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Grass-legume mixtures impact soil N, species recruitment, and productivity in temperate steppe grassland

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Abstract

Aims In this study, we aimed to understand effects and mechanisms of legume species and their relative abundance on soil N, species recruitment, and productivity in mixed grassland. We also assessed the utilization prospect of several legumes in natural grassland.

Methods In 2006, grass-legume combinations (GLCs) with different ratios (GLR) were established using Leymus chinensis and four legume species (Medicago ruthenica, Lespedeza daurica, Medicago falcata, and Medicago sativa). In 2009, plant and soil samples were taken to examine the differences in soil nitrogen (N) properties, biological N fixation (BNF), recruitment characteristics, and aboveground biomass of the community under different mixture patterns.

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Results Soil total N and available N concentration increased when legume component increased from GLR 1:0 (grass monoculture) to GLR 1:1 (grass:legume 1:1) but decreased as legume component increased further from GLR 1:3 (grass:legume 1:3) to GLR 0:1 (legume monoculture). GLR 1:1 had the highest BNF in most GLCs except for L. chinensis-M. falcata. For any GLR with legume, *L. chinensis-M. sativa* combination (L-MS) had higher soil total N, available N, water content, and BNF. The legume combinations with a GLR lower than 1:1 resulted in more species recruitments compared with the grass monoculture, and L-MS induced relatively more species recruitments than other GLCs, which were correlated with improved soil water and $NO₃⁻-N$ status. The grassland productivity increased as the GLR decreased; in any GLRs with legume, higher productivity was found under L-MS combination, and current results showed that grassland productivity was positively linked to soil $NO₃⁻$. N availability influenced by grass-legume mixture. Conclusions We concluded that the species and relative abundance of legumes had great impact on soil N status, species recruitment, and productivity in this temperate grassland ecosystem. In this study site, legume introduction can be considered as an alternative to N fertili-

zation for increasing grassland productivity. M. sativa has the greatest economic and ecological potential species to be mixed into natural temperate steppe grassland, and grass-legume ratio 1:1 is the best combination in mixed communities.

Keywords Legume \cdot N fertilization \cdot N fixation \cdot Species recruitment . Productivity. Facilitation effect

Introduction

As a key element to support plant growth and build proteins, nitrogen (N) is a primary limiting factor for forage production and quality in grazing ecosystems (LeBauer and Treseder [2008](#page-13-0); Karki et al. [2009](#page-13-0)). Traditionally, N fertilization has frequently been used to increase productivity and improve forage quality in grasslands (Schellberg et al. [1999](#page-14-0)). However, chemical fertilizers require a large monetary investment and do not benefit all grassland species due to interspecific differences in nutrient capture and utilization efficiency. Thus, chronic or intensive N fertilization can cause loss of plant species diversity, which maintains grassland ecosystem structure and functional stability (Suding et al. [2005;](#page-14-0) Clark and Tilman [2008;](#page-13-0) Bai et al. [2010\)](#page-13-0). Historical and experimental evidence shows a trend of declining species diversity with N deposition and artificial N addition in native grassland ecosystems (Wedin and Tilman [1996](#page-14-0); Gough et al. [2000](#page-13-0)).

Legumes are generally considered high quality forage due to their low fiber and high protein content (Mortenson et al. [2004\)](#page-14-0). More importantly, it is wellknown that legumes can increase N availability in soils by fixing atmospheric N in an available form (Spehn et al. [2002](#page-14-0)). The fixed N can be transferred to coexisting plants for utilization through root exudation (Lory et al. [1992;](#page-14-0) Paynel et al. [2008\)](#page-14-0) and decomposition of dead roots (Trannin et al. [2000\)](#page-14-0). Compared with N fertilization, N fixation from legumes can provide a more slow-released and persistent N source. This may reduce the competitive advantage of more efficient N consuming species and thus be beneficial for species coexistence. In addition, many legumes have welldeveloped shoot and root systems, which may maintain soil water content by increasing canopy shade and water retention (Armstrong et al. [1999\)](#page-13-0). Moreover, legumes decompose quickly due to their high litter quality, which accelerates nutrient cycling and can improve soil productivity (Thomas and Asakawa [1993](#page-14-0); Spehn et al. [2002;](#page-14-0) Milcu et al. [2008](#page-14-0)). Improved soil fertility can provide more niches for other plants to establish, which can be advantageous for maintaining higher species diversity and increasing resource complementarity (van Ruijven and Berendse [2005](#page-14-0)). Therefore, legumes can have huge economic and ecological potential as N donors to native grasslands.

Currently, inter-seeding legumes in native grasslands and establishing grass-legume mixed grasslands have been widely adopted to overcome the forage N deficit in grazing systems. Research across many sites has revealed that the introduction of legumes into grasslands leads to more soil N accumulation and higher productivity (Sleugh et al. [2000;](#page-14-0) Mortenson et al. [2004,](#page-14-0) [2005;](#page-14-0) Rao et al. [2007\)](#page-14-0). These benefits from grass-legume mixtures depend on the legume species (Spehn et al. [2002](#page-14-0); Fornara and Tilman [2008](#page-13-0)) and the trade-off with competition and facilitation effects for limiting resources between legumes and other plants (Lithourgidis et al. [2006\)](#page-14-0). The introduction of different legume species and their relative abundance in mixed grasslands may influence soil N properties and community productivity due to the differences in N fixation and competitive effects of different species and mixtures. Kirwan et al. [\(2007\)](#page-13-0) showed that grasslegume mixtures improved productivity greatly compared to grass only monocultures and that the effect of mixtures depended on the relative abundance of legumes based on results from 28 sites in European grassland. Nyfeler et al. ([2009\)](#page-14-0) found that mixtures fertilized with 50 kg N ha⁻¹ year⁻¹ produced equivalent forage yields to grass monocultures fertilized with $450 \text{ kg N} \text{ ha}^{-1} \text{ year}^{-1}$ if the legume proportion was about 50 to 70 %. However, previous studies on mixed grasslands either had artificially controlled community structure (Spehn et al. [2002](#page-14-0); Fornara and Tilman [2008](#page-13-0)) or only considered the effect of legumes on soil N status and productivity (Mortenson et al. [2004](#page-14-0), [2005;](#page-14-0) Lithourgidis et al. [2006\)](#page-14-0). Few studies revealed effects of legume presence, in particular legume species and their relative abundance, on species recruitment, coexistence, and subsequent species diversity in mixed grasslands (Carino and Daehler [2002\)](#page-13-0), although this information would be very useful for introducing legumes as a resource in native grasslands.

Therefore, in this study, we aimed to a) understand effects and mechanisms of legume species and their relative density in grass-legume mixed grassland on soil N, species recruitment, and productivity and b) assess the utilization prospect of several legumes in natural grasslands. Our hypotheses were as follows: a) soil N and water status would be improved with increased legume density in mixed grasslands, b) increased relative density of legumes in mixed grasslands could promote species recruitment and enhance productivity through improving soil N and water status, and c) the benefit of grass-legume mixtures would depend on the legume species included in the mixture.

Materials and methods

Study site

This study was conducted at the Grassland Farming Research Station (E123°31′, N44°33′, elevation 145 m) of the Northeast Institute of Geography and Agroecology, Chinese Academy of Sciences, which is located at the eastern edge of the Eurasian steppe and is characterized as Eurasian continental temperate grassland. The site has a semi-arid continental climate. Mean annual temperature is 4.9 °C; annual precipitation is approximately 410 mm, with 70 % falling from June to September. The soil type is meadow chernozem soil. Soil pH is 8.1 with 16 g kg^{-1} organic matter and 1.1 g kg⁻¹ total N in the 0–30 cm depth. The climax vegetation type is meadow steppe dominated by Leymus *chinensis*, a perennial C_3 rhizome grass, which accounts for 80 % of the aboveground biomass. However, from 1980s, part of the L. chinensis meadow had been reclaimed as cropland due to the increasing economic benefit from grain production. Our experiment was conducted in a section of the meadow that was converted to cropland in 1995, but subsequently abandoned in 2002.

Field experiment design

The experiment was a split-plot design with four grasslegume combinations (GLCs) as main plots and five grass-legume mixture ratios as sub-plots randomized allocated in the main plots, with four replicates. The four GLCs were L. chinensis-Medicago ruthenica (L-MR), L. chinensis-Lespedeza daurica (L-LD), L. chinensis-Medicago falcata (L-MF), and L. chinensis-Medicago sativa (L-MS). Under each L. chinensis-legume combination, total plant density was 600 plant individuals per square meter, which reflected the mean plant density of the natural L. chinensis grassland in this region. The grass-legume ratios (GLRs) in the mixtures were 1:0 (grass monoculture), 3:1 (3 grass:1 legume), 1:1 (1 grass:1 legume), 1:3 (1 grass:3 legume), and 0:1 (legume monoculture), respectively, by altering the sowing rate and thinning the seedlings later.

On 5 July 2006, after vegetation was removed and the soil was turned over at a depth of 0–30 cm, all 80 plots with 3×3 m in size were sown with the proper number of L. chinensis and legume seeds adjusted with actual seed germination rate. In each plot, all sown seeds were mixed then uniformly broadcasted into plots in strips with row space of 15 cm. Before sowing, all legume seeds were soaked in 98 $\%$ H₂SO₄ for 0.5 h to soften seed coats and ensure viable germination. No inoculant was applied when sown assuming the soil contains enough rhizobia as the site had history of legume growing which was confirmed by the pot experiment described below. Plots were irrigated when necessary to ensure emergence and survival of the sown species. On 5 August 2006, we thinned the plots to preserve the targeted plant species densities and GLRs. From seeding date to June 2007, plots were kept weedfree by hand-weeding to ensure successful establishment of L. chinensis and the sown legumes and to eliminate the impact of the original soil seed bank on species recruitment.

Pot experiment for legume nodulation

Nodulation of the four legume species were confirmed by growing them in pots with soil from the same field as the experimental grassland. Soil at 0–40 cm depth was collected in field experiment site and then to be passed through a 2-mm sieve. A 15-kg sieved soil was filled into a 30×35 cm (diameter \times height) pot, with a total of 24 pots. Each legume species was planted into randomly selected six plots with 40 acid treatment seeds in each pot. Pots were randomly placed in greenhouse, and their positions were exchanged every 3 days. For keeping soil moisture, equal water was sprayed in each pot every 3 days if needed. Emerged seedlings were thinned to retain 15 plants in each pot. At full flowering, the soil in each pot was washed to harvest shoot and root of each plant. Total and effective (pink or light pink) nodule number were counted, and nodule fresh weight was determined in each pot.

Plant sampling and analysis

In early August 2009, a 1×1 m quadrat was arranged in the center of each plot, and plants were counted and identified to species in each quadrat. The aboveground part of each species was clipped at the soil surface and taken back to the laboratory in paper bags. All newly present species in the plots that were not from the initially sown species pool were defined as recruited species. Plant materials were oven-dried at 65 °C for 48 h and weighed to determine the dry biomass. Shoot samples of sown legumes and non-legume species L. chinensis, Hierochloe glahra, Setaria viridis, and Kalimeris integrifoliain in grass monocultures were finely ground to determine the biological N fixation (BNF) rate using $15N$ natural abundance technique (Unkovich et al. [2008](#page-14-0)). BNF from sown legumes was estimated using the 15 N natural abundance method. The percent of N derived from the atmosphere (%Ndfa) in legume biomass was estimated using the following formula (Unkovich et al. [2008](#page-14-0)):

$$
\% \text{Ndfa} = 100 \times \left(\frac{\delta^{15} \text{N reference plant-} \delta^{15} \text{N legume}}{\delta^{15} \text{N reference plant-B}} \right) (1)
$$

To avoid possible N transfer between legumes and non-legumes species, we used L. chinensis, H. glahra, S. viridis, and K. integrifolia in grass monocultures as the reference plants. These four species were all common species in mixture and grass monoculture communities. They represent three function groups including C_3 grass, C_4 grass, and forb and likely reflect different N uptake characteristics. The B value for legume species were determined by growing them under N-free medium (sand+perlite+nutrient solution) inoculated with a soil suspension from experiment site in temperaturecontrolled glasshouse and harvesting shoot material for ¹⁵N analysis at a stage of growth that is same as our field analyses (Unkovich et al. [2008](#page-14-0)). Total BNF was calculated based on % N from fixation, legume shoot biomass, and shoot N concentration. Total N was analyzed using the continuous flow Isotope Ratio Mass Spectrometer (Stable Isotope Facility, UC Davis).

Soil sampling and analysis

After all plant material was collected, we used an auger (4 cm in diameter) to collect three soil samples at depths of 0–40 cm in each quadrat. These soil samples were sealed into three previously weighed aluminum containers and then taken back to laboratory to determine their fresh weight and then the dry weight after being oven-dried at 105 °C for 48 h. Soil water content was calculated based on the soil's fresh and dry weight. Three more soil cores were collected and bulked at depths of 0–40 cm in each quadrat. With visible plant materials and other debris removed, part of mixed soil samples were frozen at 0° C, and the remaining samples were air-dried in the dark and then ground to pass through a 0.2-mm sieve for subsequent analysis. NH_4^+ and NO_3 [–] were extracted with 2 M KCl from frozen soil

samples (Miller and Keeney [1982\)](#page-14-0), and their concentrations were determined by a Bran-Luebbe AA3 autoanalyzer (Bran and Luebbe, Hamburg, Germany). Soil total nitrogen (TN) was measured by the Kjeldahl method using air-dried soil samples (Sparks et al. [1996\)](#page-14-0). All soil nutrient traits were calculated based on units of dry soil weight.

Data analysis

All data were assessed to verify model assumptions of normality and equality of variance. Three-way ANOVA based on a split-plot design was conducted to determine the main and interactive effects of GLC and GLR with block as random factor. Means comparisons were conducted by one-way ANOVA, followed by Duncan's HSD test. Stepwise multiple regression analysis was conducted to examine the dependence of species recruitment on soil N availability and water content, with species recruitment characteristics as dependent variables and soil NH_4^+ -N, NO_3^- -N, and water content as independent variables. A similar stepwise multiple regression analysis was performed to examine the dependency of aboveground biomass (AGB) on species number, soil N availability, and water content. Significance for all statistical tests were evaluated at $P=0.05$. All data analyses were conducted with the SPSS 16.0 software (Chicago, IL, USA).

Results

Soil N and water properties

Both the GLC and GLR had significant impact on the soil available N, NO_3 ⁻-N, and water content. In addition, GLR had a significant effect on soil NH_4^+ -N and total N (Table [1\)](#page-4-0). Across all GLCs, the soil total N, available N, and its components at the depth of 0– 40 cm decreased in the order: GLR 1:1>GLR 1:3> GLR 0:1>GLR 3:1>GLR 1:0. Soil available N and its components had significantly higher values under GLR 1:1 compared with the GLR 1:0, GLR 3:1, and GLR 0:1 treatments (Table [2](#page-5-0)). In all GLCs, the mixture treatments had higher soil water content than monoculture, particularly GLR 3:1, under which soil water content was significantly higher than both grass and legume monocultures (Table [2](#page-5-0)). For any GLR treatment with legumes, the highest values of soil total N, available N,

Table 1 The effects $(P \text{ value})$ of grass-legume combination (GLC), grass-legume ratio (GLR), and their interaction on measured variables by general lineal model procedure based on a splitplot design

	Block GLC		GLR	$GLC \times GLR$
Soil total N	0.385	0.284	0.002	0.999
Soil available N	0.253	0.002	< 0.0001	0.361
Soil NH_4^+ -N	0.390	0.080	< 0.0001	0.901
Soil NO_3 ⁻ -N	0.314	0.001	< 0.0001	0.218
Soil water content	0.396	0.008	< 0.0001	0.752
Mean %Ndfa	0.269	0.017	< 0.0001	0.082
Mean BNF	0.423	< 0.0001	< 0.0001	0.059
Plant density of sown legume	0.796	0.346	< 0.0001	0.999
Plant density $of L$ chinensis	0.492	0.503	< 0.0001	0.991
Aboveground biomass of sown legume	0.539	0.007	< 0.0001	0.736
Aboveground biomass of L. chinensis	0.962	< 0.0001	< 0.0001	0.112
Number of recruited species	0.558	< 0.0001	< 0.0001	0.025
Plant density of recruited species	0.255	0.002	< 0.0001	0.251
Aboveground biomass of recruited species	0.486	0.001	< 0.0001	0.007
Aboveground biomass of community	0.734	< 0.0001	< 0.0001	0.334

and water content were found in L-MS and L-MR combinations, and the lowest values in L-MF combination (Table [2](#page-5-0)).

Legume nodulation

In pot experiment, the MS had significant higher total and effective nodule number compared with other three legume species (Fig. [1a](#page-6-0)). The total nodule number was 29 %, 45 %, and 90 % higher in MS than MR, LD, and MF, respectively, with the effective nodule number being 27 %, 48 %, and 103 % higher in MS than MR, LD, and MF, respectively. The fresh weight of total and effective nodule showed similar interspecific difference with nodule number (Fig. [1b\)](#page-6-0). However, there was no significant difference in total plant biomass between four legume species (Fig. [1c](#page-6-0)), indicating that the enhanced nodulation in MS compared to the other legumes was specific nodulation response but did not correlate with enhanced growth.

Shoot $\delta^{15}N$ of legumes and BNF

The δ^{15} N values of legume species increased with enhanced proportion of legumes in the mixtures for all GLCs. The $\delta^{15}N$ values of *L. chinensis* in mixture communities were significantly lower than grass monoculture with the lowest values in GLR 1:1, indicating that L . *chinensis* taken up N transferred from legumes when grown together with legumes, and more N was transferred from legume to L. chinensis under GLR 1:1 compared to other GLRs. The %Ndfa and BNF of legumes showed similar trend for GLC and GLR with different reference species. Compared with reference species, L. chinensis and K. integrifolia had relatively higher estimate of %Ndfa and BNF compared with H. glahra and S. viridis (Table [3](#page-7-0)). Increasing relative legume density significantly decreased %Ndfa of legumes. Averaged across GLC, the GLR 1:1 fixed significantly more nitrogen than GLR 3:1 treatments, and GLR 1:1 had the highest BNF compared to other GLRs in most GLCs except L-MF. For any given GLR, L-MS combination fixed more atmosphere N, especially compared to L-MF; L-MS combination had significant higher BNF (Tables 1 and [3](#page-7-0)).

L. chinensis and sown legumes

The plant density of *L. chinensis* significantly declined with decreased GLR, while the plant density of sown legumes significantly increased in all communities (Tables 1 and [4\)](#page-8-0). Moreover, the plant density of L. chinensis significantly increased under GLR 1:0, 3:1, and 1:1 treatments compared with initial sowing density in all L. chinensis-legume combinations (all $P<0.05$; Table [4](#page-8-0)). The aboveground biomass of L. chinensis and sown legumes also depended on GLR and GLC (Table 1). The AGB of L. chinensis significantly declined in communities as GLR decreased, in contrast to a statistically significant increase of AGB of sown legume species (Table [4](#page-8-0)). Under mixed treatments, the aboveground biomass of L. chinensis was higher in L-MS and L-MR than L-LD and L-MF combinations.

Species recruitment

GLC, GLR, and their interaction had statistically significant impacts on species recruitment as measured by the number of recruited species and their aboveground

Table 2 The soil N and water properties under different L. chinensis-legume combinations and L. chinensis-legume ratios

	GLC	GLR				
		1:0	3:1	1:1	1:3	0:1
Soil total N $(g kg^{-1})$	L-MR	1.14 ± 0.04^{Aa}	1.23 ± 0.06^{Aa}	1.38 ± 0.10^{Aa}	1.32 ± 0.05^{Aa}	1.26 ± 0.10^{Aa}
	L-LD	1.19 ± 0.04 ^{Aa}	1.20 ± 0.04 ^{Aa}	1.31 ± 0.05 ^{Aa}	1.26 ± 0.07 ^{Aa}	1.22 ± 0.07 ^{Aa}
	L-MF	1.15 ± 0.04^{Aa}	1.17 ± 0.06 ^{Aa}	1.28 ± 0.05^{Aa}	1.24 ± 0.09 ^{Aa}	1.20 ± 0.08^{Aa}
	L-MS	1.18 ± 0.03^{Ab}	1.23 ± 0.05 ^{Aab}	1.41 ± 0.07 ^{Aa}	1.35 ± 0.08 ^{Aab}	1.29 ± 0.09 ^{Aab}
Soil available N (mg kg^{-1})	L-MR	21.41 ± 0.17^{Ad}	22.69 ± 0.14 ^{Ac}	24.65 ± 0.15 ^{ABa}	23.75 ± 0.24 ^{ABb}	22.84 ± 0.31 ^{Ac}
	L-LD	$21.40 \pm 0.17^{\rm Ac}$	22.48 ± 0.18^{Ab}	23.91 ± 0.50 ^{ABa}	$23.17 \pm 0.17^{\text{BCab}}$	22.60 ± 0.23 ^{ABb}
	L-MF	21.32 ± 0.44^{Ab}	21.93 ± 0.20^{Bb}	23.63 ± 0.43 ^{Ba}	22.98 ± 0.25 ^{Ca}	22.00 ± 0.20^{Bb}
	L-MS	21.33 ± 0.30 ^{Ad}	22.74 ± 0.16^{Ac}	25.04 ± 0.23 ^{Aa}	24.29 ± 0.25^{Ab}	23.06 ± 0.24 ^{Ac}
Soil NH_4^+ -N (mg kg ⁻¹)	L-MR	3.43 ± 0.08 ^{Ac}	3.56 ± 0.07 ^{Ac}	4.13 ± 0.08 ^{Aa}	3.87 ± 0.07^{Ab}	3.56 ± 0.08 ^{Ac}
	L-LD	3.42 ± 0.06^{Ab}	3.52 ± 0.08^{Ab}	$3.90{\pm}0.07^\text{ABa}$	3.79 ± 0.07 ^{Aa}	3.53 ± 0.06^{Ab}
	$L-MF$	3.42 ± 0.06^{Ab}	3.50 ± 0.05^{Ab}	3.86 ± 0.09 ^{Ba}	$3.76{\pm}0.07^{\mathrm{Aa}}$	3.50 ± 0.07^{Ab}
	L-MS	3.46 ± 0.08^{Ab}	3.52 ± 0.06^{Ab}	4.07 ± 0.07 ^{ABa}	3.95 ± 0.09 ^{Aa}	$3.62{\pm}0.06^{\text{Ab}}$
Soil $NO3--N (mg kg-1)$	L-MR	$17.98{\pm}0.09^{\text{Ad}}$	19.12 ± 0.07 ^{Ac}	20.52 ± 0.10^{ABa}	19.89 ± 0.18^{Ab}	19.28 ± 0.24 ^{Ac}
	L-LD	17.98 ± 0.11 ^{Ac}	18.96 ± 0.16^{Ab}	20.01 ± 0.43 ^{Ba}	19.38 ± 0.10^{Bab}	19.07 ± 0.17 ^{ABb}
	L-MF	17.90 ± 0.39 ^{Ac}	18.43 ± 0.17 ^{Bbc}	$19.77 \pm 0.35^{\text{Ba}}$	19.22 ± 0.18 ^{Bab}	$18.50 \pm 0.14^{\rm Bbc}$
	L-MS	17.87 ± 0.23 ^{Ad}	19.22 ± 0.10^{Ac}	20.97 ± 0.16 ^{Aa}	20.34 ± 0.18^{Ab}	19.44 ± 0.19 ^{Ac}
Soil water content (%)	L-MR	8.29 ± 0.12^{Ab}	9.06 ± 0.11 ^{ABa}	8.95 ± 0.06 ^{Aa}	8.78 ± 0.12 ^{Aa}	8.43 ± 0.07^{Ab}
	L-LD	8.33 ± 0.07^{Ab}	8.77 ± 0.16 ^{ABa}	8.84 ± 0.08 ^{ABa}	8.61 ± 0.22 ^{Aab}	$8.31\!\pm\!0.07^{\text{Ab}}$
	L-MF	8.33 ± 0.04^{Abc}	8.69 ± 0.14^{Ba}	8.65 ± 0.08 ^{Bab}	8.51 ± 0.17 ^{Aabc}	$8.24\!\pm\!0.03^{\rm Ac}$
	L-MS	8.33 ± 0.08^{Ab}	9.14 ± 0.11 ^{Aa}	9.08 ± 0.08 ^{Aa}	8.87 ± 0.14 ^{Aa}	8.44 ± 0.07^{Ab}

Values are shown as means±SE

GLC grass-legume combination, GLR grass-legume ratio, L-MR L. chinensis-Medicago ruthenica, L-LD L. chinensis-Lespedeza daurica, L-MF L. chinensis-Medicago falcata, L-MS L. chinensis-Medicago sativa

a,b,c Different lowercase letters for each line indicate significant differences $(P<0.05)$ between grass-legume ratios

 A,B,C Different uppercase letters in each column indicate significant differences ($P < 0.05$) between grass-legume combinations

biomass (Table [1](#page-4-0)). The recruitment of legume species only occurred in communities with initial GLR≤1, and the total plant density of recruited legumes was less than 4 plants m−² (Table [5](#page-9-0)). All recruited species are commonly found in the natural steppe grassland, but none of the species recruited were dominated species. For all GLCs, the number of recruited species significantly increased from treatment with GLR 1:0 to GLR 1:1 and then significantly decreased from GLR 1:1 to GLR 0:1 (Table [6\)](#page-10-0). The plant density and AGB of recruited species were significantly higher under GLR 3:1 and GLR 1:1 compared with other GLR treatments in most GLCs excluding L-MF (Table [6](#page-10-0)). For all GLR treatments with legumes, plant density and AGB of recruited species were lower in the L-MF combination compared with other combinations (Table [6\)](#page-10-0). Stepwise multiple regression analysis showed that the number of recruited species positively correlated to soil water and $NO₃⁻-N$ content. The combination of soil water and $NO₃⁻-N$ content explained the 44.5 % variation in the number of recruited species (Table [7](#page-10-0)). The plant density and AGB of recruited species only had a significant correlation with soil water content, which explained 35.8 % and 39.5 % of the variations in plant density and AGB of recruited species, respectively (Table [7](#page-10-0)).

Aboveground biomass of the community

Aboveground biomass (AGB) varied greatly under dif-ferent GLC and GLR (Table [1\)](#page-4-0). Across all *L. chinensis*legume combinations, the AGB significantly increased with decreased GLR. However, no significant difference in AGB was found between GLR 1:1 and 0:1 (Fig. [2\)](#page-11-0). In all GLR treatments with legumes, AGB of the

Fig. 1 Nodule number (a) and fresh weight (b) of four legume species and plant biomass (c) of four legume species in plots experiment. Values are shown as means+SE. Different lowercase letters above bars indicate significant differences of effective

community decreased in the following order: L-MS>L-LD>L-MR>L-MF (Fig. [2](#page-11-0)). AGB was positively correlated to soil $NO₃⁻-N$ content, which explained 44.3 % variations in AGB of the community (Table [7](#page-10-0)).

Discussion

Effects of grass-legume ratios on soil N

Until now, the impact of legumes proportion on soil N in a mixed temperate steppe grassland was not well-documented, even though several studies examined effects of legume proportion on N transfer, productivity, soil moisture, and soil carbon flux (Nyfeler et al. [2009,](#page-14-0) [2011](#page-14-0); Zhang et al. [2013](#page-14-0)). Our results showed that soil total N and available N increased with GLR ranging from 1:0 to 1:1 and then decreased as the proportion of legumes further increased. This indicates that effects of grasslegumes mixtures on soil N depend on the relative abundance of legumes in this grassland, and that GLR 1:1 was the most ideal density relationship between L. chinensis and legumes for improving soil N status.

nodule $(P<0.05)$ between legume species. Different *uppercase* letters above bars indicate significant differences of total nodule $(P<0.05)$ between legume species in Fig. 1a, b. NS no significant difference

The difference in BNF is the major driving factor for the changed soil N status under the same site condition and fertilization management. Our results showed that BNF and soil N had similar responses to different GLRs. One of the possible explanations for the change of soil N status with GLR was that, in grass-legume mixed grasslands, decreased GLR may initially enhance BNF due to increased legume density and stimulating effects of grass species by N transfer (see δ^{15} N values of L. chinensis in Table [3](#page-7-0)) (Hartwig [1998](#page-13-0); Ehrenfeld [2003](#page-13-0); Schipanski and Drinkwater [2012\)](#page-14-0). For example, Ledgard et al. ([1985\)](#page-13-0) showed that the N fixation of Trifolium repens increased by 20 % when mixed into ryegrass grassland. However, when legume density is excessive, the intraspecific competition increased, resulting in less interaction between grass and legume in N utilization, which may inhibit BNF from legume (Ledgard and Steele [1992](#page-13-0)), and induce the physiological transformation from fixation of atmospheric N to uptake of soil N, consequently decreasing N fixation and soil N status (Tables [2](#page-5-0) and [3\)](#page-7-0). Furthermore, soil water content showed similar patterns as BNF in response to GLR in our experiment. Thus, GLR may impact BNF indirectly by regulating soil water content. Kowalenko and

GLC	GLR	Legume species		Reference species													
		B value	$N_{\rm ST}$	L. chinensis			H. glahra			S. viridis			K. integrifolia			Mean %Ndfa Mean BNF	
				N_{S1}	%Ndfa	D 区	N_{S1} ⁵	%Ndfa	$B\overline{B}$	$N_{\rm ST}$	%Ndfa	BNF	$N_{\rm{eff}}$	%Ndfa	BNF		
L-MR 1:0		$-0.61 + 0.03$		$3.94\pm0.03^{\mathrm{Aa}}$			$3.78 + 0.03$			3.74 ± 0.05			$4.09 + 0.03$				
	$\ddot{3}$:			1.47 ± 0.12^d 3.27 ± 0.08 ^{Bb}	å 55 ± 2^A	18.79 ± 0.34 ^{Ac}		53 ± 2^{ABa}	18.22 ± 0.35^{Bb}		$52\pm3\,^{\mathrm{ABa}}$	18.05 ± 0.41^{Ab}		$56\!\pm\!2^{\mathrm{ABa}}$	19.30 ± 0.32^{Ab}	55 ± 2^{ABa}	18.59 ± 0.34^{Ab}
	Ξ		$2.30\pm0.11^{\circ}$	$3.01\pm0.09^{\rm Be}$	37 ± 2^{Ab}	23.55 ± 0.71^{Ba}		34 ± 2^{ABb}	22.07 ± 0.78 ^{Ba}		34 ± 2^{Ab}	$21.66\!\pm\!0.73^{\mathrm{Ba}}$		$39{\pm}2^{\mathrm{A}\mathrm{B}\mathrm{b}}$	$24.85\pm0.84^{\mathrm{Ba}}$	$36\pm2^\text{Ab}$	23.04 ± 0.75^{Ba}
	Ξ		$2.87{\pm}0.04^{\mathrm{a}}$	$3.22\pm0.06^\mathrm{Bb}$	24 ± 1^{Ac}	21.91 ± 0.69 ^{Bab}		21 ± 1^{Ac}	19.42 ± 0.62^{Bb}		$20{\pm}2^{\rm Ac}$	$18.73\!\pm\!1.05^{\text{ABab}}$		$26\!\pm\!1^{\rm Ac}$	$24.11\!\pm\!0.61^{\mathrm{ABa}}$	23 ± 1^{Ac}	21.04 ± 0.71 ^{Bab}
	Ξ		$3.07 + 0.04a$	$\mathbb T$	$20{\pm}1^{\rm Ac}$	21.30 ± 0.76^{ABb}		$17{\pm}1^{\rm Ac}$	18.17 ± 0.57^{Ab}		$16{\pm}1^{\rm Ac}$	17.31 ± 0.96^{Ab}		$22\pm1^{\rm Ac}$	$24.08{\pm0.96}^{\mathrm{ABa}}$	$19\!\pm\!1^{\rm Ac}$	20.22 ± 0.73 ^{ABb}
QTT	1:0	$-0.47 + 0.04$	$\overline{}$	$3.97{\pm}0.06^{\mathrm{ABa}}$			$3.77 + 0.06$			$3.77 + 0.05$			$4.07 + 0.06$				
	3:1			$1.71 \pm 0.24^{\circ}$ 3.44 $\pm 0.05^{\text{Abc}}$	51 ± 5^{BCa}	18.66 ± 0.59^{Ab}		49 ± 5^{BCa}	17.77 ± 0.72 ^{Bb}		49 ± 5^{BCa}	$17.80{\pm}0.87^{\mathrm{Aa}}$		52 ± 5^{BCa}	19.08 ± 0.54^{Ab}	$50{\pm}5^{\text{BCa}}$	18.33 ± 0.68^{Ab}
	Ξ		2.45 ± 0.15^b	3.27 ± 0.09^{Ac}	34 ± 3^{Ab}	22.86 ± 0.66^{Ba}		31 ± 3^{Bb}	20.75 ± 0.74^{Ba}		31 ± 3^{Ab}	$20.68\!\pm\!0.84^{\mathrm{BCa}}$		$36\pm3^{\rm Bb}$	$23.86\pm0.66^{\mathrm{BCa}}$	$33 + 3^{Ab}$	22.04 ± 0.66^{Ba}
	$\ddot{1}3$		2.88 ± 0.10^{a}	3.49 ± 0.04^{Ab}	24 ± 1^{Ac}	$21.79{\pm}0.58^{\text{Ba}}$		21 ± 1^{Ac}	$18.60\pm0.60^{\rm B}{\rm C}$		21 ± 2^{Ac}	18.55 ± 1.29^{ABa}		$26\pm1^{\rm Ac}$	$23.29\!\pm\!0.62^\mathrm{Ba}$	23 ± 2^{Ac}	20.56 ± 0.73 ^{Bab}
	Ξ		$3.09{\pm}0.04^{\mathrm{a}}$		$20\!\pm\!1^{\rm Ac}$	$21.03 + 0.54^{ABa}$		$17{\pm}1^{\rm Ac}$	17.87 ± 0.50 ^{Ab}		$16{\pm}1^{\rm Ac}$	$17.07{\pm}1.37^{\mathrm{Aa}}$		22 ± 1^{Ac}	22.91 ± 0.60^{ABa}	$19{\pm}1^{\rm Ac}$	19.72 ± 0.72 ^{ABab}
L-MF	1:0	$-0.66 + 0.03$		$3.98\!\pm\!0.10^{A\mathrm{a}}$			$3.75 + 0.09$			$3.73 + 0.12$			4.13 ± 0.04				
	3:1			$1.87 \pm 0.11^{\circ}$ 3.49 $\pm 0.07^{\text{Ab}}$	$46 \pm 1^{\circ}$	14.81 ± 0.80^{Bb}		$43\pm1^{\rm Ga}$	13.85 ± 0.78 ^{Cb}		$42\pm2^{\rm Ca}$	13.79 ± 0.90^{Bb}		$47\pm1^{\rm Ca}$	15.40 ± 0.92 ^{Bb}	$45 \pm 1^{\rm Ga}$	14.47 ± 0.83 ^{Bb}
	Ξ		2.37 ± 0.16^{b}	$3.31\!\pm\!0.03^{\mathrm{Ac}}$	$35\!\pm\!1^{\mathrm{Ab}}$	19.98 ± 0.68 ^{Ca}		32 ± 1^{Bb}	17.94 ± 0.65 ^{Ca}		31 ± 1^{Ab}	17.84 ± 0.46 ^{Ca}		$37{\pm}2^{ABb}$	21.17 ± 1.36 ^{Ca}	34 ± 1^{Ab}	19.23 ± 0.77 ^{Ca}
	$\ddot{ }$:1		2.80 ± 0.15^a	$3.54\pm0.05^{\mathrm{Ab}}$	$26\!\pm\!1^{\mathrm{Ac}}$	$20.58 + 0.97^{Ba}$		22 ± 1^{Ac}	17.28 ± 0.76 ^{Ca}		21 ± 1^{Ac}	17.11 ± 1.22^{Bab}		$28\pm1^{\rm Ac}$	22.57 ± 1.77^{Ba}	24 ± 1^{Ac}	19.39 ± 1.04^{Ba}
	$\overline{0}$:1		3.01 ± 0.15^a		21 ± 1^{Ac}	19.28 ± 0.99^{Ba}		$17{\pm}1^{\rm Ac}$	15.33 ± 0.74^{Bb}		$16{\pm}1^{\rm Ac}$	15.14 ± 1.26 ^{Aab}		23 ± 1^{Ac}	21.59 ± 1.85^{Ba}	$19{\pm}1^{\rm Ac}$	17.83 ± 1.03^{Ba}
L-MS	1:0	$-0.78 + 0.05$	$\,$ $\,$	3.97 ± 0.09^{Aa}			$3.82 + 0.08$			3.79±0.08			$4.19 + 0.07$				
	3:1		1.26 ± 0.21 °	3.20 ± 0.04 ^{Bb}	57 ± 2^{Aa}	20.72 ± 0.48 ^{Ac}		56 ± 3^{Aa}	20.20 ± 0.45 ^{Abc}		55 ± 2^{Aa}	20.14 ± 0.43^{Ab}		59 ± 3^{Aa}	21.45 ± 0.60^{Ab}	57 ± 3^{Aa}	20.63 ± 0.46 ^{Ac}
	Ξ		2.10 ± 0.22^b	$2.95\!\pm\!0.06^{\mathrm{Be}}$	40 ± 1^{Ab}	26.91 ± 0.33^{Aa}		$38{\pm}1^{\mathrm{Ab}}$	25.55 ± 0.39^{Aa}		37 ± 2^{Ab}	25.26 ± 0.96^{Aa}		42 ± 2^{Ab}	28.67 ± 0.86 ^{Aa}	$39{\pm}2^{\mathrm{Ab}}$	$26.60{\pm}0.59^{\text{Aa}}$
	$\ddot{1}3$		2.79 ± 0.19^a	3.17 ± 0.09 ^{Bb}	25 ± 1^{Ac}	24.15 ± 0.34^{Ab}		23 ± 2^{Ac}	21.74 ± 0.44^{Ab}		22 ± 1^{Ac}	21.48 ± 1.52^{Ab}		$28\!\pm\!2^{\rm Ac}$	27.42 ± 1.52^{Aa}	24 ± 1^{Ac}	23.70 ± 0.86 ^{Ab}
	Ξ		3.08 ± 0.22 ^a		19 ± 1^{Ac}	21.81 ± 0.50^{Ac}		$16{\pm}1^{\rm Ac}$	$18.70{\pm}0.61^{\rm Ac}$		16 ± 2^{Ac}	17.95 ± 1.91^{Ab}		22 ± 2^{Ac}	25.82 ± 1.58^{Aa}	$18{\pm}1^{\rm Ac}$	$21.07{\pm}1.00^{\text{Ac}}$

^{- (}en dash) not measured, GLC grass-legume combination, GLR grass-legume ratio, L-MRL chinenica, L-JD L. chinensis-Lespedeza daurica, L-MF L. chinensis-Medicago $-(en \,dash)$ not measured, GLC grass-legume combination, GLR grass-legume ratio, L-MR L. chinensis-Medicago ruthenica, L-LD L. chinensis-Lespedeza daurica, L-MF L. chinensis-Medicago falcata, L-MS L. chinensis-Medicago sativa falcata, L-MS L. chinensis-Medicago sativa

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ab/s Different lowercase letters for each column indicate significant differences (P<0.05) between grass-legume ratios a,b,cDifferent lowercase letters for each column indicate significant differences ($P<0.05$) between grass-legume ratios

A,B,C Different uppercase letters in each column indicate significant differences (P<0.05) between grass-legume combinations A ,BC Different uppercase letters in each column indicate significant differences (P<0.05) between grass-legume combinations

Table 4 Plant density, individual and population biomass of L. chinensis, and sown legume under different L. chinensis-legume combinations and L. chinensis-legume ratios

	GLC	GLR				
		1:0	3:1	1:1	1:3	0:1
Plant density of L. chinensis (plants m ⁻²)	L-MR	673 ± 16^{Aa}	515 ± 16^{Ab}	362 ± 20^{Ac}	164 ± 19^{Ad}	
	L-LD	681 ± 22^{Aa}	495 ± 17^{Ab}	349 ± 18^{Ac}	157 ± 14^{Ad}	
	$L-MF$	684 ± 17^{Aa}	495 ± 20^{Ab}	345 ± 15^{Ac}	154 ± 14^{Ad}	
	L-MS	678 ± 18^{Aa}	513 ± 14^{Ab}	377 ± 15^{Ac}	$174\pm11^{\rm Ad}$	
Plant density of sown legume (plants/ m^{-2})	L-MR		168 ± 10^{Ad}	335 ± 14^{Ac}	462 ± 13^{Ab}	605 ± 15^{Aa}
	L-LD		167 ± 8^{Ad}	333 ± 14^{Ac}	454 ± 16^{Ab}	607 ± 14^{Aa}
	L-MF		160 ± 13^{Ad}	319 ± 12^{Ac}	443 ± 18^{Ab}	596 ± 15^{Aa}
	L-MS		164 ± 11^{Ad}	330 ± 14^{Ac}	468 ± 15^{Ab}	612 ± 13^{Aa}
Aboveground biomass of L. chinensis (g m ⁻²)	L-MR	320 ± 9^{Aa}	243 ± 13^{Ab}	169 ± 8^{Ac}	70 ± 9^{Ad}	
	L-LD	319 ± 11^{Aa}	226 ± 11^{Ab}	145 ± 5^{Bc}	63 ± 6^{Ad}	
	L-MF	318 ± 11^{Aa}	226 ± 8^{Ab}	129 ± 7^{Bc}	60 ± 7^{Ad}	
	L-MS	314 ± 4^{Aa}	251 ± 6^{Ab}	179 ± 7 ^{Ac}	77 ± 6^{Ad}	$\overline{}$
Aboveground biomass	L-MR		166 ± 7^{Ad}	311 ± 13^{Ac}	433 ± 10^{Ab}	515 ± 13^{Aa}
of sown legume $(g m^{-2})$	$L-LD$		184 ± 11^{Ad}	334 ± 13^{Ac}	445 ± 18^{Ab}	528 ± 10^{Aa}
	L-MF		160 ± 12^{Ad}	296 ± 15^{Ac}	418 ± 14^{Ab}	$480{\pm}14^{\text{Ba}}$
	L-MS		173 ± 9^{Ad}	320 ± 8 ^{Ac}	441 ± 12^{Ab}	530 ± 7^{Aa}

 $-(en dash)$ not shown here as the values are included in recruitment characteristics

Cameron ([1976](#page-13-0)) and Austin et al. ([2004](#page-13-0)) reported that improved soil water status can enhance BNF of legumes. Therefore, initially increased legume density may promote BNF by increasing soil water content (Armstrong et al. [1999\)](#page-13-0). However, relatively excessive legume density may result in more water transpiration due to larger shoot and leaf structures of the four legumes associated with L. chinensis. This could cause soil water to become a limiting factor for BNF, consequently inhibiting the potential increase in soil N (Sprent and Bradford [1977\)](#page-14-0).

Effects of grass-legume ratios on species recruitment

In arid and semi-arid ecosystems, soil water may also be an important limiting factor for species recruitment (Brown and Archer [1990](#page-13-0); Carino and Daehler [2002\)](#page-13-0). With regards to the modifying effects of legumes on soil N and water status (Armstrong et al. [1999\)](#page-13-0), we expected that increased relative abundance of legumes in this grassland would promote species recruitment. Our results showed that the introduction of legumes into this grassland increased all recruitment characteristics values when the GLR was lower than one in all *L. chinensis*- legume combinations; however, the legume monocultures showed inhibition effects on plant density and aboveground biomass of recruited species. Further regression analysis indicated that soil water content was the only control factor for all recruitment characteristics. This means that in this semi-arid grassland ecosystem, soil water content rather than soil N may be a more important environment factor to determine which species can successfully establish in a community, and legume abundance may promote or restrict species recruitment through regulating soil water status when introduced into grasslands (Brown and Archer [1990;](#page-13-0) Armstrong et al. [1999](#page-13-0)). However, when soil water content was excluded, we still found that soil $NO₃⁻-N$ significantly correlated with the number of recruited species. Experimental studies have shown that a higher NO₃⁻-N availability promotes recruitment, while communities with lower levels of $NO₃⁻-N$ are often more resistant to the establishment of recruited species (Fargione and Tilman [2005;](#page-13-0) Roscher et al. [2009\)](#page-14-0). In addition, previous studies showed that N enrichment may enhance plant tolerance to limited water through physiological regulation (Saneoka et al. [2004](#page-14-0); Wu et al. [2008](#page-14-0)). Therefore, soil $NO₃⁻-N$ may indirectly impact

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Table 6 The number, plant density, and aboveground biomass (AGB) of recruited species under different L. chinensis-legume combinations and L. chinensis-legume ratios

	GLC	GLR				
		1:0	3:1	1:1	1:3	0:1
Number of recruited species	L-MR	$5.0\pm0.4^{\mathrm{Ab}}$	$6.0{\pm}0.7^{\text{Bab}}$	7.3 ± 1.0 ^{ABa}	5.3 ± 0.3^{Ab}	5.0 ± 0.4^{Ab}
	L-LD	4.5 ± 0.3 ^{Ac}	$5.5 \pm 0.3^{\rm Bbc}$	7.3 ± 0.3 ^{ABa}	5.8 ± 0.3^{Ab}	4.8 ± 0.5^{Abc}
	L-MF	4.5 ± 0.3 ^{Ac}	$6.0 \pm 0.3^{\rm Bab}$	$6.3 \pm 0.3^{\text{Ba}}$	5.5 ± 0.3 ^{Aab}	4.8 ± 0.3 ^{Abc}
	L-MS	5.0 ± 0.4^{Ab}	8.8 ± 0.5^{Aa}	8.5 ± 0.3 ^{Aa}	6.0 ± 0.4^{Ab}	5.5 ± 0.3^{Ab}
Plant density of recruited	L-MR	26 ± 2^{Ab}	48 ± 5 ^{Aa}	44 ± 5^{ABA}	30 ± 3^{Ab}	24 ± 6^{Ab}
species (plants m^{-2})	L-LD	26 ± 1^{Ab}	57 ± 5 ^{Aa}	53 ± 6^{Aa}	30 ± 3^{Ab}	23 ± 2^{Ab}
	L-MF	$22 \pm 1^{\text{Abc}}$	31 ± 5^{Bab}	37 ± 6 ^{Ba}	$19 \pm 3^{\rm Bbc}$	15 ± 3^{Ac}
	L-MS	27 ± 3^{Ac}	57 ± 5^{Aa}	46 ± 1^{ABb}	26 ± 4 ^{ABc}	21 ± 2^{Ac}
AGB of recruited species $(g m^{-2})$	L-MR	11.9 ± 1.0^{Ab}	21.4 ± 2.0^{Ba}	21.3 ± 1.8 ^{Aa}	14.5 ± 1.5^{Ab}	11.2 ± 2.6^{Ab}
	L-LD	12.9 ± 0.6^{Ab}	26.5 ± 2.3^{ABa}	25.1 ± 2.4 ^{Aa}	14.7 ± 1.2^{Ab}	10.6 ± 0.5^{ABb}
	L-MF	10.9 ± 2.0 ^{Aab}	$12.1 \pm 2.2^{\text{Cab}}$	13.9 ± 2.6^{Ba}	7.3 ± 1.3^{Bb}	$5.8 \pm 1.1^{\rm Bb}$
	L-MS	12.4 ± 1.1^{Ac}	30.4 ± 3.2 ^{Aa}	24.3 ± 1.5^{Ab}	12.6 ± 2.1 ^{Ac}	9.3 ± 0.8 ^{ABc}

community recruitment by regulating the drought adaptability of recruited communities in this grassland ecosystem.

In our study, the density of sown legumes and L. chinensis in all mixed treatments remained steady following recruitment of new species, as compared with initial values, and significantly increased under GLR 3:1 and 1:1 in all GLCs. This indicates that the replacement of L. chinensis with legumes does not induce the passively competitive outcome for *L. chinensis* and legumes, and appropriate GLR may promote the development of a sustainable L. chinensis and legume population, while increasing species richness.

Effects of grass-legume ratios on grassland productivity

Nitrogen is a main limiting factor for grassland productivity (Karki et al. [2009](#page-13-0)). It has been widely reported that introducing legumes into grasslands leads to higher soil N availability and productivity (Sleugh et al. [2000;](#page-14-0) Mortenson et al. [2004](#page-14-0); Rao et al. [2007](#page-14-0)). Our study demonstrated that the increased relative abundance of legumes persistently increased aboveground productivity of a semi-arid grassland, which was partly attributed to higher individual biomass of legumes compared with L. chinensis (Table [4\)](#page-8-0), and also resulted from BNF by legumes (Table [3](#page-7-0)). The increased N input from legumes

Table 7 Dependence of species recruitment characteristics and aboveground biomass (AGB) of community on soil water content, NO_3^- -N, NH4 + -N, and species number as determined by four separate multiple regressions (one for each response variable)

			Regression parameters for each dependent variable				
Response variable	Intercept	Soil water content $\binom{0}{0}$	Soil NO_3 ⁻ -N $(mg kg^{-1})$	Soil NH_4^+ -N $(mg kg^{-1})$	Species number	Overall R^2	Overall F value
Number of recruited species	$-18.424***$	$1827***$	$0.442**$	-0.073^{NS}		0.445	$30.910***$
Plant density of recruited species (individual m^{-2})	$-180.707***$	24.760***	-0.067^{NS}	-0.090^{NS}	-	0.358	$43.563***$
AGB of recruited species $(g m^{-2})$ AGB of community (g m ⁻²)	$-101.244***$ $-586.651***$	$13.516***$ -0.127^{NS}	-0.057^{NS} 54.401***	-0.088^{NS} -0.320^{NS}	-0.129^{NS}	0.395 0.443	50.838*** $62.045***$

 NS not significant, $-$ (*en dash*) not included into regression analysis

 $*p<0.01$; $**p<0.0001$

Fig. 2 Aboveground biomass (AGB) of the community under different L. chinensis-legume combinations and L. chinensislegume ratios. Values are shown as means+SE

improved soil N availability and ultimately increased grassland productivity (Table [2\)](#page-5-0). Nyfeler et al. [\(2009\)](#page-14-0) similarly showed that the introduction of legumes induced consistently higher production compared with grass monocultures in European grassland. The facilitation effect of a grass-legume mixture on grassland productivity was more intense under low soil N availability (Nyfeler et al. [2009](#page-14-0)).

In the current study, the GLR1:1 treatment had 50–60 % greater aboveground production than a grass monoculture. Alternatively, grassland production can also be improved by N fertilization. In the same steppe region of the current study, Bai et al. ([2010](#page-13-0)) showed that 3 years of N fertilization at 10.5 g m⁻² year⁻¹ in this mature grassland dominated by L. chinensis increased aboveground production approximately 50 % compared with no fertilization (the AGB was around 300 g m^{-2}). However, in grassland ecosystems, species diversity loss induced by N enrichment has become a concern (Wedin and Tilman [1996;](#page-14-0) Gough et al. [2000;](#page-13-0) Suding et al. [2005](#page-14-0); Clark and Tilman [2008](#page-13-0)). Increased productivity by applying N fertilizer would be at a cost of loss of species richness. However, if the same forage production is obtained by introducing legumes into the grassland, the species richness may increase because of the facilitation effect of legumes on species recruitment. Therefore, as an alternative to N fertilization, introducing legumes in a temperate steppe grassland with GLR 1:1 can be a better choice to maximize the benefits of forage production and diversity conservation.

Species difference in grass-legume mixtures

Different species have different impacts on community recruitment and productivity as certain species drives specialized interspecific interaction (Vilà and Weiner [2004](#page-14-0); Lankau et al. [2009\)](#page-13-0), influences soil nutrient cycling (Liao et al. [2008\)](#page-13-0), and possibly results in a sampling effect (Vilà et al. [2011](#page-14-0)). Our results showed that GLC had a significant impact on species recruitment and grassland productivity, which can partly be attributed to different effects of the four legume species on BNF, soil water, and N status (Table [2](#page-5-0); Table [3;](#page-7-0) Table [7](#page-10-0)). The current experiment showed that L-MS had a higher N fixation compared with other GLCs which may result from relatively higher nodulation response and population biomass of *M. sativa* (Fig. [1;](#page-6-0) Table [4](#page-8-0)). Meanwhile, M. sativa showed stronger water retention capacity than the other three legumes when mixed with L. chinensis (Table [2](#page-5-0)). These factors helped to promote species recruitment and grassland productivity by improving soil N and water status. Considering the total effects of GLC on soil N, species recruitment, and grassland productivity, we suggest that M. sativa has more advantages for introduction into grasslands to enhance forage production and maintain species diversity, with L. *daurica* and M. *ruthenica* following, with M. falcata performing the worst.

Critical reflection about the assessment method of legume nodulation and %Ndfa

We used soil from the field site to assess legume nodulation in a pot experiment as it would be very hard to collect intact legume root for nodulation measurement in field without destructive harvest of the sample plots. However, the growth conditions in greenhouse indeed can never perfectly mimic the conditions in field regarding factors such as temperature, moisture, and competition for soil nutrient and water depending on species composition and plant development. Therefore, our nodulation data may not absolutely represent actual field values. In our pot experiment, we kept lower plant density and higher soil moisture compared with field condition. The attained legume nodulation data may reflect the inherent differences of nodulation response for four legume species in non-competition condition. Therefore, the species specific pattern of these nodulation data can be valuable for understanding the similar difference of BNF between legume species.

 15 N natural abundance method is commonly used to estimate %Ndfa (Carlsson et al. [2009](#page-13-0); Frankow-Lindberg and Dahlin [2013](#page-13-0)). The accuracy of the natural abundance technique ultimately depends upon the magnitude of $\delta^{15}N$ in the plant-available soil N pool which can be reflected by $\delta^{15}N$ value of reference plant (Unkovich et al. [2008\)](#page-14-0). Generally, higher $\delta^{15}N$ concentration of plant-available soil N causes lower BNF estimate error (Ledgard and Steele [1992](#page-13-0)). Unkovich et al. ([1994](#page-14-0)) recommended 2‰ as the lowest $\delta^{15}N$ value of reference plant for detecting a reliable %Ndfa. In our results, the $\delta^{15}N$ values of reference plants ranged 3.74– 4.09‰, which met the lowest $\delta^{15}N$ value requirement suggested by Unkovich et al. ([1994](#page-14-0)). The use of correct B value also is crucial for the accuracy of the 15 N natural abundance method (Unkovich et al. [2008](#page-14-0); Carlsson et al. 2009). Some published *B* values can be chosen for Medicago spp. but these B values frequently had difference in different literatures (Unkovich et al. [2008\)](#page-14-0). To attain accurate values of %Ndfa, here we generated our own B values for all four legume species in a glasshouse experiment. An assumption of the 15 N natural abundance method is that the $\delta^{15}N$ of reference plants identically reflects the $\delta^{15}N$ of soil N utilized by the legume. Chosen reference species influences the calculation of %Ndfa, especially when the $\delta^{15}N$ of soil mineral N changes markedly with depth or fluctuates

during the growing season because different reference species may have different N uptake and isotopic fractionation characteristics during utilization of soil N (Ledgard and Steele [1992;](#page-13-0) Unkovich et al. [2008](#page-14-0)). Since it is nearly impossible to know the "true" $\delta^{15}N$ of the soil N available to the legume and equally difficult to identify one optimal reference species, it has been suggested that the best choice is to use several reference species (Unkovich et al. [2008;](#page-14-0) Carlsson et al. [2009\)](#page-13-0). In current study, we selected four species as reference plants which likely reflected different N uptake and isotopic fractionation characteristics. We calculated the %Ndf using each reference species and also calculated a mean %Ndf (Table [3](#page-7-0)). The data showed some differences in %Ndfa when different reference δ^{15} N values were used, especially when %Ndfa was at low level. If direct N transfer from legumes to non N-fixing plants occurs, a choice of reference plants growing in close association with legumes may greatly underestimate the %Ndfa estimates (Ledgard and Steele [1992](#page-13-0); Unkovich et al. [2008](#page-14-0)). In a previous study at the same site, it appeared that the direct N transfer from legume to neighboring plants had occurred (unpublished data). In the current study, the significant difference in δ^{15} N of L. chinensis between monoculture and mixture plots also indicated that direct N transfer from legume to grass had likely occurred; therefore, we used plants in grass monocultures as reference plants. However, this may impose other errors since it fails to accurately mimic the $\delta^{15}N$ of the soil N utilized by legume (Unkovich et al. [2008;](#page-14-0) Carlsson et al. [2009\)](#page-13-0). In fact, even though there is evidence for direct N transfer from legumes to reference plants in study sites, it does not exclude the possibility for N transfer in other directions, which is an argument for using reference plants grown together with the legume (Carlsson et al. [2009\)](#page-13-0). For example, Carlsson and Huss-Danell ([2014](#page-13-0)) indicated that N transfer not only occur from legumes to neighboring nonlegume plants but also in the opposite direction, from non-legumes to legumes and partial N transferred from a legume plant should be considered equally available for re-uptake by the legume plant itself as for neighboring legume and non-legume plants. This makes the error weak for using non-legume plants in mixtures to reflect ¹⁵N availability in soil for neighboring legume. In contrast, N_2 fixation may be overestimated by using a reference plant in pure stand (Carlsson and Huss-Danell [2014](#page-13-0)). Based on these studies, during N_2 fixation estimate, a precise choice for reference plants should be done after considering the directions and amount of N transfer.

Conclusions

Introducing legumes in a temperate steppe grassland dominated by L. chinensis can improve soil N and water status. However, effects of grass-legume mixtures depend on the relative abundance of legumes, with the GLR 1:1 inducing the most positive effects on soil N and water contents. Appropriate GLR can promote species recruitment by enhancing soil water and NO_3 ⁻-N content, consequently increasing the species diversity of a grassland. Aboveground production of plant communities persistently increased with relative abundance of legumes, which was correlated to increased soil $NO₃⁻-N$ availability. The different legume species tested had different effects on soil N, species recruitment, and grassland productivity in the grass-legume mixtures. In general, M. sativa had the most positive effects on soil N, species recruitment, and grassland productivity when mixed with L. chinensis.

In conclusion, in this semi-arid steppe grassland, legume introduction can be considered an alternative to nitrogen fertilization for increasing grassland productivity without compromising grassland species diversity. M. sativa has the greatest economic and ecological potential to be mixed into grasslands, and 1:1 ratio for grass and legume is the best ratio in mixed communities.

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