

The effects of invasive rainbow trout on habitat use and diel locomotor activity in the South American Creole perch: an experimental approach

María Gabriela Otturi · Miguel
Ángel Battini · Juan Pablo Barriga

Received: 20 November 2015 / Revised: 18 April 2016 / Accepted: 19 April 2016 / Published online: 4 May 2016
© Springer International Publishing Switzerland 2016

Abstract The effects of the introduced rainbow trout (*Oncorhynchus mykiss*) on both habitat use and diel locomotor activity of the native Creole perch (*Percichthys trucha*) were assessed. Experiments were performed with juveniles of both species during a 24-h period. Habitat use was measured in channels with different habitat categories, and diel locomotor activity was registered with an infrared sensor. Both species used mainly cobbles. In the absence of *O. mykiss*, *P. trucha* used macrophytes and gravel secondarily, during dusk and midnight, respectively. *Oncorhynchus mykiss* used gravel and macrophytes secondarily, during midday. In the presence of *O. mykiss*, *P. trucha* increased its use of macrophytes and decreased its use of less structured habitats. *Percichthys trucha* juveniles were more active during night. This pattern did not change in the presence of *O. mykiss*, nor was the intensity of movement affected. These results reveal the importance of structured environments for juveniles of both species. In particular, substrates of large size with interstitial spaces in

between serve as shelter and reduce visual detection between individuals. Both the association of *P. trucha* juveniles with structured habitats and their nocturnal activity reduce predation risk.

Keywords Interspecific interactions · Habitat use · Diel activity · Introduced salmonids · *Percichthys trucha* · *Oncorhynchus mykiss*

Introduction

Several factors influence the habitat use of freshwater fishes (Jackson et al., 2001). Among these, substrate type, particle size, competition, and predation are important (Chase et al., 2002; Van Liefferinge et al., 2005). In rivers, habitat heterogeneity and patches of different substrates influence the spatial distribution of the fishes (Petty & Grossman, 1996; Dixon & Vokoun, 2009). Biotic factors are more conspicuous in small-scale studies and often underestimated in large-scale studies (Grossman et al., 2006).

Competition with introduced fish can lead to a decline in native fish populations (e.g., VanderZanden et al., 1999). There is evidence of asymmetric competition where the outcome is the habitat displacement of native by non-native species (Jonsson et al., 2008; Amundsen et al., 2010; Grabowska et al., 2016). An example of this is the habitat shift of *Etheostoma olmstedii* Storer, 1842 in sympatry with

Handling editor: Odd Terje Sandlund

M. G. Otturi (✉) · M. Á. Battini · J. P. Barriga
Laboratorio de Ictiología y Acuicultura Experimental,
Instituto de Investigaciones en Biodiversidad y
Medioambiente (INIBIOMA), Consejo Nacional de
Investigaciones Científicas y Técnicas (CONICET) -
Universidad Nacional del Comahue, Quintral 1250,
8400 San Carlos de Bariloche, Río Negro, Argentina
e-mail: gabyotturi@yahoo.com.ar

the non-native *E. zonale* (Copp, 1868), where *E. olmstedii* is restricted to shallower pools or marginal zones of streams (Van Snik Gray et al., 2016). A more complex situation was reported by Hasegawa & Maekawa (2006), when *Oncorhynchus mykiss* (Walbaum, 1792) displaces the white-spotted charr, *Salvelinus leucomaenis* (Pallas, 1814) forcing it to compete with masu salmon *Oncorhynchus masou* (Brevoort, 1856), another native species. In any case, the displacement to a suboptimal habitat implies a reduction in fitness (e.g., Blanchet et al., 2007).

Competition impact can be reduced by temporal segregation (Schoener, 1974; Brännäs & Alanära, 1997; Kronfeld-Schor & Dayan, 2003). That is, different individuals can use the same resources at different times of the day, thus lessening the time they interact. This shift in the temporal use of resources has been shown at both inter and intraspecific level. Alanära et al. (2001) pointed out that *Salmo trutta* Linnaeus, 1758 dominant individuals foraged mainly at optimal times while subordinates of the same species fed in suboptimal periods. Competition between juvenile exotic rainbow trout and native *Salmo salar* Linnaeus, 1758 forced Atlantic salmon to be more active during the day, thereby increasing the risk of being depredated (Blanchet et al., 2008). Therefore, as biotic variables may differ between day and night (Schlosser, 1988; Copp et al., 2005), habitat use and interspecific interaction studies must be performed over a complete daily cycle (Copp, 2008).

Studies of the diel activity of coexisting species are relevant for the evaluation of the dynamics of competitive interactions. This activity can be synchronized by factors such as light, temperature (López-Olmeda et al., 2006), feeding (Landless, 1976) or social interactions. Since locomotor activity is related to the light–dark cycle (Thorpe, 1978), measuring swimming activity contributes to the understanding of the daily rhythm of a species.

Salmonids are the main invasive fish species in Patagonia. They were introduced at the beginning of the 20th century (Marini, 1936) and *O. mykiss* in particular has become widely dispersed since 1904 (García de Leaniz et al., 2010; Macchi & Vigliano, 2014). It has successfully invaded several environments outside its original distribution range and is considered to be amongst the 100 most invasive species (ISSG, 2012). Among native Patagonian species, the most ecologically similar to *O. mykiss* is

the Creole perch, *Percichthys trucha* (Valenciennes, 1833). This species is widely distributed in Argentina and Chile, both in lentic and lotic environments (Menni, 2004; Ruzzante et al., 2006). It presents remarkable intraspecific phenotypic variation (Ruzzante et al., 1998, 2003), associated to differences in predation risk (Ruzzante et al., 2011) and trophic resource use (Crichigno et al., 2014).

The use of lotic environments by juveniles of both *O. mykiss* and *P. trucha* (Barriga et al., 2007) would indicate that the probability of interspecific interactions is high in these habitats. Juvenile *O. mykiss* have been found from fast, shallower waters to slower, deeper habitats and also using the interstitial space between rocks (Johnson & Douglass, 2009). *Percichthys trucha* juveniles, in turn, have been registered in slow water habitats with aquatic vegetation in rivers with *O. mykiss* presence (Barriga et al., 2013). The diel activity of *P. trucha* juveniles is currently unknown. The main goals of this study were (1) to determine *P. trucha* and *O. mykiss* habitat use during a diel cycle and to evaluate whether *P. trucha* habitat use varies in the presence of *O. mykiss*; (2) to determine the diel locomotor activity of *P. trucha* and to evaluate whether it is affected by the presence of *O. mykiss*. The study hypotheses were as follows: (1) Juveniles of *P. trucha* use vegetated areas whereas juveniles of *O. mykiss* select habitats of coarse substrate. (2) A shift in *P. trucha* habitat use is evident when *O. mykiss* is present. (3) *Percichthys trucha* locomotor activity diel pattern is affected by the presence of *O. mykiss*.

Materials and methods

Fish collection

Fish were collected during the fall of 2013, in two locations on the Limay River Basin: Pichileufu River (41°05'S, 70°49'W, 926 m a.s.l.) and Piedra del Águila Reservoir (40°27'S, 70°39'W, 590 m a.s.l.). The ichthyofauna of the Basin includes native species: *Galaxias maculatus* (Jenyns, 1842), *G. platei* Steindachner, 1898, *P. trucha*, *Hatcheria macraei* (Girard, 1855), *Trichomycterus areolatus* Valenciennes, 1846, *Olivaichthys viedmensis* (MacDonagh, 1931) and *Odontesthes hatcheri* (Eigenmann, 1909), and introduced salmonids: *O. mykiss*, *S. trutta* and *Salvelinus fontinalis* (Mitchill, 1814). Fish were captured using a

24 V DC backpack electrofishing unit, model 12-B (Smith-Root, Inc., Vancouver, WA, USA) and a seine net (20-m-long, 1.6-m-deep, 12-mm mesh size, with a 10-m-long central section of 1-mm mesh size). A total of 100 *P. trucha* juveniles and 80 *O. mykiss* juveniles were captured and transported in heat insulated containers to the *Centro de Salmonicultura Bariloche* of the *Universidad Nacional del Comahue* in San Carlos de Bariloche city, Río Negro province, Argentina.

Habitat use experiment

Habitat use experiments were performed indoors in four flow-regulated channels (2.85 m long \times 0.3 m wide \times 0.27 m depth). Each channel was subdivided into four equal-sized sections, and different categories of habitat were placed in each one. Water was supplied from Gutierrez Stream, constituting an open system with an average flow of 90 ml/s, temperature: 11.2°C; pH: 7.43; O₂: 5.9 mg/l and conductivity: 70.5 μ S/cm. The natural photoperiod was imitated artificially using an electronic device. Dawn was simulated by means of a progressive lineal increment of light intensity from 0 to 85 lx (or lumen m⁻²) in one hour. The inverse situation was generated to simulate dusk, namely a lineal decrement from 85 to 0 lx in one hour. The beginning of each crepuscule was set to approximate the natural cycle; thus, the 24-h cycle was divided as follows: 1 h of dawn, 11 h of maximum light intensity (at 85 lx measured at water surface level), 1 h of dusk, and 11 h of darkness (at 0 lx). Lamps were placed in a position that ensured homogeneous distribution of light over the entire surface of the channel. Four habitat categories were used: sand (<4 mm in diameter), gravel (>4 and <64 mm), cobbles (>64 and <150 mm), and 20-cm-tall green artificial aquarium plants (Sunny Aquarium Company, Singapore) to mimic *Myriophyllum* sp., the dominant macrophyte species found in the river. Habitat categories were placed in each channel in different positions relative to the water inlet, so as to discern whether fish distribution was related to habitat or to position within the channel (Fig. 1). Channels were covered with black opaque plastic both above and at the sides, to prevent infiltration of external light and fish disturbance. Fish were not fed either during acclimation or experimental periods. In addition, macroinvertebrate colonization was prevented using mechanical filters placed under

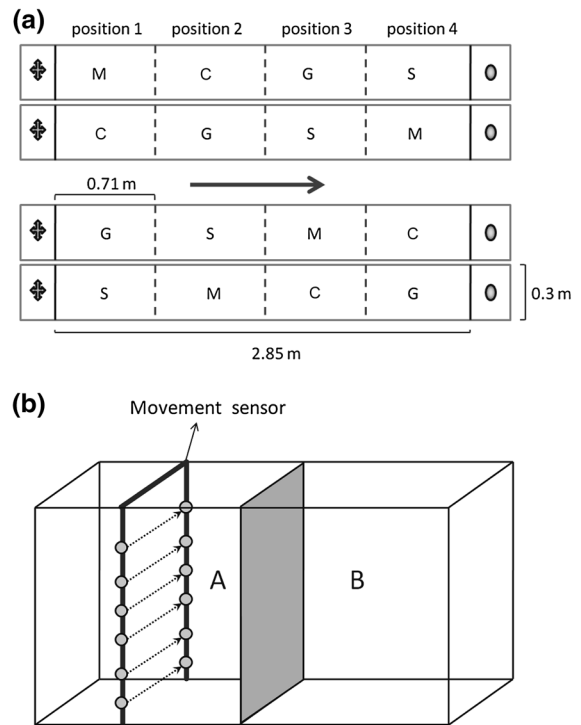


Fig. 1 **a** Scheme of experimental channels and distribution of habitats: *S* sand, *G* gravel, *C* cobbles, and *M* macrophytes. The arrow indicates water flow direction. Left crosses and right circles indicate water inlet and outlet, respectively. Dotted lines represent the sluices. **b** Scheme of experimental aquarium and movement sensor (dotted arrows represent infrared beams). *A* Compartment where locomotor activity of *P. trucha* juvenile was registered in both treatments. *B* compartment where *O. mykiss* was placed during the second treatment

the inlet water tap of each channel, to exclude the possible effect of non-random distribution of prey on fish behavior.

Three series of habitat use experiments were performed, all following the same procedure. First, fish were acclimated for 24 h in the experimental channels, allowing them to swim freely through the four compartments. Following this, the location of individuals was recorded every 6 h (coinciding with dawn, dusk, midday, and midnight) for a period of 24 h. Before counting, each substrate was isolated using a mobile sluice that prevented fish moving from one section to another as a result of the presence of the observer. The first and second series of experiments were conducted with the presence of only one species per experiment (*P. trucha* or *O. mykiss*, respectively), using 4 individuals per channel, with 4 repetitions (series 1—*P. trucha*: $N = 64$, 4.45 ± 0.83 cm TL;

series 2—*O. mykiss*: $N = 64$, 7.98 ± 1.25 cm TL). The third series combined the presence of both species, with 4 individuals of each species per channel and 4 repetitions (64 *P. trucha* of 4.32 ± 0.79 cm TL and 64 *O. mykiss* 7.89 ± 1.28 cm TL).

Differences in habitat use were analyzed with Kruskal–Wallis test (KW), since data normality and homoscedasticity assumptions failed. Following this, a multiple comparison procedure (Tukey test) was used to detect differences between pairs of habitats. The same procedure was used to check for differences in the use of any position of the channel, with regard to water inlet. A Spearman correlation test was used to establish whether there was a correlation between *P. trucha* and *O. mykiss* presence in the third series of experiments.

Diel locomotor activity experiment

This experiment was performed in a 30-l aquarium, divided into two compartments by a transparent glass sheet that allowed visual fish detection without water exchange. One compartment (A) was equipped with a movement sensor to register *P. trucha* locomotor activity. The equipment comprised 6 vertically disposed infrared beams (with a distance of 1.5 cm between each one) which crossed the aquarium transversely (Fig. 1). The interruption of any of these beams was registered by a computer every second for 48 h (i.e., the duration of the experiment). The other compartment (B) was destined for the *O. mykiss* individual.

The aquarium was placed in an incubator (Ingelab, Modelo I.501 PF) to provide a stable temperature of 10°C. The natural photoperiod was imitated using the same device as in the habitat use experiment. Light intensity varied from 0 to 97 lx measured at water surface level. All fish were acclimated for a period of at least 10 days prior to experimentation, in two aquariums placed in the incubator, under identical experimental conditions (i.e., temperature and photoperiod). During this period fish were fed with *Tubifex* sp.

In the first treatment, one individual of *P. trucha* was placed in compartment A and acclimated for 24 h before its locomotor activity was registered for the following 24 h. For the second treatment, an individual of *O. mykiss* was placed in compartment B. After 24 h of measuring *P. trucha* activity in the presence of *O. mykiss*, both fish were removed from the aquarium.

A total of 10 replicates were performed, using 10 *P. trucha* individuals of 5.02 ± 0.58 cm TL and the same individual of *O. mykiss* (10.53 cm TL).

Although *P. trucha* and *O. mykiss* had no contact (visual or olfactory) during night (0 lx period) in the second treatment, the rationale of this experiment included possible *P. trucha* stress due to the *O. mykiss* presence during the previous day, and its putative impact on *P. trucha* activity the following night.

To compare data of movement per individual between day, night and crepuscules, the relative percentage of movement was calculated as $M = R_H \times R_T^{-1} \cdot 100$, where M is the percentage movement, R_H is the number of registers per hour and R_T is the total number of registers during the 24-h cycle. The movement of every individual was calculated separately for each treatment, then the average movement during day or night hours per individual (e.g., average of 11 h of the day for perch 1 = day % movement of perch 1) was used. Since data of each crepuscule corresponded only to one hour there was no need to obtain the average. Following this, a Rank Transformation (Conover & Iman, 1981) was applied to achieve normality and homoscedasticity before performing a two-way ANOVA. A multiple comparison procedure (Tukey test) was used to detect the factors that differed. To compare the intensity of movement between treatments (*P. trucha* alone vs. *P. trucha* with *O. mykiss*) the same procedure was followed, using the total data of movement (i.e., 11 h of day and 11 h of night, without averaging).

To detect any relationship between light intensity and *P. trucha* movement, data from crepuscules were analyzed. Percentage of crepuscular movement was calculated as $M_c = R_{10m} \times R_c^{-1} \cdot 100$, where M_c is the percentage movement during crepuscule, R_{10m} is the number of registers per 10 min and R_c is the total number of registers during that crepuscule (i.e., one hour). A Spearman Correlation test was used to check the relationship between M_c and light intensity.

Results

Habitat use

There was no clear use by fish of any particular position regarding water inlet in the experimental channels. *Percichthys trucha* showed differences

during dusk (KW; $H = 8.97$; $df = 3$; $P = 0.030$) and midnight (KW; $H = 8.39$; $df = 3$; $P = 0.039$); however, this was only between the use of positions 1 and 3 (Tukey, $P > 0.05$) in the first treatment.

The general pattern using the total number of registers indicates that cobbles was the habitat most frequented by *P. trucha* when experimenting with this species alone (KW; $H = 103.54$; $df = 3$; $P < 0.001$) (Fig. 2). Use of cobbles was significantly higher from the other habitats during both dawn and midday (Tukey, $P < 0.05$), but no differences from macrophytes were presented during dusk, or with gravel during midnight (Tukey, $P > 0.05$). These results suggest that at these two moments, macrophytes and gravel, are used secondarily by *P. trucha* (Fig. 3).

Similar to *P. trucha*, *O. mykiss* preferred the cobbles habitat when it was alone (KW; $H = 116.65$; $df = 3$; $P < 0.001$). However, *O. mykiss* had a slightly different order of preference when analyzing the total number of registers; it used mostly cobbles, and secondarily, macrophytes and gravel, whereas sand was the least used (Tukey, $P < 0.05$) (Fig. 2). The same pattern was observed during the day (KW; $H = 46.78$; $df = 3$; $P < 0.001$), when sand was the least-used habitat (Tukey, $P < 0.05$). In addition, for all moments of the cycle *O. mykiss* juveniles preferred cobbles (Tukey, $P < 0.05$) as also registered in the first experiment for *P. trucha* (Fig. 3).

Differences were found in *P. trucha* habitat use when the two species were together (KW; $H = 115.40$;

$df = 3$; $P < 0.001$), with a preference for cobbles and secondarily, macrophytes. The least-used habitats were sand and gravel. This differed from the general pattern observed when *P. trucha* was alone, when macrophytes were as infrequently used as sand and gravel (Fig. 2). In particular, when habitat use was analyzed by moment of the day, differences were found during dawn (KW; $H = 41.11$; $df = 3$; $P < 0.001$), midday (KW; $H = 39.29$; $df = 3$; $P < 0.001$), dusk (KW; $H = 32.93$; $df = 3$; $P < 0.001$), and midnight (KW; $H = 9.09$; $df = 3$; $P = 0.028$). However, at midnight no differences were found when applying an a posteriori test (Tukey, $P > 0.05$). Macrophytes and cobbles were the most used during midday and dusk, whereas sand and gravel were less used. The preference order during dawn was, from most to least used, cobbles, macrophytes, gravel, and sand (Fig. 3).

There was also a negative correlation between the presence of *O. mykiss* and the presence of *P. trucha* in sand or gravel (Spearman, $\rho = -0.377$, $P = 0.00512$). In addition, the only habitat in which the 8 individuals (i.e., third treatment) were found together was cobbles. Finally, this habitat was more frequently used during midday, whereas sand and gravel were used more often during midnight.

Diel locomotor activity

Percichthys trucha were more active during night in both treatments. In general, no differences were found

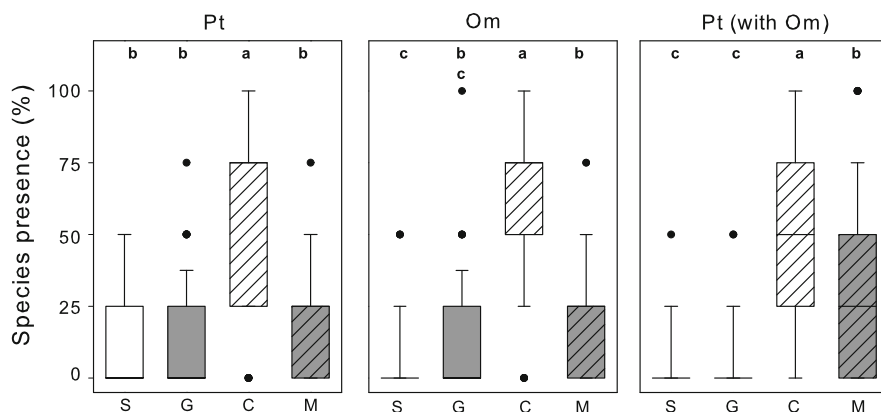
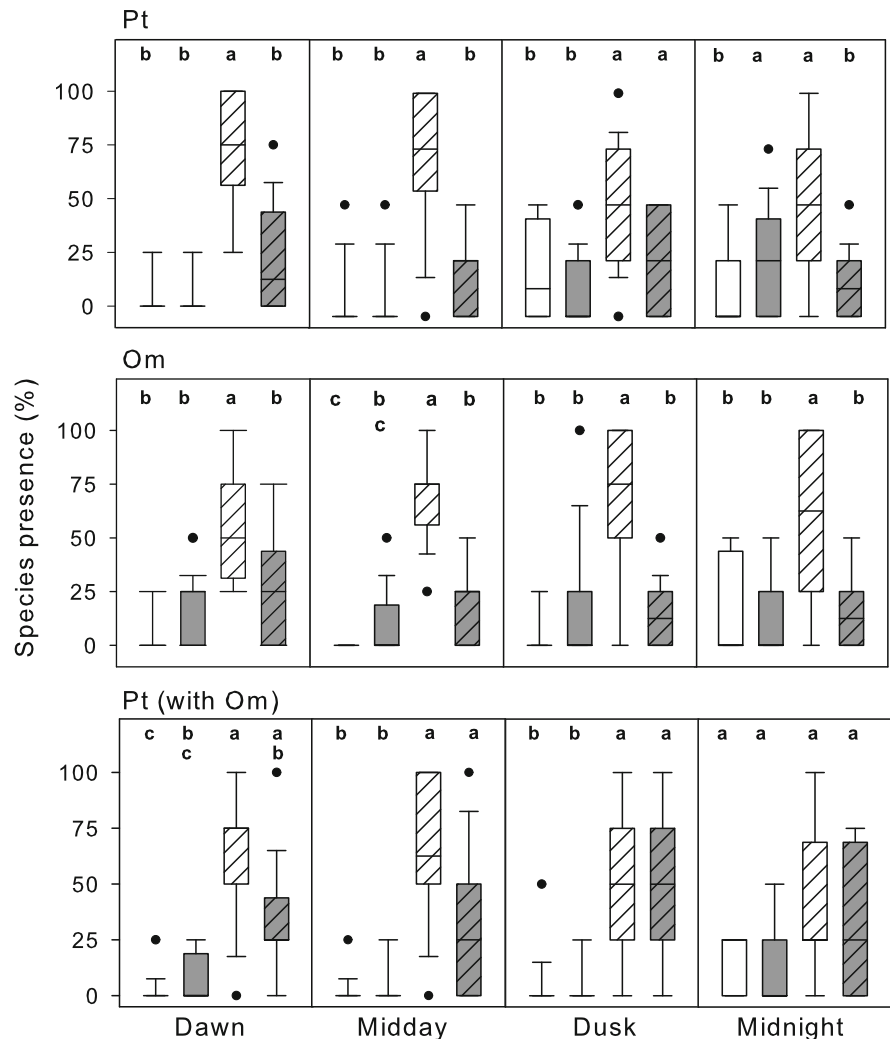


Fig. 2 General pattern of *P. trucha* and *O. mykiss* habitat selection as percentage of individuals present in each habitat. Pt: *P. trucha* alone (experiment 1). Om: *O. mykiss* alone (experiment 2). Pt (with Om): *P. trucha* in the presence of *O. mykiss* (experiment 3). Median, quartiles and data outside 10th

and 90th percentiles are indicated for sand (S), gravel (G), cobbles (C) and macrophytes (M). Significant differences between substrates are indicated by different letters (Tukey, $P < 0.05$)

Fig. 3 Habitat selection of *P. trucha* and *O. mykiss* per moment of the cycle as percentage of individuals present in each habitat. Pt: *P. trucha* alone (experiment 1). Om: *O. mykiss* alone (experiment 2). Pt (with Om): *P. trucha* in the presence of *O. mykiss* (experiment 3). Median, quartiles, and data outside 10th and 90th percentiles are indicated for sand (white), gravel (gray), cobbles (striped white), and macrophytes (striped gray). Significant differences between substrates within each time of the day (dawn, midday, dusk, and midnight) are indicated by different letters (Tukey, $P < 0.05$)



between treatments in the activity of *P. trucha* per hour (KW; $H < 1.12$; $df = 1$; $P > 0.23$). The ANOVA performed with the relative movement data showed differences between moments of the day ($F = 14.1$; $P < 0.001$), showing that *P. trucha* is more active during night (Tukey, $P < 0.05$) (Fig. 4).

Regarding the intensity of movement (total registers), there was no difference between treatments during crepuscules (ANOVA; $F = 0.040$, $P = 0.84$) or between dawn and dusk (ANOVA; $F = 0.150$, $P = 0.70$). Nor did the interaction between factors show differences (ANOVA; $F = 0.011$, $P = 0.92$). When nighttime and daytime were compared, the activity of *P. trucha* depended on time of day ($F = 19.170$, $P < 0.001$), movement levels being higher during night (Tukey, $P < 0.05$), without

differences between treatments ($F = 0.066$, $P = 0.80$). That is, no differences were found regarding the intensity of activity of *P. trucha* with or without trout (Fig. 5).

Finally, *P. trucha* activity was correlated with light intensity during crepuscules (dawn treatment 1: Spearman; $\rho = -0.558$; $P < 0.001$; dusk treatment 1: $\rho = -0.433$; $P < 0.001$; dusk treatment 2: $\rho = 0.526$; $P < 0.001$), with the exception of dawn in the second treatment (Fig. 6).

Discussion

Fish distribution was related to habitat category rather than channel position. The habitat results highlight

that *Percichthys trucha* juveniles, in the absence of *Oncorhynchus mykiss*, mainly used environments that provide refuge: mostly cobbles, and secondarily macrophytes and gravel during dusk and midnight, respectively. Sand was the most exposed (i.e., least structured) habitat and least used by *P. trucha*, only during midnight or dusk. *Oncorhynchus mykiss* juveniles displayed a similar pattern to *P. trucha*: cobbles were preferred during all periods of the cycle and sand was the least used during midday. As predicted, *O. mykiss* used mainly coarse substrate. However, *P. trucha* juveniles were expected to use mainly

macrophytes, but instead they were found to use mainly cobbles. Lapointe & Corkum (2007) pointed out that small-sized fish are strongly associated with habitat variables that reduce predation risk. Small fishes are important food items for adults of *P. trucha*, *O. mykiss* and *Odontesthes hatcheri* in Negro River (Alvear et al., 2007). In addition, the Neotropical Cormorant *Phalacrocorax brasilianus* (Gmelin, 1789) also preys on small fishes, as has been inferred by its diet (Alarcón et al., 2012). Structured environments like cobbles may provide low-risk areas, inaccessible to predators because of the formation of interstitial spaces that can be used by small fish. This substrate is commonly found in erosional zones of rivers, where high water velocity prevents the deposit of fine particles between larger rocks.

In the presence of *O. mykiss*, *P. trucha* increased the use of macrophytes and decreased the use of sand. This supports the second hypothesis of the study, since a shift in *P. trucha* habitat use was evident in the presence of *O. mykiss*. Moreover, the increased use of macrophytes agrees with field observations. Barriga et al. (2013) found a higher abundance of *P. trucha* juveniles associated with macrophytes in Pichileufu River, where the abundance of salmonids, mainly *O. mykiss* and *Salmo trutta*, was high. Other ecologically similar species of the northern hemisphere, such as *Perca fluviatilis* Linnaeus, 1758, also use vegetated areas as an antipredatory strategy during early life stages (Byström et al., 2003; Lewin et al., 2004;

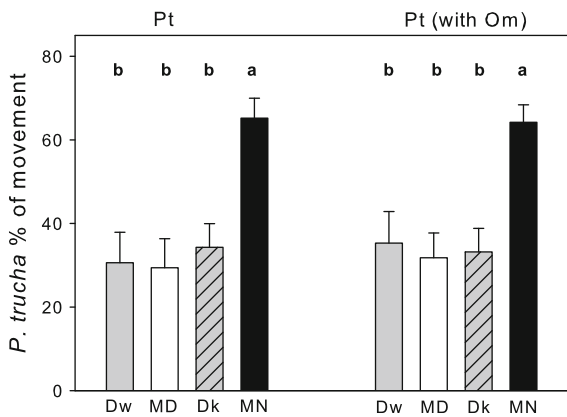


Fig. 4 *P. trucha* percentage of movement (in ranks) per moment of the 24-h cycle. Significant differences between pairs are indicated by different letters (Tukey, $P < 0.05$). Dw: dawn, MD: midday, Dk: dusk, MN: midnight

Fig. 5 *P. trucha* intensity of movement (N° of registers, in ranks) per moment of the 24-h cycle. *Left* Crepuscules. *Right* Day and Night. Significant differences between pairs in each panel are indicated by different letters (Tukey, $P < 0.05$). Dw dawn, MD midday, Dk dusk, MN midnight

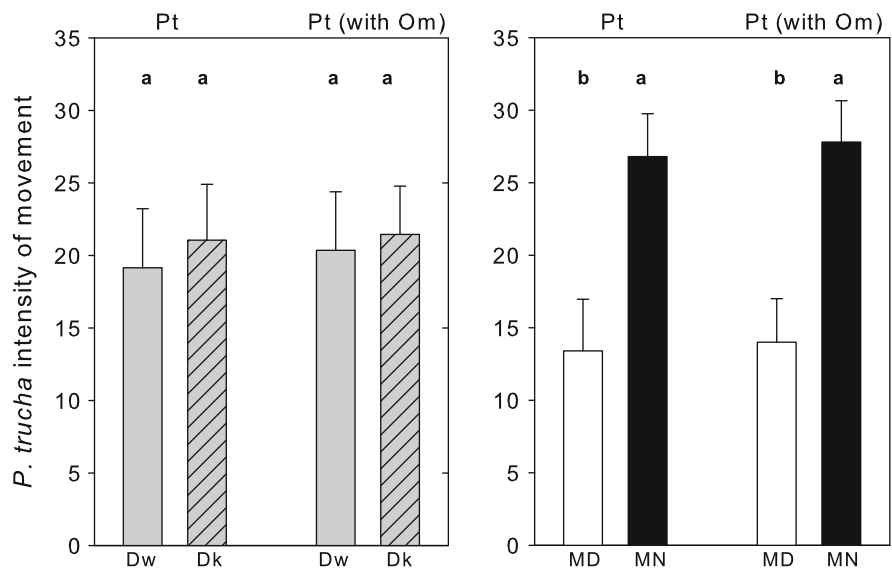
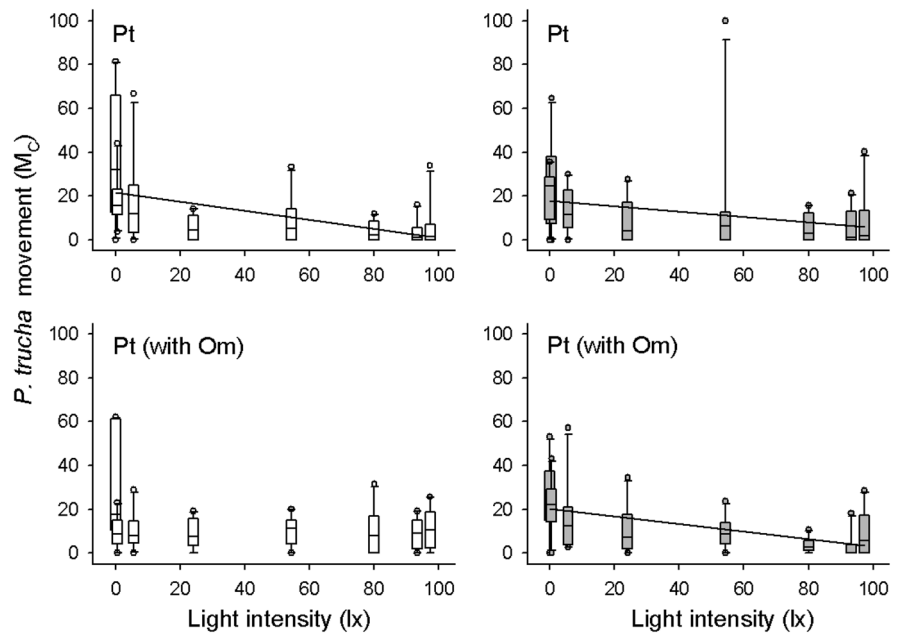


Fig. 6 Crepuscular movement of *P. trucha* (M_C) in relation to light intensity (lx) with and without *O. mykiss* presence. Median, quartiles and data outside 10th and 90th percentiles are indicated for dawn (white) and dusk (gray). Lines indicate negative correlations (Spearman, $P < 0.001$)



Snickars et al., 2004). Similar results were shown by Fullerton & Lamberti (2006), in habitat use experiments performed with the native *Perca flavescens* (Mitchill, 1814) and the introduced *Gymnocephalus cernua* (Linnaeus, 1758) in North America. Both species used macrophytes and secondarily cobbles over mud during day. During night, *G. cernua* used mud more often whereas *P. flavescens* showed the same pattern as during day. Differences between *P. trucha* and *P. flavescens* in the order of preference of the two structured habitats (macrophytes and cobbles) might be related to the fact that Fullerton & Lamberti (2006) employed a higher density of macrophytes. Regardless, their results also indicate that other juveniles of this family use structured habitats during day, while risky zones are used during nighttime.

Fine substrates constitute areas of high predation risk because fish cannot use shelters. These risky zones were used more often by *P. trucha* juveniles during midnight. The response of an individual to predation risk is shaped by the trade-off between the need to avoid predators and benefits of foraging in appropriate places (Barbosa & Castellanos, 2005). *Perca* juveniles undergo starvation during winter, which causes high mortality (Huss et al., 2008). Therefore, the use of risky habitats might be related to food search since feeding efficiency is diminished in complex environments (Crowder & Cooper, 1982; Starry et al., 1998; Kolar et al., 2002).

In the absence of refuge, *P. trucha* and *O. mykiss* juveniles segregate spatially. This is supported by the negative correlation found between the presence of the two species in the less structured habitats. The only habitat in which the total number of fish was found at the same moment was cobbles. Hence, the size of the interstitial space is key in providing shelter for small fish. Similarly, Hegggenes et al. (2013) found that Atlantic salmon could easily move within the substrate and that fish size and substrate harshness were the parameters that limited their movement. As *P. trucha* and *O. mykiss* juveniles share structured habitats in nature, and considering their similar diets (Cussac et al., 1998; Ruzzante et al., 1998; Macchi et al., 1999; Lattuca et al., 2008), the potential for competition is very high if one of these resources is limited (Fullerton & Lamberti, 2006). In this context, the ability of *P. trucha* to increase the use of macrophytes could represent an advantage, allowing niche segregation to avoid competition with *O. mykiss* juveniles.

Concerning diel activity, *P. trucha* juveniles exhibited a nocturnal movement pattern, showing that photoperiod is an important synchronizer of activity in this species. Fullerton & Lamberti (2006) found a similar pattern for *G. cernua*, which was very active during nighttime but remained almost still during daytime. Barriga et al. (2016) also registered the same diel pattern in the small stream catfish *Hatcheria macraei*. Moreover, the negative correlation between

the activity of *P. trucha* and light intensity during crepuscules is consistent with a nocturnal habit.

Contrary to the last hypothesis, the presence of *O. mykiss* did not affect the movement pattern of *P. trucha*. Neither was movement intensity affected. At this point it is important to note that although *P. trucha* was unable to detect *O. mykiss* during night, *P. trucha* behavior could have been affected during nighttime by the previous exposition to *O. mykiss* presence during the light period. However, this was not the case, ruling out a possible stress effect generated previously by the presence of *O. mykiss*. Nonetheless, it was confirmed that *P. trucha* did detect the presence of *O. mykiss* during the light period, since defensive responses were repeatedly observed. This behavior entailed an erection of the dorsal fin spines and an unusual opening of the operculum (Otturi, personal observation). This response was also observed by Ylönen et al. (2007) in *P. fluviatilis* and *G. cernua*, while they evaluated response to predator odor. Furthermore, predation pressure effects can be so important that they can even induce morphological changes in prey, such as enlargement of defensive structures. For example, Ruzzante et al. (2011) found a direct relationship between the dorsal spine length of *P. trucha* and the density of potential predators.

The introduction of exotic species in an ecosystem entails a potential ecological risk (Gozlan & Newton, 2009), resulting in possible detrimental effects for native species or ecosystem function (Gozlan et al., 2010). Both the habitat use and diel activity registered in this study suggest that predation is a key factor in early life stages of *P. trucha*. This factor is likely to have increased in importance since the introduction of salmonids in Patagonia, although this is difficult to prove due to lack of data prior to introduction (Macchi et al., 2007). However, there is consistent evidence of the deleterious effects that salmonids have had on native populations in other countries (Tilzey, 1976; McDowall, 1990, 2006; Glova et al., 1992; McIntosh et al., 1992). For example, competition for space and food has been proposed as the cause of decline in the *Galaxias vulgaris* Stokell, 1949 population in New Zealand streams (McIntosh et al., 1992). In addition, introduced salmonids resulted in a 26% decline in Galaxiidae populations in Chile (Habit et al., 2010), and Young et al. (2010) showed a decline in the growth rate of *Aplocheilichthys zebra* Jenyns, 1842 in artificial enclosures. Elgueta et al. (2013) suggested

that the absence of *A. zebra* in rivers invaded with *S. trutta* may reflect competitive exclusion in trophic resource use. Penaluna et al. (2009) registered a shift in the mesohabitat use of three native species when salmonids were removed from a river section.

In conclusion, the trade-off between feeding in appropriate but dangerous areas and sheltering in areas with less food (Ydenberg & Dill, 1986) is directly related to diel cycle. Light intensity appears to be the variable that regulates this balance in order to minimize the risk of being preyed upon. Most animals experience a certain degree of predation risk associated with most of their activities (Abrahams, 2005). *P. trucha* juveniles minimize this risk by sheltering in structured areas during daytime and exploring less structured areas during nighttime, when this species is more active. However, sharing the same habitat with other individuals, such as *O. mykiss* juveniles, could entail resource competition during shelter use periods. The ability of *P. trucha* to use macrophytes could solve this problem for juveniles of this species. In any case, the use of structured habitats is crucial for juvenile fish. In this sense, the removal of substratum from a riverbed or the construction of channels or dams could have a severe impact on river fish communities. Therefore, the structural factors of a river must be considered when developing management strategies to improve the conservation status of native fish species.

Acknowledgments We would like to thank two anonymous reviewers for their comments that helped to improve this work. We also thank *Dirección de Pesca Continental* of Río Negro Province for permission to collect native fishes. This study was partially funded by *Agencia Nacional de Promoción Científica y Tecnológica, Argentina* (ANPCyT, PICT 2010 No. 0262 and PICT 2013 No. 1387) and *Consejo Nacional de Investigaciones Científicas y Técnicas, Argentina* (CONICET, PIP No. 11220120100063CO). All handling, care, and experimental procedures used in this research complied with the animal welfare laws stated by the Government of Argentina (Law n° 14346).

References

- Abrahams, M., 2005. The physiology of antipredator behaviour: what you do with what you've got. *Behaviour and Physiology of Fish* 24: 79–108.
- Alanärä, A., M. D. Burns & N. B. Metcalfe, 2001. Intraspecific resource partitioning in brown trout: the temporal distribution of foraging is determined by social rank. *Journal of Animal Ecology* 70: 980–986.
- Alarcón, P. A. E., P. J. Macchi, A. Trejo & M. F. Alonso, 2012. Diet of the Neotropical Cormorant (*Phalacrocorax*

- brasilianus*) in a Patagonian freshwater environment invaded by exotic fish. *Waterbirds* 35: 149–153.
- Alvear, P. A., M. Rechencq, P. J. Macchi, M. F. Alonso, G. E. Lippolt, M. A. Denegri, G. Navone, E. E. Zattara, M. I. García Asorey & P. H. Vigliano, 2007. Composición, distribución y relaciones tróficas de la ictiofauna del Río Negro, Patagonia Argentina. *Ecología Austral* 17: 231–246.
- Amundsen, P.-A., R. Knudsen & H. T. Bryhni, 2010. Niche use and resource partitioning of Arctic charr, European whitefish and grayling in a subarctic lake. *Hydrobiologia* 650: 3–14.
- Barbosa, P. & I. Castellanos, 2005. *Ecology of Predator-Prey Interactions*. Oxford University Press, New York.
- Barriga, J., M. Battini & V. Cussac, 2007. Annual dynamics variation of landlocked *Galaxias maculatus* (Jenyns 1842) population in a river of Northern Patagonia: occurrence of juvenile upstream migration. *Journal of Applied Ichthyology* 23: 128–135.
- Barriga, J. P., N. A. Espinós, M. Chiarello-Sosa & M. A. Battini, 2013. The importance of substrate size and interstitial space in the microhabitat selection by the stream-dwelling catfish *Hatcheria macraei* (Actinopterygii, Trichomycteridae). *Hydrobiologia* 705: 191–206.
- Barriga, J. P., N. A. Espinós, W. D. Torres & M. A. Battini, 2016. Diel locomotor activity and shelter use in the Patagonian catfish *Hatcheria macraei* under experimental conditions. *Journal of Ethology* 34: 53–58.
- Blanchet, S., G. Loot, G. Grenouillet & S. Brosse, 2007. Competitive interactions between native and exotic salmonids: a combined field and laboratory demonstration. *Ecology of Freshwater Fish* 16: 133–143.
- Blanchet, S., G. Loot & J. J. Dodson, 2008. Competition, predation and flow rate as mediators of direct and indirect effects in a stream food chain. *Oecologia* 157: 93–104.
- Brännäs, E. & A. Alanära, 1997. Is diel dualism in feeding activity influenced by competition between individuals? *Canadian Journal of Zoology* 75: 661–669.
- Byström, P., L. Persson, E. Wahlström & E. Westman, 2003. Size- and density-dependent habitat use in predators: consequences for habitat shifts in young fish. *Journal of Animal Ecology* 72: 156–168.
- Chase, J., P. Abrams, J. Grover, S. Diehl, P. Chesson, R. Holt, S. Richards, R. Nisbet & T. Case, 2002. The interaction between predation and competition: a review and synthesis. *Ecological Letters* 5: 302–315.
- Conover, W. J. & R. L. Iman, 1981. Range Transformation as a bridge between parametric and nonparametric statistics. *The American Statistician* 35: 124–129.
- Copp, G., 1868. Putting multi-dimensionality back into niche: diel vs. day-only niche breadth separation in stream fishes. *Fundamental and Applied Limnology* 170: 273–280.
- Copp, G., S. Spathari & M. Turmel, 2005. Consistency of diel behaviour and interactions of stream fishes and invertebrates during summer. *River Research and Applications* 21: 75–90.
- Crichigno, S. A., M. A. Battini & V. E. Cussac, 2014. Natural and diet induced morphological variation in *Percichthys trucha* (Valenciennes, 1833) (Perciformes, Percichthyidae) in Patagonia: the importance of phenotypic plasticity. *Zoologischer Anzeiger* 253: 192–202.
- Crowder, L. B. & W. E. Cooper, 1982. Habitat structural complexity and the interaction between bluegills and their prey. *Ecology* 63: 1802–1813.
- Cussac, V. E., D. E. Ruzzante, S. J. Walde, P. J. Macchi, V. Ojeda, M. F. Alonso & M. A. Denegri, 1998. Body-shape variation of three species of *Percichthys* in relation to their coexistence in the Limay river basin, in Northern Patagonia. *Environmental Biology of Fishes* 53: 143–153.
- Dixon, C. J. & J. C. Vokoun, 2009. Burbot resource selection in small streams near the southern extent of the species range. *Ecology of Freshwater Fish* 18: 234–246.
- Elgueta, A., J. González, D. Ruzzante, S. Walde & E. Habit, 2013. Trophic interference by *Salmo trutta* on *Aplochiton zebra* and *Aplochiton taeniatus* in southern Patagonian lakes. *Journal of Fish Biology* 82: 430–443.
- Fullerton, A. H. & G. A. Lamberti, 2006. A comparison of habitat use and habitat-specific feeding efficiency by Eurasian ruffe (*Gymnocephalus cernuus*) and yellow perch (*Perca flavescens*). *Ecology of Freshwater Fish* 15: 1–9.
- García de Leaniz, C., G. Gajardo & S. Consuegra, 2010. From best to pest: changing perspectives on the impact of exotic salmonids in the southern hemisphere. *Systematics and Biodiversity* 8: 447–459.
- Glova, G., P. Sagar & I. Naslund, 1992. Interaction for food and space between populations of *Galaxias vulgaris* Stokell and juvenile *Salmo trutta* L. in a New Zealand stream. *Journal of Fish Biology* 41: 909–925.
- Gozlan, R. & A. Newton, 2009. Biological invasions: benefits versus risks. *Science* 324: 1015–1016.
- Gozlan, R., J. Britton, I. Cowx & G. Copp, 2010. Current knowledge on non-native freshwater fish introductions. *Journal of Fish Biology* 76: 751–786.
- Grabowska, J., T. Kakareko, D. Błonska, M. Przybylski, J. Kobak, Ł. Jermacz & G. H. Copp, 2016. Interspecific competition for a shelter between non-native racer goby and native European bullhead under experimental conditions – effects of season, fish size and light conditions. *Limnologica* 56: 30–38.
- Grossman, G. D., R. E. Ratajczak Jr, J. T. Petty, M. D. Hunter, J. T. Peterson & G. Grenouillet, 2006. Population dynamics of mottled sculpin (pisces) in a variable environment: information theoretic approaches. *Ecological Monographs* 76: 217–234.
- Habit, E., P. Piedra, D. Ruzzante, S. Walde, M. Belk, V. Cussac, J. Gonzalez & N. Colin, 2010. Changes in the distribution of native fishes in response to introduced species and other anthropogenic effects. *Global Ecology and Biogeography* 19: 697–710.
- Hasegawa, K. & K. Maekawa, 2006. The effects of introduced salmonids on two native stream-dwelling salmonids through interspecific competition. *Journal of Fish Biology* 68: 1123–1132.
- Heggenes, J., G. Bremset & A. Brabrand, 2013. Visiting the hyporheic zone: young Atlantic salmon move through the substratum. *Freshwater Biology* 58: 1720–1728.
- Huss, M., P. R. Byström, A. S. Strand, L. O. Eriksson & L. Persson, 2008. Influence of growth history on the accumulation of energy reserves and winter mortality in young fish. *Canadian journal of fisheries and aquatic sciences* 65: 2149–2156.

- ISSG, 2012. Invasive Species Specialist Group: www.issg.org/database/species/ecology.sp?fr=1&si=103.
- Jackson, D., P. Peres-Neto & J. Olden, 2001. What controls who is where in freshwater fish communities – the roles of biotic, abiotic, and spatial factors. *Canadian Journal of Fisheries and Aquatic Sciences* 58: 157–170.
- Johnson, J. H. & K. A. Douglass, 2009. Diurnal stream habitat use of juvenile Atlantic salmon, brown trout and rainbow trout in winter. *Fisheries Management and Ecology* 16: 352–359.
- Jonsson, B., N. Jonsson, K. Hindar, T. G. Northcote & S. Engen, 2008. Asymmetric competition drives lake use of coexisting salmonids. *Oecologia* 157: 553–560.
- Kolar, C. S., A. H. Fullerton, K. M. Martin & G. A. Lamberti, 2002. Interactions among zebra mussel shells, invertebrate prey, and Eurasian ruffe or yellow perch. *Journal of Great Lakes Research* 28: 664–673.
- Kronfeld-Schor, N. & T. Dayan, 2003. Partitioning of time as an ecological resource. *Annual Review of Ecology, Evolution and Systematics* 34: 153–181.
- Landless, P. J., 1976. Demand-feeding behavior of rainbow trout. *Aquaculture* 7: 11–25.
- Lapointe, N. W. R. & L. D. Corkum, 2007. Seasonal and ontogenic shifts in microhabitat selection by fishes in the shallow waters of the Detroit River, a large connecting channel. *Transactions of the American Fisheries Society* 136: 155–166.
- Lattuca, M., M. A. Battini & P. Macchi, 2008. Trophic interactions among native and introduced fishes in a northern Patagonian oligotrophic lake. *Journal of Fish Biology* 72: 1306–1320.
- Lewin, W. C., N. Okun & T. Mehner, 2004. Determinants of the distribution of juvenile fish in the littoral area of a shallow lake. *Freshwater Biology* 49: 410–424.
- López-Olmeda, J., J. Madrid & F. Sánchez-Vázquez, 2006. Light and temperature cycles as zeitgebers of zebrafish (*Danio rerio*) circadian activity rhythms. *Chronobiology International* 23: 537–550.
- Macchi, P. J. & P. H. Vigliano, 2014. Salmonid introduction in Patagonia: the ghost of past, present and future management. *Ecología Austral* 24: 162–172.
- Macchi, P. J., V. Cussac, M. Alonso & M. A. Denegri, 1999. Predation relationships between introduced salmonids and native fish fauna in lakes and reservoirs of northern Patagonia. *Ecology of Freshwater Fish* 8: 227–236.
- Macchi, P. J., M. A. Pascual & P. H. Vigliano, 2007. Differential piscivory of the native *Percichthys trucha* and exotic salmonids upon the native forage fish *Galaxias maculatus* in Patagonian Andean lakes. *Limnologica* 37: 76–87.
- Marini, T. L., 1936. Los salmónidos en nuestros Parque Nacional de Nahuel Huapi. *Anales de la Sociedad Científica Argentina* 121: 1–24.
- McDowall, R. M., 1990. When galaxiid and salmonid fishes meet – a family reunion in New Zealand. *Journal of Fish Biology* 37: 35–43.
- McDowall, R. M., 2006. Crying wolf, crying foul, or crying shame: alien salmonids and a biodiversity crisis in the southern cool-temperate galaxioid fishes? *Reviews in Fish Biology and Fisheries* 16: 233–422.
- McIntosh, A. R., C. R. Townsend & T. A. Crowl, 1992. Competition for space between introduced Brown trout (*Salmo trutta* L.) and a native galaxiid (*Galaxias vulgaris* Stokell) in a New-Zealand stream. *Journal of Fish Biology* 41: 63–81.
- Menni, R. C., 2004. Peces y ambientes en la Argentina continental. *Monografías del Museo Argentino de Ciencias Naturales* 5: 1–316.
- Penaluna, B., I. Arismendi & D. Soto, 2009. Evidence of interactive segregation between introduced trout and native fishes in northern Patagonian Rivers, Chile. *Transactions of the American Fisheries Society* 138: 839–845.
- Petty, T. & G. Grossman, 1996. Patch selection by mottled sculpin (Pisces: Cottidae) in a southern Appalachian stream. *Freshwater Biology* 35: 261–276.
- Ruzzante, D., S. Walde, V. Cussac, P. Macchi & M. Alonso, 1998. Trophic polymorphism, habitat and diet segregation in *Percichthys trucha* (Pisces: Percichthyidae) in the Andes. *Biological Journal of the Linnean Society* 65: 191–214.
- Ruzzante, D., S. Walde, V. Cussac, P. Macchi, M. Alonso & M. Battini, 2003. Resource polymorphism in a Patagonian fish *Percichthys trucha* (Percichthyidae): phenotypic evidence for interlake pattern variation. *Biological Journal of the Linnean Society* 78: 497–515.
- Ruzzante, D., S. Walde, V. Cussac, M. Dalebout, J. Seibert, S. Ortubay & E. Habit, 2006. Phylogeography of the Percichthyidae in Patagonia: roles of orogeny, glaciation, and volcanism. *Molecular Ecology* 15: 2949–2968.
- Ruzzante, D., S. Walde, V. Cussac, P. Macchi, M. Alonso & J. Barriga, 2011. Phylogeography and phenotypic diversification in the Patagonian fish *Percichthys trucha*: the roles of Quaternary glacial cycles and natural selection. *Biological Journal of the Linnean Society* 103: 514–529.
- Schlosser, I. J., 1988. Predation risk and habitat selection by two size classes of stream cyprinids: experimental test of a hypothesis. *Oikos* 52: 36–40.
- Schoener, T. W., 1974. Resource partitioning in ecological communities. *Science* 185: 27–39.
- Snickars, M., A. Sandström & J. Mattila, 2004. Antipredator behaviour of 0+ year *Perca fluviatilis*: effect of vegetation density and turbidity. *Journal of Fish Biology* 65: 1604–1613.
- Starry, O., J. Wanzenboeck & D. L. Danielopol, 1998. Tendency of the amphipod *Gammarus roeseli* Gervais to colonize coarse sediment habitats under fish predation pressure. *International Review of Hydrobiology* 83: 5–6.
- Tilzey, R. D. J., 1976. Observations on interactions between indigenous Galaxiidae and introduced Salmonidae in the Lake Eucumbene catchment, New South Wales. *Australian Journal of Marine and Freshwater Research* 27: 551–564.
- Thorpe, J. E., 1978. *Rhythmic Activity of Fishes*. Academic Press, London.
- Van Liefferinge, C., P. Seeuws, P. Meire & R. Verheyen, 2005. Habitat use and preferences of the endangered *Cottus gobio* in the River Voer, Belgium. *Journal of Fish Biology* 67: 897–909.
- Van Snik Gray, E., K. A. Kellogg & J. R. Stauffer Jr, 2016. Habitat shift of a native darter *Etheostoma olmstedii* (Teleostei: Percidae) in sympatry with a non-native darter

- Etheostoma zonale*. The American Midland Naturalist 154: 166–177.
- Vander-Zanden, J., M. Casselman & J. B. Rasmussen, 1999. Stable isotope evidence for food web shifts following species invasions of lakes. Nature 401: 464–467.
- Ydenberg, R. C. & L. M. Dill, 1986. The economics of fleeing from predators. Advances in the Study of Behavior 16: 229–249.
- Ylönen, H., R. Kortet, J. Myntti & A. Vainikka, 2007. Predator odor recognition and antipredatory response in fish: does the prey know the predator diel rhythm? Acta Oecologica 31: 1–7.
- Young, K., J. Dunham, J. Stephenson, A. Terreau, A. Thailly, G. Gajardo & C. Garcia de Leaniz, 2010. A trial of two trouts: comparing the impacts of rainbow and brown trout on a native galaxiid. Animal Conservation 13: 399–410.