Thermogravimetric and kinetic analysis of the decomposition of solid recovered fuel from municipal solid waste

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Abstract The thermal decomposition of a solid recovered fuel has been studied using thermogravimetry, in order to get information about the main steps in the decomposition of such material. The study comprises two different atmospheres: inert and oxidative. The kinetics of decomposition is determined at three different heating rates using the same kinetic constants and model for both atmospheres at all the heating rates simultaneously. A good correlation of the TG data is obtained using three nth-order parallel reactions.

Keywords Kinetics · Municipal solid wastes · Pyrolysis · Combustion · Solid recovered fuel

Introduction

In Spain, as in the European Union (EU) countries, there has been an increase in per capita waste generation with the growth of the economy. Proper management prevents environmental problems, by converting waste into resources that contribute to the saving of raw materials and energy, so that good management practices together with source reduction are two of the cornerstones of policy environment.

In 2010, total waste generation in the EU-27 amounted to 2.5 billion Mg, an average of 4,986 kg per EU inhabitant. Nine hundred and twenty-seven million Mg of this total is covered by the indicator "generation of waste excluding

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Department of Chemical Engineering, University of Alicante, P.O. Box 99, 03080 Alicante, Spain e-mail: ja.conesa@ua.es major mineral wastes," corresponding to 1,847 kg inhabitant⁻¹ and to 37 % of the generated waste total [1]. In Spain, in 2011, each person generated 490 kg of municipal waste, a value well below the 662 kg produced in the year 2000 [2]. This value is close to the average municipal waste per capita in the EU-27 (499 kg person⁻¹) in the 2010.

Energy recovery from waste has become the main option for the recovery of resources contained in waste. In this sense, an effort to reuse or composting the municipal solid waste should be given, and a final energy recovery of non-reusable and non-compostable matter should be done.

Solid recovered fuel (SRF) is a waste-derived fuel made from non-hazardous household waste. It is feasible to recover energy with an important reduction in the amount of waste landfilled. The SRF is the non-recyclable fraction of municipal waste and consists of approximately 30 % of paper, 20 % wood, 35 % plastic and 15 % textile waste. It is a lightweight material with <20 % moisture and formed by fragments of about 4 cm. It is being tested as an alternative fuel for cement industry, but its calorific value is much lower than the petroleum coke, so that to achieve high-energy substitutions must be fed a great deal more than conventional fuels.

The SRF should have some characteristics in order to be used as a fuel for cement kilns. The most developed standard, but by all means not the only one, is the CEN/TC 343 regulation, which stipulates several kinds of SRF depending on the calorific power and the presence of different pollutants, chlorine and mercury among them, being very important parameters in the operation of cement factories.

The most important characteristics of SRF destined to the cement sector are:

• High calorific power, about $16-18 \text{ kJ g}^{-1}$, although some clinker plants demand higher values

- Reduced amount of chlorine (inferior to 0.5 %)
- Reduced amount of mercury (inferior to 10 mg kg⁻¹, on a dry base)

This specification corresponds to high-quality SRF based on the CEN/TC 343 regulation (European Committee for Standardization). In order to be able to comply with the emission restrictions established by the legislation, the most important limitations are related to the chlorine load, because of the stable operation of the installation, as well as mercury and heavy metals presence in the SRF.

In the present paper, it is studied the thermal decomposition of a SRF destined to the cement industry. It has been studied using thermogravimetry, in order to get information about the main steps in the decomposition. The study comprises two different atmospheres: inert (pyrolysis) and oxidative (combustion).

Materials and methods

SRF characterization

The SRF used for the development of this work was provided by the company CEMEX ESPAÑA S.A. and was obtained from three different suppliers. The SRF has a split appearance, with diverse colors because of the heterogeneity in composition. It has also a strong unpleasant odor. Prior to the characterization of the material, a representative amount of the three samples were mixed, homogenized and grounded to an average size of 1 mm. Analyses described below were performed on this representative sample of the waste (see in Table 1, the results):

- Determination of moisture content: Moisture of the sample was obtained from mass loss suffered by drying in an oven at 105 °C until constant mass.
- Determination of ash: The solid residue obtained by calcining the sample in a muffle furnace at 850 °C.
- Elemental analysis: Analysis of the major components (carbon, hydrogen, nitrogen and sulfur) is performed by oxidation of the sample to 1,000 °C and subsequent detection of combustion products (CO₂, H₂O, N₂ and SO₂). The equipment used was a Perkin-Elmer 2400 (Perkin-Elmer, UK).
- Calorific value (NCV): determined by a calorimeter bomb AC-350 Leco Instruments.
- Determination of biomass content by analyzing the content of hemicellulose, cellulose and lignin. The contents of hemicellulose, cellulose and lignin for this material were determined according to Rowell et al. [3] and test methods T12, T222 and T203 of the "Technical Association for the Pulp and Paper Industries" [4].

Table 1 SRF characterization

Moisture/mass%	16.67
NCV/kJ kg ⁻¹	1,014.2
Ash/mass%	10.23
Element/mass%	
Ν	1.07
С	40.85
Н	5.33
S	0.15
O (by difference)	42.37
Metal/mg kg ⁻¹	
V	15
Cr	78
Mn	354
Со	4
Ni	52
Cu	240
Zn	658
As	25
Cd	2
Sn	18
Sb	32
Tl	4
Hg	0.2
Те	2
Pb	235
Biomass in SRF/%	86.5
Lignin in biomass/%	31.3
Cellulose in biomass/%	61.1
Hemicellulose in biomass/%	7.6

Thermobalance

SRF has been subjected to a thermal decomposition study as part of the characterization of their properties. To this end, we have carried out a series of thermogravimetric analysis in two different atmospheres: inert atmosphere of N₂ and oxidative atmosphere N₂–O₂ = 9:1 (10 % of oxygen) and three different heating rates every atmosphere (5, 15 and 30 °C min⁻¹). Analyses were performed by simultaneous TG–DTA equipment brand METTLER TOLEDO TGA/ SDTA851e/LF/1600 model that can work between room temperature and 1,600 °C. On this equipment, the specimen holder and the oven are arranged horizontally.

Results

TG runs and kinetics

Six runs were performed in dynamic conditions by combining two reaction atmospheres (N₂, N₂–O₂ = 9:1) with three heating rates (5, 15 and 30 °C min⁻¹). All experiments were performed with an initial mass of about 5 mg sample, the carrier gas flow was 100 mL min⁻¹, and the temperature range studied was from 25 to 1,000 °C.

Figures 1 and 2 represent the experimental curves of mass loss versus temperature in dynamic experiments at three heating rates for each of the atmospheres of reaction studied. In the graphs, *w* is defined as the mass fraction of solid (including both the residue formed and the unreacted solid reactant), i.e., represents the ratio between the total mass of solid at any instant (*m*) with respect to the mass initial solid (m_0). Figure 3 compares the effect of the atmosphere of reaction at 5 °C min⁻¹.

The points represented, which have been those used for the kinetic analysis, have been selected according to techniques recommended by Caballero and Conesa [5] so that:

- The derivative of the points is calculated accurately and correctly.
- The points are equally spaced on a representation dW/ dT versus temperature.
- The fitting is simultaneous, with no variation of the kinetic constants, for at least three different heating rates.

Besides this, we use numerical techniques for solving differential equations and optimization methods prior definition of a proper objective function [5, 6].

In the previous graphs, we can see the effect of the heating rate on TG curves. It can be appreciated that increasing heating rate curves is shifted to the right, i.e., that the mass losses occur at increasing temperatures. This behavior has been described by several researchers and can be explained using different arguments [7, 8]. Some authors

argue that this behavior is due to changes in reaction mechanism caused by increased heating rate or changes in the apparent activation energy [9]. Furthermore, poor heat transmission to the sample in the oven may cause increasing differences with increasing heating rate between the nominal and real temperature of the sample. It could also be due to different rates of heat dissipation or absorption of the reaction at different heating rate. However, the observed shift can be simply explained by the mathematical form of the kinetic laws [7, 10], which can provide a shift of the curves at higher temperatures with increasing reaction rate with the same kinetic constants.

Obviously, a kinetic model that represents a set of experiments should be able to explain such movements at various heating rates. Several authors have shown that some TG curves can be fitted to different kinetic models, providing very different values of the kinetic parameters, depending on the models used. Therefore, only models capable of explaining the shift in the TG with heating rate, without changing the kinetic parameters can be considered as potentially correct. However, if the heating rates used are very high, it might be better to include heat transfer effects. In any case, the kinetic models obtained should be considered as models of data correlation, away from the claim for mechanistic models.

In Fig. 1, we can see that in the inert atmosphere, the final solid residue at any heating rate is similar and close to 20 %, whereas in the oxidative atmosphere, the amount of final solid residue is much lower, on the order of 10 % (as the ash analysis in Table already indicated). From Fig. 3, it is clear that the presence of oxygen accelerates the thermal decomposition of SRF, as it occurs with other materials [11–14], producing the decomposition at lower temperatures.







Fig. 3 TG curves at 5 K min⁻¹ under an atmosphere of N_2 and $N_2:O_2 = 9:1$

Kinetic model

Figure 1 represents the pyrolytic decomposition of the SRF. It can be distinguish three regions of mass loss. Each of these three stages would be centered at 540, 640 and 700 K, approximately. Due to this, the better results for the kinetic modeling will assume three different organic fractions that would decompose simultaneously in parallel reactions [8, 15]. Each organic fraction would decompose into volatile and a carbonaceous solid residue.

Using this technique, pseudo reaction mechanisms are constructed, in which each reaction includes one or more elementary reactions, since gases and tars cannot be treated separately and are grouped as volatile matter. The kinetic parameters obtained are representative of each overall reaction. The kinetic model proposed for the pyrolysis of SRF could be interpreted considering the material formed by three independent parts, each one following an independent reaction, as follows:

$$w_{s_{10}}$$
Solid₁ $\xrightarrow{1}$ $(w_{s_{10}} - v_{1\infty})$ Char₁ + $v_{1\infty}$ Volatiles₁ (1)

$$w_{s_{20}}$$
Solid₂ $\xrightarrow{2}$ $(w_{s_{20}} - v_{2\infty})$ Char₂ + $v_{2\infty}$ Volatiles₂ (2)

$$w_{s_{30}}$$
Solid₃ $\xrightarrow{3}$ $(w_{s_{30}} - v_{3\infty})$ Char₃ + $v_{3\infty}$ Volatiles₃ (3)

In the previous reactions, Solid₁, Solid₂ and Solid₃ refer to different fractions or components of the original material, "Volatiles_i" are the gases and condensable volatiles evolved in the corresponding reactions (i = 1-3), and "Char_i" is the char formed in the decomposition of each Solid_i (i = 1-3). On the other hand, the small letters represent the yield coefficients representative of each reaction and consequently, and it is considered not to be changing with time and with the extension of the reaction. Moreover, each fraction has a yield coefficient that represents the maximum mass fractions obtainable by each reaction. In this way, $v_{i\infty}$ is the yield coefficient for the Volatiles_i and v_i is the mass fraction of volatiles. The different initial mass fraction of the components ($w_{s_{i0}}$) is related so the following must be fulfilled:

$$\sum_{i=1}^{3} w_{\rm s_{i0}} = 1 \tag{4}$$

Considering an *n*th-order kinetic decomposition, the kinetic equations for the pyrolysis runs can be expressed as followed:

$$-\frac{\mathrm{d}w_{\mathrm{s}_{\mathrm{i}}}}{\mathrm{d}t} = k_{\mathrm{i}}w_{\mathrm{s}_{\mathrm{i}}}^{n_{\mathrm{i}}} \tag{5}$$

or

$$\frac{\mathrm{d}V_{\mathrm{i}}}{\mathrm{d}t} = k_{\mathrm{i}} (V_{\mathrm{i}\infty} - V_{\mathrm{i}})^{n_{\mathrm{i}}} \tag{6}$$

with the kinetic constants following the Arrhenius equation:

$$k_{\rm i} = k_{\rm i0} \, \exp\left(-\frac{E_{\rm i}}{\rm RT}\right) \quad i = 1 \text{ to } 3$$

$$\tag{7}$$

In the equations, V_i and $V_{i\infty}$ represent the volatiles evolved from the decomposition of each fraction, and the maximum yield of volatiles that can be obtained, respectively.

For the optimization of three TG runs (approximately 300 experimental points), 11 parameters were obtained $(3 \times E_i, 3 \times k_{i0}, 3 \times n_i \text{ and } 2 \times w_{s_{i0}}).$

On the other hand, the model proposed to explain the thermal decomposition of SRF in the presence of oxygen is the same that has been considered in the case of pyrolysis runs, but with different values of the kinetic constants.

In order to obtain a single set of parameters for the combustion of the material, all the runs were correlated with the same set of parameters by a similar procedure explained with the correlation of the pyrolysis data. For all the fractions (1–3), the same values of apparent activation energy and reaction order obtained under a nitrogen atmosphere have been considered, but the pre-exponential factor could change due to the presence of the oxygen. This type of model satisfactorily fitted the decomposition of other materials [16, 17]. With all these considerations, acceptable correlations of the data are obtained.

The calculated values were obtained by integration of the differential equations presented in the kinetic model, by the explicit Euler method, but considering and testing that the intervals of time are small enough so the errors introduced are negligible. The optimization method of the function Solver in a Microsoft Excel spreadsheet was used to minimize the differences between experimental and calculated mass loss and their derivatives. The objective function (OF) to minimize was the sum of the square differences between experimental and calculated mass loss values:

$$O.F. = \sum_{j} \sum_{k} \left[w_{\exp_{kj}} - w_{\operatorname{calc}_{kj}} \right]^2$$
(8)

where 'k' represents the experimental data at time 't' in the experiment with a heating rate 'j'. The value of $w_{\exp_{k_j}}$ or $w_{\operatorname{cal}_{k_j}}$ represents the mass loss fraction in the experimental and calculated data, respectively.

The kinetic parameters for decomposition of each fraction are shown in Table 2. As mentioned above, the same activation energy and reaction order are assumed for pyrolysis and combustion processes. Figure 4 shows the loss curves of experimental and calculated mass, at all heating rates studied and in both inert and oxidative atmosphere. It is clear that best fits would be obtained if all parameters are allowed to vary for each run, but they would be less representative of the overall process.

Discussion

The values of activation energies obtained from the fitting are 324.9, 98.1 and 274.3 kJ mol⁻¹ and correspond to the fractions discussed above.

 Table 2 Kinetic parameters for pyrolysis and combustion (all heating rates)

	Pyrolysis	Combustion
Fraction 1		
$k_0/{\rm min}^{-1}$	7.15×10^{20}	1.63×10^{21}
$E/kJ \text{ mol}^{-1}$	324.9	
W _{i0}	0.19	
Ν	0.70	
Fraction 2		
$k_0/{\rm min}^{-1}$	2.23×10^{6}	6.92×10^{6}
$E/kJ \text{ mol}^{-1}$	98.1	
W _{i0}	0.48	
Ν	0.76	
Fraction 3		
$k_0/{\rm min}^{-1}$	1.58×10^{18}	4.13×10^{19}
$E/kJ \text{ mol}^{-1}$	274.3	
w_{i0} (by difference)	0.33	
n	2.87	





Fig. 5 Decomposition of each fraction assumed in the kinetic modeling at the different experimental conditions



Figures 5a, b shows the decomposition of each fraction calculated for the pyrolysis and combustion processes at the three heating rates considered.

The first fraction, comparable to hemicellulose, decomposes in a temperature range of 200-250 °C for the combustion and 250-300 °C for the pyrolysis process. The data are consistent with the literature consulted [18]. The reaction order for this fraction is close to unity, as it is accepted for this type of material [8, 15, 19, 20].

The second fraction, mainly composed of cellulose, decomposes in a temperature range of 300–350 °C for pyrolysis and reaches 400 °C for high speed heating in combustion because the main reactions involve breakage of glycosidic linkages with the consequent partial depolymerization of the cellulosic component of wood [21, 22].

The third fraction, similar to lignin, is the most refractory and difficult to degrade by thermal route. The highest percentage of fixed carbon present in wood with higher percentages of lignin, which are attributed to the lignin being the most resistant to thermal decomposition when compared with cellulose and hemicellulose, due to its highly complex structure. This fraction decomposes between 400 and 900 K in both pyrolysis and combustion, presenting the highest range of temperature decomposition. The reaction order is high as previously reported for ligninic materials [8].

The values of the pre-exponential factors of all the organic fractions considered increase with the partial pressure of oxygen. This indicates that exists an acceleration of the process in the presence of oxygen, as pointed out previously. From the data presented in Table 2, it is possible to calculate the following "Enhancement factor":

$$EF = \frac{(k_{i0})_{\text{combustion}}}{(k_{i0})_{\text{pyrolysis}}}$$
(9)

that will give an idea of the effect of oxygen in the decomposition rate. The values of this EF are 2.2, 3.1 and 26.1 for fractions 1, 2 and 3, respectively, denoting that the fraction that most increases the decomposition rate in the presence of oxygen is fraction 3, similar to lignin.

Conclusions

A thermogravimetric study on the decomposition of a SRF has been done at different heating rates and atmospheres. A kinetic model for decomposition in inert and oxidizing atmosphere is proposed. The model assumes the presence of three organic fractions in the SRF, which will be assimilated to cellulosic, hemicellulosic and ligninic species. The decomposition of each fraction is produced by an nth-order kinetic law. The mathematical treatment of the data permits to fit simultaneously pyrolysis and combustion experimental data. The combustion is much more rapid than the pyrolysis, and the ligninic fraction is the most affected by the presence of oxygen.

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