



Trophic ecology of tadpoles in floodplain wetlands: combining gut contents, selectivity, and stable isotopes to study feeding segregation of syntopic species

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Abstract Little consensus exists on the role of food partitioning in the organization of tadpole assemblages. We studied trophic ecology of syntopic tadpoles through the analysis of gut contents, selectivity, and stable isotopes to assess the occurrence of food partitioning in tadpole assemblages. Tadpoles ($n = 194$) were collected in three wetlands and corresponded to four species: *Elachistocleis bicolor* (Eb), *Scinax nasicus* (Sn), *Physalaemus albonotatus* (Pa), and *Dendropsophus* sp. (D); and belonged to four ecomorphological groups (EMGs): suspension feeders (Eb), nektonic (Sn), benthic (Pa) and macrophagous (D). *Sn*, and *Pa* showed low selective diet and a wider trophic spectrum than *Eb* and *D*, which mainly consumed one or two food categories. Diet overlap

was higher between *Sn* and *Pa*. Still, *Sn* and *Pa* presented some differences in the food resources consumed. Stable isotopes analysis showed that *Eb*, *Sn*, and *Pa* had a lower trophic position than *D*, explained by the high contribution of animal food oligochaete in *D* diet, in contrast to the importance of algae in the diet of *Eb*, *Sn*, and *Pa*. Diet specialization of some species, combined with the low dietary and isotopic overlap among the ecomorphological groups, suggests that trophic partitioning facilitates coexistence of syntopic tadpoles.

Keywords Amphibians · Ecological guild · Food partitioning · Food selection · Stable isotopes

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Introduction

Anuran tadpoles are found in a variety of freshwater habitats in which different species can coexist

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(Lajmanovich, 2000; Altig et al., 2007). Tadpoles are transitory consumers and, in the absence of fish, their density is usually high enough to affect primary production, flux of nutrients, and competitive interactions (Seale & Beckvar, 1980; Waringer-Löschenkohl & Schagerl, 2001). Thus, understanding their feeding ecology is essential to disentangling the energy and matter pathways within freshwater systems as well as between water and land (Altig et al., 2007).

Historically, tadpoles have been considered as predominantly non-selective phytophagous (Heyer, 1973; Duellman & Trueb, 1986), exhibiting a narrow differentiation in the trophic niche. However, later studies highlighted the importance of the trophic selectivity of some species (Kupferberg, 1997; Sabagh et al., 2012) and of the animal food in the diet (de Rossa-Feres et al., 2004; Ghioca-Robrecht & Smith, 2010; Asrafuzzaman et al., 2018). Gaps in the knowledge of the trophic ecology of most species still exist because, for example, few studies have analyzed food selection considering food availability (Diaz-Paniagua, 1985; Huckembeck et al., 2014; Vassilieva et al., 2017; Kloh et al., 2019), a limitation that could lead to erroneous interpretations of trophic ecology (Maneyro & da Rosa, 2004; López et al. 2009).

Traditionally, studies on trophic ecology have been based only on the analysis of gut contents, inferring ecological differences among species from a snapshot of their diet (i.e., resources ingested at a given time). Instead, stable isotope analyses allow the study of diet in a wider temporal window than the analysis of gut contents because stable isotope signatures reflect the resources assimilated in the consumer tissue (Post, 2002; Fry, 2007; Schiesari et al., 2009; Schalk et al., 2014; Dalu et al., 2015). Stable isotope analyses have shown that trophic position of tadpoles can vary from primary consumers to relatively high trophic levels (Vander Zanden & Rasmussen, 1999; Schiesari et al., 2009). Ecomorphological analysis also represents a useful approach for studying resource use and niche overlap (e.g., Pianka, 1986; Ricklefs & Miles, 1994; Quiroga et al., 2018; Vassilieva et al., 2017). Tadpoles' morphology may reflect different feeding habits (Wassersug, 1980; Altig & Johnston, 1989; Vera Candioti, 2007), thereby providing insights on the degree of trophic overlap between species.

At present, little consensus exists regarding the occurrence of food partitioning and its role in the organization of tadpole assemblages (de Rossa-Feres

et al., 2004). Some studies have suggested that trophic segregation is not important to tadpole coexistence (Heyer, 1974; Pavignano, 1990; Lajmanovich, 2000), while other more recent and isotopic-based studies have highlighted the role of trophic segregation to explain the coexistence of different species (e.g., Hunte-Brown, 2006; Verburg et al., 2007; Schiesari et al., 2009; Vassilieva et al., 2017). Thus, a more detailed and precise research on trophic ecology of tadpole assemblages is necessary to determine the importance of trophic segregation in organization of tadpole assemblages (Altig et al., 2007).

Here, we study trophic ecology and mechanisms of coexistence of tadpoles of four species, belonging to different ecomorphological groups (Altig & Johnston, 1989), that co-occur in temporal floodplain wetlands of the Paraná River. We combined different analytical approaches: (a) gut content analysis; (b) field surveys of food availability to estimate food selectivity; and (c) nitrogen and carbon stable isotopes ratios to determine trophic position. We expect trophic niche partitioning among species, evidencing that trophic segregation is an important mechanism in the structuring of the assemblages of tadpoles.

Materials and methods

Field sampling

We selected three temporary marshes of the floodplain of the Middle Paraná River, near Santa Fe city, Argentina (site 1: 31° 37' 28" S–60° 37' 23" W; site 2: 31° 37' 35" S–60° 36' 51" W; site 3: 31° 38' 29" S–60° 40' 22" W). Sites area varied between 2000 and 7300 m² and they were covered by emergent macrophytes (*Ludwigia* sp., *Panicum* sp., *Echinochloa* sp. and *Paspalum* sp.). The climate of the area is temperate-subtropical, the mean air temperature is 19 °C, and the annual rainfall is approximately 1000 mm, occurring mainly from October to April (Rojas & Saluso, 1987).

Sites were sampled during November 2011 and November 2014 collecting tadpoles in stages 38 to 40 *sensu* Gosner (1960) at 25 to 45 cm depth using a dip net. Collected tadpoles ($n = 194$) were kept in ice to later analysis in laboratory.

Availability of food resources was sampled simultaneously to collection of tadpoles. In each site, food resources were sampled by triplicate. Phytoplankton

was sampled using a 100-ml bottle. Adhered algae was sampled with the scraping of a pool of three submerged leaves of *Ludwigia* sp. (Marker & Bolas, 1984, leaves of similar sizes were selected), a plant species present in the three sites. Zooplankton was sampled by filtering 6 L of wetland water through a net 50- μm mesh (Paggi et al., 2001). Samples of sedimentary organic matter were collected with a Core-type tubular sampler 8 cm in diameter (Anderson et al., 2013).

Samples for isotopic analysis were also taken simultaneously to tadpoles' collection. Measuring the trophic position of consumers requires the estimation of an isotopic baseline; thus, we sampled basal organisms according to the most consumed food categories for each tadpole species. For the adhered algae, we used artificial substrate samplers (green Christmas garlands), which were placed one month before the collection of tadpoles. Sedimentary organic matter and zooplankton were sampled such during food resource sampling as described above. Benthonic invertebrates were collected with a Core sampler (8 cm diameter) that was driven about 15 cm into the bottom sediment (Batzler & Resh, 1992).

Laboratory

Collected tadpoles ($n = 194$) were identified and assigned to ecomorphological groups (EMGs; Fig. 1) following Altig and Johnston (1989): *Elachistocleis bicolor* (Guérin-Méneville, 1838) (Microhylidae, suspension feeder EMG, $n = 47$ tadpoles), *Scinax nasicus* (Cope, 1862) (Hylidae, nektonic EMG, $n = 49$), *Physalaemus albonotatus* (Steindachner, 1864) (Lepidodactylidae, benthic EMG, $n = 52$ tadpoles), and *Dendropsophus* sp. (Hylidae, macrophagous EMG, $n = 46$ tadpoles).

We analyzed between 45 and 50 tadpoles per species (14–17 tadpoles per site) to study diet through gut contents. For diet analysis, we extracted the digestive tract of each tadpole and the contents were analyzed and identified under an inverted microscope at $400\times$ (Nikon TS100 Eclipse). Phytoplankton and adhered algae were classified according to seven morphologically based functional groups (MBFG; Kruk et al., 2010; see Table 1 of Appendix for details): Small algae (MBFG1; up to $134\ \mu\text{m}^2$); Algae with silica (MBFG2); Filamentous algae (MBFG3); Medium-size algae (MBFG4; up to $791\ \mu\text{m}^2$); Flagellated

unicellular algae (MBFG5); Diatomean (MBFG6); and Colonial with mucilage (MBFG7; up to $791\ \mu\text{m}^2$). Food items with an advanced degree of digestion, but still with identifiable structures, such as pieces of chitinous exoskeleton and locomotor appendages, were classified as animal remains. Fragments of vegetation were classified as plant remains (Huckembeck et al., 2014). Unidentified items, which formed a mass of organic material, were classified as remaining sedimentary organic matter (i.e., detritus) (Asrafuzman et al., 2018).

For food availability, we estimated relative abundance of each food category in the environment. Phytoplankton and adhered algae were classified according to seven morphologically based functional groups as in diet analysis and were quantified using an inverted microscope following the Utermöhl (1958) technique. Zooplankton and sedimentary organic matter were quantified under a binocular microscope (Molina et al., 2010).

For stable isotopes analysis, we extracted tail muscle tissue from tadpoles collected in sites 1 and 2 (Verburg et al., 2007; Dalu et al., 2015). Ten 1 mg dry weight samples per species were prepared, coming each sample from the tail of one to three tadpoles, depending on amount of tissue in specimens. Among the food resources available in the environment, six 1 mg dry weight samples of each food item (phytoplankton, adhered algae, sedimentary organic matter, cladocerans and oligochaetes) were prepared. We considered only cladocerans of zooplankton and only oligochaetes of benthonic invertebrates because these were the only consumed items within these communities.

Analyses of stable isotopes were performed in a mass spectrometer (IRMS Finnigan MATDelta S) coupled to an elemental analyser (CATNAS, Facultad de Agronomía, Universidad de la República, Uruguay). The ratio of stable isotopes was expressed by convention in delta (δ) notation: $\delta X = \{(R_{\text{sample}}/R_{\text{standard}}) - 1\} \times 1000$, where δX is the isotope ratio of the sample relative to a standard, R_{sample} and R_{standard} are the fractions of heavy to light isotopes in the sample and the standard, respectively. The standard values for $\delta^{15}\text{N}$ were atmospheric Nitrogen and for $\delta^{13}\text{C}$ was Pee Dee Belemnite. As high lipid levels (indicated by a high C:N ratio) may drive $\delta^{13}\text{C}$ values in a negative direction (Matthews & Mazumder, 2005), when C:N ratios were higher than 3.5, the $\delta^{13}\text{C}$ values were

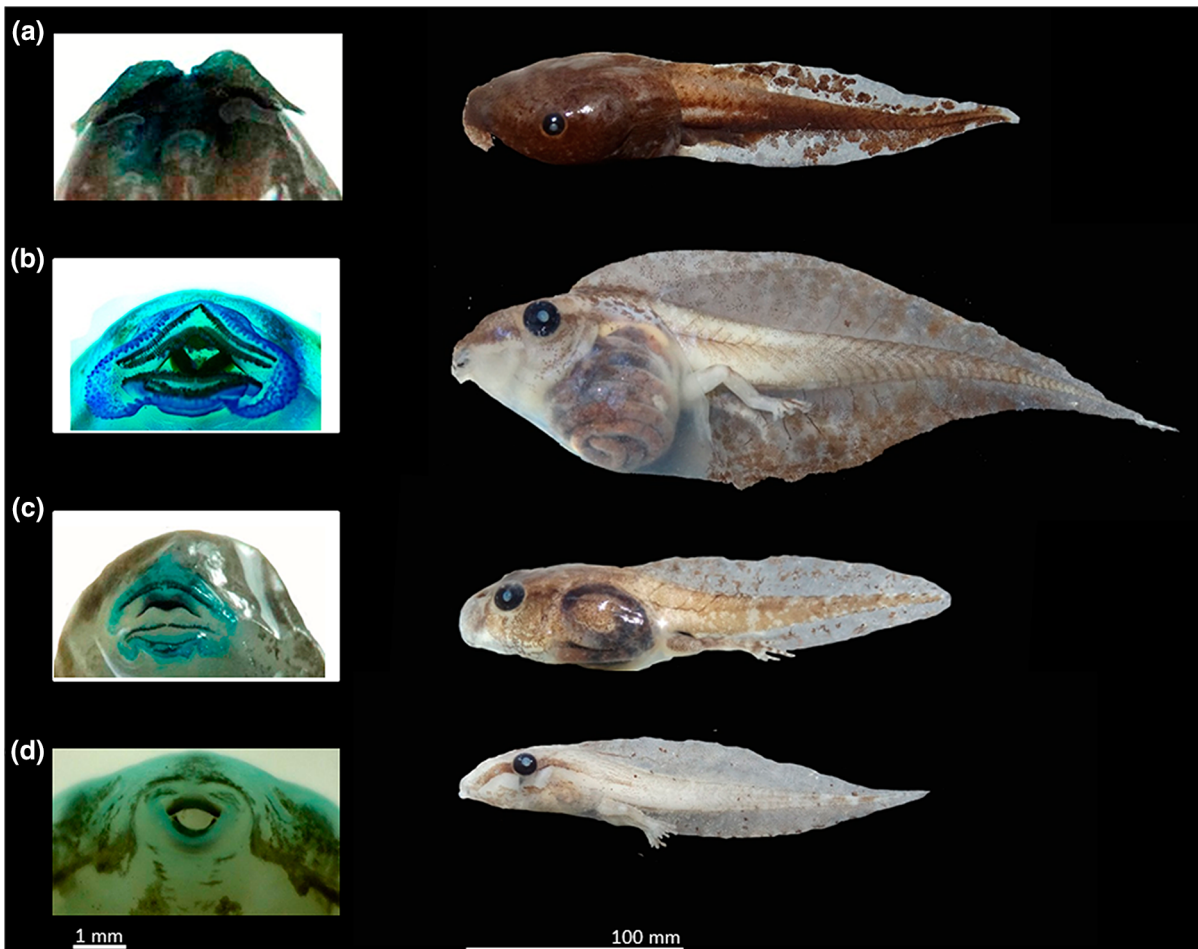


Fig. 1 Assignment of tadpoles to ecomorphological groups (EMGs) following Altig and Johnston (1989). Species oral disk and body side view: **a** *E. bicolor* (suspension feeder EMG); **b** *S.*

nasicus (nektonic EMG); **c** *P. albonotatus* (benthic EMG); **d** *Dendropsophus* sp. (macrophagous EMG)

normalized using the following equation: $\delta^{13}\text{C}_{\text{corrected}} = \delta^{13}\text{C} - 3.32 + 0.99 \times \text{C:N}$ (Post et al., 2007).

Data analyses

Collected tadpoles among the three sampled sites were combined for data analysis. The Index of Relative Importance (IRI%) proposed by Pinkas et al. (1971) was used to determine the importance of each food category in the diet of each species. This index combines abundance, frequency of occurrence and volume in a single estimator of relative importance of each food category in the diet (Hart & Christensen, 2002). Volume of ingested items was calculated by approximation to regular geometric shapes. When we found chaeta of oligochaetes in tadpoles' guts, we

considered these as equivalent to only one individual for data analysis.

An overlap index (Pianka, 1973) was calculated to assess differences (and similarities) of diet among species. We tested whether overlap indices were different from those expected by chance using the RA2 Overlap Randomization Algorithm through the *EcoSimR* package in R (R Core Team, 2016) with IRI% values.

A Permutational Multivariate Analysis (PERMANOVA) was performed to assess whether the composition of the diet of the tadpoles (i.e., relative abundance of each food category) differed among the species. To assess these differences (and similarities) among species graphically, we used a Non-metric Multidimensional Scaling (NMDS). The

Table 1 Percentage of index of relative importance (IRI%) of food category the tadpoles

Food category (IRI%)	<i>E. bicolor</i>	<i>S. nasicus</i>	<i>P. albonotatus</i>	<i>Dendropsophus</i> sp.
Algae				
Small algae	47.25	11.80	4.59	0.96
Algae with silica	< 0.01	< 0.01	< 0.01	–
Filamentous algae	4.88	39.56	14.03	8.54
Medium-size algae	1.57	10.64	3.35	1.34
Flagellated unicellular algae	6.01	4.51	8.02	21.64
Diatomean	4.91	3.52	16.01	4.04
Colonial with mucilage	0.37	2.46	0.28	0.63
Vegetation				
Plant remains	–	0.09	0.15	0.05
Animals				
Rotifera	0.56	3.49	0.79	–
Cladocera	25.78	23.56	33.19	1.53
Nematoda	–	0.03	0.02	–
Gastrotricha	–	< 0.01	–	–
Oligochaeta	–	–	–	60.97
Insecta	–	–	< 0.01	–
Protozoa	0.08	0.13	0.18	0.30
Sedimentary organic matter	8.58	0.21	19.33	–
Undetermined	< 0.01	< 0.01	0.05	–

Bold letters indicate the most important food items of each species. In bold, the most important item stands out for each species

NMDS ordination was evaluated by the coefficient of determination based on stress (R^2 ; Oksanen, 2011). The matrix of food relative abundances to PERMANOVA and NMDS was double standardized through the Wisconsin method, where food categories are first standardized by maxima category and then by total categories (Oksanen, 2011). The trophic dissimilarity matrix was based on the Bray–Curtis index. The PERMANOVA and the NMDS ordination were created using the ‘adonis’ and ‘metaMDS’ functions, respectively, of the *vegan* package for R (Oksanen et al., 2015; R Core Team, 2016). In addition, we performed an Indicator Species Analysis (ISA) to determine which food categories were particularly characteristic (indicative) of each tadpole species. This analysis was performed using the *indicspecies* package in R (Cáceres & Legendre, 2009; R Core Team, 2016).

We tested for selectivity of tadpoles on the more important food categories by comparing their relative abundances between guts contents and the environment using a two-sample permutation test through the *perm* package in R (Fay, 2010; R Core Team, 2016). Feeding selectivity was calculated as the linear index of food selection (Li) proposed by Strauss (1979). Li

ranges from -1 to $+1$. $Li = 0$ indicates a lack of selection (i.e., consumption of the category in the same proportion as it occurs in the environment); $Li < 0$ indicates negative selection, avoidance or inaccessibility (i.e., less of food category i occurs in the diet than expected from random feeding); and $Li > 0$ indicates positive selection or preference (more of food category i occurs in the diet than expected from its environmental abundance).

We assessed the differences in isotopic signatures among the tadpoles by a Kruskal–Wallis test. The baseline for each species was selected according to their most important food categories: phytoplankton and sedimentary organic matter for *E. bicolor*, phytoplankton and algae adhered for *S. nasicus* and *P. albonotatus*, and oligochaete for *Dendropsophus* sp. To estimate the trophic position of *Dendropsophus* sp. we used a single-end member equation: $TP_{\text{consumer}} = \{(\delta^{15}N_{\text{consumer}} - \delta^{15}N_{\text{base}}) - TP_{\text{base}}\} / \text{TEF}$, where TP_{consumer} is the trophic position of the tadpole and TP_{base} is the trophic position of the baseline (in this case, 1). For the other species, we applied a two-end member model proposed by Post (2002): $FCL = (\delta^{15}N_{\text{consumer}} - (\delta^{15}N_{\text{base1}} \times \alpha + \delta^{15}N_{\text{base2}} \times (1 - \alpha))) /$

TEF, and $\alpha = (\delta^{13}\text{C}_{\text{consumer}} - \delta^{13}\text{C}_{\text{base1}})/(\delta^{13}\text{C}_{\text{base1}} - \delta^{13}\text{C}_{\text{base2}})$, where TEF is the trophic enrichment factor for $\delta^{15}\text{N}$. We set the TEF as 2.22, which is the mean value reported for tadpoles (Schiesari et al., 2009; Caut et al., 2013; San Sebastián et al., 2015).

Results

Diet of *E. bicolor* was composed mainly of unicellular algae (Table 1), mostly represented by *Chlorella* sp. (IRI = 47.25%), and Cladocera, represented by individuals of the family Chydoridae (IRI = 25.78%). Diet of *S. nasicus* was composed of filamentous algae of the genus *Oedogonium* sp. (IRI = 39.56%), Cladocera (IRI = 23.56%), unicellular algae (IRI = 11.80%) and medium-size algae (IRI = 10.64%). Diet of *P. albonotatus* was composed mainly of Cladocera (IRI = 33.19%), sedimentary organic matter (IRI = 19.33%), diatoms (IRI = 16.01%) and filamentous algae (IRI = 14.03%). Diet of *Dendropsophus* sp. was mainly composed of Oligochaeta (IRI = 60.97%).

Diet composition differed among species (PERMANOVA: $F_{3, 189} = 101.01$, $P < 0.001$, $R^2 = 0.62$). NMDS showed a greater separation in the diet between *Dendropsophus* sp. from the rest of the species along the first dimension, related to *Dendropsophus* sp. predation upon Oligochaeta (Fig. 2). A different proportion of fine-particulate sedimentary organic matter and unicellular algae in the diet of *E. bicolor* in relation to *S. nasicus* and *P. albonotatus* accounted for the separation between these species in the second dimension (Fig. 2). Oligochaeta were more abundant and/or frequent in the diet of *Dendropsophus* sp. (ISA: $P = 0.005$); sedimentary organic matter and algae with silica were more important in the diet of *P. albonotatus* (ISA: $P < 0.01$); sedimentary organic matter was more abundant and/or frequent in the diet of *E. bicolor* (ISA: $P < 0.05$); and plant remains and Nematoda were more common in the diet of *P. albonotatus* and *S. nasicus* in relation to *Dendropsophus* sp. and *E. bicolor* (ISA: $P < 0.05$). Moreover, all categories of food (unicellular algae, medium-size algae, diatomean, colonial with mucilage Protozoa and Rotifera) were more abundant and frequent in *P. albonotatus*, *E. bicolor* and *S. nasicus* than in of *Dendropsophus* sp. (ISA: $P < 0.001$).

Trophic overlap was less than expected by chance among *Dendropsophus* sp. and the rest of the species and between *S. nasicus* and *P. albonotatus*, and was not different than expected by chance between *E. bicolor* and *S. nasicus* and between *E. bicolor* and *P. albonotatus* (Table 2).

Elachistocleis bicolor selectively fed on unicellular algae, while the consumption of Cladocera was proportional to the environmental availability (Fig. 3A). *Scinax nasicus* avoided the medium-size algae and *P. albonotatus* selectively fed on diatomean (Fig. 3B and 3C). The selection of Oligochaeta by *Dendropsophus* sp. was not measured because of the difficulty mentioned in estimating the number of worms ingested (see methods section).

The isotopic signature of $\delta^{13}\text{C}$ varied among species (KW = 27.5; $P < 0.001$), with differences between *E. bicolor* and *Dendropsophus* sp. (Dunn: $P = 0.006$), *P. albonotatus* and *Dendropsophus* sp. (Dunn: $P = 0.021$), *P. albonotatus* and *E. bicolor* (Dunn: $P < 0.001$), *S. nasicus* and *Dendropsophus* sp. (Dunn: $P = 0.021$), *S. nasicus* and *E. bicolor* (Dunn: $P < 0.001$). The isotopic signature of $\delta^{15}\text{N}$ also varied among species (KW = 8.45, $P < 0.037$), with differences between *P. albonotatus* and *Dendropsophus* sp. (Dunn: $P < 0.036$), *S. nasicus* and *Dendropsophus* sp. (Dunn: $P < 0.002$) (Table 3). *E. bicolor*, *S. nasicus* and *P. albonotatus* had a lower trophic position than *Dendropsophus* sp. (Fig. 4).

Discussion

Our results show that co-occurring tadpoles of *E. bicolor*, *S. nasicus*, *P. albonotatus*, and *Dendropsophus* sp. exploit different food resources in the floodplain wetlands of the Paraná River, selecting and feeding upon different aquatic communities. These results reinforce the conclusions of other studies that have also found differences in the diet among tadpoles of syntopic species (e.g., de Rossa-Feres et al., 2004; Pollo et al., 2015; Santos et al., 2015; Vassilieva et al., 2017; Schmidt et al., 2017), suggesting that trophic segregation may represent a mechanism of coexistence in tadpole assemblages.

Tadpoles of the four species differed in diet composition, and these differences reflected the ecomorphological differences among them. *Dendropsophus* sp. showed the greatest diet differences. The

Fig. 2 NMDS based on the relative abundance of food category in gut contents of tadpoles of *Dendropsophus* sp. (black circles), *Elachistocleis bicolor* (white squares), *Scinax nasicus* (grey rhombus), and *Physalaemus albonotatus* (black triangles)

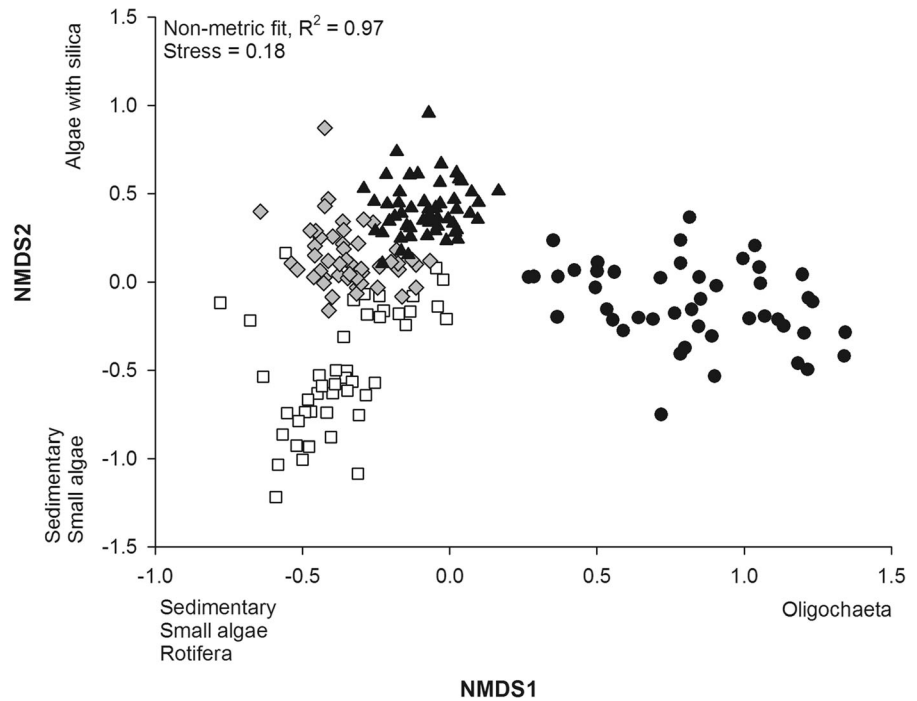


Table 2 Trophic overlap among the four tadpoles’ species, *Elachistocleis bicolor* (Eb), *Scinax nasicus* (Sn), *Physalaemus albonotatus* (Pa) and *Dendropsophus* sp. (D), using Pianka (1973) niche overlap index

	Observed	Bootstrap		Standardized Effect size
		Mean ± SD	95% CIs	
Eb-Sn	0.52	0.61 ± 0.01	0.41–0.79	1.41
Eb-Pa	0.51	0.60 ± 0.01	0.39–0.79	– 0.92
Eb-D	0.05	0.64 ± 0.01	0.44–0.80	– 5.48
Sn-Pa	0.68	0.70 ± 0.01	0.50–0.86	– 0.12
Sn-D	0.10	0.67 ± 0.01	0.50–0.82	– 5.82
Pa-D	0.09	0.67 ± 0.01	0.51–0.81	– 6.20

Observed index and bootstrap-based mean ± SD and 95% intervals of confidence (95% CIs) are showed. Values in bold indicate trophic overlaps different than expected by chance

elevated trophic segregation of *Dendropsophus* sp. seems to be a characteristic extended throughout its wide distribution (de Rossa-Feres et al., 2004). *Dendropsophus* sp. can ingest larger food items than any other syntopic species because of its macrophagous buccal structure, allowing ingestion of large preys by suction, such as benthos oligochaetes as long as tadpoles size (Vera Candioti, 2007). The branchial

basket is relatively reduced in some genus *Dendropsophus* (Vera Candioti, 2007), limiting the ingestion of very small items by filtration. In contrast, the highly developed branchial basket of the suspension feeder tadpoles of *E. bicolor* (Vera Candioti, 2007), likely allows for filter feeding on very small items such as the microalgae ($\approx 20\mu\text{m}$). Similarly, Echeverría et al., (2007) reported the consumption of relatively small items such as desmid planktonic algae (medium-size $50\mu\text{m}$) in tadpoles of *E. bicolor*. However, the highly developed branchial basket of *E. bicolor* not necessarily precludes from ingesting large prey, since we also found cladocerans of $\approx 250\mu\text{m}$ in their gut contents, as also reported Vera Candioti (2007). Similarly, the consume of relatively large prey such as zooplankton, insect larvae, and tadpoles was observed in other species with suspension feeder tadpoles as in genus *Rhinophrynus* (Starrett, 1960; Wassersug, 1972).

The diets of the tadpoles of *S. nasicus* and *P. albonotatus* were more diverse in comparison with that of *Dendropsophus* sp. and *E. bicolor*. Diet overlap between *S. nasicus* and *P. albonotatus* was the higher within the assemblage. Tadpoles of these two species have a similar configuration of the hyobranchial apparatus, with around 50% of the total area occupied

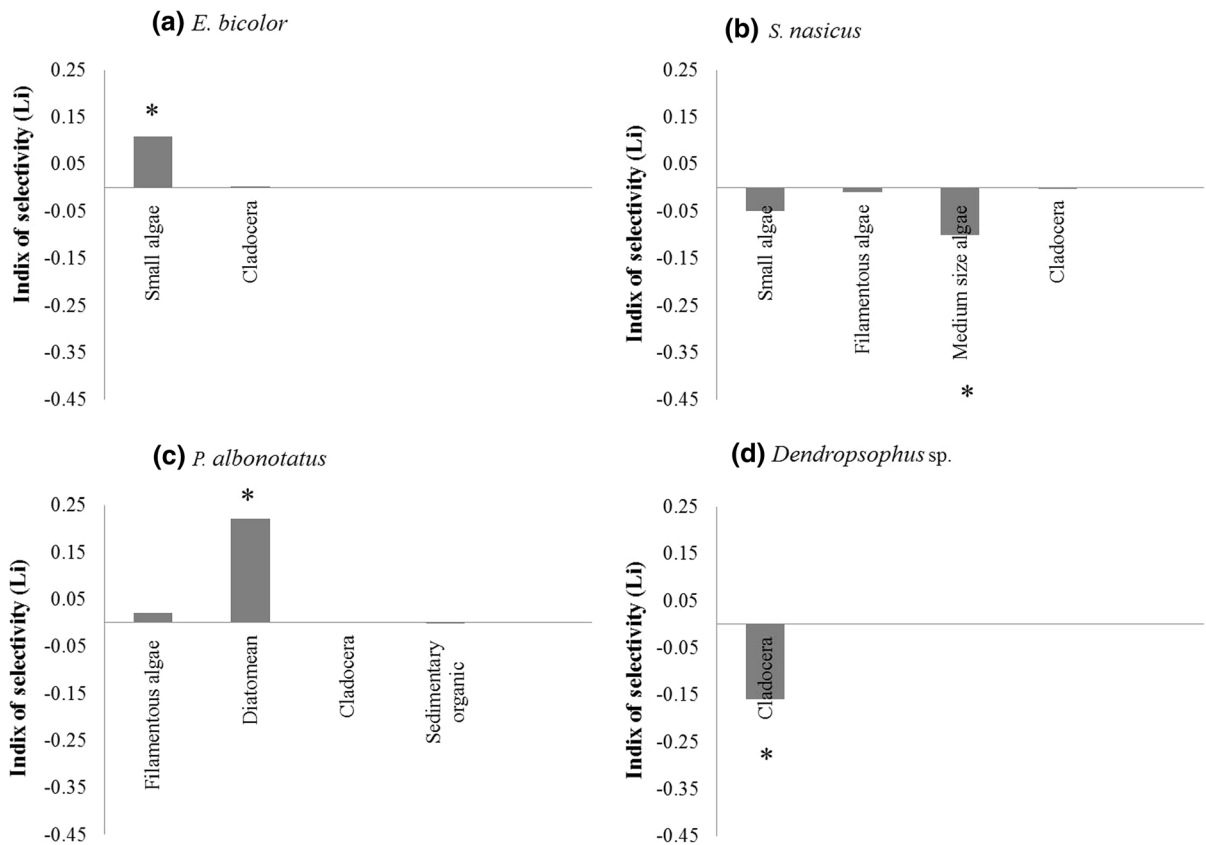


Fig. 3 Strauss's linear index of selectivity (L_i) for *E. bicolor* (A), *S. nasicus* (B), *P. albonotatus* (C) and *Dendropsophus* sp (D), applied to the relative abundance of trophic items. $L_i > 0$:

positive selection, $L_i < 0$: negative selection, $L_i = 0$ absence of selectivity. Asterisk (*) indicates statistical difference ($P < 0.05$) from zero (two-sample permutation tests, two-tailed)

Table 3 Isotopic signature ($\delta^{13}\text{C}$ and $\delta^{15}\text{N}$) of tadpoles and food category of the wetlands

	$\delta^{13}\text{C}$	$\delta^{15}\text{N}$
<i>E. bicolor</i>	-20.375 ± 1.071	3.441 ± 2.702
<i>S. nasicus</i>	-26.329 ± 1.112	2.978 ± 0.349
<i>P. albonotatus</i>	-26.396 ± 01.165	3.445 ± 0.737
<i>Dendropsophus</i> sp.	-24.692 ± 1.257	4.234 ± 0.342
Phytoplankton	-25.797 ± 1.448	2.886 ± 0.560
Algae adhered	-25.928 ± 0.794	1.473 ± 0.423
Zooplankton	-29.457 ± 0.872	2.318 ± 0.529
Oligochaete	-27.167 ± 3.463	2.952 ± 0.80
Sedimentary organic matter	-25.322 ± 0.299	3.123 ± 0.219

Mean and standard deviation is indicated

by the branchial basket (Vera Candiotti, 2007), and a similar buccal apparatus adapted to scrape surfaces. This allows them to consume adhered algae (mainly filamentous algae), as it has also been found in other species included within the benthic and nektonic ecomorphological groups (de Rossa-Feres et al., 2004;

Vera Candiotti et al., 2004; de Sousa Filho et al., 2007). We found that tadpoles of both species prey on microcrustaceans (cladocerans and rotiferans), as it has been reported in previous studies for tadpoles of these species and other congeneric species (de Rossa-Feres et al., 2004, de Sousa Filho et al., 2007, Vera

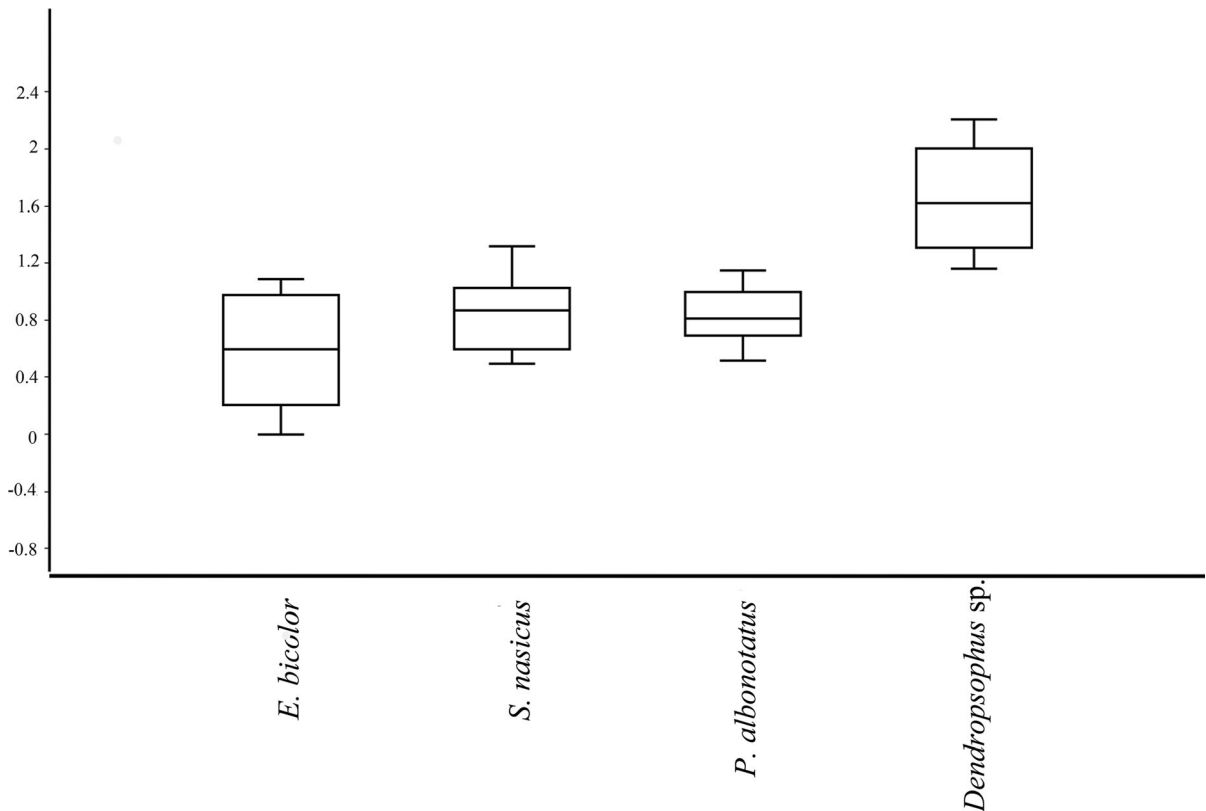


Fig. 4 Boxplot of the trophic position the tadpoles of the four species studied: *E. bicolor*, *S. nasicus*, *P. albonotatus*, and *Dendropsophus sp.*

Candiotti (2007). Beyond these similarities, *S. nasicus* and *P. albonotatus* also presented some differences in their diets. *Physalaemus albonotatus* consumed a higher proportion of sedimentary organic matter and a lower proportion of filamentous algae than *S. nasicus*. The higher importance of sedimentary organic matter in the diet of *P. albonotatus* could be related to benthic habits of their tadpoles, in contrast to more nektonic habits of tadpoles of *S. nasicus* (Peltzer & Lajmanovich, 2004). Pollo et al. (2015) also studied diet of nektonic and benthic tadpoles and suggested that differences in diet were related to behavioral differences for food acquisition in the water column between species. An alternative – or complementary – explanation to the lower consumption of filamentous algae by *P. albonotatus*, is the lower number of labial teeth than in *S. nasicus* (Kehr & Duré, 1995; Kehr et al., 2004; Vera Candiotti 2007). Studies have shown that the efficiency in the consume of adhered algae is lower in tadpoles with a lower number of labial teeth (Venesky et al., 2010). Thus, subtle differences in

morphology and microhabitat exploitation contributes diet segregation (Kehr & Duré, 1995; Kehr et al., 2004; Peltzer & Lajmanovich, 2004; Vera Candiotti 2007), even for tadpoles inhabiting the same shallow pond where the water column is less than half a meter deep.

The degree to which tadpoles of different species select their food or not remains little known. Both, selectivity (Johnson 1991; Kupferberg, 1997) and non-selectivity (Seale & Beckvar, 1980; Chen et al., 2008), have been reported for feeding of tadpoles of different species, although the number of studies is limited, in part because of the difficulty of assessing in detail environmental food availability. We found a selectivity gradient among studied species. Tadpoles of the nektonic and benthic ecomorphological groups, *S. nasicus* and *P. albonotatus*, showed a wider trophic spectrum and did not select their most important food categories. In contrast tadpoles of the suspension feeder and macrophagous ecomorphological groups, *E. bicolor* and *Dendropsophus sp.* respectively, could

be considered specialized consumers, because they largely and selectively consumed one or two food categories, which accounted for more than 80% of their diet: small unicellular algae in the case of *E. bicolor* and oligochaetes in *Dendropsophus* sp.

A less studied approach to understanding the trophic ecology of tadpoles' assemblages is the analysis of stable carbon and nitrogen isotopes. In recent studies, authors have suggested different positions in food webs for tadpoles (Verburg et al., 2007; Huckembeck et al., 2014; Schmidt et al., 2017). In accordance with these suggestions, we found a low trophic position for *E. bicolor*, *S. nasicus* and *P. albonotatus* tadpoles and a higher trophic position for *Dendropsophus* sp. Thus, our results show that suspension feeder, nektonic and benthic tadpoles, are consumers of the first levels of the food web, while microphagous tadpoles have a higher trophic position, one level above in the food web.

Schalk et al. (2017) found differences between the trophic position of *S. nasicus* and *P. albonotatus*, assigning a higher trophic position to *S. nasicus* (TP = 5.7) than to *P. albonotatus* (TP = 3). Differences with our results could be due to differences in food resources availability together with the trophic plasticity of tadpoles (López et al., 2015). The selection of different sources of carbon and nitrogen for isotopic analysis by Schalk et al. (2017) could also account for the differences with our study. In both species, we found a broad trophic spectrum composed by algae and, in a lower proportion by zooplankton. The low consumption of animal prey would not be enough to increase their trophic position. In relation to *E. bicolor*, its narrow trophic spectrum focused in the consumption of microalgae explain this species low trophic position in the food web. Although *S. nasicus*, *P. albonotatus* and *E. bicolor* are at the lowest trophic level within consumers of the food web, our multi-approaches strategy to the study of the trophic ecology of the assembly allowed us to recognize trophic segregation among these species, indicating that the three fulfil different functions in the ecosystem. Lastly, the predation on oligochaete by *Dendropsophus* sp., the largest prey registered for the hole tadpoles assemblage, allows to the macrophagous tadpoles the incorporation of a great amount of energy per prey and the access to nutrients not available to other syntopic tadpoles, and explain its higher trophic position in the food web.

Conclusion

Our integrative approach provided a more in-depth view of the trophic ecology of neotropical tadpoles assemblages. Using a combination of techniques and a detailed analysis of food categories, we found a relatively high specialization of the diet among tadpoles of different species, showing the importance of using an integrative approach to study the trophic ecology. Our results suggest that the partitioning of the trophic niche is an important mechanism for the segregation and coexistence of tadpoles in subtropical floodplain wetlands. Finally, the absence of ecological redundancy among neotropical tadpoles assemblage highlights the importance of conserving the diversity of species to maintain the identity and balance of aquatic ecosystems they inhabit.

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