Development of an extractive spectrophotometric method for uranium using MWCNTs as solid phase and arsenazo(III) as chromophore

Faheem Shah • Mustafa Soylak • Tasneem Gul Kazi • Hassan Imran Afridi

Received: 26 September 2012 / Published online: 29 January 2013 © Akadémiai Kiadó, Budapest, Hungary 2013

Abstract Multiwalled carbon nanotubes (MWCNTs) based sorptive extraction method for uranium (U) from aqueous solutions has been developed. The proposed method was optimized by evaluating the analytical parameters including pH, eluent type, flow rates of sample and eluent, etc. The adsorption capacity of MWCNTs was found to be 9.80 µg g^{-1} , while the detection limit based on 3σ criterion was 1.9 μ g L⁻¹. The presented method was applied for the estimation of U in ore sample. Effect of potentially interfering ions was also studied and were found to inert not interfering with U during the analysis. The results suggest that MWCNTs can be used as reliable solid phase for preconcentration and arsenazo-III as chromophore for U spectrometric determination from aqueous solutions.

Keywords Carbon nanotubes - Uranium - Solid phase - Pre-concentration - Arsenazo-III

Introduction

Due to its chemical toxicity and radioactive nature, uranium (U) is one of the most hazardous heavy metals in

F. Shah $(\boxtimes) \cdot M$. Soylak Department of Chemistry, Erciyes University, Fen Fakultesi, 38039 Kayseri, Turkey e-mail: shah_ceac@yahoo.com

M. Soylak e-mail: soylak@erciyes.edu.tr

F. Shah · T. G. Kazi · H. I. Afridi National Center of Excellence in Analytical Chemistry, University of Sindh, Jamshoro 76080, Pakistan e-mail: tgkazi@yahoo.com

H. I. Afridi e-mail: hassanimranafridi@yahoo.com the environment. U in trace amount is found almost everywhere in water, soil and rocks [[1\]](#page-5-0). It is also known as responsible to cause toxicological effects in mammals and is potential occupational carcinogen [[2\]](#page-5-0).

Uranium assessment in trace quantity is important task to estimate the performance of different nuclear processes and thus its determination in environmental samples is a matter of concern for safety reasons [[3\]](#page-5-0). A number of various methods have been developed for U determination in different samples. These various techniques are laser fluorimetry [[4\]](#page-5-0), gamma spectrometry [[5\]](#page-5-0), stripping voltammetry [\[6](#page-5-0)], fluorimetry [\[7](#page-5-0)], potentiometry [[8\]](#page-5-0), polarography [\[9](#page-5-0)], X-ray fluorescence [\[10](#page-5-0)], inductively coupled plasma mass spectrometry $[11, 12]$ $[11, 12]$ $[11, 12]$ $[11, 12]$ $[11, 12]$, neutron activation analysis $[13]$ $[13]$, and alpha spectrometry [[4,](#page-5-0) [14](#page-5-0)]. Although all these techniques are sensitive and capable of fast measurement, but requires expensive instruments, perfect experimental conditions and complicated sample-pretreatment steps. For U determination spectrophotometric methods are widely because it is quite simple and reliable accuracy [[7,](#page-5-0) [8\]](#page-5-0). Spectrophotometry is still a requisite technique for the determination of trace U and a variety of different chromogenic agents have been used in this regard. The chromogenic reagents commonly used for U determination are arsenazo-III (A-III) [\[15](#page-5-0)], 4,4-diaminophenylmethane [[16\]](#page-5-0), p-carboxychlorophosphonazo [\[17](#page-5-0)], chromazurol S [[18\]](#page-5-0), 4-(2-pyridylazo) resorcinol [\[19](#page-5-0)], meloxicam [\[20](#page-5-0)], dimethylphenylazosalicylfluorone [\[21](#page-5-0)], 8-quinolinol [[22\]](#page-5-0), anthranilic acid and rhodamine 6G [\[23](#page-5-0)], pyrocatechol [\[24](#page-5-0)], 2-(5-bromo-2-pyridylazo-5-diethylaminophenol) [[25\]](#page-5-0), dibenzoylmethane [[26\]](#page-5-0), and thiocyanate [[27\]](#page-5-0), although most of them are not selective.

A number of different methods have been developed for the removal of uranium from aqueous solutions based on various adsorbents such as activated carbon [[28\]](#page-5-0), olivinerock [[29\]](#page-5-0), montmorillonite [[30\]](#page-5-0), carbon fiber [[31\]](#page-6-0), algae

and biomass of fungus [\[32](#page-6-0)], attapulgite [[33](#page-6-0)], manganese coated zeolite [\[34\]](#page-6-0) and carbon nanotubes (CNTs) [[35\]](#page-6-0).

Carbon nanotubes (CNTs) have been used earlier as sorbents in solid phase extraction [\[36\]](#page-6-0). Depending on the layers, CNTs are divided into single walled (SWCNTs) and multiwalled carbon nanotubes (MWCNTs). CNTs have a great potential as advanced sorbents for the extraction and preconcentarion of different contaminants like phenols [\[37\]](#page-6-0), dyes [\[38](#page-6-0)], pesticides [[39\]](#page-6-0), metal ions [[40\]](#page-6-0) drugs [\[41,](#page-6-0) [42](#page-6-0)], natural organic matter [\[43,](#page-6-0) [44\]](#page-6-0) and many other chemicals [\[45](#page-6-0)].

Because of its dynamic oxidation state, U has a unique nature as compared to many other metals. U has the capability to be a part of positive, neutral and negatively charged complexes, nearly at neutral pH. Due to the strong association of U with other elements, its determination needs selectivity. Because of simplicity and selectivity, organic dyes have been widely investigated and reported for the spectrophotometric determination of U [[45–47\]](#page-6-0). Among a variety of these organic reagents which are largely based on azo-dyes for U determination, sodium salt of A-III is reported as a responsive chromogenic reagent. Because of U-(A-III) high stability its quantitative determination even in very low pH became possible, where neither hydrolysis, nor the formation of polynuclear species can take place [\[48](#page-6-0)].

In this work, we studied the sorption of uranium on MWCNTs as function of the following parameters: pH, sample volume, eluent nature and volume, flow rate and A-III concentration. The basic objectives of present work are to investigate the sorption behavior of uranium on MWCNTs and its spectrophotometric determination using A-III as chromogenic reagent.

Experimental

Apparatus

Absorbances were measured using Hitachi Model 150-20 UV–Vis double beam spectrophotometer carrying a 10 mm optical path cell. For pH adjustments, pH meter, Nel pH 900 (Ankara-Turkey) Model glass-electrode was used. To measure the concentrations of interfering ions Perkin-Elmer Model 3110 atomic absorption spectrometer (Norwalk, CT, USA) was used.

Reagents

Standard solutions of U were prepared by the dilution of certified standard solution (1000 mg L^{-1}) HIGH-PURITY standards (Charleston, SC). Arsenazo-III (3,6-bis[(2-arsonophenyl)azo]-4,5-dihydroxy-2,7-naphthalenedisulfonic acid) (Fluka Chemie, Buchs, Switzerland), solution was prepared by dissolving 0.1 g of the reagent in 100 mL distilled water. Working solutions were prepared by adequate dilution of the stock solution. All reagents and solvents used in experimental work were of analytical grade.

Sample preparation

0.5 g of ore samples were weighed into a PTFE Teflon beakers of 100 mL capacity and adding 10 mL of $HNO₃$ followed by heating on hot plate for 3 h at 70 \degree C. 5 mL of HClO4 were added to this mixture and the temperature was raised to 90 °C and left overnight. Further $HNO₃$ (10 mL), $HClO₄$ (5 mL) and HF (5 mL) were added again and the mixture was heated for 24 h at 70 $^{\circ}$ C until a semi dried mass was obtained. The obtained residue was redissolved in 15 mL of acid mixture of $HNO₃$, HCl and HF (1:1:1) and heated overnight at 70° C. At last, the residue was dissolved in 10 mL of 1 N HNO₃ and diluted to the desired sample volume with deionized water [[49](#page-6-0)].

Adsorption studies

In batch adsorption experiments, 0.2 g of MWCNTs was introduced into 100 mL Erlenmeyer flasks, containing 50 mL of 150 μ g mL⁻¹ stock solution at pH 5. Flasks were shaken in a mechanical shaker at 150 rpm for 24 h on a pre-settled temperature. Then the mixture was filtered and the maximum amount of U thus pre-concentrated was eluted with 1.5 N HCl and determined spectrophotometrically adding A-III. The capacity of MWCNTs was found to be 9.80 mg g^{-1} .

Adsorption capacity (q) was calculated as:

 $q = V(C_0 - C_e)/W$

where C_0 is initial concentration and C_e is the equilibrium concentration of U (μ g mL⁻¹). *V* represents the volume of the solution (mL), and W is the mass of adsorbent, 0.2 g in our case.

Procedure

In column study, column with 1 cm diameter and 15 cm length was loaded with 0.2 g of MWCNTs. An aliquot of U solution containing $10-150 \mu$ g of the metal ion was taken in a beaker and diluted in the range of 20–500 mL with deionized water. The test solution, whose pH was already adjusted to 5 by using acetate buffer, was passed through the MWCNTs column at $1-5$ mL min⁻¹ flow rate. The effluent carrying the analyte was collected in a volumetric flask and washing the column three times with 5 mL of acetate buffer. A mixture of acid and methanol was used for the elution of U retained in the column. For rapid and quantitative elution of U from the column the best choice is a mixture of HCl and methanol and quantitative elution

for retained U was observed with 10 mL of methanol containing 1 mL of 1 N HCl [[50\]](#page-6-0).

The column was washed with deionized water and the washings were also collected in a flask where U was determined spectrophotometrically at 649.1 nm by adding A-III. Deionized water was used for repeatedly washing the column and in the usual mode the U retained in the column was eluted out and determined. For 10 min the solutions were kept until the complete color development and followed by measuring the absorbance at 649.1 nm against a blank. The calibration curve was fitted by taking average of five similar measurements (samples with different concentration) of each standard solution.

Results and discussion

To attain maximum U recovery from aqueous solution and its best determination, it is obligatory to optimize the experimental parameters like pH, sample and eluent volume, eluent nature and flow rates and effects of interfering ions, etc.

Effect of pH

The pH has a key role in the determination of U retained on MWCNTs because the changing the pH affects both the stability of the U and the MWCNTs surface morphology (e.g. protonation of the functional groups causing the sorption). The sorption competence of MWCNTs for U was strongly influenced by the initial pH of aqueous solution. Initially the effect of pH on the adsorption of U ions on MWCNTs was studied in the pH range of 1–7 for the U. The results regarding the pH effect on the adsorption yield of the U are shown in Fig. 1. When pH is lower $(pH = 1)$, the functional groups on MWCNTs will be protonated and may not freely interact with U ions. At pH 5, due to deprotonation of the metal-binding sites, the negative charge density on the surface of MWCNTs increases resulting in the increase of attraction between metal ions and negative charge, which allows the sorption on the MWCNTs surface [[7,](#page-5-0) [8](#page-5-0), [14,](#page-5-0) [43](#page-6-0)]. It was observed that adsorption efficacy increased as pH was increased from 1 to 5. Maximum adsorption was observed at pH 5 and this was used for adsorption for U in column.

The absorbance of the A-III complex is dependent on pH. The optimum pH where maximum absorbance occurs is dependent on the ionic medium used. Finally after elution pH was maintained at 1.5 with HCl for spectrometric determination step as the U-(A-III) complex is stable at this pH [\[47](#page-6-0)].

Effect of sample volume

To get high enhancement factor, the effect of sample volume on U sorption in the solid phase extraction system was

Fig. 1 Effect of pH on % recovery of Uranium

Fig. 2 Effect of sample volume (mL) on % recovery of Uranium

Table 1 Effect of eluents combination on desorption of U

$Eluent$ (Acid Concentration=1 N)	Ratio	% Recovery
Methanol		70
Methanol+HCl	1:1	82
Methanol+HCl	1:2	91
Methanol+HCl	1:3	100
Methanol $+HNO3$	1:1	79
Methanol $+HNO3$	1:2	83
Methanol $+HNO3$	1:3	91
Methanol $+H_2SO_4$	1:1	69
Methanol $+H_2SO_4$	1:2	78
Methanol $+H_2SO_4$	1:3	82

investigated by passing varying volume of solution containing 10 μ g L⁻¹ of U through MWCNTs under the optimum conditions. Uranium was recovered from different volumes of aqueous solutions of different concentrations. In this study, initial volume of the solution was selected as 400 mL. When the sample volume increases a gradual decrease in the recoveries can be seen, at 500 mL of sample volume only 82 % of recovery was obtained (Fig. 2). Final

volume of the solution was considered as 10 mL, a 40-fold pre-concentration factor was obtained for the U. The results have shown that U can be quantitatively recover even from a small volume of solution while recovery greater than 96 % can be obtained from a solution volume of 400 mL.

Effect of type and volume of the eluent

Acidified methanol can be used for the quantitative elution of U retained in the column. Though, the recoveries were dependent on the nature and concentration of eluent. The elution of sorbed U, MWCNTs was also studied by using methanol, HCl, HNO₃ and $H₂SO₄$ of various concentrations in methanol. An aliquot of 100 mL containing 10 μ g of U was taken in beaker and after the adjustment of its pH was passed through the column at optimum flow rate. Elution was carried out with HCl, $HNO₃$ and $H₂SO₄$ with methanol in different ratios. The results were given in Table [1](#page-2-0). The recovery increased with the increase in concentration of the eluent. Quantitative recovery (100 %) of U was achievable while using a mixture of and 1 N HCl in methanol but declined when $HNO₃$ or $H₂SO₄$ were used in place of HCl. Maximum recovery was obtained when 10 mL of I N HCl with methanol in 1:3 was used as eluent (Fig. 3). Similarly another problem with $HNO₃$ as elutant is its high blank values in the spectrophotometric determination of U while using A–III.

Flow rates of sample and eluent

Sample solutions can be passed through the MWCNTs column at a higher flow rate without affecting the adsorption as a bulk volume of sample solution is required in solid phase extraction procedures. The effect of the flow rate of the sample and eluent solutions on the adsorption was examined in the wide range of $1-10$ mL min⁻¹. The results are given in Fig. 4. The effect of flow rate of the sample solutions had no obvious effect on the adsorption of U at the range of $1-5 \text{ mL min}^{-1}$, while the case of eluent was quite different. At lower flow rate of eluent $(1-3 \text{ mL min}^{-1})$, the recoveries were quantitative while at higher flow rates the recoveries decreased. Thus 5 mL min^{-1} was chosen as the flow rate of the sample and 2 mL min^{-1} for eluent solutions in the following experiments.

Effect of reagent concentration

Uranium reacts with A-III in a wide pH range to form a 1:1 complex [[51\]](#page-6-0). The effect of A-III concentration on the absorbance of the complex was investigated with changing the volume of 0.1 % concentration. Effect of ligand was studied in the range of $200-1,000 \mu L$; at $200 \mu L$ the

Fig. 3 Effect of eluent volume (mL) on % recovery of Uranium

Fig. 4 Effect of flow rates of sample and eluent $(mL min^{-1})$

Fig. 5 Effect of 0.1 % (w/v) A-III volume (μ L) on % recovery of Uranium

recovery was about 85 % which increased with increasing the volume. It is clear from Fig. 5 that the maximum recovery was attained with 500 μ L of 0.1 % A-III; above this volume up to $1,000 \mu L$ the absorbance remained

Table 2 Influence of interfering ions in the determination of U

Foreign ions	Amount (μg)	% Recovery
Na^{+} , Mg^{+2}	6000	99
Ca^{+2}	3000	99
Fe^{+3} , Al^{+3} , Cu^{+2}	2500	100
Ni^{+2} , Zn^{+2} , Mn^{+2} , Pb^{+2} , Co^{+2} , Cd^{+2}	1000	97
Th^{+4}	500	97
Cl^{-}	4000	100
F^-	3000	98
NO_3^- , CO_3^-	1000	100
PO_4^- , SO_4^-	500	100

Table 3 Accuracy of analytical method based on added/observed concentration of U (μ g g⁻¹) in ore samples (*n* = 5)

unchanged. Therefore, 500 μ L of 0.1 % A-III was used in all further experiments.

Effect of diverse ions on the recovery

The tolerance for diverse ions on U determination was scrutinized separately by adding a rather higher concentration of different foreign ions. The probable interference

expected from various ions, which may be present in mineral samples, was examined by adding them in the aqueous solutions of U and following the above discussed procedure. We found that a number of foreign ions were not interfering in the analysis of U even at higher concentrations and the recoveries were quantitative (Table 2).

Analytical features of the method

The accuracy of the proposed procedure was evaluated by U determination in ore sample solutions collected from Sivas region of Turkey. Ore samples were digested and spiked with a known amount of U. 10 mL of sample was taken in calibrated flask and the proposed procedure was applied.

Linear plot was obtained between the measured absorbance and the U concentrations. The concentration range of 6–250 μ g L⁻¹ of U was selected as linear calibration range. Pre-concentration factor of 80 is achievable if sample volume of 400 mL and eluent of 5 mL is passed through column. The detection limit (LOD) and quantification limit (LOQ) were 1.9 and 6.0 μ g L⁻¹, respectively. The accuracy of our developed method was checked by determining U in samples by standard addition method (Table 3). Different authors have used different combinations of solid phases and detection procedures for U pre-concentration followed by determination. So far our proposed procedure for U pre-concentration and determination is simple, cost effective and quicker and additionally, analytical features of our proposed methods are comparable with them (Table 4).

* LIF Laser induced fluorimeter

Conclusions

Though a number of different techniques have been reported for U determination, but our proposed procedure is relatively rapid, convenient, low cost, sensitive and selective for U determination at trace amount. Additionally, because of eliminating the synthesis and use of any complexing ligand in column for sorption has reduced the risk of contamination to a higher extent. Furthermore, the detection limit and relative standard deviation of our proposed method are better than those techniques which were developed earlier. The possible interference of foreign ions like, alkaline, earth alkaline and other ions was investigated and no prominent effect was encountered. The accuracy was authenticated by analyzing the spiked ore samples.

Acknowledgments The authors are grateful to the Scientific and Technological Research Council of Turkey (TÜBİTAK) for "2216 Research Fellowship Programme for Foreign Citizens'' and financial supports. The authors are also grateful for the financial support of the Unit of the Scientific Research Project of Erciyes University (Project no: FBA-12-3822).

References

- 1. Yemalyanov VS, Yevstyukhin AI (1969) The Metallurgy of Nuclear Fuel. Atomizdat, Moscow
- 2. Agency for toxic substances and disease registry, US public health service (2000), Chapman and Hall, New York
- 3. Suresh A, Patre DK, Srinivasan TG, Rao PRV (2002) A new procedure for the spectrophotometric determination of uranium(VI) in the presence of a large excess of thorium(IV). Spectrochimica Acta Part A 58:341–347
- 4. Rathore DPS (2008) Advances in technologies for the measurement of uranium in diverse matrices. Talanta 77:9–20
- 5. Sundar U, Ramamurthy V, Buche V, Rao DN, Sivadasan PC, Yadav RB (2007) Rapid measurements of concentrations of natural uranium in process stream samples via gamma spectrometry at an extraction facility. Talanta 73:476–482
- 6. Piech R, Bas B, Kubiak WW (2007) The cyclic renewable mercury film silver based electrode for determination of uranium(VI) traces using adsorptive stripping voltammetry. Electroanalysis 19:2343–2350
- 7. Nivens DA, Zhang Y, Angel SM (2002) Detection of uranyl ion via fluorescence quenching and photochemical oxidation of calcein. J Photochem Photobiol A Chem 152:167–173
- 8. Anwar M, Mohammad D (1989) Potentiometric determination of free acidity and uranium in uranyl nitrate solutions. J Radioanal Nucl Chem 134:45–51
- 9. Lakar MM, Branica M (1989) Stripping voltammetric determination of trace levels of uranium by synergic adsorption. Anal Chim Acta 221:279–287
- 10. McMahon AW (1993) Application of analytical methods based on X-ray spectroscopy to the determination of radio nuclides. Sci Total Environ 130:285–295
- 11. Aydin FA, Soylak M (2007) Solid phase extraction and preconcentration of uranium(VI) and thorium(IV) on Duolite XAD761 prior to their inductively coupled plasma mass spectrometric determination. Talanta 72:192–197
- 12. Rozmaric M, Ivsic AG, Grahek Z (2007) Determination of uranium and thorium in complex samples using chromatographic separation, ICP-MS and spectrophotometric detection. Talanta 80:352–362
- 13. Shinotsuka K, Ebihara M (1997) Precise determination of rare earth elements, thorium and uranium in chondritic meteorites by inductively coupled plasma mass spectrometry a comparative study with radiochemical neutron activation analysis. Anal Chim Acta 338:237–246
- 14. Benedik L, Vasile M, Spasova Y, Wätjen U (2009) Sequential determination of 210 Po and uranium radioisotopes in drinking water by alpha-particle spectrometry. Appl Radiat Isot 67:770– 775
- 15. Khan MH, Warwick P, Evans N (2006) Spectrophotometric determination of uranium with arsenazo-III in perchloric acid. Chemosphere 63:1165–1169
- 16. Hosseini MS, Raissi H, Yavari HR (2006) Synergistic flotation of U(VI)–alizarin complex with some diamines followed by spectrophotometric determination of U(VI) using 4,4-diaminophenylmethane. Anal Chim Acta 559:181–185
- 17. Ru Y, Yan L, Guilan S, Tao W, Jiaomai P (1995) Spectrophotometric determination of uranium in natural water with the new chromogenic reagent p-carboxychlorophosphonazo. Anal Chim Acta 314:95–99
- 18. Leong CL, Florence TM, Farrar Y (1973) Spectrophotometric determination of uranium(VI) with chromazurol S and cetylpyridinium bromide. Anal Chem 45:201–203
- 19. Florence TM, Farrar YJ (1963) Spectrophotometric determination of uranium with 4-(2-pyridylazo) resorcinol. Anal Chem 35: 1613–1616
- 20. Lutfullah Alam MN, Rahman N, Azmi SNH (2008) Optimized and validated spectrophotometric method for the determination of uranium(VI) via complexation with meloxicam. J Hazard Mater 155:261–268
- 21. Jie C, Zaijun L, Ming L (2008) Spectrophotometric determination of ultra trace uranium(VI) in seawater after extractive pre-concentration with ionic liquid and dimethylphenylazosalicylfluorone. Int J Environ Anal Chem 88:583–590
- 22. Motojima K, Yoshida H, Izawa K (1960) Spectrophotometric determination of small amounts of uranium with 8-quinolinol. Anal Chem 32:1083–1085
- 23. Ramakrishna TV, Murthy RSS (1980) Spectrophotometric determination of uranium with anthranilic acid and rhodamine 6G. Talanta 27:442–444
- 24. Madrakian T, Afkhami A, Mousavi A (2007) Spectrophotometric determination of trace amounts of uranium(VI) in water samples after mixed micelle-mediated extraction. Talanta 71:610–614
- 25. Das SK, Kedari CS, Tripathi SC (2010) Spectrophotometric determination of trace amount of uranium (VI) in different aqueous and organic streams of nuclear fuel processing using 2-(5-bromo-2-pyridylazo-5-diethylaminophenol). J Radioanal Nucl Chem 285:675–681
- 26. Shemirani F, Kozani RR, Jamali MR, Assadi Y, Milani SMR (2005) Micelle-mediated extraction for direct spectrophotometric determination of trace uranium(VI) in water samples. Sep Sci Technol 40:2527–2537
- 27. Currah JE, Beamish FE (1947) Colorimetric determination of uranium with thiocyanate. Anal Chem 19:609–612
- 28. Faust SD, Aly OM (1987) Adsorption processes for water treatment. Butterworth, Boston
- 29. El Aamrani FZ, Duro L, de Pablo J, Bruno J (2002) Experimental study and modeling of the sorption of uranium (VI) onto olivinerock. Appl Geochem 17:399–408
- 30. Sylwester ER, Hudson EA, Allen PG (2000) The structure of uranium(VI) sorption complexes on silica, alumina, and montmorillonite. Geochim Cosmochim Acta 64:2431–2438
- 31. Xu Y, Zondlo JW, Finklea HO, Brennsteiner A (2000) Electrosorption of uranium on carbon fibers as a means of environmental remediation. Fuel Process Technol 68:189–208
- 32. Akhtar K, Akhtar MW, Khalid AM (2007) Removal and recovery of uranium from aqueous solutions by Trichoderma harzianum. Water Res 41:1366–1378
- 33. Niu Z, Fan Q, Wang W, Xu J, Chen L, Wu W (2009) Effect of pH, ionic strength and humic acid on the sorption of uranium(VI) to attapulgite. Appl Radiat Isot 67:1582–1590
- 34. Han R, Zou W, Wang Y, Zhu L (2007) Removal of uranium(VI) from aqueous solutions by manganese oxide coated zeolite: discussion of adsorption isotherms and pH effect. J Environ Radioact 93:127–143
- 35. Fasfous II, Dawoud JN (2012) Uranium(VI) Sorption by multiwalled carbon nanotubes from aqueous solution. Appl Surf Sci 259:433–440
- 36. Mauter MS, Elimelech M (2008) Environmental applications of carbon-based nanomaterials. Environ Sci Technol 42:5843–5859
- 37. Salam MA, Burk R (2010) Thermodynamics and kinetics studies of pentachlorophenol adsorption from aqueous solutions by multi-walled carbon nanotubes. Water Air Soil Pollut 210: 101–111
- 38. Machado FM, Bergmann CP, Fernandes THM, Lima EC, Royer B, Calvete T, Fagan SB (2011) Adsorption of Reactive Red M-2BE dye from water solutions by multi-walled carbon nanotubes and activated carbon. J Hazard Mater 192:1122–1131
- 39. Chen GC, Shan XQ, Zhou YQ, Shen X, Huang HL, Khan SU (2009) Adsorption kinetics, isotherms and thermodynamics of atrazine on surface oxidized multiwalled carbon nanotubes. J Hazard Mater 169:912–918
- 40. Rao GP, Lu C, Su F (2007) Sorption of divalent metal ions from aqueous solution by carbon nanotubes: a review. Sep Purif Technol 58:224–231
- 41. Suárez B, Simonet BM, Cárdenas S, Valcárcel M (2007) Determination of non-steroidal anti-inflammatory drugs in urine by combining an immobilized carboxylated carbon nanotubes mini column for solid-phase extraction with capillary electrophoresismass spectrometry. J Chromatogr A 1159:203–207
- 42. Carabineiro SAC, Thavorn-amornsri T, Pereira MFR, Serp P, Figueiredo JL (2012) Comparison between activated carbon, carbon xerogel and carbon nanotubes for the adsorption of the antibiotic ciprofloxacin. Catal Today 186:29–34
- 43. Lu C, Su F (2007) Adsorption of natural organic matter by carbon nanotubes. Sep Purif Technol 58:113–121
- 44. Yang K, Xing B (2009) Adsorption of fulvic acid by carbon nanotubes from water. Environ Pollut 157:1095–1100
- 45. Yan H, Gong A, He H, Zhou J, Wei Y, Lv L (2006) Adsorption of micro cystines by carbon nanotubes. Chemosphere 62:142–148
- 46. Khan MH, Hasany SM, Khan MA (1994) Spectrophotometric determination of micro amounts of thorium with disodium salt of 2-(2-hydroxy-3,6-disulfo-1-naphthylazo) benzenearsonic acid (thorin) as a chromogenic reagent. J Radioanal Nucl Chem 188:341–353
- 47. Rohwer H, Rheeder N, Hosten E (1997) Interactions of uranium and thorium with arsenazo-III in an aqueous medium. Anal Chim Acta 341:263–268
- 48. Khan MH, Ali A, Khan NN (2001) Spectrophotometric determination of thorium with disodium salt of arsenazo-III in perchloric acid. J Radioanal Nucl Chem 353:353–357
- 49. Goossens J, Moens L, Dams R (1995) Inductively coupled plasma mass spectrometric determination of heavy metals in soil and sludge candidate reference materials. Anal Chim Acta 304:307–315
- 50. Singh BN, Maiti B (2006) Separation and pre-concentration of U(VI) on XAD-4 modified with 8-hydroxy quinoline. Talanta 69:393–396
- 51. Hosten E, Rohwer HE (1997) Complexation reactions of uranyl with arsenazo III. Anal Chim Acta 355:95–100
- 52. Behpour M, Ghoreishi SM, Nikkhah Qamsari Z, Samiei M, Soltani N (2010) Solid phase extraction of uranium by naphthalenemethyltrioctylammonium chloride and arsenazo(III) adsorbent and subsequent spectrophotometric determination. Chin J Chem 28:1457–1462
- 53. Aydin FA, Soylak M (2007) Solid phase extraction and preconcentration of uranium(VI) and thorium(IV) on Duolite XAD761 prior to their inductively coupled plasma mass spectrometric determination. Talanta 72:187–192
- 54. Cyriac B, Balaji BK (2010) A novel method of synthesizing solid phase adsorbent silica modified with xylenol orange: application for separation, pre-concentration and determination of uranium in calcium rich hydro-geochemical samples and sea water—Part 1. Microchim Acta 171:33–40
- 55. Kazeraninejad M, Haji Shabani AM, Dadfarnia S, Ahmadi SH (2011) Solid phase extraction of trace amounts of U(VI) from water samples using 8-hydroxyquinoline immobilized on surfactant coated alumina. J Anal Chem 66:11–15
- 56. Amin AS (2012) Solid-phase extraction using polymer-based cartridge modified with 2-(2-benzothiazolylazo)-3-hydroxyphenol for pre-concentration of U(VI) ions from water and real samples. Spectrosc Lett 45:246–255
- 57. Sadeghi S, Sheikhzadeh E (2008) Solid phase extraction using silica gel functionalized with sulfasalazine for pre-concentration of U(VI) ions from water samples. Microchim Acta 163:313–320
- 58. Gladis JM, Rao TP (2002) Solid phase-extractive pre-concentration of uranium on to 5,7-dichloroquinoline-8-ol modified napthalene. Anal Lett 35:501–515
- 59. Liu Y, Cao X, Le Z, Luo M, Xub W, Huang G (2010) Preconcentration and determination of trace U(VI) in environments using ion-imprinted chitosan resin via solid phase extraction. J Braz Chem Soc 21:533–540
- 60. Ozdemir S, Kilinc E (2012) Geobacillus thermoleovorans immobilized on Amberlite XAD-4 resin as a biosorbent for solid phase extraction of U(VI) prior to its spectrophotometric determination. Microchim Acta. doi:[10.1007/s00604-012-0841-2](http://dx.doi.org/10.1007/s00604-012-0841-2)