been established that additives like polyaniline can be used as the face-controlling and/or linker agent to accomplish in situ growth of ultra-dispersed $TiO₂$ nanoparticles on the surface of $TiO₂$ -graphene aerogels; however, the loading amount of $TiO₂$ nanoparticles was small and difficult to control [[36](#page-14-0)]. Therefore, it is still a challenge and immensely desired to accomplish in situ growth of $TiO₂$ nanocrystals on a graphene sheet with sturdy binding interactions and well-regulated loading concentrations.

Here, we report a simple surfactant-free facile one-step solvothermal method for the synthesis of mesoporous $TiO₂$ nanocrystals on graphene. In the present work, the $TiO₂$ nanoparticles nucleated, grew and anchored on GO sheets using an inexpensive and stable titanium sulfate as a $TiO₂$ source in a water and ethylene glycol system with a trivial amount of ammonia. The resultant mesoporous $TiO₂$ rGO nanocomposites not only possess highly dispersed small-size anatase nanoparticles (\sim 5–10 nm) but also exhibits enhanced visible light photocatalytic performance for the degradation of rhodamine B dye and phenol.

Experimental

Materials used

Graphite powder ($\langle 20 \mu m$, synthetic), titanium sulfate (Ti $(SO_4)_2$), ethylene glycol, NaNO₃ (99.0%), H₂SO₄ (95.0%), KMnO₄ (99.3%), H₂O₂ (30%) and NH₄OH were purchased from Sinopharm Chemical Regent (Shanghai). Degussa supplied the commercial P25. Deionized water was used in all the experiments.

Synthesis of graphene oxide (GO)

GO was prepared by a modified Hummer method $[10, 37]$ $[10, 37]$ $[10, 37]$. In a typical preparation protocol, 2 g graphite (500 mesh), 1 g NaNO₃ and 50 mL concentrated H_2SO_4 were put into a 250-mL three-necked flask in an ice bath. After continuous stirring for 2 h, 7.3 g KMnO4, 7 mL hydrogen peroxide (30 wt%) and 150 mL DI water were added into the above reaction mixture. The resulting yellow–brown product was obtained after centrifugation and washed multiple times with 200 mL warm (nearly 40 °C) dilute HCl (3 wt%) and finally dried at 60 °C for 24 h in a vacuum oven. The as-prepared GO was then dispersed ultrasonically for 1 h in 500 mL deionized water at room temperature. The unexfoliated GO in the suspension was extracted by successive centrifugation at 4500 rpm for 30 min. The as-obtained solid was dried in vacuum at 60° C overnight and finally pulverized into powdered GO.

Scheme1 The synthesis of $TiO₂$ –rGO nanocomposites

Synthesis of $TiO₂-rGO$ nanocomposites

The synthesis of mesoporous $TiO₂$ –rGO composites is illustrated in Scheme 1. In a typical synthesis, different amounts of GO were dispersed in a mixed equimolar 60 mL solution of water and ethylene glycol (EG), followed by ultrasonication for 1 h. Then, Ti (SO_4) (2 mmol) was added to the GO dispersion under magnetic stirring and pursued by the addition of a small amount of ammonia solution $(100 \mu L)$. After stirring for 30 min, the resultant suspension was transferred to a 100-mL Teflon-lined stainless steel autoclave for solvothermal treatment for 12 h at 180 C. In this case, the EG was used as both reducing agent and solvent. The ammonia solution adjusts the alkalinity to assist the reduction of GO and help in the subsequent deposit of $TiO₂$ nanoparticles on the graphene. After the autoclave was cooled to room temperature, the resulting products were collected and thoroughly washed with DI water and ethanol for several times to eliminate undecorated $TiO₂$ nanoparticles, unreacted chemicals, and residual EG. Finally, the as-prepared nanocomposite was dried at 60° C under vacuum condition overnight before characterization. The proposed mass ratio of GO to TiO₂ was maintained as 0, 2.5, 5 and 10.0 wt% and the resultant final products are represented as TG0, TG2.5, TG5 and TG10%, respectively.

Material characterization

XRD investigation of the prepared photocatalysts was conducted at room temperature with a Rigaku D/MAX 2550 VB/PC apparatus using Cu Ka radiation $(\lambda = 1.5406 \text{ A})$ and a graphite monochromator, operated at 40 kV and 30 mA. Diffraction patterns were recorded in the angular range of $5^{\circ}-80^{\circ}$ with a step width of 0.02°. The ultraviolet–visible diffuse reflectance spectra (UV–Vis DRS) were obtained using a scan UV–Vis spectrophotometer (Varian Cary 500) equipped with an integrating sphere assembly, using $BaSO₄$ as the reflectance

sample. The spectra were obtained at STP within the range of 200–800 nm. The surface morphologies and particle sizes were examined by transmission electron microscopy (TEM; JEM2000EX), using an accelerating voltage of 200 kV. The chemical states of the nanocomposite were inspected by using X-ray photoelectron spectroscopy (XPS) with a Perkin-Elmer PHI 5000 C ESCA System with Al $K\alpha$ radiation functioning at 250 W. The shift of binding energy owing to relative surface charging was corrected using the C 1s level at 284.6 eV as an internal standard. The Fourier transform infrared (FT-IR) spectra were executed with KBr disks holding the powder sample with an FT-IR spectrometer (Nicolet Magna 550). Raman spectra were obtained using a micro-Raman spectrometer (Renishaw In Via) with 514 nm excitation light. The recombination of electron-holes in the samples were studied with a spectrofluorophotometer (Shimadzu, RF–5301) by using an Xe lamp as the excitation source at room temperature with the excitation wavelength of 320 nm, and the slit widths at the excitation and emission of the spectrofluorimeter were 5 and 2 nm, respectively. Brunauer–Emmett–Teller (BET) surface area measurements were conducted by N_2 adsorption at 77 K using an ASAP2020 instrument. All the samples were degassed at 180 $^{\circ}$ C prior to analysis and the BET surface area was verified by adsorption data in the relative pressure (p/p_0) range of 0.06–0.2.

Photocatalytic activity measurement

The photocatalytic performance was assessed by photodegradation of RhB and phenol solution under visible light irradiation. The photodegradation of RhB was carried out using a homemade photo-reactor furnished with 500-W halogen lamp and 420-nm cut-off glass optical filter, which used as a light source to provide visible light irradiation. In a typical photocatalytic activity experiment, 50 mg of the photocatalyst was added into 50 mL of the above stated 10 mg L^{-1} RhB solution. Before starting the light source, the reaction mixture was stirred for 30 min in the dark with the purpose of adsorption/desorption equilibrium between the dye and the photocatalyst. At a given time interval of visible light irradiation, a sample of 3 mL was withdrawn and then centrifuged to essentially separate all the photocatalyst. The concentrations of the remaining dye were spectrophotometrically examined by using a UV–Vis spectrophotometer (Shimadzu UV-2450) and evaluating the absorbance of the solutions at 554 nm in the course of the photodegradation process. The photocatalytic activity of phenol was evaluated using a 300-W Xe lamp under simulated solar light (with an AM1.5 air mass filter). Typically, the as-prepared photocatalyst (50 mg) was placed in a cylindrical Pyrex glass tube having 50 mL phenol aqueous solution (10 mg L^{-1}), which was first magnetically stirred in the dark for 30 min in order to accomplish adsorption/desorption equilibrium and then exposed to visible light. At given time intervals, 3-mL aliquots of the irradiated solution were sampled, centrifuged, and analyzed by using high-performance liquid chromatography.

Results and discussion

Characterization of $TiO₂-rGO$ nanocomposites

Transmission electron microscopy (TEM) examined the morphological and structural features of the synthesized composite. Figure 1a discloses that a large number of $TiO₂$ nanocrystals are highly dispersed on the rGO sheet and the majority of TiO₂ nanocrystals range from 5 to 10 nm. The size distribution curve of the TiO₂ nanocrystals demonstrates a mean diameter centered at \sim 8 nm (Fig. 1a, inset). Notably, intra-particle aggregation constitutes a mesoporous structure, which is further proven by the nitrogen adsorption–desorption isotherm. The HRTEM image (Fig. 1b) portrays a distinct crystalline lattice with a d spacing of 0.35 nm corresponding to the (101) plane of anatase $TiO₂$ [\[38](#page-14-0)].

Figure [2a](#page-6-0) demonstrates the XRD patterns of the bare $TiO₂$ and $TiO₂$ –rGO nanocomposites, As depicted from the figure, only anatase phase $TiO₂$ (JCPDS No. 21-1272) was observed in the bare $TiO₂$ sample (TG0%) and $TiO₂$ -rGO nanocomposites photocatalyst (samples TG2.5, TG5 and TG10%). Above all, no distinctive diffraction peaks of GO or graphene were detected in all the $TiO₂-rGO$ nanocomposites, signifying that the aggregation of graphene sheets through the reduction route can be efficiently barred. This can be attributed to the idea that the systematic stacking pattern of GO was shattered under the solvothermal reduction conditions and could not be resolved using the XRD pattern.

UV–Vis DRS has been recognized as an efficient optical characterization approach to investigate the probable modification in the band gap and interaction between $TiO₂$ and rGO as shown in Fig. [2b](#page-6-0). The $TiO₂$ -rGO nanocomposites show an enhanced red-shift in the absorption edge as compared to its bare $TiO₂$ nanoparticle (TG0%) counterparts which show typical and intense absorption in the UV region, resulting in electron transition from the valence band (VB) to the conduction band (CB). This broad circumstantial absorption of $TiO₂$ –rGO in visible

Fig. 1 a TEM images of the TiO₂–rGO nanocomposite (TG5%). Inset in (a) is the corresponding particle size distribution of the loaded $TiO₂$ nanoparticles derived from 100 of $TiO₂$ particles shown in (a). **b** HRTEM image showing nanosized $TiO₂$

Fig. 2 XRD pattern (a), UV–Visible absorption spectra (b) of samples $TG0\%$ (a), $TG5\%$ (b), $TG5\%$ (c) , TG10% (d)

light could be attributed to the existence of reduced GO in the as-prepared titania nanocomposites and this red-shift is helpful for the degradation of RhB dye and phenol. Nevertheless, any specific tendency or quantifiable variation in the shift of the absorption edge has not been seen. Zhang et al. also [[39\]](#page-14-0) affirmed similar results. According to the above results and the shift of the absorption edge of $TiO₂-rGO$ nanocomposites was probably attributed to the interaction between the $TiO₂$ nanocrystals and the reduced GO, and it is anticipated that the Ti–O–C bond between $TiO₂$ and rGO was formed which was also soundly established by XPS examination.

X-ray photoelectron spectroscopy (XPS) was used to investigate the interactions and successful functionalization of $TiO₂$ nanocrystals on the graphene sheet. Figure [3](#page-7-0)a, b demonstrate the high-resolution C 1s XPS spectra of GO and the $TiO₂$ rGO nanocomposite (sample TG 5%), respectively. The C 1s XPS spectrum of the GO nanosheets clearly reveals the existence of two kinds of carbon bonds. The peak positioned at 285 eV is typically assigned to the $Sp₂$ hybridized carbon of the graphene sheet while an other peak at 287.4 eV represented the carbonaceous band (C–OH), proposing a substantial degree of oxidation for the GO nanosheets and providing plenty of active sites for direct linking of the $TiO₂$ nanoparticles. After solvothermal treatment of the GO nanosheet in the $TiO₂$ -rGO nanocomposite, there is a decrease in the XPS peak intensity, suggesting the effective deoxygenation of the GO nanosheet and emphasizing that the GO nanosheet in the $TiO₂$ –GO scheme can be in situ reduced to produce $TiO₂$ – FGO nanocomposites.

Raman spectroscopy is an excellent non-destructive technique to describe the crystalline feature of carbon and provide an additional insight into the local order of the products. The substantial structural changes that occurred during the solvothermal reaction route from GO to graphene are also echoed in the Raman spectra. Figure [4](#page-7-0)a demonstrates a comparison of the Raman spectra of GO and the $TiO₂$ rGO nanocomposite (sample TG5%). In the case of TG5% (Fig. [4a](#page-7-0), spectrum b), numerous characteristic bands at 146, 397, 516 and 637 cm⁻¹ correspond to the Eg(1), B1 g (1), A1 g + B1 g(2) and Eg(2) modes of anatase, respectively There are two typical Raman bands assigned for GO (Fig. [4a](#page-7-0), spectrum a). The band at

Fig. 4 (a) Raman spectra of the GO (a) and (b) TG5% samples, (b) nitrogen adsorption–desorption isotherms and the corresponding pore-size distribution curves (inset) of the TG5%

about 1352 cm⁻¹ normally called the D-band and is assigned to disordered $Sp₂$ carbon. While the other band near 1587 cm^{-1} is adjacent to the well-ordered graphite domain and is frequently called the G-band $[40, 41]$ $[40, 41]$ $[40, 41]$ $[40, 41]$. For the TiO₂-rGO (TG5%) composite, with all Raman bands for anatase along with two distinctive peaks at nearly 1360 and 1604 cm^{-1} for the D- and G-bands, respectively, are also found, indicating the successful incorporation of graphene into the hybrids. The D/G intensity ratio of GO is 0.996, but this value rises to 1.06 in TiO₂-rGO (TG5%).

This rise in the intensity of D/G ratio results in a reduction in the average size of the sp₂ domains prepared through the solvothermal course of reaction $[41]$ $[41]$.

The specific surface area and porous structure of the prepared $TiO₂-rGO$ and $TiO₂$ nanocomposite were determined by the multi-point BET techniques. Figure [4](#page-7-0)b demonstrates the nitrogen adsorption–desorption isotherms and the corresponding pore size distribution curve (inset in Fig. [4](#page-7-0)b) for the TG5% sample, which can be regarded as type IV isotherm with distinct mesopores and H3 type hysteresis loop relating to the IUPAC classification [[42\]](#page-14-0). The predominant average pore size exists between 2 and 10 nm and all prepared $TiO₂$ –rGO nanocomposites possess a mesoporous texture. Table 1 displays that all $TiO₂-rGO$ nanocomposite samples have larger specific surface areas than pure $TiO₂$ and P25. The introduction of rGO in $TiO₂$ not only itself proliferates the BET surface area, but also the existing rGO decreases the crystal size of $TiO₂$, which ultimately results in the augmentation of surface area. This is due to the existence of graphene in the nanocomposites, which has an exceptionally high surface area (theoretical value of 2600 m^2 g⁻¹). But at a certain limit with rising contents of GO in the nanocomposite, an increasing trend of specific surface area became apparent and the graphene content reaching 10% then declining in the specific surface area were observed. This may possibly be caused by aggregation of GO, which happens at higher concentrations or $TiO₂$ nanoparticles, are coated by a graphene nanosheet. Consequently, when the surface area of $TiO₂$ rGO nanocomposites is calculated, its size is largely governed by surface properties of anatase $TiO₂$ nanoparticles.

FT-IR spectra of the as-prepared nanocomposites were recorded, as demonstrated in Fig. [5a](#page-9-0), spectrum a, GO was found to display numerous typical absorption bands of oxygen-containing groups. The absorption peaks were at 1052 cm^{-1} for C–O stretching, 1220 cm⁻¹ for phenolic C–OH stretching $[42]$ $[42]$, 1625 cm⁻¹ for the -OH groups of molecular H₂O [\[43](#page-14-0)] and 1729 cm⁻¹ for C=O stretching vibration of residual COOH groups [[44\]](#page-14-0). The minor peaks positioned at 2856 and 2920 cm^{-1} in the spectrum correspond to the $CH₂$ stretching vibration mode. The FT-IR results for the TiO₂–rGO nanocomposite (TG[5](#page-9-0)%) are shown in Fig. 5 (spectrum b). It is apparent that the intensities of oxygen-containing functional groups, for instance, C –O (1052 cm⁻¹), vividly diminished. The phenolic and carbonyl C=O bands at

Samples	$GO(wt\%)$			S_{BET} (m ² g ⁻¹) ^a Avg. pore size (nm) ^b Pore volume (cm ³ g ⁻¹) ^c
P ₂₅		39.0	27.9	0.13
TG 0%		111	10.8	0.19
TG 2.5%	2.5	121	8.9	0.23
TG 5%		147	6.6	0.24
TG 10%	10	125	8.1	0.25

Table 1 Parameters obtained from N_2 desorption isotherm measurement

^a BET specific surface area was calculated from the linear part of the corresponding BET plot

^b Average pore diameter was assessed from the BJH formula

^c BJH desorption cumulative pore volume between 1.7 and 300.0 nm diameter

Fig. 5 a FTIR spectra of the GO (a) and (b) TG5% samples, (b) comparison of the photoluminescence spectra of samples TG0% (a), TG2.5% (b), TG10% (c), TG5% (d)

1220 and 1729 cm⁻¹, respectively, also disappeared in the TiO₂-rGO (TG5%) nanocomposite. The spectra also display sturdy absorption bands at $400-900 \text{ cm}^{-1}$, signifying the existence of the Ti–O–Ti stretching vibration mode in TiO₂ [\[45](#page-14-0), [46\]](#page-14-0).

In a photocatalytic process, the photogenerated electron and hole recombination is one of the many traits mainly liable for the low photocatalytic efficiency. It is well documented that graphene is a viable candidate for the electron acceptor material owing to its two-dimensional *p*-conjugation edifice $[47]$ $[47]$, and has been efficiently applied to the $TiO₂$ –graphene scheme that retards the recombination of photogen-erated electrons in TiO₂ [[48,](#page-14-0) [49](#page-14-0)]. The photoluminescence (PL) emission spectra have been broadly used to explore the efficacy of radiative charge carrier recombination, migration and transfer, and to understand the lifespan of photogenerated charge carriers in the semiconductors [\[50](#page-14-0)]. Figure 5b exhibits the PL for pure $TiO₂ (TG0%)$ and well-adorned $TiO₂$ -nanocrystals on reduced GO nanocomposites with different mass ratios of GO. It is detected that the PL intensity of TG0% (bare $TiO₂$) diminished significantly after successful introduction of rGO. This noteworthy PL quenching is due to the effectual electron transfer from $TiO₂$ to graphene sheets. In addition, the uppermost PL quenching happens in TG5%, which suggests that there is an optimum amount of GO in order to obtain the paramount and efficient electron transportation. However, if the amount of GO is in excess up to 10% (TG10%), the PL quenching intensity increased instead of decreasing, because the number of active sites capturing the photo-induced electrons declined. Moreover, excessive graphene can wrap the surface of $TiO₂$ nanoparticles, leading to a drop-off in the concentration of photogenerated charge carriers and the photocatalytic activity of the photocatalyst.

Photocatalytic measurement of $TiO₂$ – FGO nanocomposites

Finally, we tested the photocatalytic performance of our mesoporous $TiO₂$ -rGO nanocomposites in the photodegradation of rhodamine B (RhB) and phenol under visible light irradiation and compared it with free $TiO₂$ nanocrystals synthesized by the same method in the absence of GO and P25 (a well-known commercial $TiO₂$ photocatalyst). It is well recognized that RhB and phenol are important chemicals and have been extensively used in industrial sectors, and so are often found as notorious environmental contaminants. Therefore, in the current work, we selected them as model organic substances to assess the photocatalytic performance of the as-prepared $TiO₂$ -rGO nanocomposites photocatalyst. Prior to photodegradation experiment, the dye and phenol solution with the photocatalyst were magnetically stirred in the dark for 30 min to ensure the absorption/desorption equilibrium of the pollutants with the photocatalyst. Figure 6 exhibits the photodegradation efficiencies and temporal evaluation of the RhB and phenol concentration in the presence of different catalysts under visible light irradiation. The photocatalytic activity of

Fig. 6 Photodegradation of RhB under visible light irradiation (a); photodegradation of phenol under simulated solar light irradiation (b); fractional removal of RhB under visible light illumination for 90 min (c); fractional removal of phenol under simulated solar light illumination for 150 min (d)

commercial TiO₂ (P25) was investigated for the comparison purpose in the study of the degradation of RhB under the same reaction conditions. Interestingly, even though the photocatalytic activity of pure $TiO₂$ nanoparticles (TG0%) was lower than P25, all the three $TiO₂$ -rGO nanocomposites still exhibited the remarkably higher photocatalytic activity as compared to P25 and its bare $TiO₂$ counterparts. The bare $TiO₂$ also showed some photocatalytic activity, which is possibly due to the mesoporous structure with a surface area of 111 m² g^{-1} .

As is depicted from Fig. [6](#page-10-0), the GO contents have a substantial effect on the photocatalytic activity of $TiO₂$ for the degradation of RhB and phenol. The sample $TG0\%$ (bare $TiO₂$) displays a low photocatalytic activity due to the rapid recombination between CB electrons and VB holes in pure $TiO₂$. However, photodegradation was enhanced remarkably after the introduction of graphene sheet in $TiO₂$ -rGO nanocomposite photocatalysts. The photocatalytic activity increased with increasing contents of GO. However, at a certain limit increase in GO contents favored the enhanced photocatalytic activity and a further increase in GO contents especially resulted in the decline of the photocatalytic activity, in the present system when the graphene contents reached to $10.0 \text{ wt}\%$, a slight decrease in photocatalytic activity has been observed. The interpretation for the reduced GO contents dependency of photocatalytic performance of the $TiO₂-rGO$ nanocomposites and the relative decrease in photocatalytic activity is perhaps attributable to the intensification in the opaqueness and light scattering of TG10%. The loading of a higher concentration of graphene resulted in shielding the $TiO₂$ from absorbing visible light and leading to rapid diminution of irradiation through the reaction suspension. Moreover, certain active sites of the photocatalyst $(TiO₂)$ was also detrimentally occupied by reduced GO, which leads to a lessening in the number of potential reaction sites of the photocatalyst, resulting in a subsequent decrease in photocatalytic activity. The photocatalytic efficiency is at a maximum at the optimum reduced GO contents in the photocatalyst. Therefore, the nanocomposite of TG5% demonstrates the most excellent photocatalytic activity for the degradation of RhB. After 90 min of visible light exposure, 85% of RhB is photocatalytically degraded by the $TiO₂$ -rGO photocatalyst. Additionally, photodegradation of phenol has also been achieved to further uncover the photocatalytic activity and universality of the as-prepared $TiO₂$ –rGO nanocomposites (Fig. [6b](#page-10-0)). In comparison with bare $TiO₂$, the $TiO₂$ -rGO nanocomposites reveal a remarkable photodegradation rate of phenol and, after 150 min exposure under the simulated solar light irradiation, 100% degradation of phenol took place, signifying that the assynthesized $TiO₂$ -rGO nanocomposites embrace not only an enhanced photocatalytic activity but also a relative generality for the photodegradation of various pollutants. This experimental observation not only understandably reveal that phenol was degraded by the catalyst but also further strengthen the finding that degradation of RhB happens because of the excitation of TiO2-reduced GO rather than merely the photoexcitation of the dye [\[39](#page-14-0), [51](#page-14-0)].

The enhanced photocatalytic activity could be largely credited to the synergistic coupling outcomes in the nanocomposites. The ultra-small $TiO₂$ nanocrystals can curtail the radial length for the diffusion of radiative electrons and holes to the $TiO₂$ surface, and the ultra-dispersed nature of $TiO₂$ nanocrystals on graphene not only improve the intimate interfacial contact but also expedite the charge separation and transfer of electrons from $TiO₂$ to the graphene nanosheet. Graphene has been described as a viable and efficient acceptor material owing to its two-dimensional π conjugation structure, and the excited electrons of $TiO₂$ could transfer from the conduction band to the graphene. Therefore, the swift transport of charge carriers could be attained and, subsequently, improved charge transportation and separation of photogenerated charge carriers is accomplished. Besides the efficient charge separation and transportation, the enhanced light absorption intensity and light absorption range of the nanocomposites resulted in the more proficient utilization of the solar spectrum and exhibited noteworthy improvements in the photodegradation of the pollutant. This visible light photocatalytic activity of the catalyst is probably helpful to expedite its use in practical environmental remediation. Another possible contributory factor is that the hybrid material has a mesoporous structure with good absorptivity for pollutants and enlarged BET surface area $(147 \text{ m}^2 \text{ g}^{-1})$ that offer more potential reaction sites for RhB dye and phenol adsorption and successive redox reactions. It is well acknowledged that samples with enlarged surface areas readily exhibit enhanced photocatalytic competency, predominantly owing to the abundant exposed active sites and large light harvesting.

Figure 7 demonstrates the anticipated mechanism of RhB dye degradation. As the photocatalyst becomes excited in the visible light irradiation, the transfer of electrons from the valance band to the conduction band of $TiO₂$ takes place where reduced GO can behave as an electron trap and photogenerated electrons are then transported to rGO. These electrons then react with the surface oxygen resulting in the formation of successive active oxygen radicals such as O_2 ⁻, which lead to a decreased rate of electrons and holes recombination. A reduced GO matrix improves the charge separation and hence the holes are made available for the formation of hydroxyl (OH) radicals. The consequential free radicals generated then play a significant role in the degradation of the dye molecule. The main reactions involved in all this process are shown below:

$$
TiO_2/rGO \to TiO_2(h^+) + rGO(e^-)
$$
 (1)

Fig. 7 Schematic of proposed mechanism of rhodamine B dye degradation

$$
rGO(e^-) + O_2 \rightarrow rGO + O_{2} \tag{2}
$$

$$
TiO2(h+) + H2O/OH- \rightarrow TiO2 + OH
$$
 (3)

$$
OH + RhB \rightarrow degradation products \qquad (4)
$$

Conclusions

In summary, we have devised a simple, scalable and environmentally benign onestep solvothermal method for producing highly dispersed $TiO₂$ nanocrystals on graphene with a small particle size (\sim 5–10 nm). Titanium sulfate is used as TiO₂ source with the assistance of ultra-low content of ammonia to accomplish in situ growth of mesoporous $TiO₂$ nanocrystals on graphene sheets. Graphene has proven to be a dynamic cocatalyst to improve the photocatalytic activity of $TiO₂$ nanoparticles, and its contents have a substantial impact on the photoactivity of $TiO₂$. Notably, the as-synthesized nanocomposites with an appealing nanoarchitecture have revealed a mesoporous structure and effective charge separation with enhanced visible light absorption, which have bestowed the nanocomposites with admirable photocatalytic performance for RhB dye and phenol degradation. We anticipate that the work presented here will open up a scalable and cost-effective approach to prepare robust materials for organic dye degradation and deleterious phenol oxidation for the purification of polluted water resources.

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