Complexation and Precipitation of Arsenate and Iron Species in Sodium Perchlorate Solutions at 25 °C

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The complexation of As(V) in aqueous solutions in the presence of iron(III) was investigated spectrophotometrically with both variable and constant ionic strengths. The determined thermodynamic and stoichiometric formation constants of the FeHAsO $_4^+$ species are $\log_{10} \circ \beta = 9.21 \pm 0.01$ and $\log_{10} I \beta$ (1.0 mol·dm⁻³ NaClO₄) = 7.78 ± 0.01, respectively. The numerical treatment of the obtained spectral data was performed with the SPECA program. The analysis required the consideration of the hydrolysis of Fe(III) and the protonation of As(V) in the pH range studied. No significant hydrolysis was observed because of the low pH values (pH < 2.5) involved. The stabilities of the solid Fe(III) arsenates was established by solubility experiments. All of the solubility experiments were performed in aqueous NaClO₄ solutions at constant ionic strength (1.0 mol·dm⁻³) and at 25 °C. The experimental data were consistent with FeAsO₄·2H₂O being the solid phase ($\log_{10} {}^{\circ}K_{so} = -24.30 \pm 0.08$). The corresponding thermodynamic constants were computed by means of the Modified Bromley's Methodology (MBM) that describes the variation of the activity coefficients of all of the ions involved in the complexation and precipitation equilibria with the medium and ionic strength. Finally, the solid phase obtained in this work was also characterized by FT-IR and FT-Raman spectroscopies, and the hydration of the solid iron arsenate was confirmed by X-ray diffraction data.

KEY WORDS: Iron(III) arsenate; complexation equilibria; precipitation equilibria; thermodynamic model; spectroscopic characterization.

1. INTRODUCTION

One of the main factors controlling the concentration of inorganic arsenic in natural aquatic systems is the presence of ligands in the sediments that react with soluble arsenate species by means of precipitation or complexation reactions.

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Along this line, some authors have emphasized that the adsorption and mobility processes of the arsenate anion into/from natural sediments is mainly controlled by the amount of the iron(III) oxides in the solid phases.⁽¹⁻⁴⁾

Furthermore, a better knowledge of arsenic immobilization mechanisms in natural systems is necessary to develop and to implement removal or protection processes. To achieve this, different processes such as reverse osmosis, coagulation and filtration, or precipitation with soluble iron(III) salts should be implemented in the treatment plants. For the control of arsenate mobility or adsorption by iron(III), and the implementation of removal processes for the inorganic arsenic, it is important to be able to model the interactions of the primary form of arsenic (arsenate) with iron(III), and to define how those interactions are affected by changes in the composition of the medium or ionic strength of the natural waters.

Published information indicates that $\text{FeAsO}_4(s)^{(5-7)}$ is the main solid phase that controls the solubility of arsenate in natural waters. In 1996 Leblanc⁽⁸⁾ described the formation of some amorphous ferric arsenates in aqueous solutions. However, there were no data reported for the variation of the precipitation constant with ionic strength and the solubility information was reported as the solubility product at infinite dilution (thermodynamic solubility product).

Therefore, it is necessary to perform solubility experiments at different ionic strengths in order to characterize the corresponding ferric arsenates as a function of the medium composition and to determine the variation of the precipitation constant with ionic strength. Then, the stoichiometric solubility constant obtained experimentally for a given ionic medium, *e.g.*, NaClO₄, can be extrapolated to infinite dilution by using the Modified Bromley's Methodology (MBM).^(9,10)

In order to perform a correct extrapolation and interpretation of the solubility data, all other possible equilibria affecting both Fe(III) and As(V) in solution (hydrolysis and/or complexation) need to be considered. For this purpose, the chemical models of Fe(III)⁽¹¹⁾ and As(V)⁽¹²⁾ can be used. These chemical models, constructed by means of the MBM, define all of the interaction parameters of Fe(III) and As(V) with the ions of the considered ionic medium, along with the corresponding thermodynamic constants for the hydrolysis of the Fe³⁺ cation and the protonation constants of the AsO₄³⁻ anion, respectively. Thus, the only reactions to be studied, besides the solubility experiments, are the formation of iron(III) arsenate complexes.

In this work, the complex formation equilibrium between As(V) and Fe(III) was studied by UV-vis spectrometry as suggested in the literature⁽¹³⁾ and the precipitation equilibrium between As(V) and Fe(III) was studied through solubility experiments. For both types of measurements, the experiments were performed in NaClO₄ solutions of constant strength. The solubility study falls into three clearly different parts: (a) preparation of the solid iron arsenate, (b) the determination of its stoichiometry, and (c) the calculation of the solubility products for the precipitate. Moreover, the spectroscopic characterization of the precipitated iron arsenate was

carried out by means of FT-IR and FT-Raman spectroscopies, along with an X-ray diffraction study.

2. EXPERIMENTAL

2.1. Reagents and Solutions

The chemicals used in the experiments were all of analytical reragent grade, with purities >99%. All solutions were prepared in freshly purified Milli-Q water.

Sodium arsenate (Aldrich, p.a.) stock solutions were standardized by acidbase titrations using bromocresol green as the indicator.⁽¹⁴⁾ Ferric nitrate (Merck, p.a.) stock solutions were standardized by titration with EDTA (Panreac, p.a.) with 5-sulphosalicilic acid as indicator.⁽¹⁴⁾

The ionic medium was sodium perchlorate (Fluka, p.a.), whose concentrations were determined gravimetrically after evaporation of aliquots of the stock solutions and drying at 110 °C.

Stock solutions of perchloric acid (Fluka, puriss, p.a.) were also used and their concentrations were standardized against tris(hydroxymethyl)aminomethane (Merck, p.a.) using bromocresol green as the indicator.⁽¹⁵⁾

2.2. Experimental Procedure

For the complexation studies, solutions were prepared with different Fe(III) $(0.001 \text{ and } 0.0005 \text{ mol} \cdot \text{dm}^{-3})$ and As(V) $(0.0001, 0.0005, \text{ and } 0.00029 \text{ mol} \cdot \text{dm}^{-3})$ concentrations. In addition, experiments were performed with solutions not containing Fe(III) but containing arsenate at the same concentrations mentioned above. The experiments were performed in both $1.0 \text{ mol} \cdot \text{dm}^{-3}$ NaClO₄ constant ionic medium and with variable ionic strength. The pH values were varied by the addition of HClO₄ at the same ionic strength. The solutions were equilibrated in an oil bath thermostat at 25 °C for 1 week. After that, the pH value was measured and the UV-vis spectrum was obtained with a diode array spectrophotometer (Hewlett Packard 8452A).

The solubility experiments were performed by generating the precipitates at a constant concentration of both iron and arsenate ions $(1/1 \text{ ratio at } 0.014 \text{ mol} \cdot \text{dm}^{-3})$. The experiments were performed in $1.0 \text{ mol} \cdot \text{dm}^{-3}$ NaClO₄ ionic medium and the pH was controlled by adding HClO₄ at the same ionic strength. In all cases, the solutions were mixed with stirring to avoid local precipitate formation and the temperature was controlled at 25 °C in a thermostatted oil bath. In order to determine when equilibrium was reached, the pH of the saturated solutions was measured periodically until constant values were obtained (Crison micropH 2000). The concentrations of As and Fe in the saturated solutions equilibrated with the solid phases were measured by ICP-AES (ARL Fisons 3410).

The spectroscopic studies were made using a FT-IR spectrometer (Nicolet Nexus), a FT-Raman spectrometer (Nicolet 950) and a Rx diffractometer (Philips PW 1710). After separation from the equilibrated saturated solutions, the precipitates were washed with water several times to eliminate any traces of NaClO₄ (from the ionic medium), Na⁺ (from the arsenate reagent) and NO₃⁻ (from the Fe(III) salt). The precipitates were dried in vacuum for one week. Finally, each precipitate was stored in a desiccator prior to the spectroscopic analyses.

3. RESULTS AND CALCULATIONS

3.1. Spectrophotometric Experimentation

In all of the experiments, the E° values of the electrolytic cell were determined using Gran's method,⁽¹⁶⁾ by means of titrations with HClO₄ in the corresponding ionic medium and ionic strength (1.0 mol·dm⁻³ NaClO₄). The Nernst equation⁽¹⁷⁾ was then used to calculate the pH values, defined as $-\log_{10}$ [H⁺], in all of the solutions. The free concentration of hydrogen ions (H⁺) was calculated from the measured *E* by using equations:

$$E = E^{\circ} + 59.16 \log_{10}[\mathrm{H}^+] + E_{i,h} \tag{1}$$

and

$$E_{j,h} = j_{ac}[\mathrm{H}^+] + j_{bas}K_{w}[\mathrm{H}^+]^{-1}$$
(2)

where j_{ac} and j_{bas} are the acidic and the basic solution liquid junction potential coefficients, respectively, and K_W is the autoprotolysis constant of water for the ionic medium at the given ionic strength. All of these parameters were determined at the given ionic strength by means of separate potentiometric titrations in the ionic medium. They were calculated using Gran's method and refined numerically with the Model Function⁽¹⁸⁾ version of the Letagrop program.⁽¹⁹⁾ Because the equation used to calculate the *h* value can not be solved analytically, an iterative procedure, the Newton–Raphson method,⁽¹⁷⁾ was used by implementing it in an Excel[®] spread sheet.⁽²⁰⁾

A numerical analysis of absorbtivity *versus* pH for each dataset was performed by means of the SPECA program,⁽²¹⁾ in order to determine both the thermodynamic (from the experiments without constant ionic medium) and stoichiometric constants (from the experiments in NaClO₄ ionic media). In addition, the absorbitivity coefficients of all species involved in the complexation equilibrium were calculated using the SPECA program.

The isosbestic point ($\lambda = 270$ nm), characteristic of the Fe³⁺ spectra (Fig. 1a), disappeared upon the addition of arsenate to the solution (Fig. 1b). This observation can only be explained by the formation of iron arsenate complexes. In order to determine the wavelengths containing the most important spectral information, a



Fig. 1. Variation of the (a) Fe(III) (0.001 mol·dm⁻³) and (b) Fe(III) (0.001 mol·dm⁻³) $+ HAsO_4^{2-}$ (0.0005 mol·dm⁻³) spectra with pH (0.9 < pH < 2.2).

computer routine in MATLAB $4.0^{(22)}$ was created in our laboratory. The results obtained indicate that 290, 300 and 310 nm were the most important wavelengths. Prior to the numerical treatment, it was verified that there were no spectral features due to the NaClO₄, HClO₄ and HAsO₄²⁻ species in the wavelength range of interest.

	$I \rightarrow 0$	$1.0 \mathrm{mol} \cdot \mathrm{dm}^{-3}$
$\log_{10}\beta$	9.21 ± 0.01	7.78 ± 0.01
E290	3888 ± 25	5771 ± 14
E300	3075 ± 22	4707 ± 16
ε_{310}	2180 ± 17	3414 ± 11

Table I. Absorbtivity Coefficients and Thermodynamicand Stoichiometric Formation Constants of the Soluble $FeHAsO_4^+$ Complex at 25 °C

The fittings of the absorbtivity *versus* pH data were performed assuming the presence of only Fe⁺³ and Fe(OH)²⁺ species for Fe(III)^{[11,23)} and the protonated forms of the arsenate anion.⁽¹²⁾ Several possible chemical models with different assumed iron arsenate complexes were examined by the SPECA program. The results show that only one arsenate complex is present in the solutions, FeHAsO₄⁺, whose formation equilibrium can be expressed as

$$\operatorname{Fe}^{3+} + \operatorname{HAsO}_4^{2-} \longleftrightarrow \operatorname{FeHAsO}_4^+$$
 (3)

This equilibrium reaction is adequate to explain the whole set of spectrometric data with a global standard deviation of $\sigma(A) = \pm 0.005$.

The stoichiometric and thermodynamic constants of equilibrium (1) and the absorbitivity coefficient of the new complex computed by SPECA are summarized in Table I. As examples, Figs. 2 and 3 show the fittings obtained from the numerical treatment.



Fig. 2. Variation of the absorbtivity *versus* pH data in solutions with an Fe³⁺ concentration of 0.0005 mol·dm⁻³ and AsO_4^{3-} concentrations of 0.001 ($\circ-\circ$) or 0.0005 ($\Box -\Box$) mol·dm⁻³ at 310 nm.



Fig. 3. Variation of the absorbtivity *versus* pH data in solutions with an Fe³⁺ concentration of 0.001 mol·dm⁻³ (\circ - \circ) and AsO₄³⁻ concentrations of 0.001 (Δ - Δ), 0.0005 (\diamond - \diamond) and 0.00029 (\Box - \Box) mol·dm⁻³ in 1.0 mol·dm⁻³ NaClO₄.

рН	$C_{\rm Fe} ({\rm mol}\cdot{\rm dm}^{-3})$	$C_{\rm As} ({\rm mol}\cdot{\rm dm}^{-3})$
0.98	0.00772	0.00982
1.15	0.00450	0.00599
1.26	0.00151	0.00225
1.36	0.00072	0.00123
1.45	0.00043	0.00089
1.49	0.00038	0.00089
1.56	0.00026	0.00080
1.64	0.00021	0.00077
1.68	0.00013	0.00071
1.73	9.5×10^{-5}	0.00051
1.80	8.2×10^{-5}	0.00060
1.82	9.7×10^{-5}	0.00062
1.85	6.9×10^{-5}	0.00061
1.91	4.7×10^{-5}	0.00058
1.95	4.5×10^{-5}	0.00061
1.99	4.0×10^{-5}	0.00064
2.17	1.5×10^{-5}	0.00497
2.83	3.1×10^{-6}	0.00080
7.75	2.3×10^{-6}	0.00303
12.33	2.0×10^{-6}	0.01050

 Table II. Raw Data (Solution Concentrations)

 Obtained in the Solubility Experiments

3.2. Solubility Data

The raw solubility data are listed in Table II. The concentration profiles obtained in the solubility experiments imply the presence of solid iron arsenate, since there was a simultaneous decrease of the concentrations of both the arsenate and metal in the range of 1 < pH < 2.5 (Fig. 4). For pH > 2.5, the concentration profiles were explained by the formation of a Fe(OH)₃(s) precipitate with the liberation of arsenate anion to the solution phase. As it can be seen in Fig. 4, the concentration profile of the total arsenate concentration did not show the theoretical redissolution behavior. This fact can be explained by the well-known adsorption of the arsenate anion on the precipitated Fe(OH)₃(s).⁽¹⁻⁴⁾

The formation of the new precipitate can be described by the general equilibrium:

$$p\mathrm{F}\mathrm{e}^{3+} + q\mathrm{H}^{+} + r\mathrm{AsO}_{4}^{3-} \longleftrightarrow (\mathrm{F}\mathrm{e}^{3+})_{p} \cdot (\mathrm{H}^{+})_{q} \cdot (\mathrm{AsO}_{4}^{3-})_{r}(s)$$
(4)

with the corresponding stoichiometric solubility product:

$$K_{s0}^{I} = [\text{Fe}^{3+}]^{p} \cdot [\text{H}^{+}]^{q} \cdot [\text{AsO}_{4}^{3-}]^{r}$$
(5)

where [] denotes the molar concentrations of the species involved in the precipitation equilibrium.



Fig. 4. Variation of the total concentration As(V) in solution. The predicted solubility of As(V) in 1.0 mol·dm⁻³ NaClO₄ is also shown. The (—) line is the theoretical solubility profile for As(V) without adsorption occurring.

The values of $[H^+]$ were calculated from the experimental pH values. The free $[AsO_4^{3-}]$ and $[Fe^{3+}]$ concentrations were obtained from the mass-balance equations by using the experimental total concentrations of As(V) and Fe(III), the pH value for each solution, and the chemical model for the soluble species: Fe³⁺, Fe(OH)²⁺, FeHAsO_4^+, H_3AsO_4, H_2AsO_4^-, HAsO_4^{2-}, and AsO_4^{3-} at 1.0 mol·dm⁻³ NaClO₄ ionic strength. Using these data, a graphical treatment was performed by considering the relationship between the concentrations of the AsO_4^{3-} and H⁺ species on one hand and between Fe³⁺ and H⁺ on the other. The reaction stoichiometries (*p:q:r*) can be determined from the slopes of the different straight lines plotted in Fig. 5a and b.

Because neither the $[Fe^{3+}]$ nor the $[AsO_4^{3-}]$ had constant values at any pH value, the saturation condition (5) could be extended to the following three equations:

$$K_{s0}^{I} = (cte_2)(cte_1) = [Fe^{3+}]^{p} \cdot [H^+]^{(q_1+q_2)} \cdot [AsO_4^{3-}]^{r}$$
(6)

with

$$cte_2 = [\mathrm{Fe}^{3+}]^p \cdot [\mathrm{H}^+]^{q_1}$$
 (7)

and

$$cte_1 = [\mathrm{H}^+]^{q_2} \cdot [\mathrm{AsO}_4^{3-}]^r$$
 (8)

Moreover, the electroneutrality condition for any solid phase requires that

$$2p + q = 3r \tag{9}$$



Fig. 5. Plot of $\log_{10} C_i$ versus pH for the free concentrations of (a) Fe^{3+} and (b) AsO_4^{3-} species; and (c) plot of $\log_{10} C_{arsenate}$ versus $\log_{10} C_{iron}$ for the free concentrations (—, theoretical line with slope = 1.0).

Complexation and Precipitation of Arsenate and Iron

with

$$q = q_1 + q_2 \tag{10}$$

Applying condition (7) to Fig. 5a, the slope of this plot (= -2.2) gives the value of the ratio q_1/p . In the same way, condition (8) gives the value of the ratio q_2/r for the slope (= 2.2) of Fig. 5b and, taking into account Eqs. (9) and (10), the following indexes were obtained: p = 1, q = 0 ($q_1 = -2.2$, $q_2 = 2.2$) and r = 1. In other words, FeAsO₄(s) is the composition obtained for the solid phase at pH < 2.5.

From the previously calculated [H⁺], [AsO₄³⁻] and [Fe³⁺] concentrations, the stoichiometric solubility constant of the precipitate was calculated at each experimental point (see Fig. 6) using Eq. (5). The obtained value for the stoichiometric solubility product of FeAsO₄(s) is $\log_{10}{}^{I}K_{s0} = -21.08 \pm 0.07$ (in 1.0 mol·dm⁻³ NaClO₄ at 25 °C). Figure 6 shows that there is no systematic drift in the $\log_{10}{}^{I}K_{s0}$ values for the whole investigated pH range.

4. DISCUSSION

The presence of hydration waters in the precipitated FeAsO₄·xH₂O(s) was confirmed by the characteristic signals of H—O—H in the spectral regions 3560 to 3400 cm⁻¹ and 1650 to 1600 cm⁻¹, which were obtained from the IR spectra (see Table III). The value of "x" in the precipitate was obtained in this work by comparison of the data obtained from X-ray diffraction with those reported in the literature. In that manner, the formulation of FeAsO₄·2H₂O was established.

Inasmuch as this solid was obtained by mixing Na_2HAsO_4 and $Fe(NO_3)_3$ salts, the IR signals of $Na_2HAsO_4 \cdot 7H_2O$ are also reported in Table III for the



Fig. 6. Stoichiometric solubility product of FeAsO₄ obtained from the solubility experiments.

	=		-
FeAsO ₄ ·2H ₂ O	Na ₂ HAsO ₄ ·7H ₂ O ^a	AsO_4^{3-b}	Assignment
3545			H-O-H stretching
3467	344.3		H-O-H stretching
3404	335.9		H-O-H stretching
1641	168.8		H-O-H bending
1618	162.9		H-O-H bending
	859	878	(v_3) AsO ₄ ³⁻
836 840 ^a	833	837	(v_1) As $O_4^{\frac{3}{4}}$
470 472 ^{<i>a</i>}		463	$(v_4) \operatorname{AsO}_4^{\overline{3}-}$

 Table III.
 FT-IR Spectra (cm⁻¹) Obtained for Iron Arsenate Precipitates Compared to Those from This Study

^aFrom Miller.⁽²⁴⁾

^bFrom Nakamoto.⁽²³⁾

purpose of confirming that disodium arsenate was absent from the solid phase. The reported frequencies for dissolved AsO₄³⁻ are also known.⁽²⁴⁾ The H–O–H stretching bands are well defined for the hydrated arsenate ion, as are the two H–O–H bending signals. These five frequencies are different from those found experimentally for Na₂HAsO₄·7H₂O. Two different characteristic arsenate bands occurring at 836 cm⁻¹ (ν_1) and 470 cm⁻¹ (ν_4) were observed for the iron arsenate obtained in this work; the frequencies for the two bands are in good agreement with those reported by Miller.⁽²⁵⁾ However, the band around 590 cm⁻¹ found for Na₂HAsO₄·7H₂O (594 cm⁻¹) by Miller⁽²⁵⁾ was not observed by us. Also, the ν_3 band of arsenate was not observed.

Three As-O-H wagging signals were obtained in the Raman spectrum (Table IV) using the 1064 cm⁻¹ laser as the excitation source. In the Raman spectrum, the v_3 band is well defined (it is not seen in the IR one) but v_1 did not appear. The v_4 vibration band was split into two very sharp bands (482 and 444 cm⁻¹). Also, a split As-O-H rocking band was obtained at 655 and

	1
FeAsO ₄ ·2H ₂ O	Assignment
1146	As-O-H wagging
1097	As-O-H wagging
1088	As-O-H wagging
888	(v_3) AsO ₄ ³⁻
655	As-O-H rocking
620	As-O-H rocking
482	$(\nu_4) \operatorname{AsO}_4^{3-}$
444	$(\nu_4) \operatorname{AsO}_4^{3-}$

Table IV. Raman Spectra (cm⁻¹) Obtained Here for the Iron Arsenate Precipitate

Complexation and Precipitation of Arsenate and Iron

 620 cm^{-1} . These arsenate frequencies are similar to those obtained in a previous work for the calcium and magnesium arsenates.⁽²⁶⁾

The thermodynamic solubility constant was calculated using the relationship with the stoichiometric solubility products, obtained experimentally, and the corresponding activity coefficients:

$$\log_{10} K_{s0}^{I} = \log_{10} K_{s0}^{\circ} - \log_{10} \gamma_{\text{Fe}^{3+}} - \log_{10} \gamma_{\text{AsO}_{4}^{3-}}$$
(11)

The values of γ_X were estimated by using the MBM expression on the molar composition scale:

$$\log_{10} \gamma_X = \frac{-A_c |Z_X|^2 I^{1/2}}{1 + I^{1/2}} + \left(\frac{(0.06 + 0.6B_{YX})|Z_Y Z_X|}{\left(1 + \frac{1.5}{|Z_Y Z_X|}I\right)^2} + B_{YX}\right) \frac{(|Z_Y| + |Z_X|)^2}{4}I$$
(12)

where *I* is the ionic strength on the molar scale and B_{YX} is the ionic interaction parameter of species *X* (AsO₄³⁻ or Fe³⁺) involved in the precipitation equilibrium with the *Y* ions in the ionic medium (Na⁺ or ClO₄⁻). The B_{YX} values in the NaClO₄ ionic medium for Fe³⁺⁽¹¹⁾ and AsO₄³⁻⁽¹²⁾ were determined previously. Therefore, the only unknown parameter, the thermodynamic solubility product, was calculated using Eqs. (11) and (12) and found to be $\log_{10} {}^{\circ}K_{s0} = -24.30 \pm 0.08$.

In the same way, it was possible to estimate the ionic interaction parameter of $FeHAsO_4^+$ with CIO_4^- using Eq. (12) and the MBM to calculate the activity coefficients of the following expression:

$$\log_{10}{}^{I}\beta = \log_{10}{}^{\circ}\beta + \log_{10}{}_{\gamma}_{\mathrm{Fe}^{3+}} + \log_{10}{}_{\gamma}_{\mathrm{HAsO}_{4}^{2-}} - \log_{10}{}_{\gamma}_{\mathrm{FeHAsO}_{4}^{+}}$$
(13)

This equation relates the values of the stoichiometric and thermodynamic formation constants reported in Table I.

The ionic interaction parameter was evaluated by direct calculation using Eq. (13): $B_{\text{FeHAsO}^+} = -0.4814 \pm 0.0009$ on the molar scale.

Finally, the influence of the ionic medium on the distribution of the iron and arsenate ions was examined. For this purpose, the MEDUSA program⁽²⁷⁾ was used to calculate the concentrations of all of the soluble arsenic and iron species as well as of the Fe(III) precipitates. All the thermodynamic formation constants were extrapolated to various environmental conditions by using the MBM to calculate the activity coefficients for each particular medium and ionic strength.

The distribution of the stable Fe(III) species as a function of pH was calculated for river water ($I < 0.1 \text{ mol} \cdot \text{dm}^{-3}$ as NaCl) and estuarine water ($I = 0.7 \text{ mol} \cdot \text{dm}^{-3}$ as NaCl), for typical concentrations of iron and arsenic ions that are



Fig. 7. Distribution diagram of Fe(III) as a function of pH at typical concentrations⁽¹⁾ (As(V) = $60 \ \mu g \cdot L^{-1}$, Fe(III) = $70 \ \mu g \cdot L^{-1}$) for river (a) and estuarine waters (b), considering the formation of FeHAsO₄⁺, Fe(OH)²⁺, Fe(OH)₃(s) and FeAsO₄(s).

found in natural systems. The respective species distribution diagrams are shown in Fig. 7a and b. As can be seen, the iron(III)-arsenate precipitate can form, depending on the pH as well as the composition of the ionic medium. It was found that the pH range for the predominance of FeAsO₄ is higher in river waters than in seawater, and that FeHAsO₄⁺ is not a predominant species at least for [Fe(III)]/[As(V)] = 1

and at the pH range of most natural waters (7.0 to 8.5). The presence of Fe(III) could control the retention of As(V) as a solid phase in river waters, but this is less probable in estuarine waters and seawaters because $Fe(OH)_3(s)$ is more stable than the FeAsO₄(s) under those conditions. For such waters with a relatively high ionic strength, the retention of As(V) in the Fe(OH)₃(s) solid phase can be explained by adsorption processes (as showed in the experimental data ploted in Fig. 4b), rather than arsenate precipitation.

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