


A geometric morphometric and microsatellite analyses of *Scaptotrigona mexicana* and *S. pectoralis* (Apidae: Meliponini) sheds light on the biodiversity of Mesoamerican stingless bees

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Received: 4 February 2016 / Accepted: 13 August 2016 / Published online: 31 August 2016
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Abstract Geometric morphometrics and molecular methods are effective tools to study the variability of stingless bee populations and species that merit protection given their worldwide decline. Based on previous evidence of cryptic lineages within the *Scaptotrigona* genus, we tested the existence of multiple evolutionary lineages within the species *S. mexicana* and we investigated the status of *S. pectoralis*. By analyzing their population structure, we found differences between the Pacific and Atlantic populations of each of these species, although geometric morphometrics of the wing only confirmed these results in *S. mexicana*. There was a tendency towards enhanced genetic differentiation over larger distances in the Atlantic populations of both species but not in the Pacific populations. These results revealed a pattern of differentiation among evolutionary units and a specific distribution of genetic diversity within these *Scaptotrigona* species in Mesoamerica, suggesting the need for future taxonomic revisions, as well as activities aimed at management and conservation.

Keywords *Scaptotrigona* · Stingless bees · Geometric morphometrics · Microsatellites · Evolutionary units

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Introduction

Stingless bees (Hymenoptera: Apidae: Meliponini) are eusocial, haplodiploid insects that carry out ecologically and economically important pollinating activity, especially in Neotropical regions (Slaa et al. 2006). Most of the diversity of the tribe Meliponini is concentrated in this region, with approximately 42 endemic genera (Camargo and Pedro 2013) that include around 400 species (Michener 2007). One of these genera is *Scaptotrigona* Moure, 1942, which consists of 24 species extending from Mexico to Argentina (Michener 2007). Three *Scaptotrigona* species are found in Mexico: *S. mexicana* Guérin, 1845 the distribution of which extends to Costa Rica; *S. pectoralis* Dalla Torre, 1896 the distribution of which reaches Panama; and the endemic *S. hellwegeri* Friese, 1900 (Ayala 1999; Camargo and Pedro 2013). These three species are managed for honey production (Manzo 2009; González-Acereto 2012; Ayala et al. 2013) but at present, only *S. mexicana* and *S. pectoralis* contribute to the pollination of crops like avocado (Ish-Am et al. 1999).

Given their important contribution to agriculture, the commercial use of these species is augmenting (Albores-González et al. 2011), potentially influencing the gene flow among populations due to hive translocation (Jaffé et al. 2016; Santiago et al. 2016). Like many other insects, these species are also being affected by extensive deforestation and the extension of monocrops in these areas (Freitas et al. 2009; Stout and Morales 2009). These phenomena may potentially drive a decline in stingless bee populations and they may provoke endogamous depression, an increase in homozygosity and the ensuing loss of genetic diversity (Zayed 2009). Hence, the lack of information about their richness, diversity, genetic status and distribution is an important obstacle to ensure their protection and to

establish appropriate conservation programs (Brown and Paxton 2009; Freitas et al. 2009). These programs should include measures avoiding the translocation of hives and to define which populations should remain isolated to preserve genetic integrity.

The existence of cryptic species of stingless bees has been proposed (Tavares et al. 2007; Francisco et al. 2008; May-Itzá et al. 2012) and thus, the total number of species could be higher than anticipated (Michener 2007). The number of cryptic species will probably increase as more molecular methods are made available and more powerful morphological analyses are employed. In the last decade, geometric morphometric analysis of wings has proven to be a good tool to clarify discrepancies within species and to identify bees (for example honeybees: Bouga et al. 2011; da Silva et al. 2015; or bumblebees: Barkan and Aytekin 2013). In stingless bees, this approach has proved to be very sensitive to characterize species and even to differentiate populations (Ferreira et al. 2011; Francoy et al. 2011; Lima Junior et al. 2012; Nunes et al. 2013; Bonatti et al. 2014), although genetic information should be incorporated to more rigorously define species and to resolve taxonomic problems (Schlick-Steiner et al. 2010). Species and population identification using genetic markers has advanced significantly, and population parameters important for the conservation of endangered species can now be derived from the use of molecular markers (Frankham et al. 2002; Hedrick 2005). Among these, microsatellite loci have good statistical resolution to characterize demographic events within populations (Lui-kart and England 1999), and they have been successfully used within the Meliponini to characterize genetic variability (Arias et al. 2006; Francisco et al. 2006; Fernandes et al. 2012) and issues like the genetic characteristics of drone congregation areas (Kraus et al. 2008; Mueller et al. 2012).

Studies of morphological and/or molecular markers in the genus *Scaptotrigona* have suggested the existence of genetically distinct evolutionary lineages (or cryptic species). A study on *S. hellwegeri* found signs of ongoing speciation, indicating that two populations distributed at different altitudes along the Trans-Mexican Volcanic Belt and the Pacific coast should be treated as separate units (Quezada-Euán et al. 2012). The existence of cryptic species was also proposed within *S. mexicana* colonies in two regions located at opposite extremes of the species distribution range in Mexico (Veracruz and Chiapas: Hurtado-Burillo et al. 2013). Recently, up to five distinct clusters (considered as units of management and conservation) with high rates of genetic diversity were described in another *Scaptotrigona* species (*S. xanthotricha*, which is widely distributed in the Brazilian Atlantic rainforest: Duarte et al. 2014).

As such, here we aimed to resolve taxonomic issues in *S. mexicana* and *S. pectoralis*, two species that share an equivalent distribution, as well as similar biological and ecological aspects, but that are managed distinctly. Specifically, we used geometric morphometry and microsatellite markers to assess the population structure and the molecular diversity of *S. mexicana* and *S. pectoralis*, and to test the existence of various evolutionary units within *S. mexicana*. In addition, these tools were used to investigate the species status of *S. pectoralis* in the light of the diversity observed in this genus.

Materials and methods

Sampling

This study was carried out in Mexico and northern Guatemala, within the distribution area of *S. mexicana* and *S. pectoralis*. Colonies were sampled at 15 sites between 2008 and 2011 (Table 1; Fig. 1), and most *S. mexicana* samples (70 colonies in total) were provided by stingless beekeepers that maintain colonies in clay pots or in their original trunks, either on their own property or in the vicinity (meliponaries). *S. pectoralis* samples (33 colonies in total) were obtained from non-managed wild colonies located in native forests. Adult worker bees were collected at the entrance of each colony and preserved in absolute ethanol at -20°C . Voucher specimens were deposited in the insect collection of the Animal Biology Laboratory at the Veterinary Faculty of the University of Murcia (Spain).

Geometric morphometric analysis

One to six specimens from 70 *S. mexicana* and 32 *S. pectoralis* colonies were analysed (Table 1). Only the right wing of each specimen was used and they were photographed with a Spot Insight Firewire digital camera (Sterling Heights, USA) adapted to a Zeiss Stemi 2000C Trinocular Zoom Stereomicroscope (Thornwood, USA). We manually plotted twelve homologous landmarks of the wing vein intersections (Fig. 2) using the tpsDig software, (version 2.17: Rohlf 2013). MorphoJ software (version 1.06c: Klingenberg 2011) was used to analyze the images, which were first Procrustes aligned (Bookstein 1991) in order to identify the points of shape variation. Mean values of the specimens from each colony were used to perform the analyses at the colony level. Principal component (PCA) and canonical variate (CVA) analyses were performed with these data.

Mahalanobis distances between groups of colonies were defined by the PCA analysis and were obtained with the

Table 1 Details of *S. mexicana* and *S. pectoralis* sampling in Mexico and Guatemala

Species	Locality, state, country	Coordinates		Map code	N1	N2	Cluster
		Latitude	Longitude				
<i>S. mexicana</i>	Cacahoatan, Chiapas, Mexico	14.996	−92.167	10	8	8	<i>Sm1</i>
<i>S. mexicana</i>	Tapachula, Chiapas, Mexico	14.979	−92.266	12	16	8	<i>Sm1</i>
<i>S. mexicana</i>	Tuxtla Chico, Chiapas, Mexico	14.906	−92.261	9	21	22	<i>Sm1</i>
<i>S. mexicana</i>	Chilcuahuta, Hidalgo, Mexico	20.331	−99.232	1	–	1	<i>Sm2</i>
<i>S. mexicana</i>	Melchor de Mencos, Peten, Guatemala	17.066	−89.150	15	11	11	<i>Sm2</i>
<i>S. mexicana</i>	Cuetzalan del Progreso, Puebla, Mexico	20.017	−97.522	4	–	1	<i>Sm2</i>
<i>S. mexicana</i>	Tuzamapan de Galeana, Puebla, Mexico	20.065	−97.576	2	–	1	<i>Sm2</i>
<i>S. mexicana</i>	Coatepec, Veracruz, Mexico	19.451	−96.959	5	6	6	<i>Sm2</i>
<i>S. mexicana</i>	Coyulta, Veracruz, Mexico	20.248	−97.658	3	8	9	<i>Sm2</i>
Total <i>S. mexicana</i>					70	67	
<i>S. pectoralis</i>	Merida, Yucatan, Mexico	20.861	−89.624	14	8	8	<i>Sp2</i>
<i>S. pectoralis</i>	Tlaltetela, Veracruz, Mexico	19.314	−96.901	6	4	5	<i>Sp2</i>
<i>S. pectoralis</i>	Palenque, Chiapas, Mexico	17.511	−91.993	13	3	3	<i>Sp2</i>
<i>S. pectoralis</i>	Tapachula, Chiapas, Mexico	14.966	−92.261	8	8	8	<i>Sp1</i>
<i>S. pectoralis</i>	Tuxtla Chico, Chiapas, Mexico	14.937	−92.167	7	9	9	<i>Sp1</i>
Total <i>S. pectoralis</i>					32	33	

Map code refers to the number labelling each locality in Fig. 1. N1 = number of colonies used for geometric morphometrics analysis; N2 = number of colonies used for microsatellite analysis. The cluster assigned to each locality is also indicated

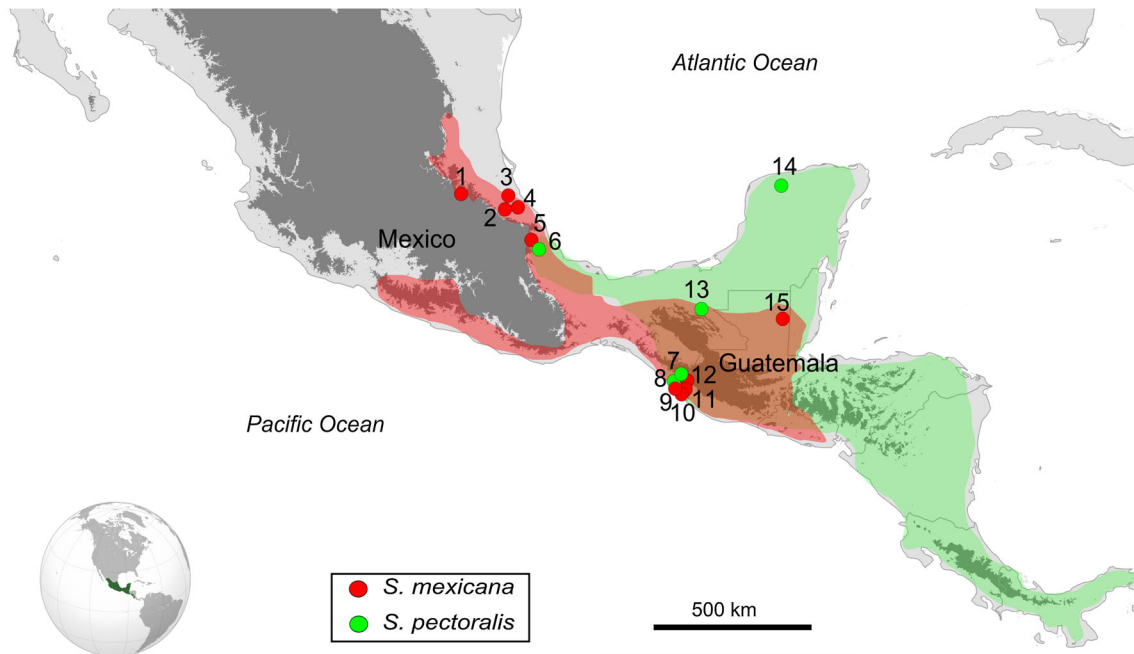


Fig. 1 Localities in Mexico and Guatemala where specimens were collected: red dots represent *S. mexicana* and green dots—*S. pectoralis* colonies. Numbers correspond to the sampling code used

in Table 1. Distribution ranges have been shadowed: in red—*S. mexicana*, in green—*S. pectoralis* and in light brown the area where the two species co-exist. (Color figure online)

MorphoJ software. We also performed a Mantel test to investigate whether the Procrustes-fitted landmark co-ordinates of the wings varied with the geographic distance in each group using Past 3.08 software (Hammer et al. 2001).

DNA extraction and microsatellite amplification

DNA was extracted from the leg of one worker bee per colony (*S. mexicana*, N = 67; *S. pectoralis* N = 33:

Fig. 2 Worker wing with 12 landmarks in the vein junctions

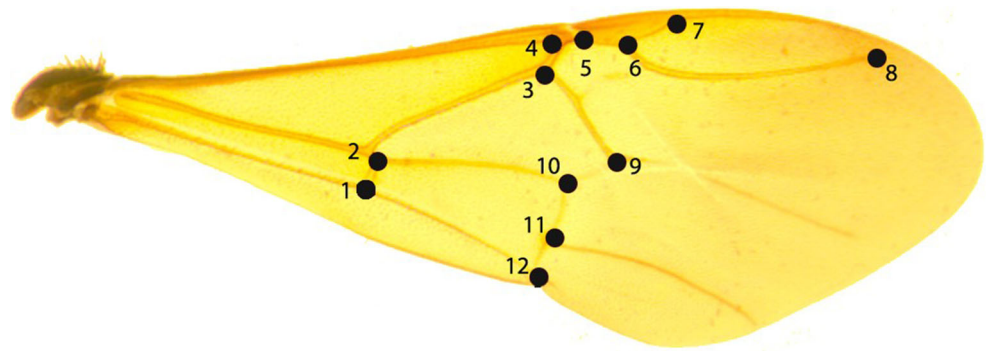


Table 1) using the DNeasy tissue kit (QIAGEN). Seven microsatellite loci were amplified in three reactions (two of them multiple), two of which were microsatellites originally described in *S. postica* (T4–171 and T7–5; Paxton et al. 1999), four in *Melipona bicolor* (Mbi278, Mbi259, Mbi254 and Mbi201; Peters et al. 1998) and one in the bumblebee *Bombus terrestris* (B124; Estoup et al. 1995). The T4–171, T7–5 and B124 loci have already been used successfully in *S. mexicana* (Kraus et al. 2008; Mueller et al. 2012) and *S. hellwegeri* (Quezada-Euán et al. 2012).

PCR reactions were carried out in 12.5 μ l with PureTaq™ Ready-To-Go™ PCR beads (GE Healthcare) in a PTC-200 Thermal Cycler (Biorad). The amplified fragments (microsatellite alleles) were detected with an ABI prism 3100 sequencer (Applied Biosystems) and scored with Genemapper software (version 3.7: Applied Biosystems).

Molecular data analyses

The number of clusters (K) was estimated with the STRUCTURE software (version 2.2: Pritchard et al. 2000) using a Bayesian model-based method, basing the results on simulations of 80,000 burn-in steps and 1,000,000 MCMC (Markov Chain Monte Carlo) algorithm iterations. Five runs for each K-value ($K = 1 - 8$) were used to estimate the most likely value of K, and the runs were performed without any prior information about the origin of samples using the “Admixture Model” and considering allele frequencies correlated among populations. The number of clusters defined by the ΔK value described previously (see Evanno et al. 2005) was inferred by STRUCTURE HARVESTER (Earl and vonHoldt 2012). CLUMPP (v1.1.2: Jakobsson and Rosenberg 2007) was used to obtain the average permuted individual and population Q-matrices of the best K value. The *distrupt* software (v1.1: Rosenberg 2004) was used to graphically display the CLUMPP results.

The dataset describing the genotypes included in each cluster was analyzed with MICRO-CHECKER (Van Oosterhout et al. 2004) to detect scoring incongruities and

possible null alleles. GENALEX (Peakall and Smouse 2006) was used to calculate population genetics parameters per locus and cluster in each species (the number of detected and private alleles, and the expected heterozygosity). Given the different number of individuals sampled for each cluster, a rarefaction analysis gave an objective estimate of allelic richness (i.e. the number of alleles: Leberg 2002) and it was carried out with the HP-RARE 1.0 program (Kalinowski 2005). Differences between clusters were estimated using the two-tailed Mann–Whitney U Test (Mann and Whitney 1947).

Hardy–Weinberg and genotypic linkage equilibria, and the population differentiation using the Fisher’s exact probability test were calculated in GENEPOP (Raymond and Rousset 1995a, b). All probability values were adjusted for multiple comparison tests using sequential Bonferroni adjustments (Rice 1989).

Pairwise R_{ST} values were calculated as a measure of genetic distance in ARLEQUIN 3.5 (Excoffier and Lischer 2010), calculating R_{ST} pairwise values to estimate the connectivity and patterns of gene flow between clusters given that F_{ST} values might underestimate the differentiation in highly structured populations (Balloux and Lugon-Moulin 2002) like those of *Scaptotrigona* species (Quezada-Euán et al. 2012; Hurtado-Burillo et al. 2013). The correlation between genetic and geographic distances within each cluster was analyzed with a Mantel test, as implemented in GENALEX, to test whether allelic frequencies varied in function of a pattern of isolation by distance (IBD).

Results

Geometric morphometrics

PCA analysis of the 12 wing landmarks generated 20 relative warp measurements. The first 11 factors of these measurements had eigenvalues greater than one and explained 95.46 % of the total variability in the data. The first two PC values explained 64.52 % of the total variability (PC1

Table 2 Microsatellite variation in *S. mexicana* (*Sm1* and *Sm2*) and *S. pectoralis* (*Sp1* and *Sp2*). Sample size (N), number of detected alleles (Na) and number of private alleles (Npa) (within brackets after rarefaction) and expected (He) heterozygosity per locus. Mean values ± SE are shown for each cluster

Cluster	Locus	N	Na	Npa	He
<i>Sm1</i>	Mbi254AAG	32	2 (1.71)	0 (0.01)	0.26
	Mbi259AAG	37	4 (2.06)	0 (0.36)	0.41
	Mbi278AAG	37	1 (1)	0 (0)	0
	Mbi201AAG	37	2 (1.46)	1 (0.46)	0.15
	T4-171	35	9 (4.24)	4 (3)	0.80
	T7-5	35	9 (3.77)	6 (2.23)	0.73
	B124	32	10 (4.3)	6 (4.02)	0.78
	Mean ± SE	35 ± 0.9	5.3 ± 1.5 (2.65 ± 0.4)	2.43 ± 1.1 (1.44 ± 0.5)	0.45 ± 0.1
<i>Sm2</i>	Mbi254AAG	24	2 (1.99)	0 (0.28)	0.50
	Mbi259AAG	28	4 (2.39)	0 (0.69)	0.43
	Mbi278AAG	28	2 (1.88)	1 (0.88)	0.38
	Mbi201AAG	28	1 (1)	0 (0)	0
	T4-171	23	6 (3.3)	1 (2.06)	0.64
	T7-5	25	7 (3.84)	4 (2.30)	0.76
	B124	27	8 (3.08)	4 (2.80)	0.57
	Mean ± SE	26.1 ± 0.8	4.3 ± 1.0 (2.50 ± 0.3)	1.43 ± 0.7 (1.29 ± 0.3)	0.47 ± 0.1
<i>Sp1</i>	Mbi254AAG	15	4 (2.41)	2 (0.57)	0.51
	Mbi259AAG	17	2 (1.83)	1 (1.29)	0.33
	Mbi278AAG	17	2 (1.71)	0 (0.05)	0.25
	Mbi201AAG	17	1 (1)	0 (0)	0
	T4-171	13	4 (3.38)	0 (1.01)	0.70
	T7-5	13	1 (1)	0 (0)	0
	B124	17	1 (1)	0 (0)	0
	Mean ± SE	15.6 ± 0.7	2.1 ± 0.5 (1.76 ± 0.3)	0.43 ± 0.3 (0.4 ± 0.2)	0.3 ± 0.1
<i>Sp2</i>	Mbi254AAG	9	2 (1.89)	0 (0.06)	0.35
	Mbi259AAG	16	3 (2.63)	2 (2.09)	0.60
	Mbi278AAG	16	2 (1.93)	0 (0.27)	0.40
	Mbi201AAG	16	1 (1)	0 (0)	0.00
	T4-171	15	10 (5.28)	6 (2.91)	0.88
	T7-5	16	2 (1.95)	1 (0.95)	0.43
	B124	16	3 (2.20)	2 (1.20)	0.52
	Mean ± SE	14.9 ± 1.0	3.3 ± 1.1 (2.41 ± 0.4)	1.58 ± 0.8 (1.1 ± 0.3)	0.5 ± 0.1

explained 49.31 % and PC2 15.21 %) and a PCA identified three groups, two within *S. mexicana* (herein named *Sm1* and *Sm2*) and one with the *S. pectoralis* colonies. The graphic representation showed that the colonies from the Pacific coast (Chiapas, *Sm1*) were placed in quadrant 4 while the rest of the colonies from the Atlantic coast (Veracruz in Mexico and Peten in Guatemala, *Sm2*) were mainly placed in quadrant 3. Conversely, the samples of *S. pectoralis* were dispersed in quadrants 1–2 (Fig. 3).

The Mahalanobis distances between the groups confirmed the differentiation between *Sm1* and *Sm2* (4.847), although this value was around half of that observed

between *S. pectoralis* and *Sm1* (8.477) or *Sm2* (8.171). In addition, the Mantel test showed no significant correlation between morphological and geographical distances in each putative species (*Sm1*: $R = -0.13$, $p = 0.99$; *Sm2*: $R = 0.012$, $p = 0.3$ and *S. pectoralis*: $R = 0.010$, $p = 0.39$).

Microsatellites

The average genotyping error rate for the seven microsatellite loci assessed in this study was <5 %. Less than four loci were amplified from two of the 67 *S.*

Fig. 3 Distribution of the average scores of *Scaptotrigona* colonies against principal components 1 and 2 of the principal component analysis (PCA) based on geometric morphometry data. *Sm* named the two clusters detected within *S. mexicana* and *Sp* referred to *S. pectoralis* colonies

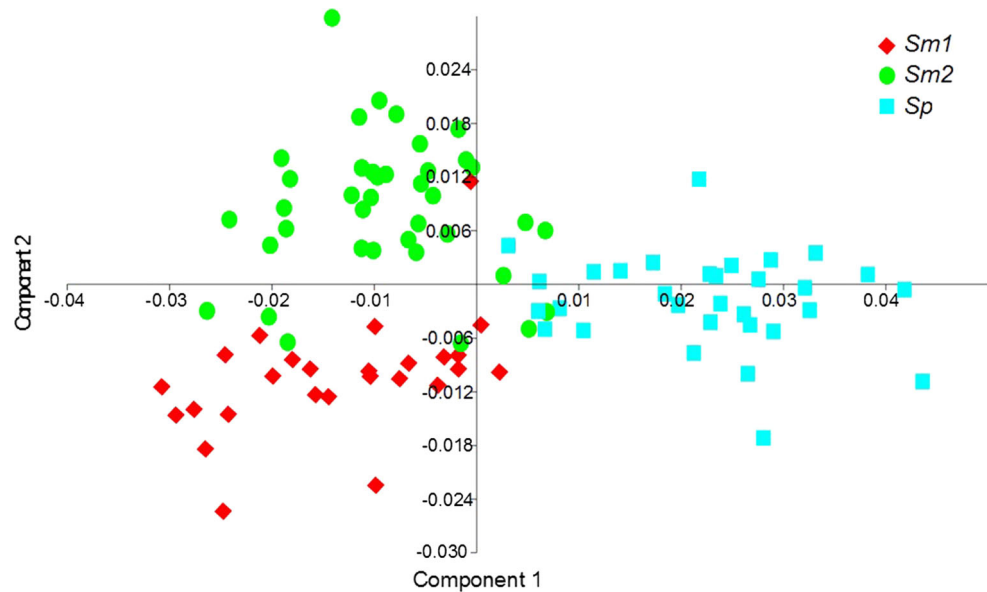
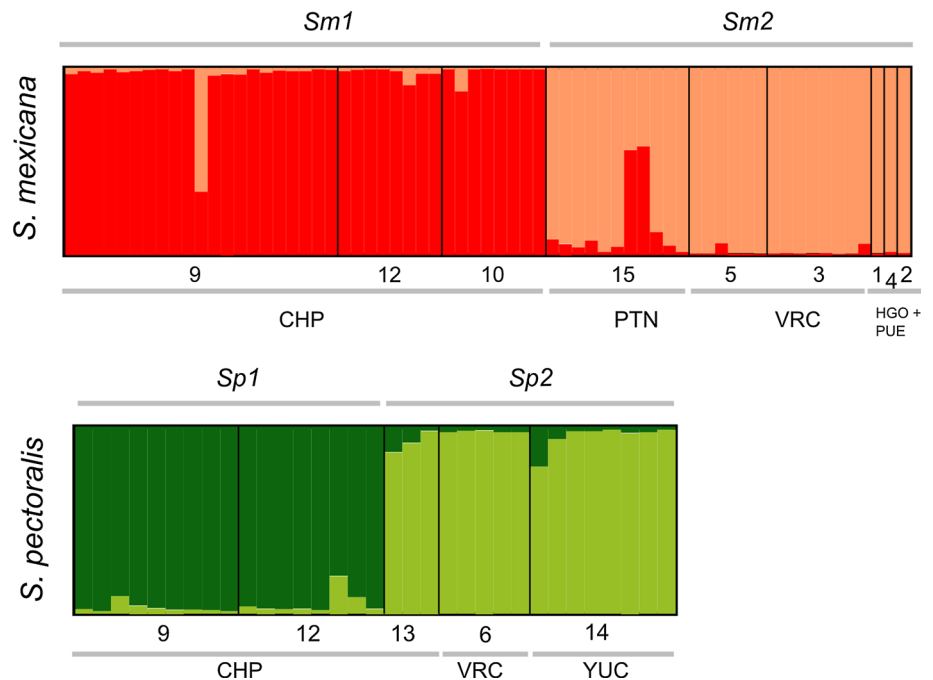


Fig. 4 Results of the Bayesian clustering showing the most probable number of clusters ($K = 2$) for the two species. Division of specimens into colored segments represents the assignment probability of that specimen to each of the K clusters (red—*Sm1*; light red—*Sm2*; dark green—*Sp1*; light green—*Sp2*). Numbers correspond to the sampling code used in Table 1 and acronyms to the name of the state (*CHP* Chiapas, *PTN* Petén, *VRC* Veracruz, *HGO* Hidalgo, *PUE* Puebla, *YUC* Yucatán). (Color figure online)



mexicana specimens and thus, they were excluded from the analysis. In a first step, we looked for signs of genetic structure in each species. The highest posterior probability of the data set was detected when assuming two clusters ($K = 2$) in both species: *S. mexicana* (*Sm1* and *Sm2*) and *S. pectoralis* (*Sp1* and *Sp2*) (Fig. 4). *S. mexicana* clusters were congruent with the geometric morphometry data: *Sm1* (colonies from Chiapas) and *Sm2* (colonies from Veracruz and Petén, including in this case three more colonies, two from Puebla and one from Hidalgo). In contrast to the

morphological results, *S. pectoralis* was divided into two clusters: *Sp1* (colonies from southern Chiapas: Tapachula and Tuxtla Chico) and *Sp2* (colonies from Veracruz, Yucatan and northern Chiapas: Palenque). These clusters showed a geographic pattern for both species, splitting Pacific (*Sm1* and *Sp1*) and Atlantic populations (*Sm2* and *Sp2*), and they were used to estimate the population genetic parameters.

There was no evidence of scoring error due to stuttering or allele dropout, yet there appeared to be null alleles at

locus T4-171 in *Sm2*, and at loci Mbi254-AAG and B124 in *Sp2*. These loci had a higher percentage of unsuccessfully genotyped individuals (3.5, 4.5 and 1.5 %, respectively), a fact that was taken into account to explain several population parameters.

The number of alleles of the seven microsatellite loci ranged from one to ten across all the samples (Table 2). The average allele number after rarefaction was higher in the *Sm1* (2.65 ± 0.4) than in *Sm2* (2.50 ± 0.3) populations, and in the *Sp2* (2.41 ± 0.4) than in *Sp1* (1.76 ± 0.3) populations, and fewer private alleles (i.e. alleles unique to each cluster, *N_{pa}*) were evident after rarefaction. The highest *N_{pa}* was detected in *Sm1* within *S. mexicana*, (1.44 ± 0.5), whereas for *S. pectoralis* the highest *N_{pa}* was detected in *Sp2* (1.1 ± 0.3). All comparisons were not statistically significant ($p > 0.05$). Gene diversity was measured as the expected heterozygosity (*H_e*) and it was higher in the *Sm2* than in the *Sm1* population, and in the *Sp2* than in the *Sp1* population, although these difference were not significant.

The *Sm1* cluster was in Hardy–Weinberg equilibrium before Bonferroni correction ($p > 0.05$ for each locus), whereas in the *Sm2* population one locus (T4-171) did not show Hardy–Weinberg equilibrium even after Bonferroni correction, highlighting a marked excess of homozygotes that was probably due to the presence of null alleles. A similar result was obtained in the *S. pectoralis* populations as all loci were in Hardy–Weinberg equilibrium in the *Sp1* population whereas *Sp2* did not show such equilibrium at those loci with null alleles (Mbi254AAG and B124). The loci were generally out of linkage disequilibrium after Bonferroni correction for each cluster, with the exceptions of 4 out of the 15 pairwise comparisons in *Sm2*.

A Fisher's exact test of population differentiation showed highly significant genetic differences between the clusters of each species. The pairwise *R_{ST}* was higher between *Sm1* and *Sm2* ($R_{ST} = 0.266$, $p < 0.0001$) than between *Sp1* and *Sp2* ($R_{ST} = 0.198$, $p < 0.0001$). The IBD was evaluated with the Mantel test for each cluster and there was a significant correlation between genetic and geographic distance in the Atlantic populations of both species (*Sm2* $r = 0.383$, $p = 0.001$; and *Sp2* $r = 0.477$, $p = 0.001$), but no correlation in the Pacific populations (*Sm1* $r = -0.034$, $p = 0.177$; and *Sp1* $r = 0.158$, $p = 0.079$).

Discussion

Our results demonstrate that diversity in the *Scaptotrigona* genus is higher than expected in Mesoamerica. The genetic and geometric morphometric analyses fully support the existence of two evolutionary significant units (ESUs)

within *S. mexicana* (*Sm1* and *Sm2*), with a Pacific and Atlantic distribution, respectively. By contrast, *S. pectoralis* population differentiation is evident through genetic analysis (*Sp1* and *Sp2*) but not so clearly by geometric morphometry, suggesting that *S. pectoralis* is in an incipient stage of population differentiation.

The clusters found within *S. mexicana* resembled two ESUs with different distributions, one dispersed along the Pacific coast of Mexico (*Sm1*) and the other along the Atlantic Mexican coast up to the North of Guatemala (*Sm2*). These data support previous results suggesting the existence of a cryptic species within *S. mexicana* based on high divergence between the populations identified by means of the barcoding method (Hurtado-Burillo et al. 2013). The results obtained here support this hypothesis, since both the molecular and morphological approaches discriminate the two ESUs. A multidisciplinary approach such as that performed here also proved to be useful in other population studies on stingless bees species (Mendes et al. 2007; Gonçalves 2010; Francoy et al. 2011; May-Itzá et al. 2012; Bonatti et al. 2014) and on honey bees, albeit at the sub-species level in the latter (Oleksa and Tofilski 2015).

The differentiation between *Sm1* and *Sm2* is evident through several population parameters. First, geometric morphometry of the wings indicates a significant phenotypic differentiation of two separate groups. Second, while both *S. mexicana* clusters share alleles, the presence of private alleles in both the Pacific *Sm1* and Atlantic *Sm2* populations, together with different allelic frequencies at some loci, indicate genetic differences between these two evolutionary units. Although microsatellite markers usually show less allelic diversity when used in species other than those they have been designed for (Borges et al. 2010), they yielded appropriate information in analyses of other *S. mexicana* populations (Kraus et al. 2008; Mueller et al. 2012) and *S. hellwegeri* (Quezada-Euán et al. 2012). Finally, the *R_{ST}* value (0.266) points to significant divergence within *S. mexicana* units. While this *R_{ST}* value could reflect an IBD effect that restricts gene flow among colonies, it is not conclusive of complete separation into two species. Introgression events between ESUs of *S. mexicana* (*Sm1* and *Sm2*; Fig. 4) have been observed in both Atlantic and Pacific populations, possibly the result of natural gene flow. However, human influence should not be ignored as *S. mexicana* is one of the species traditionally managed in Mesoamerican cultures and our samples were obtained from stingless beekeepers. Although movements of hives over large distances are uncommon in Mesoamerica (May-Itzá et al. 2012; González-Acereto pers. comm.), genetic traces of human-mediated transportation have been observed elsewhere for other stingless bees (Byatt et al. 2016).

Genetic differentiation was also observed for *S. pectoralis*, with microsatellite analysis defining two clusters (*Sp1* and *Sp2*) and confirming the data obtained with mitochondrial markers (Hurtado-Burillo et al. 2013). The clustering obtained previously through the barcoding approach also suggested some degree of differentiation within *S. pectoralis*, although the intraspecific divergence did not fully support the existence of distinct genetic lineages. Since extant geometric morphometric data does not corroborate this differentiation, the two clusters in *S. pectoralis* may not be ESUs but rather, units of management and conservation, as in the case of Brazilian populations of *Scaptotrigona xanthotricha* (Duarte et al. 2014). In *S. pectoralis*, the failure to detect introgression suggests there has been no recent gene flow between the two clusters. This could be due to the limited sampling ($n = 33$) and/or the fact that the colonies sampled were wild, unmanaged colonies located in native forests.

The homogeneous genetic structure detected in distant populations of Atlantic clusters, despite the presence of three introgressed *S. mexicana* specimens (one from Chiapas in *Sm1* and two from Petén in *Sm2*), and especially in the Pacific clusters of both species, may be related to the reproductive behavior of this group of bees. This behavior favors gene flow within nearby populations due to the migration of males from several colonies to form male congregation areas. This reproductive strategy is used by *S. mexicana* (Kraus et al. 2008) and it is an effective mechanism to avoid inbreeding (Mueller et al. 2012). Such homogeneity contrasts with the population structure observed in the Pacific and Atlantic clusters of each species, which may be explained by the influence of a geographical barrier to their distribution, the Sierra Madre, as seen in other Meliponini (May-Itzá et al. 2010, 2012), Hemiptera (Dorn et al. 2009), Coleoptera (Anducho-Reyes et al. 2008) and even in terrestrial birds within the same area (Álvarez and Morrone 2004; Yáñez-Ordóñez et al. 2008). Furthermore, molecular data indicate that recent evolutionary processes like IBD also affect the genetic diversity observed within the *Sm2* and *Sp2* colonies sampled at two extremes of the Atlantic axis. By contrast, the lack of IBD in Pacific *Sm1* and *Sp1* is probably due to the close proximity of the colonies. In addition, the lack of a significant relationship between geographic and morphometric distances suggests that wing shape may not be a neutral marker and that it is affected by selective pressures (Reed and Frankham 2001). In these stingless bees, not only the presence of geographic barriers but also the limited dispersion of the colonies influences the differentiation between populations.

Stingless bees are important elements for the preservation of ecosystems as they are the most frequent visitors of many native plants, including economically important local

crops in Mesoamerica (Russell et al. 2005; Morandin and Winston 2006; Ayala et al. 2013). A major threat to stingless bees is the loss of tropical forests, which implies a reduction of potential nesting sites and foraging areas (Foley et al. 2005; Venturieri 2009; May-Itzá et al. 2010). Deforestation may drastically affect the effective size of bee populations as they remain isolated in fragments (Brown and Albrecht 2001). The effect of landscape fragmentation may be more drastic on stingless bees as these insects have some of the lowest dispersal rates, reflecting the mother–daughter colony attachment that persists in the process of swarming (Engels and Imperatriz-Fonseca 1990). This feature of stingless bee reproduction could become maladaptive when habitats are destroyed, as colonies and individuals may be unable to bridge the deteriorated landscapes (Zayed 2009). Indiscriminate use of pesticides is another factor that may affect stingless bee populations, especially those that frequently visit crops (Valdovinos-Núñez et al. 2009). The effect of human-induced translocation of colonies is also particularly relevant to species exploited for honey production (Jaffé et al. 2016), as is the case of *Scaptotrigona*. Avoiding the movement of colonies is crucial to preserve the genetic integrity of ecotypes adapted to particular geographic areas. Thus, it is also important to educate farmers and other individuals through conservation programs that promote the preservation of natural habitats for the reproduction of colonies.

In conclusion, the combined use of genetic and morphological techniques indicates that only the *S. mexicana* colonies from both coasts are distinct evolutionary significant units. A requisite to establish valid species affirmation is congruence among several independent lines of evidence (Schlick-Steiner et al. 2010). Accordingly, further studies that include data from more samples that more extensively cover the distribution area of these two ESUs and that address other aspects of their biology will be necessary to confirm their taxonomic status. In any case, the results presented here favor the promotion of conservation measures not only for the *Sm1* and *Sm2* populations but also, for the *Sp1* and *Sp2* populations, which should be treated as separate units in order to avoid inbreeding and a loss of diversity (González-Acereto et al. 2006). Programs and strategies focused on maintaining the diversity of these bees should be set-up in order to preserve their genetic diversity.

Acknowledgments We appreciate the stingless beekeepers and Dr. Remy Vandame that provided samples. We also thank the editor and three anonymous reviewers for their comments which helped improve the manuscript and BiomedRed for English editing. This research has been funded by the Regional Government of Murcia (Fundación Séneca; Grant Number 19908/GERM/2015) and Conacyt (“CONSERVACION DE LAS ABEJAS SIN AGUIJON DE MEXICO”

Grant Number 103341). Dr. De la Rúa is presently member and receives support from COST Action FA1307, Sustainable pollination in Europe: joint research on bees and other pollinators, SUPER-B (http://www.cost.eu/COST_Actions/fa/Actions/FA1307).

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