

Novel AgCl nanoparticles coupling with PbBiO₂Br nanosheets for green and efficient degradation of antibiotic oxytetracycline hydrochloride under visible-light irradiation

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Abstract

In this work, novel Ag/AgCl/PbBiO₂Br photocatalysts were synthesized via a hydrothermal and in situ photoreaction method. The microstructure, morphology, composition, electrochemical, and optical properties of the synthesized catalysts were investigated by multiple techniques. The obtained Ag/AgCl, PbBiO₂Br, and Ag/AgCl/PbBiO₂Br composites were evaluated via degradation of oxytetracycline (OTC) hydrochloride antibiotic under visible-light irradiation. The results show that the Ag/AgCl/PbBiO₂Br composites are composed of Ag/AgCl nanoparticles (NPs) and PbBiO₂Br nanosheets. The Ag/AgCl/PbBiO₂Br (20.4%) composite exhibits the highest visible-light absorption and best photogenerated charge separation efficiency. The photocatalytic degradation experiments show that all Ag/AgCl/PbBiO₂Br composites exhibit an enhanced degradation activity under visible-light irradiation, and maintain good stability in the photocatalytic process. The Ag/AgCl/PbBiO₂Br (20.4%) composite has the highest degradation activity, which is 1.82 and 2.11 times higher than that of Ag/AgCl and PbBiO₂Br, respectively. The enhanced photocatalytic activity of Ag/AgCl/PbBiO₂Br can be mainly attributed to the fact that the loading of Ag NPs on the surface of the AgCl promotes the separation efficiency of photoinduced charge and enhance the visible-light absorption. Additionally, active species trapping experiments confirm that superoxide radicals ($\cdot O_2^-$), Cl⁰ and holes (h⁺) play an very important role in the degradation process.

1 Introduction

Nowadays, the widespread usage of antibiotics has received the increasing attention because they flow into the water system and cause the water pollution [1]. Oxytetracycline (OTC) hydrochloride is one of the very important antibiotics, extensively used in human and veterinary medicine [2].

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Worryingly, the most of OTC is only partially metabolized in humans or animals and hardly biologically degraded, ultimately released into lakes and rivers. These antibiotics have harmful effects on water environment and human health [3]. Hence numbers of research attempts have been made in the past decades to eliminate these refractory antibiotics, such as electrochemical treatments [4], photoelectron-Fenton [5] and advanced oxidation treatment [6]. However, most of the methods use expensive oxidants. Therefore, it is indispensable to develop novel visible-light-driven photocatalysts [7], which are environmentally friendly and efficient approach to remove OTC from waters.

PbBiO₂Br is an n-type visible-light-driven semiconductor and has attracted more and more attention in recent years, owing to its physicochemical stability, highly anisotropic layered structure, and outstanding photocatalytic performance [8–11]. Unfortunately, the photocatalytic activity of bare PbBiO₂Br is still unsatisfactory owing to its fast recombination rate of photoexcited electron-hole (e⁻/h⁺) pairs [12, 13]. To overcome the above mentioned drawback of PbBiO₂Br and improve the degradation efficiency, constructing semiconductor composites is effective in improving the separation efficiency of photoinduced e^-/h^+ pairs. So far, studies have reported on g-C₃N₄/PbBiO₂Br [14], NbSe₂/PbBiO₂Br [15], PbBiO₂Br/UiO-66-NH₂ [16] Cu₂O/PbBiO₂Br [17], *p*-Ag₂O/*n*-PbBiO₂Br [18], PbBiO₂Br/ BiOBr composites [19], and etc. These heterojunction composites were found to exhibit superior photocatalytic activity. Despite many PbBiO₂Br-based materials have been reported, it is still necessary to be committed to the exploitation of more efficient visible-light-driven PbBiO₂Br-based photocatalysts for making the best use of the solar energies.

The surface plasmon resonance (SPR) strategy is widely used in fabricating efficient visible-light-driven photocatalysts [20]. Because of SPR of noble metal nanoparticles (NPs), the absorption range of visible-light region can be expanded, resulting in the enhanced degradation performance of photocatalysts [21]. Recently, Ag/AgCl has been widely considered as a promising photocatalyst due to its being a p-type SPR structure semiconductor [22]. Furthermore, Ag⁰ NPs dispersed on the surface of AgCl can not only effectively absorb visible light, but also can accelerate the transfer of photo-carriers. By now, a number of Ag/ AgCl-based photocatalysts have been successfully synthesized, such as Ag/AgCl/NaTaO₃ [23], and BiVO₄/MWCNT/ Ag@AgCl [24]. These composite photocatalysts exhibited superior photocatalytic performances. To the best of our knowledge, the coupling of PbBiO2Br nanosheets with Ag/ AgCl NPs has not been reported yet. Hence, we expect that the new Ag/AgCl/PbBiO₂Br composites not only improve the utilization rate of solar energy, but also enhance photocatalytic ability.

Inspired by previous studies, we have successfully fabricated a series of Ag/AgCl/PbBiO₂Br composites by a hydrothermal and in situ photoreaction method. Morphology and microstructure, elements chemical states, optical and electrochemical properties of the Ag/AgCl/PbBiO₂Br composites were systematically studied. Their photocatalytic performances were investigated by the degradation of OTC under visible-light irradiation. The possible enhanced photocatalytic mechanism was also proposed.

2 Experimental

2.1 Preparation of the photocatalysts

PbBiO₂Br nanosheets were synthesized via a facile hydrothermal method [17]. Detailed experimental process was given in Supporting Information (S1). Ag/AgCl/PbBiO₂Br composites were prepared via a photoreduction method. The preparation process was as follows: 1 mmol of PbBiO₂Br was dispersed in deionized water, stirred for 20 min to form uniform suspension A. Then, 1 mmol AgNO₃ was added into the suspension A and stirred for 20 min. Subsequently, 1 mmol NaCl was transferred into the suspension A under strong stirring for 30 min. The resulting mixture was illuminated under a 500 W xenon lamp for 30 min so that the Ag⁺ NPs on the surface of AgCl/PbBiO₂Br were reduced to Ag⁰ NPs. Eventually, the precipitate was filtered, rinsed with deionized water and ethanol, and dried at 80 °C for 24 h. The obtained product, in which the mass ratio of Ag to PbBiO₂Br was 20.4%, was designated as Ag/AgCl/PbBiO₂Br (20.4%). Ag/AgCl/PbBiO₂Br (13.6%), Ag/AgCl/PbBiO₂Br (40.8%), and bare Ag/AgCl were also obtained with the same conditions by changing the content of PbBiO₂Br.

2.2 Characterization and photocatalytic evaluation

The synthesized catalysts were investigated in detail by multiple instruments analyses. The photocatalytic activities of the Ag/AgCl, PbBiO₂Br, and Ag/AgCl/PbBiO₂Br composites were evaluated via the degradation of the antibiotic OTC under visible-light irradiation. Detailed experimental process was given in Supporting Information (S2).

2.3 Photoelectrochemical measurements

The electrochemical properties of as-prepared samples were investigated on a electrochemical workstation (CS350H, wuhan sikete instrument Co., Ltd, China) with standard calomel electrode (SCE). Preparation of the working electrodes and detailed experimental process were given in Supporting Information (S3).

3 Results and discussion

3.1 XRD analysis

The crystal structures of as-synthesized PbBiO₂Br, Ag/ AgCl, and Ag/AgCl/PbBiO2Br (20.4%) composite were analyzed using X-ray diffraction (XRD), as presented in Fig. 1a. It can be seen that the XRD pattern of bare PbBiO₂Br was consistent with the standard spectrum of tetragonal phase PbBiO₂Br (PDF#38-1008). The strong peak located at 30.6° corresponds to the (103) plane of PbBiO₂Br, indicating that the obtained catalyst is well-crystallized [25]. For the Ag/AgCl, the peaks at $2\theta = 27.7^{\circ}$, 32.1° , 46.1° , 54.7° , and 57.3° correspond to the (111), (200), (220), (311), and (222) planes of cubic AgCl (PDF#31-1238), respectively [26]. In addition, the diffraction peaks at $2\theta = 37.9^{\circ}$, 44.1° , 64.3° , and 77.2° match with the (111), (200), (220), and (311) facets of Ag crystal (PDF# 65-2871), respectively [27]. Additionally, as for the Ag/AgCl/PbBiO₂Br (20.4%) photocatalyst, all the diffraction peaks correspond to PbBiO2Br and Ag/AgCl, and no additional crystal phases can be detected, which indicates the formation of Ag/AgCl/PbBiO₂B composites.



Fig. 1 a XRD patterns of Ag/AgCl, PbBiO₂Br, and Ag/AgCl/PbBiO₂Br (20.4%) composite. XPS spectra of Ag/AgCl/PbBiO₂Br (20.4%) composite; b the XPS survey spectrum, c Ag 3d, d Pb 4f, e Bi 4f, f Cl 2p, g Br 3d and h O 1s

3.2 XPS analysis

The elemental valence states of the as-synthesized Ag/ AgCl/PbBiO₂Br (20.4%) were detected by X-ray photoelectron spectroscopy (XPS) technology, and the obtained results are illustrated in Fig 1. The main peaks in the XPS survey spectrum of Ag/AgCl/PbBiO₂Br (20.4%) composite (Fig 1b) correspond to Br 3d, Pb 4f, Bi 4f, Cl 2p, Ag 3d, and O 1s. Figure 1c displays the XPS spectrum of Ag 3d, where the peaks at 367.93 and 373.60 eV are assigned to Ag⁰, and the other two strong peaks at 367.69 and 373.87 eV are ascribed to Ag 3d_{5/2} and Ag 3d_{3/2} of Ag⁺ in Ag/AgCl, respectively. This result is consistent with other reports in literatures [28]. In Fig. 1d, the binding energy peaks at 138.27 and 143.02 eV are corresponding to Pb 4f_{7/2} and Pb 4f_{5/2}, respectively [29]. Figure 1e shows

that the XPS spectrum of Bi element, where the peaks at 157.09 and 164.32 eV are attributed to Bi $4f_{7/2}$ and Bi $4f_{5/2}$, respectively [30–34], indicating that Bi³⁺ ions exist in PbBiO₂Br. Furthermore, in Fig 1f, two typical peaks at 199.61 and 197.99 eV can be attributed to Cl $2p_{3/2}$ and Cl $2p_{1/2}$, indicating that Cl⁻ ions exist in AgCl phase [34]. Figure 1g shows the XPS spectrum of Br 3d, where the binding energy peaks at 68.5 and 69.4 eV are corresponding to Br 3d_{3/2} and Br 3d_{5/2}, respectively [35]. The O 1s XPS spectrum of Ag/AgCl/PbBiO₂Br (20.4%) composite (Fig. 1h) is composed of fitted peaks at 529.66 and 531.04 eV, which could be attributed to the lattice oxygen of PbBiO₂Br and surface-adsorbed oxygen species, respectively [36, 37]. From the XPS analysis, it is clear that Ag/ AgCl/PbBiO₂Br is a composite sample composed of Ag/ AgCl and PbBiO₂Br.

3.3 Scanning electron microscope (SEM) and EDX analysis

Figure 2 shows the morphologies of the PbBiO₂Br, Ag/ AgCl, and Ag/AgCl/PbBiO₂Br (20.4%) composite. Figure 2a reveals that the AgCl consists of cubic-like NPs with grain size of 300–500 nm and Ag NPs are dispersed on the surfaces of AgCl cubes. Figure 2b shows that the as-prepared PbBiO₂Br has a sheet-like morphology with thickness about 30 nm. As seen from Fig. 2c, the Ag/AgCl NPs are attached on the surface of PbBiO₂Br nanosheets. Moreover, we also notice that compared with the pure Ag/AgCl, the particle size of Ag/AgCl in the Ag/AgCl/PbBiO₂Br (20.4%) composite undergoes significant change, which could be due to the fact that PbBiO₂Br could influence the surface energy of Ag/AgCl and thus impede their growth. Additionally, energy-disperse X-ray (EDX) spectroscopy analysis of Ag/AgCl/PbBiO₂Br (20.4%) composite was carried out, and the obtained result is illustrated in Fig. 2d. From the EDX spectrum, the peaks belonging to Ag, Cl, Pb, Bi, O, C, Au, and Br are observed (C and Au element come from the test instrument). The atomic ratio of Ag/Pb equals to 1:1, which is in good agreement with the Ag/Pb atomic ratio of Ag/ AgCl/PbBiO₂Br (20.4%) composite.

3.4 TEM analysis

To further obtain more detailed structure information of Ag/ AgCl/PbBiO₂Br (20.4%) composite, field emission transmission electron microscopy (TEM) and high resolution TEM (HRTEM) images were carried out. As shown in Fig. 2e, the Ag/AgCl NPs are formed on the surface of PbBiO₂Br nanosheets. From Fig. 2f, it is clearly seen that the lattice fringes of 0.235 and 0.277 nm are corresponding to the (111)

Fig. 2 SEM images of **a** Ag/ AgCl, **b** PbBiO₂Br, and **c** Ag/ AgCl/PbBiO₂Br (20.4%); **d** EDS spectrum, **e** TEM and **f** HRTEM images of as-prepared Ag/AgCl/PbBiO₂Br (20.4%) composite



and (220) planes of Ag and AgCl, respectively [38]. The lattice fringes of 0.291 nm are correlated with the (103) plane of $PbBiO_2Br$ [39].

3.5 Optical properties of the photocatalysts

The optical properties of photocatalysts are very important for their photocatalytic application in the degradation of the antibiotics. Therefore, the optical properties of as-obtained bare PbBiO₂Br, Ag/AgCl, and different Ag/AgCl/PbBiO₂Br composites were investigated via ultraviolet–visible diffuse reflectance spectra (UV–Vis DRS) measurement, as shown in Fig. 3a. The bare PbBiO₂Br exhibits the absorption edge at 500 nm, which is in agreement with the previous results in literatures [17, 40]. It can be seen that Ag/AgCl exhibits a strong absorption in the visible-light region. It is also obvious that the absorption intensities of Ag/AgCl/PbBiO₂Br composites are stronger than that of bare PbBiO₂Br in the visible-light regions, which can be attributed to the Ag SPR strategy [41].

3.6 FT-IR analysis

Figure 3b shows the Fourier transform infrared spectroscopy (FT-IR) spectra of the samples. For the pristine Ag/ AgCl sample, the peak at 1044.1 cm⁻¹ is attributed to the stretching vibration of Ag–Cl [42]. Furthermore, the stretching vibration of Ag NPs bond can be also observed at 2790.1 and 2908.7 cm⁻¹ [43]. For pure PbBiO₂Br, the peaks at 1388.4 and 1600.1 cm⁻¹ are attributed to the bending vibrations of the Pb–O bond and the Bi–O bond, respectively [44, 45]. The broad absorption bands on the right side from 3250 to 3425 cm⁻¹ are corresponding to the stretching vibration O–H band by the absorbed H₂O [46, 47]. As for the Ag/ AgCl/PbBiO₂Br composites, all the absorption peaks are from Ag/AgCl and PbBiO₂Br. The analysis results indicate that Ag/AgCl/PbBiO₂Br are successfully fabricated.

3.7 Nitrogen adsorption analysis

According to the previously reported literature [47–49], the photocatalytic efficiency of the catalyst is largely dependent on its specific surface area, so the Brunauer–Emmett–Teller (BET) specific surface areas of the as-prepared samples were measured using nitrogen adsorption–desorption measurements. The BET specific surface areas of pure PbBiO₂Br, Ag/AgCl, and Ag/AgCl/PbBiO₂Br composites are summarized in Table 1. It is found that the BET specific surface area of the Ag/AgCl/PbBiO₂Br (20.4%) is measured to be 37.16 m²/g, which is 3.06 times higher than that of pure PbBiO₂Br (12.12 m²/g). The much larger surface area facilitates the contaminant contact with the catalyst and enhances the photocatalytic performance.

3.8 Photocatalytic activity

The removal of OTC was used to evaluate the photocatalytic properties of the obtained photocatalysts under visible-light irradiation, and the attained results are given in Fig. 4a. No apparent OTC degradation is detected without photocatalyst under visible-light irradiation, indicating that the direct

Table 1 BET specific surface areas of the as-prepared samples

Samples	BET(m ² /g)
Pure PbBiO ₂ Br	12.12
Ag/AgCl/PbBiO ₂ Br (13.6%)	25.82
Ag/AgCl/PbBiO ₂ Br (20.4%)	37.16
Ag/AgCl/PbBiO ₂ Br (40.8%)	32.01
Ag/AgCl	26.14



Fig. 3 a UV–Vis absorption spectra and b FT-IR spectra of bare PbBiO₂Br, Ag/AgCl and different Ag/AgCl/PbBiO₂Br composites





Fig.4 a OTC photodegradation and $\mathbf{b} - \ln(C_t/C_0)$ vs. time plots for photodegradation of OTC by the obtained catalysts; **c** Cycling degradation efficiency of Ag/AgCl/PbBiO₂Br (20.4%) composite and **d**

photolysis of OTC can be almost neglected. It can be observed that 44% and 51% of OTC solution is removed within 80 min visible-light irradiation for bare PbBiO₂Br and Ag/AgCl, respectively. However, the Ag/AgCl/PbBiO₂Br composites exhibit enhanced photocatalytic activity in comparison to pure PbBiO₂Br and Ag/AgCl under identical experimental conditions. The degradation percentage of OTC solution reaches 72%, 93.2%, and 84% for Ag/AgCl/PbBiO₂Br (13.6%), Ag/ AgCl/PbBiO₂Br (20.4%), and Ag/AgCl/PbBiO₂Br (40.8%) composites within 80 min visible-light irradiation, respectively. It is worth noting that the Ag/AgCl/PbBiO₂Br (40.8%) photocatalyst has a higher mass ratio of Ag than the Ag/AgCl/ PbBiO₂Br (20.4%) photocatalyst, however, the photocatalytic activity of the former is lower than that of the latter. The reason may be that Ag NPs are loaded on the surface of the photocatalyst, which not only motivate the SPR, but also promote separation of electrons and holes. However, excessive Ag NPs covering on surface of the AgCl could inhibit the light absorption and decrease the separation efficiency of e^{-}/h^{+} pairs, thus leading to decreased photocatalytic activity. In addition, to get further insight into the reaction kinetic behaviors, the

XRD patterns of Ag/AgCl/PbBiO₂Br (20.4%) composite before and after the recycling photocatalytic experiment

photocatalytic degradation rates are calculated using the following equation [50–52]:

$$\ln\left(C_t/C_0\right) = K_{\rm app}t\tag{1}$$

where, k_{app} stands for degradation rates constant [53]. The results are drawn and displayed in Fig. 4b. The obtained rate constants k_{app} are 1.06×10^{-2} , 7.25×10^{-3} , 1.71×10^{-2} , 3.25×10^{-1} and 2.21×10^{-1} min⁻¹ for Ag/AgCl, PbBiO₂Br, Ag/AgCl/PbBiO₂Br (13.6%), Ag/AgCl/PbBiO₂Br (20.4%), and Ag/AgCl/PbBiO₂Br(40.8%), respectively. It is clear that the rate constant k_{app} of Ag/AgCl/PbBiO₂Br (20.4%) is 3.04 and 4.48 times higher than that of Ag/AgCl and PbBiO₂Br, respectively. These results confirm that Ag/AgCl/PbBiO₂Br composites accelerate the degradation of OTC in photocatalytic progress.

3.9 Cyclic experiments

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In order to investigate the structural stability and practical application of Ag/AgCl/PbBiO₂Br (20.4%) composite, recycling experiments were performed under the same condition, as shown in Fig. 4c. It can be observed that after the 4th run recycle experiment, the removal efficiency of Ag/AgCl/PbBiO₂Br (20.4%) photocatalyst decreases from 93.2 to 92.8%. This implies that the decrease of the degradation efficiency can be negligible. Furthermore, Fig. 4d exhibits the XRD patterns of Ag/AgCl/PbBiO₂Br (20.4%) photocatalyst before and after photodegradation recycling. It is clearly observed that all diffraction peaks undergo no change, indicating no any change in crystalline structure. Above results further confirm the stability of the Ag/AgCl/ PbBiO₂Br composites during photocatalytic process.

3.10 Possible photocatalytic mechanism

As we all know, superoxide radicals ($\cdot O_2^-$), hydroxyl radicals ($\cdot OH$) and holes (h⁺) are involved in the photocatalytic reaction system as the main radical species [54]. To explore the role of the active species, the radical trapping experiments was implemented by separately adding 10 mM ethylene diaminetetraacetic acid disodium salt (EDTA-2Na), 10 mM



Fig.5 The degradation of OTC on the Ag/AgCl/PbBiO₂Br (20.4%) composite in presence of various scavengers

Fig. 6 a PL spectra (excited at 325 nm) and b EIS Nyquist plots of the as-synthesized samples isopropanol (IPA) and 1 mM benzoquinone (BQ) into the photocatalytic reaction system, which act as the h⁺, •OH and $\cdot O_2^-$ scavengers, respectively. As depicted in Fig. 5, it is evident that with adding IPA, the degradation rate of Ag/AgCl/ PbBiO₂Br (20.4%) decreases slightly to 90.5%, demonstrating that there are almost no •OH radicals generated in the degradation process. However, when adding BQ or EDTA-2Na, the degradation efficiency sharply decreases from 93.2 to 17 and 26%, respectively, demonstrating that $\cdot O_2^-$ and h⁺ play very important role in the degradation process. Furthermore, in this study, considering that the Cl⁻ could be oxidized by holes to Cl⁰ atoms and the antibiotics OTC could be oxidized [55]. Cl⁰ atoms are considered to be another actual active species in the photocatalytic degradation process [56].

Many researches indicate that photoluminescence (PL) emission spectra can be induced by the recombination between photogenerated electrons and holes [57]. The lower the PL emission peaks, the less the recombination of photoexcited charge carriers. Therefore, the charge transfer and recombination processes in the photodegradation experiment can be investigated by PL spectra. Figure 6a shows the PL spectra of bare PbBiO₂Br, Ag/AgCl, and different Ag/AgCl/ PbBiO₂Br composites in the range of 420–620 nm under excitation at 325 nm, which arise due to the recombination of photogenerated electrons and holes. It is observed that the emission spectrum intensity of bare PbBiO₂Br is the strongest. However, after the coupling of Ag/AgCl NPs with PbBiO₂Br nanosheets, the intensity of the PL emission spectra is decreased, indicating that the charge separation rate of Ag/AgCl/PbBiO₂Br composites is more efficient than that of bare PbBiO₂Br and Ag/AgCl nanoplates. It is noteworthy that the Ag/AgCl/PbBiO₂Br (20.4%) composite exhibits the weakest intensity, suggesting that it has the highest separation efficiency of photoexcited charge carriers [58]. In order to further understand the charge transfer in the photocatalytic process. Electrochemical impedance spectroscopy (EIS) measurement was also carried out for the bare PbBiO₂Br, Ag/AgCl, and Ag/AgCl/PbBiO₂Br (20.4%) composite. As shown in Fig. 6b, it is found that Ag/AgCl/ PbBiO₂Br (20.4%) composite owns the smallest semicircle



radius. It is commonly recognized that the curvature radius serves as an indicator of charge-transfer resistance, and a smaller semicircle radius implies higher charge transfer efficiency [59, 60].

The positions of the conduction band (CB) and valence band (VB) of obtained PbBiO₂Br are about -1.0 and 1.5 eV (vs.NHE), respectively, according to our previously reported results [17, 18]. Furthermore, in the light of the literature, the positions of the CB and VB of the AgCl are located at -0.09 and 3.16 eV (vs. NHE), respectively [23, 25].

In the light of above experimental results, a possible photocatalysis mechanism is proposed to explain the charge transfer behaviors of Ag/AgCl/PbBiO₂Br composite in the photocatalytic process. As shown in Fig. 7, the PbBiO₂Br and metallic Ag NPs are photoexcited to generate e^{-}/h^{+} under visible-light irradiation (Eqs. 2, 3). The AgCl is difficult to be stimulated under visible-light irradiation due to its broad bandgap. Since the SPR of Ag NPs is energetic enough to the photoexcited electrons and can be easily injected into the E_{CB} of AgCl or PbBiO₂Br (Eq. 4). These accumulated electrons on the E_{CB} of AgCl could not reduce oxygen to form $\cdot O_2^-$, due to the $E_{\rm CB}$ potential of AgCl (-0.09 eV) more positive than the standard reduction potential of $E_0(O_2/O_2^-) = -0.33 \text{ eV vs. NHE}$ [61-64]. These accumulated charges could react with O_2 on the surface of PbBiO₂Br to form \cdot O₂⁻ due to the E_{CB} potential of $PbBiO_2Br (-1.0 \text{ eV})$ more negative than the standard reduction potential of $E_0(O_2/O_2^-) = -0.33 \text{ eV}$ vs. NHE (Eq. 5). Above radical trapping experimental results verify that the $\cdot O_2^-$ is one of the main active species in the photocatalytic process. Meanwhile, the residual h⁺ at Ag NPs migrates to the $E_{\rm VB}$ of AgCl surface to oxidize the Cl⁻ ion to form Cl⁰ atoms (Eqs. 6, 7). The Cl⁰ atoms are reactive radical species. After that, Cl⁰ atoms oxidize OTC and hence are reduced to Cl⁻ again (Eq. 8) [65, 66]. Thus, the Ag/AgCl/PbBiO₂Br can maintain good catalytic performance and stability. On the other hand, from a thermodynamic point of view, the photogenerated h⁺ cannot react with OH⁻ or H₂O to produce •OH since the VB potential of PbBiO₂Br is more negative than the redox potentials of $E^0(OH^-/\cdot OH)$ (1.99 eV vs. NHE) and $E^0(H_2O/\cdot OH)$ (2.38 eV vs. NHE), indicating that the h⁺ can directly oxide OTC [67–69]. The produced active species (h⁺,·O₂⁻) can efficiently decompose OTC into intermediate products and finally into H₂O and CO₂ (Eq. 9). The above discussion suggests that the Ag/AgCl/PbBiO₂Br composites can improve the separation of photogenerated e^-/h^+ , finally leading to the enhancement of photocatalytic activity.

$$Ag + h\nu \rightarrow Ag(e^{-})_{CB} + Ag(h^{+})_{VB}$$
 (2)

$$PbBiO_2Br + h\nu \rightarrow PbBiO_2Br(e^{-})_{CB} + PbBiO_2Br(h^{+})_{VB}$$
(3)

$$Ag(e^{-})_{CB} + PbBiO_2Br \rightarrow PbBiO_2Br(e^{-})_{CB} + Ag$$
(4)

$$PbBiO_2Br(e^-)_{CB} + O_2 \rightarrow O_2^-$$
(5)

$$Ag(h^{+})_{VB} + AgCl \rightarrow AgCl(h^{+})_{VB} + Ag$$
 (6)

$$\operatorname{AgCl}(h^{+})_{VB} + \operatorname{Cl}^{-} \to \operatorname{Cl}^{0} + \operatorname{AgCl}$$

$$\tag{7}$$

$$Cl^0 + OTC \rightarrow products + Cl^-$$
 (8)

$$O_2^-/h^+ + OTC \rightarrow \text{products}$$
 (9)



Fig. 7 Proposed photocatalytic reaction processes and charge separation of Ag/AgCl/PbBiO₂Br composites under visible-light irradiation

4 Conclusions

In this study, visible-light-driven novel Ag/AgCl/PbBiO₂Br composites were successfully synthesized through hydrothermal and in situ photoreaction method. The UV-Vis absorption spectra confirm that the as-obtained Ag/AgCl/ PbBiO₂Br composites exhibit remarkable photo-absorption property in the visible-light region as compared to PbBiO₂Br nanosheets, which could be due to the surface Ag resonance. The Ag/AgCl/PbBiO₂Br (20.4%) composite exhibits the strongest capacity for degradation of the antibiotic OTC under visible-light irradiation, which can be mainly attributed to strong visible-light absorbance and the efficiently separation of photoexcited charge. The recycling experiments demonstrate that the Ag/AgCl/PbBiO₂Br composites possess good stability. In addition, active species trapping experiments confirm that O_2^- , Cl^0 and h^+ play an very important role in the degradation process. This work provides a way to design an excellent environmental purification material.

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Availability of data and materials All data are fully available without restriction.

Compliance with ethical standards

Conflict of interest The authors declare that they have no competing interests.

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